

Machine-learning-based estimates of global natural vegetated wetland methane emissions (2000-2025)

Mengze Li^{1,2}, Robert B. Jackson^{2,3}, Marielle Saunois⁴, Philippe Ciais⁴, Ben Poulter⁵, Josep G. Canadell⁶, Prabir K. Patra^{7,8}, Hanqin Tian^{9,10}, Zhen Zhang¹¹, Etienne Fluet-Chouinard¹², Zutao Ouyang¹³, Ting Zhang¹, David J. Beerling¹⁴, Dmitry A. Belikov¹⁵, Philippe Bousquet⁴, Danilo Custodio⁴, Naveen Chandra⁷, Xinyu Dou², Nicola Gedney¹⁶, Peter O. Hopcroft¹⁷, Alison M. Hoyt², Kazuhito Ichii^{15,18}, Akihito Ito¹⁹, Atul K. Jain²⁰, Katherine Jensen²¹, Fortunat Joos²², Thomas Kleinen²³, Masayuki Kondo^{8,24}, Fa Li², Tingting Li²⁵, Xiangyu Liu²⁶, Shamil Maksyutov²⁷, Avni Malhotra²⁸, Adrien Martinez⁴, Kyle McDonald²¹, Joe R. Melton²⁹, Jurek Müller²², Yosuke Niwa^{27,30}, Shufen Pan⁹, Shushi Peng³¹, Changhui Peng^{32,33}, Zhangcai Qin³⁴, Peter Raymond³⁵, William Riley³⁶, Arjo Segers³⁷, Rona L. Thompson¹⁶, Aki Tsuruta³⁸, Xi Yi⁴, Kunxiaojia Yuan³⁹, Wenxin Zhang³⁰, Bo Zheng^{40,41}, Qing Zhu³⁶, Qiuhan Zhu⁴², Qianlai Zhuang²⁶

1 Department of Geography, National University of Singapore, Singapore.

2 Stanford Doerr School of Sustainability, Department of Earth System Science, Stanford, CA, USA.

3 Department of Earth System Science, Woods Institute for the Environment, and Precourt Institute for Energy, Stanford University, Stanford, CA, USA.

4 Laboratoire des Sciences du Climat et de l'Environnement (LSCE), CEA, CNRS, UVSQ, Université Paris Saclay, Gif-sur-Yvette, France.

5 Spark Climate Solutions, San Francisco, CA, USA.

6 Global Carbon Project, CSIRO Environment, ACT 2601, Australia.

7 Research Institute for Global Change, JAMSTEC, 3173-25 Showa-machi, Kanazawa, Yokohama, 236-0001, Japan.

8 Seto Inland Sea Carbon Neutral Research Center (S-CNC), Hiroshima University, Higashi-Hiroshima, Hiroshima 739-8529, Japan.

9 Center for Earth System Science and Global Sustainability, Schiller Institute for Integrated Science and Society, Boston College, Chestnut Hill, MA, USA.

10 Department of Earth and Environmental Sciences, Boston College, Chestnut Hill, USA.

11 State Key Laboratory of Tibetan Plateau Earth System, Environment and Resources, and Institute of Tibetan Plateau Research, Chinese Academy of Sciences, Beijing, China.

12 Earth System Science Division, Pacific Northwest National Laboratory, Richland, WA, USA.

13 College of Forestry, Wildlife and Environment, Auburn University, Auburn, AL, USA.

14 School of Biosciences, University of Sheffield, U.K.

15 Center for Environmental Remote Sensing, Chiba University, Chiba, Japan.

16 MetOffice Hadley Centre, Joint Centre for Hydrometeorological Research, Wallingford, U.K.

17 School of Geography, Earth & Environmental Sciences, University of Birmingham, U.K.

18 Graduate School of Science and Engineering, Chiba University, 1-33, Yayoi-cho, Inage-ku, Japan.

19 Graduate School of Agricultural and Life Sciences, The University of Tokyo, Tokyo, Japan.

20 Department of Climate, Meteorology, and Atmospheric Sciences, University of Illinois at Urbana-Champaign, Urbana, IL, USA.

21 Department of Earth and Atmospheric Sciences, City University of New York, New York, NY, USA.

22 Climate and Environmental Physics, Physics Institute and Oeschger Centre for Climate Change Research, University of Bern, Bern, Switzerland.

23 Max Planck Institute for Meteorology, Hamburg, Germany.

- 45 24 The IDEC Institute - Center for Peaceful and Sustainable Futures (CEPEAS), Network for Education and Research on
46 Peace and Sustainability (NERPS), Graduate School of Humanities and Social Sciences International Economic
47 Development Program (IEDP), Graduate School of Innovation and Practice for Smart Society (SmaSo), Hiroshima
48 University, Higashi-Hiroshima, Hiroshima 739-8529, Japan.
49 25 Key Laboratory of Atmospheric Environment and Extreme Meteorology, Institute of Atmospheric Physics, Chinese
50 Academy of Sciences, Beijing, China.
51 26 Department of Earth, Atmospheric, and Planetary Sciences, Department of Agronomy, Purdue University, West
52 Lafayette, IN, USA.
53 27 Center for Global Environmental Research, National Institute for Environmental Studies, Tsukuba, Ibaraki, Japan.
54 28 Biological Sciences Division, Pacific Northwest National Laboratory, Richland, WA, USA.
55 29 Climate Research Division, Environment and Climate Change Canada, Victoria, BC, Canada.
56 30 Meteorological Research Institute (MRI), Nagamine 1-1, Tsukuba, Ibaraki 305-0052, Japan.
57 31 Sino-French Institute for Earth System Science, College of Urban and Environmental Sciences, Peking University,
58 Beijing 100871, China
59 32 Department of Biology Sciences, University of Quebec at Montreal, C.P. 8888, Succ. Centre-Ville, Montreal, QC H3C
60 3P8, Canada.
61 33 College of Geographic Science, Hunan Normal University, Changsha 410081, China.
62 34 School of Atmospheric Sciences, Sun Yat-sen University, and Southern Marine Science and Engineering Guangdong
63 Laboratory (Zhuhai), Zhuhai 519000, China.
64 35 School of the Environment, Yale University, New Haven, CT, USA.
65 36 Climate and Ecosystem Sciences Division, Lawrence Berkeley National Laboratory, Berkeley, CA, USA.
66 37 TNO, Department of Climate Air & Sustainability, P.O. Box 80015, NL-3508-TA, Utrecht, the Netherlands.
67 38 Finnish Meteorological Institute, P.O. Box 503, 00101, Helsinki, Finland.
68 39 Department of Earth and Atmospheric Sciences, University of Houston, Houston, TX, USA.
69 40 Institute of Environment and Ecology, Tsinghua Shenzhen International Graduate School, Tsinghua University, Shenzhen
70 518055, China.
71 41 State Environmental Protection Key Laboratory of Sources and Control of Air Pollution Complex, Beijing 100084,
72 China.
73 42 College of Hydrology and Water Resources, Hohai University, Nanjing 210098, China.

74
75 *Correspondence to:* Mengze Li (mengze@nus.edu.sg)
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81 **Abstract.** Wetlands are the largest natural source of atmospheric methane (CH₄), yet comprehensive
82 global budgets are typically delayed by years, preventing a timely understanding of CH₄ sources, sinks,
83 and trends. To reduce this delay, we present a model emulator-driven framework and accompanying
84 workflow that enable timely, continuous emission updates using a machine-learning emulator to
85 reconstruct spatially explicit monthly emission fields at 1°x1° resolution. We apply this framework to a
86 global dataset of natural vegetated wetland CH₄ emissions to extend the most recent Global Methane
87 Budget (GMB; Saunio et al., 2025) record that covers the 2000-2020 emissions through 2025. In the
88 test data (~30% of the total dataset), the emulator achieved a global R² of 0.65 ± 0.003 (mean ± 95% CI,
89 hereafter) and an RMSE of 5.49 ± 0.12 × 10⁻³ Tg CH₄/year. The emulator is trained on 35 GMB model
90 estimates, including 22 process-based models and 13 atmospheric inversions, paired with 10 ensemble
91 realizations of 11 gridded climate predictor variables from atmospheric reanalyses. Our results show

92 that the global mean predicted wetland CH₄ emissions for 2021-2025 (157.8 ± 2.4 Tg CH₄/year) are not
93 significantly higher (~ 0.05 Tg CH₄/year) than the 2000-2020 baseline. However, this stability masks a
94 significant hemispheric redistribution of emissions. We detect an increase in Northern Hemisphere
95 (NH) emissions in 2021-2025, with mid- and high-latitudes increasing by 0.76 ± 0.07 and 0.35 ± 0.03
96 Tg CH₄/year, respectively, while the tropics and Southern Hemisphere (SH) extratropics show
97 offsetting negative trends (-0.95 ± 0.19 and -0.11 ± 0.02 Tg CH₄/year, respectively). The predicted
98 emissions are able to capture the low emissions in 2023 in South America linked to El Niño-related
99 drought, as reported by recent studies (Ciais et al., 2026; Quinn et al., 2025). Furthermore, we identify a
100 distinct seasonal amplification of global emission trends that peaks in late boreal summer. This new
101 modeled dataset and operational framework bridge the gap between the latest updated budgets and low-
102 latency monitoring, providing a scalable capacity to frequently update global emission estimates and
103 critical early warnings of regional wetland feedback loops. The data are publicly available at
104 <https://doi.org/10.5281/zenodo.18870108> (Li et al., 2026a).
105

106 **1 Introduction**

107 Wetlands are the largest natural source of methane (CH₄), a potent greenhouse gas, yet they remain one
108 of the most uncertain components of the global CH₄ budget (Saunois et al., 2025; Zhu et al., 2025).
109 They account for 25-30% of global CH₄ emissions and strongly influence the global carbon cycle
110 (Jackson et al., 2024; Li et al., 2026b; Peng et al., 2022; Saunois et al., 2025; Zhang et al., 2023).
111 Quantifying how wetland CH₄ emissions respond to climate variability is essential for interpreting
112 interannual variability in the global CH₄ budget and for understanding recent atmospheric CH₄ changes
113 (Ciais et al., 2026; Gedney et al., 2019a; Parker et al., 2018; Stavert et al., 2022). Community efforts
114 such as the Global Methane Budget (GMB) synthesize process-based “bottom-up” (BU) models and
115 atmospheric “top-down” (TD) inversions to estimate all CH₄ emission sources and sinks (Saunois et al.,
116 2020, 2025). However, the substantial effort required to assemble these products into a reliable
117 ensemble typically leads to lags of multiple years. For instance, the most recent GMB published in 2025
118 (Saunois et al., 2025) covered the period 2000-2020. Although low-latency estimates exist for some
119 individual models (Quinn et al., 2025), accelerating the assembly of a trustworthy multi-model
120 synthesis would better support timely emission monitoring and evidence-based assessment of mitigation
121 progress.
122

123 Wetland CH₄ emissions combine biogeochemical and hydrometeorological controls that regulate CH₄
124 production with oxidation processes within sediments and the water column. Temperature exerts a
125 strong influence on microbial methanogenesis and ecosystem respiration (Knox et al. 2019; Bansal et
126 al., 2023; Li et al., 2025; Yvon-Durocher et al., 2014), whereas soil thermal and free-thaw conditions
127 can modulate growing-season length and cold-season CH₄ dynamics (Hyvärinen et al., 2025; Zona et
128 al., 2016). Hydrology, often summarized by water table depth, soil moisture content, and inundation
129 extent, controls oxygen availability and redox conditions and influences wetland CH₄ production and
130 oxidation (Cui et al., 2024; He et al., 2025; Knox et al., 2021). Precipitation, lateral water transport, and

131 surface energy fluxes jointly constrain water balance through inputs and evapotranspiration (Aalto et al.,
132 2025; Helbig et al., 2020; Tyystjärvi et al., 2024). Furthermore, radiation and vegetation state provide
133 proxies of substrate supply and plant-mediated transport pathways that connect below-ground CH₄
134 production to the atmosphere (Helfter et al., 2022; McNicol et al., 2023).

135
136 Currently, global wetland CH₄ emissions are primarily estimated using two approaches: BU and TD
137 models. Process-based BU models explicitly simulate the complex biogeochemical mechanisms driven
138 by environmental conditions (Zhang et al., 2025). TD atmospheric inversions estimate surface fluxes by
139 optimising prior emission inventories to match observed atmospheric CH₄ concentrations (Patra et al.,
140 2018). Recent studies have increasingly turned to data-driven machine-learning models. For instance,
141 McNicol et al. (2023) upscaled eddy covariance CH₄ fluxes globally using random forest algorithms;
142 Bernard et al. (2025) introduced a satellite observation-based model to simulate temporal emission
143 variability. These machine-learning approaches offer some advantages, including rapid operational
144 speeds and lower computational costs. However, they are constrained by the sparsity of training data,
145 particularly in the tropics and high latitudes.

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147 Building on these well-established controls and the advancements in data-driven modeling, we provide
148 a low-latency continuation of the GMB gridded wetland CH₄ emissions by extending monthly
149 emissions from 2000-2020 through 2025 at 1°x1° resolution. Our approach is supported by the fact that
150 many key climate, hydrometeorological and vegetation drivers are routinely updated as global gridded
151 datasets, enabling a practical emulator-based pathway for timely wetland CH₄ emission estimates. In
152 this study, wetlands are defined following the GMB natural vegetated wetland category, excluding
153 lakes, rivers, reservoirs, coastal waters and managed sources (Saunois et al., 2025). We emulate each
154 BU and TD ensemble member from GMB using machine-learning models trained on the 2000-2020
155 GMB wetland emission fields and climate predictor variables (e.g., soil temperature and precipitation)
156 from ERA5 reanalysis to yield spatially explicit monthly emissions for 2000-2025 with uncertainty
157 derived from both GMB ensemble spread (35 runs) and ERA5 monthly reanalysis (10 ensemble
158 members). Using the resulting 2000-2025 wetland CH₄ emission data, we then quantify recent emission
159 changes in 2021-2025 relative to 2000-2020, characterize interannual variability in annual regional and
160 latitudinal emission time series, and diagnose long-term growth rates and their seasonality.

161
162 Our proposed framework and dataset provide a lower-latency, scalable tool for spatially explicit
163 wetland CH₄ emission estimates. It provides support for early-warning diagnostics, attribution of recent
164 atmospheric CH₄ anomalies, and provision of timely priors for atmospheric inversions.
165

166 2 Methods

167 2.1 Input datasets

168 2.1.1 Wetland CH₄ flux data

169 We used monthly net CH₄ flux estimates for natural vegetated wetlands from the most recent Global
170 Methane Budget (GMB) synthesis for 2000-2020, including both process-based BU models and
171 atmospheric TD inversions (Saunois et al., 2025; Zhang et al., 2025). In this study, the natural vegetated
172 wetland definition excludes lakes, rivers, reservoirs, coastal waters, and managed sources, following the
173 underlying GMB definition (Saunois et al., 2025). Detailed descriptions of each model and inversion
174 framework are provided in the cited GMB publications; here we summarize the key elements relevant to
175 this study.

176
177 The BU wetland biogeochemical models (22 estimates, Table S1) included in the analysis are from 11
178 models: CLASSIC (Arora et al., 2018; Melton and Arora, 2016), ELM-ECA (Riley et al., 2011), ISAM
179 (Shu et al., 2020; Xu et al., 2021), JSBACH (Kleinen et al., 2020, 2021, 2023), JULES (Gedney et al.,
180 2019b), LPJ-MPI (Kleinen et al., 2012), LPJ-WSL (Zhang et al., 2016), LPX-Bern (Spahni et al., 2011;
181 Stocker et al., 2014), ORCHIDEE (Ringeval et al., 2011), SDGVM (Beerling and Woodward, 2001;
182 Hopcroft et al., 2011, 2020), and VISIT (Ito and Inatomi, 2012). Each model was run under two global
183 climate forcings (CRU and GSWP3-W5E5). All BU simulations are prognostic, i.e. each model
184 computes wetland extent internally rather than relying on a shared wetland extent dataset. Differences in
185 simulated wetland extent therefore contribute to inter-model spread in CH₄ emissions and may explain a
186 substantial fraction of discrepancies across regions.

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188 The TD inversion products (13 estimates, Table S2) were constrained by surface or satellite
189 observations over the 2000-2020 period and they are from seven systems: CarbonTracker Europe-CH₄
190 (Tsuruta et al., 2017), LMDz-CIF (Thanwerdas et al., 2022), LMDz-PYVAR (Zheng et al., 2018a, b,
191 2019), MIROC4-ACTM (Chandra et al., 2021; Patra et al., 2018), NISMOM-CH₄ (Niwa et al., 2022,
192 2025), and NIES-TM-FLEXPART (NFVAR) (Maksyutov et al., 2021; Wang et al., 2019). Two
193 alternative anthropogenic prior inventories, EDGAR v6 (Crippa et al., 2021) and GAINS (Höglund-
194 Isaksson et al., 2020), were used to represent anthropogenic emissions. Wetland and inland freshwater
195 emissions were prescribed as separate prior fluxes in the inversion framework. Wetland priors were
196 derived from the ensemble mean of dynamic process-based BU wetland models used in the GMB
197 estimates, whereas inland freshwater emissions were based on independent BU estimates. We used
198 posterior flux estimates throughout this study.

199
200 All BU and TD outputs provide monthly wetland CH₄ flux as emission per grid cell area, typically at
201 1°x1° resolution in the regridded GMB output products. Coarser products (CLASSIC models at 1.85°
202 and JSBACH models at 2.8°) were remapped to a 1°x1° grid for spatial consistency across datasets. In
203 total, we included 22 BU estimates and 13 TD estimates (35 estimates overall). Tables S1 and S2
204 summarize the BU and TD estimates.

206 **2.1.2 Predictor variables from ERA5 reanalysis**

207 We assembled 10 ensemble members of global ERA5 monthly averaged data at $0.5^\circ \times 0.5^\circ$ spatial
208 resolution (Hersbach et al., 2023) for 11 climate predictor variables: t2m (near-surface air temperature),
209 tp (total precipitation), ssrd (downward shortwave radiation at surface), slhf (surface latent heat flux),
210 sshf (surface sensible heat flux), swv11 (volumetric soil water in layers 1, 0-7 cm below surface), swv12
211 (volumetric soil water in layers 2, 7-28 cm below surface), stl1 (soil temperature in layer 1, 0-7 cm
212 below surface), stl2 (soil temperature in layer 2, 7-28 cm below surface), lai_hv (high-vegetation leaf
213 area index), and lai_lv (low-vegetation leaf area index). LAI (leaf area index) from ERA5 is prescribed
214 in the ECMWF land-surface model as a seasonally varying, monthly climatology derived from satellite
215 LAI (Roberts et al., 2018). We chose ERA5 data in part because it is operationally updated regularly,
216 which enables continuous extension of predictor fields and low-latency updating of emission
217 reconstructions as new months become available. ERA5 data has also undergone extensive evaluation
218 and is widely used due to its high spatiotemporal consistency and generally strong performance
219 (Hersbach et al., 2020), but notably regional biases remain, for example, a weaker performance in parts
220 of Asia/Africa for hydrology (Gebrechorkos et al., 2024) and temperature in the Arctic (Tian et al.,
221 2024). The 10 ensemble members correspond to the ERA5 Ensemble of Data Assimilations, which
222 provides multiple physically consistent realizations of the reanalysis by perturbing the data assimilation
223 system to sample uncertainty in atmospheric and land-surface states. The ensemble spread can be
224 interpreted as an indicator of reanalysis uncertainty.

225
226 The 11 variables included in this analysis represent first-order climatic and biophysical controls on
227 wetland CH₄ production, oxidation, and transport. Their relevance to wetland CH₄ emissions is well
228 established in the literature (Knox et al., 2021; Li et al., 2024, 2025; McNicol et al., 2023; Pu et al.,
229 2024; Toet et al., 2011; Yuan et al., 2022, 2024; Zhang et al., 2025). To represent the ecosystem
230 memory effect in response to environmental conditions (Chen et al., 2025), we also included 1-month
231 and 2-month lagged versions of each variable. While introducing short lags would increase collinearity
232 in the machine-learning model, the risk of overfitting is reduced by our model architecture (Section
233 2.2.2, including feature screening, regularization, subsampling, and early stopping). The data are
234 regridded using bilinear interpolation to $1^\circ \times 1^\circ$ resolution to match the wetland CH₄ flux data from the
235 regridded GMB models for further machine-learning modeling.

237 **2.2 XGBoost model**

238 **2.2.1 Data assembly and filtering**

239 For each of 350 model-ensemble pairs (35 GMB model estimates \times 10 ERA5 ensemble members), we
240 constructed gridded monthly time series pairing GMB data with a set of 35 predictor variables: 11
241 contemporaneous ERA5 variables, 22 lagged variables (1- and 2-month lags for each of the 11 ERA5
242 variables), and two seasonal encoding terms:

$$m_{\sin}=\sin(2\pi m/12), m_{\cos}=\cos(2\pi m/12) \quad (\text{Eq.1})$$

Where:

m=month of year (1-12, 1=January and 12=December);

π =the constant pi;

m_{\sin} and m_{\cos} are the sine/cosine transforms of m and represent seasonality.

We optimised our analyses to reduce noise and computational burden by retaining only grid cells with substantial wetland emissions, defined as those maintaining a mean flux of $\geq 10^{-15}$ kg CH₄/m²/s across all GMB model estimates over 2000-2020 (30,000 times lower than the global mean flux).

2.2.2 Model architecture

To systematically reconstruct wetland CH₄ fluxes, we developed a machine-learning pipeline, training independent Extreme Gradient Boosting (XGBoost) models (Chen and Guestrin, 2016) for each grid cell and model-ensemble pair. XGBoost has been applied as a surrogate model for ensemble estimates of GMB BU models with high performance (Zhu et al., 2024). The overall modeling workflow consisted of four stages: (1) data splitting, (2) feature screening, (3) hyperparameter running, and (4) final model fitting and prediction. Each stage is introduced as follows.

Each model-ensemble pair dataset was temporally split into a training set and two testing sets to assess reconstruction capability across different climatic periods at monthly scale. The training data covers January 2003 to December 2017. The testing data covers two periods: January 2000-December 2002 and January 2018-December 2020. Two test windows were used only for out-of-sample evaluation and were not seen during fitting or tuning. The final 24 months of the training period (January 2016 to December 2017) were reserved as a validation set for early stopping and hyperparameter selection. This splitting strategy is to test the model's ability to capture long-term climate variabilities in early/late periods and reduce overfitting.

Prior to XGBoost training, we applied Boruta feature screening (Kursa and Rudnicki, 2010) on the training subset to reduce the effective dimensionality of the predictor set and mitigate overfitting risk arising from the relatively large number of candidate predictors. Boruta augments the predictor matrix with 'shadow' features formed by randomly permuting each predictor across samples, thereby preserving each predictor's marginal distribution while separating its association with CH₄ flux. A random forest regressor is fit to the combined real/shadow feature set and feature importances are compared. Predictor variables were retained if their random forest importance exceeded the maximum shadow-feature importance in more than 20% of 10 iterations (Boruta-style screening), while seasonal terms were always retained. The resulting per-grid cell feature subset was then used for both XGBoost fitting and subsequent reconstructions, ensuring that training and prediction used an identical, cell-specific predictor variable set.

282 After the feature screening, we performed a grid search over six hyperparameter configurations
283 (learning rate, maximum tree depth, minimum child weight, subsampling ratio, column subsampling per
284 tree, and L2 regularization) (Table S3). For each configuration, training stopped if the validation Root
285 Mean Squared Error (RMSE) did not improve for 50 consecutive boosting rounds. Once the optimal
286 hyperparameter set was identified for a grid cell, the model was refitted on the full training data using
287 the optimal number of boosting iterations selected during early stopping.

288
289 After that, the saved booster was then applied to predict wetland CH₄ flux 2000-2025 (including 2021-
290 2025, which extends beyond the most recent GMB estimates period). All the predicted wetland fluxes
291 (unit of kg CH₄/m²/s) were converted to emissions (Tg CH₄/month) by multiplying the fluxes by grid-
292 cell area (m²) and total number of seconds (s) in each month.

293
294 We recognize that many of these variables are multicollinear, an issue that XGBoost is considered
295 robust against. Nevertheless, we minimized overfitting by constraining model complexity (feature
296 screening, regularized trees with subsampling), using a held-out validation window with early stopping
297 to select the optimal number of boosting iterations, and evaluating performance on temporally separated
298 test periods (2000-2002, 2018-2020) not seen during training. In addition, using the 10 ERA5 ensemble
299 members allows us to propagate reanalysis uncertainty in the predictor fields through the reconstruction.

301 **2.2.3 Model evaluation**

302 Model performance was evaluated for each grid cell per model-ensemble pair by computing the R²,
303 RMSE, and NRMSE (normalised RMSE, computed by dividing the RMSE by the standard deviation of
304 emissions) values on the withheld test periods. The mean R², RMSE and NRMSE of pairs per grid cell
305 are applied to evaluate model performance at the grid cell level (Figure 1). To assess the model
306 reconstruction skills for grid cells with different emission magnitudes, we grouped grid cells into (1)
307 emission percentile bins based on their total wetland CH₄ emissions during 2000-2020 from GMB
308 estimates, (2) 18 geographical regions defined in Saunio et al. (2025) (they are USA, Canada, Central
309 America, Northern South America, Brazil, Southwest South America, Europe, Northern Africa,
310 Equatorial Africa, Southern Africa, Russia, Central Asia, Middle East, China, Korea-Japan, South Asia,
311 Southeast Asia, and Australasia), and (3) five latitude bands: global, Northern Hemisphere (NH) high-
312 latitudes (60°N-90°N), NH mid-latitudes (30°N-60°N), tropics (30°N-30°S), and Southern Hemisphere
313 (SH) extratropics (30°S-90°S). We then compared the mean R² and RMSE values of each emission bin
314 (0-10% denotes the highest emitting grid cells) and 18 geographical regions (Figure S1, S2). Monthly
315 mean R² and RMSE at five latitude bands over test periods are presented to study the temporal variation
316 of model performance (Figure S3, S4). Detrended monthly global emissions from GMB and predictions
317 were compared to assess the model performance in capturing interannual variability (Figure S5).
318 Detrended anomalies were calculated by fitting a linear least-squares trend to each monthly time series
319 over 2000-2020 and subtracting the corresponding best-fit linear trend from the original series. We
320 generated out-of-sample wetland CH₄ emission predictions (CH₄_pred) for the test periods to compare

321 with the emissions from GMB models (CH₄_GMB) (Figure 2). This ensured that the model predicted
322 CH₄_pred based on the trained model without having seen the CH₄_GMB in the test data.
323

324 **2.3 Emission trend analyses**

325 We quantified spatial patterns, regional aggregates and temporal emission trends in reconstructed
326 wetland CH₄ emissions for 2000-2025, and compared them with corresponding GMB estimates over
327 2000-2020. We performed analyses for 18 geographic regions and for five latitude bands.
328

329 We analyzed recent emission changes by comparing 2021-2025 against 2000-2020 baseline at both
330 grid-cell and regional scales (Figures 3, 4). For Figure 3, gridded monthly emissions were aggregated to
331 per-grid mean emissions for each period (2021-2025 and 2000-2020). Emissions presented in Figure 4
332 were summed across grid cells within each region. Emission changes between the two periods were
333 calculated as paired differences for each grid cell and model–ensemble pair. Therefore, the reported
334 mean and 95% confidence interval for emission changes reflect the uncertainty of the paired differences
335 themselves. We examined interannual variability and long-term annual emission changes by
336 aggregating monthly emissions to annual totals for both the latitude bands and 18 geographic regions
337 (Figures 5, S6, S8). Ordinary least squares regression was applied to monthly emissions over 2000-2025
338 to estimate emission trends (Figure 6).
339

340 **2.4 Uncertainty analysis**

341 Emulator skill uncertainty was evaluated using held-out test period R², RMSE, NRMSE and prediction
342 residuals relative to the GMB fields (Sections 2.2.3 and 3.1). Uncertainty in reconstructed wetland CH₄
343 emissions was estimated from the ensemble spread across all reconstructions generated from the 35
344 GMB estimates and 10 ERA5 ensemble members. This propagated two sources of uncertainty through
345 the framework: (i) spread in the underlying GMB estimates and (ii) uncertainty in the predictors
346 represented by the ERA5 ensemble. We use 95% confidence interval (CI) to represent this emission
347 uncertainty.
348

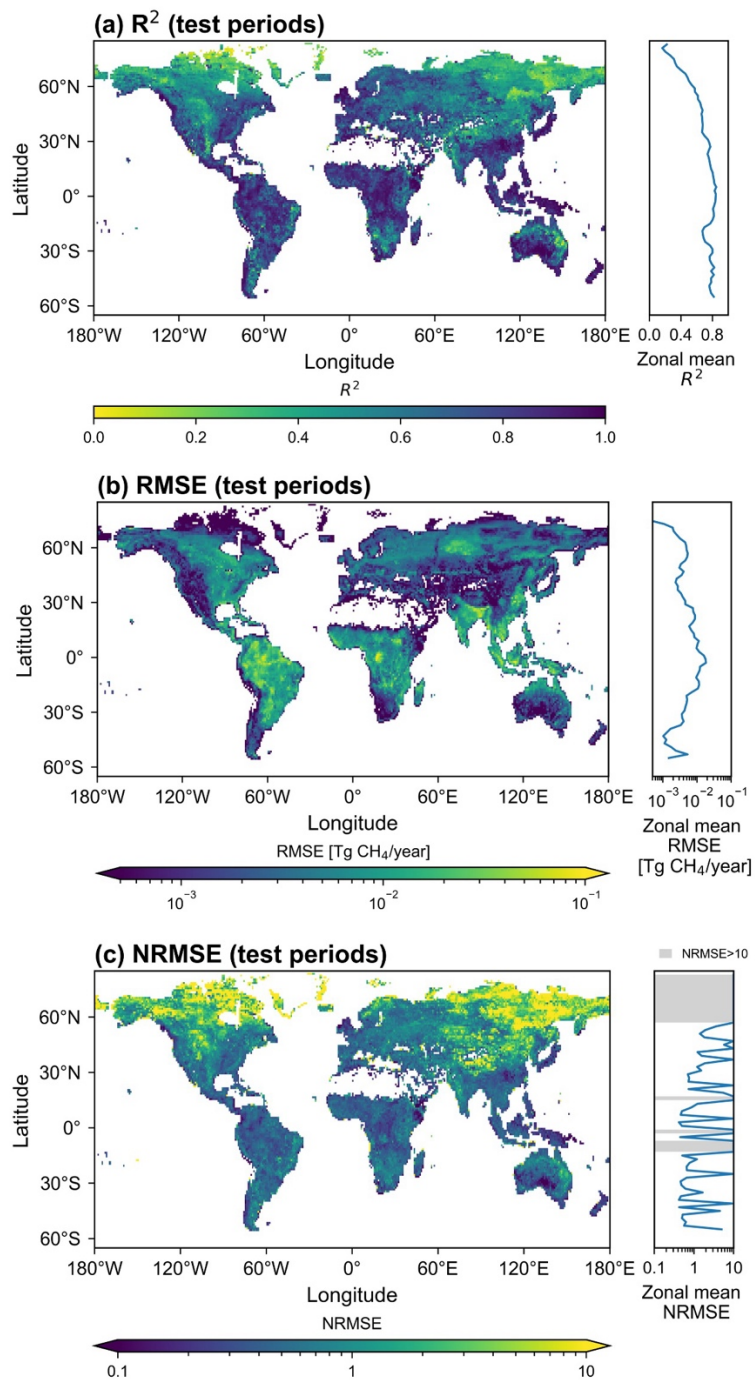
349 **3 Results and Discussion**

350 **3.1 Model performance evaluation**

351 We begin by evaluating model performance at the spatial scales most relevant to CH₄ budgeting.
352 Aggregated annual wetland CH₄ emissions at the global and latitudinal-band scales show that the
353 emulator reproduces the overall magnitude, temporal variability, and broad zonal structure of the GMB
354 estimates over 2000-2020 (Figure S6), providing a large-scale context for the finer grid-cell evaluation
355 presented below.

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Across the two test windows (2000-2002 and 2018-2020), the XGBoost model reproduces monthly wetland CH₄ emissions well. Globally, the coefficient of determination R² of 0.65 ± 0.003 (mean \pm 95% CI, hereafter; Figure 1a), which refers to the mean grid-cell R² across all global land cells over the two test periods, with the 95% confidence interval calculated across grid cells. This reported reconstructive skill indicates strong predictability from predictor variables. However, the model performance varies geographically. Mean R² is higher in the Southern Hemisphere (SH) than in the Northern Hemisphere (NH) (0.77 ± 0.006 compared with 0.62 ± 0.004 , respectively), and higher in the tropics than in the extratropics (0.78 ± 0.004 and 0.58 ± 0.004 , respectively). In the top emitting regions, mean R² values are generally higher than in regions with lower emissions (Figure S1a). Global RMSE over test periods is $5.49 \pm 0.12 \times 10^{-3}$ Tg CH₄/year (Figure 1b). RMSE for the tropics is about twice that of the RMSE for the extratropics ($9.37 \pm 0.27 \times 10^{-3}$ and $3.35 \pm 0.08 \times 10^{-3}$ Tg CH₄/year, respectively). Top emitting regions exhibit higher RMSE, e.g., RMSE doubled for top 10% emitting regions compared with top 10-20% emitting regions (Figure S1b). Regionally, mean R² values are highest in Southeast Asia and Korea-Japan (>0.8), and lowest in Russia and Canada (~ 0.5) (Figure S2a). Mean RMSE is highest in the Americas (Northern South America and Central America), and lowest in the Middle East and Central Asia. Generally, NRMSE is inversely correlated to R², indicating a better performance of our model in top emitting regions (Figure 1c). Temporal variation of model performance reveals the lowest R² in Jan/Feb in NH high latitudes (Figure S3), and the highest RMSE in July-Sep in the tropics (Figure S4). We further evaluate how these spatial and temporal differences in model skill influence long-term variability in annual emissions in Section 3.3.

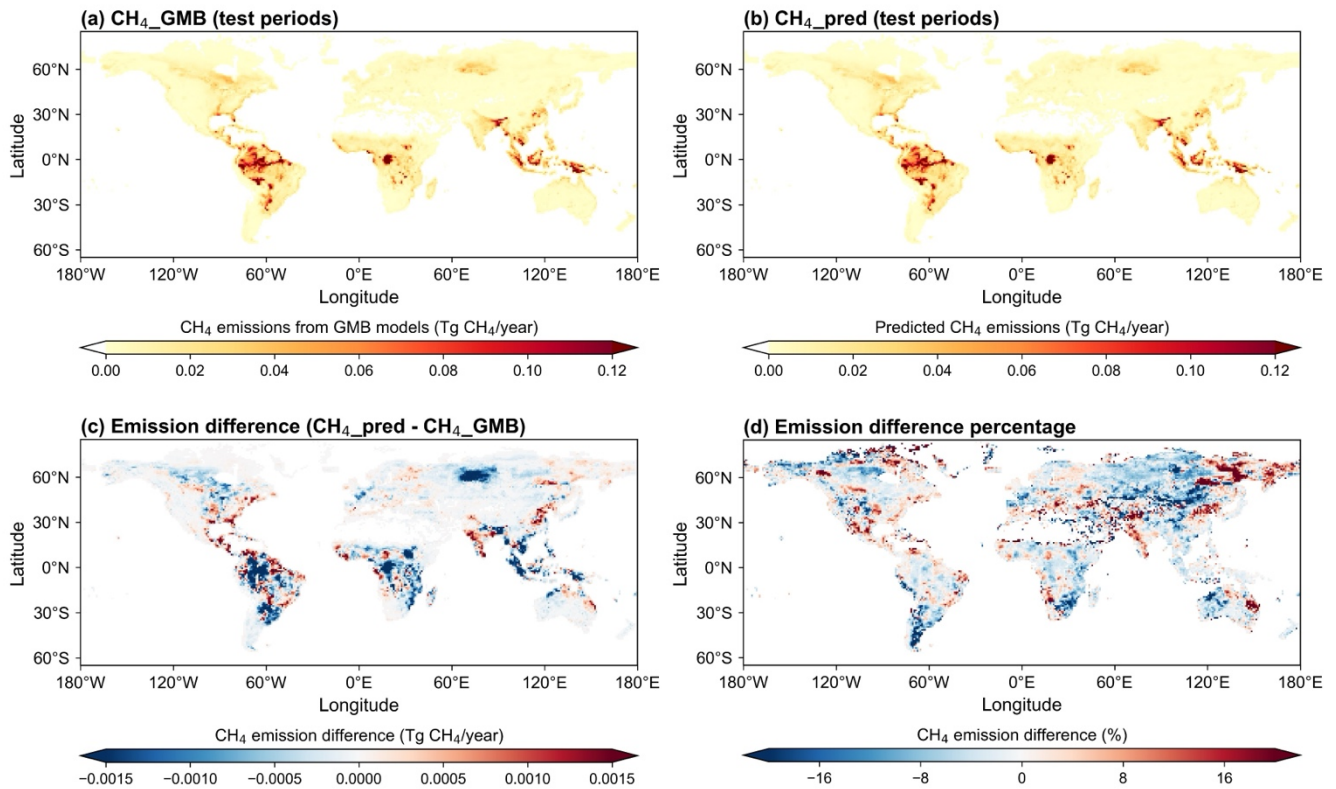


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Figure 1. Predictive skill of the XGBoost wetland CH_4 model (per-grid-cell R^2 , RMSE, and NRMSE over the test periods). Insets indicate the zonal mean R^2 , RMSE and NRMSE in 2° latitude bands.

382 To examine predictive uncertainty, we compare wetland CH₄ emissions from the GMB models
383 (CH₄_GMB; Figure 2a) with XGBoost predictions (CH₄_pred; Figure 2b) over the two test periods.
384 Globally, CH₄_pred reproduces both the magnitude and spatial pattern of CH₄_GMB, including the
385 major hotspots concentrated in the tropics, such as Amazon Basin (central-western Amazon floodplains,
386 Pantanal and Moxos plains), Congo Basin, the Sudd wetlands in South Sudan, and Southeast Asia
387 (Sumatra, Borneo, Peninsular Malaysia, and New Guinea), with additional hotspots in NH high
388 latitudes, such as West Siberian Plain and Hudson/James Bay Lowlands.

389
390 The differences in emissions (Figure 2c) and the corresponding percent emission differences (Figure 2d)
391 between CH₄_GMB and CH₄_pred highlight spatial biases. During the test periods, the global mean
392 predicted wetland CH₄ emission is 158.5 ± 2.1 Tg CH₄/year, which is 2.3 Tg CH₄/year lower than the
393 global mean of CH₄_GMB (160.7 ± 2.2 Tg CH₄/year), corresponding to a 1.4% underestimation (Figure
394 2c). This difference is within the estimated uncertainty range of the two global means. Underestimation
395 is strongest over high-emission systems, such as the central-western Amazon Basin, Southeast Asia,
396 Congo Basin and West Siberian Plain (Figure 2c), where our model predicted emissions are lower than
397 GMB estimates, but percent differences are smaller because the denominator (CH₄_GMB) is large. For
398 example, in Southeast Asia, the mean bias is -0.44 Tg CH₄/year, yet the relative bias is only -1.7%
399 given a high regional mean emission (CH₄_GMB of 25.2 Tg CH₄/year). Monthly anomalies in global
400 emissions reveal small disagreements between the predictions and GMB estimates ($\sim \pm 4$ Tg CH₄/year)
401 (Figure S5). Largest discrepancies occur in July/August 2020 (underestimation of 15 Tg CH₄/year).
402



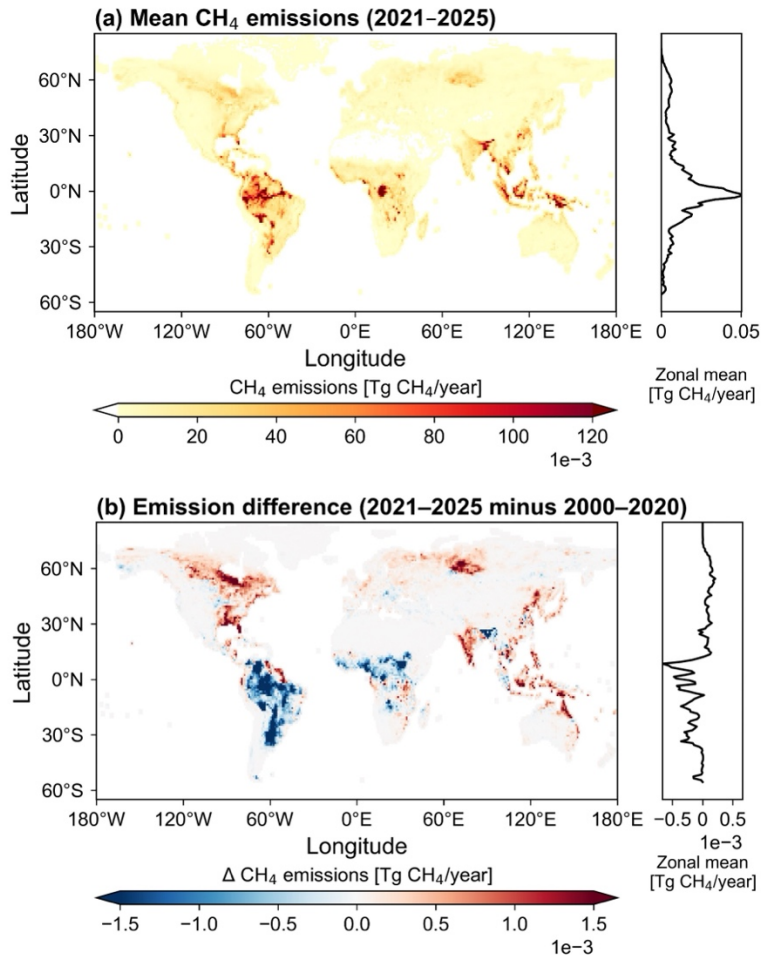
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Figure 2. Wetland CH₄ emissions (per 1°x1° grid cell) from GMB models and XGBoost predictions for two test periods combined. (a) Mean CH₄_GMB (GMB emissions) over all test months. (b) Mean CH₄_pred (XGBoost prediction) over all test months. (c) Differences in mean CH₄_GMB and mean CH₄_pred. (d) Percent emission difference between mean CH₄_GMB and mean CH₄_pred. Reds in (c) and (d) indicate higher emissions in CH₄_pred, blues lower emissions.

412 3.2 Emission trends and predicted anomalies in recent years (2021-2025)

413 Figure 3 shows spatial maps of global predicted mean wetland CH₄ emissions for 2021-2025 (Figure
414 3a) and for 2000-2020 (Figure 3b). The 2021-2025 mean emission is 157.8 ± 2.4 Tg CH₄/year ($z=3.72$,
415 with the z-score indicating how far each region's mean emission estimate deviates from the cross-region
416 mean, measured in standard deviation units), representing a non-significant increase of 0.05 Tg
417 CH₄/year relative to the 2000-2020 mean. However, changes are more pronounced across latitudes. NH
418 mid- and high-latitude wetland emissions increase during 2021-2025 by 0.76 ± 0.07 (from 29.9 ± 1.1
419 in 2000-2020 to 30.7 ± 1.2 in 2021-2025, $z=2.21$) and 0.35 ± 0.03 (from 11.0 ± 0.7 to 11.4 ± 0.7 Tg
420 CH₄/year, $z=1.01$), respectively, compared with 2000-2020 emissions. These changes correspond to
421 $\sim 2.5\%$ and $\sim 0.8\%$ increase. The continued growth in these regions in 2021-2025 (relative to 2000-

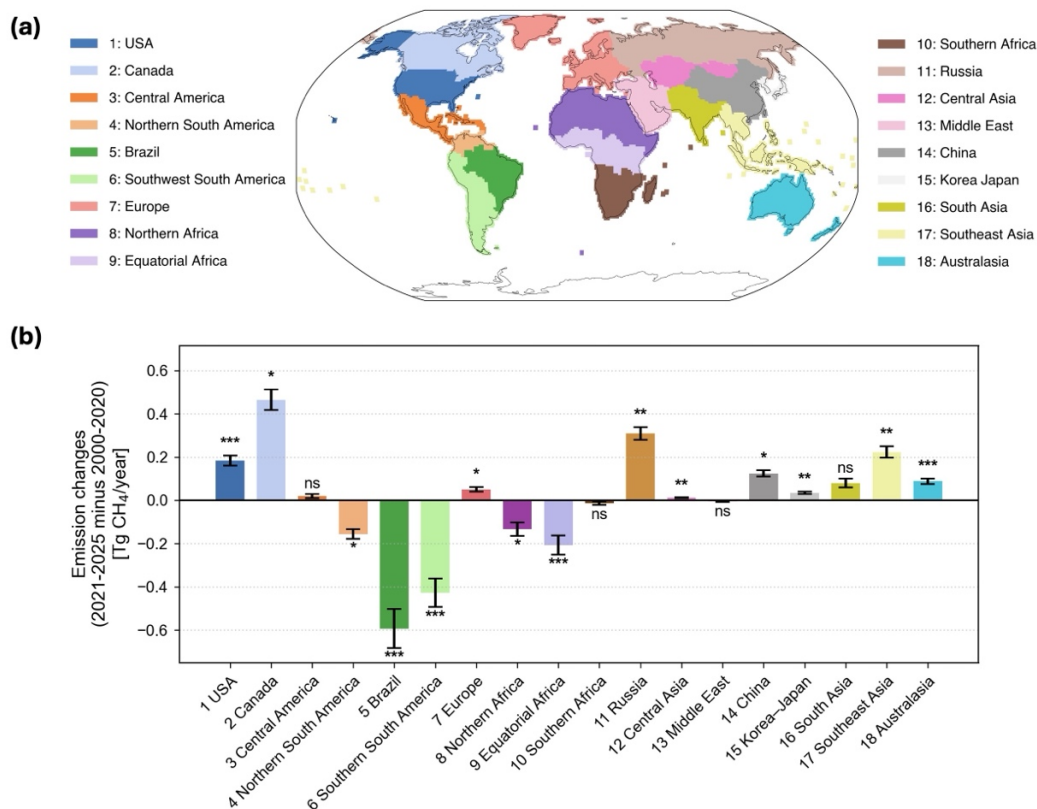
422 2020), together with increases reported for 2010-2018 relative to 1981-1989 (Feron et al., 2024),
 423 suggests an increasing role of boreal and temperate wetlands in recent decades of global warming. In
 424 contrast, emissions decrease in the tropics and SH extratropics during 2021-2025. The tropics exhibit a
 425 0.95 ± 0.19 Tg CH₄/year emission decrease ($z=-2.81$), from 113.9 ± 2.5 (2000-2020) to 112.9 ± 2.5 Tg
 426 CH₄/year (2021-2025), while SH extratropics emissions decline by 3.5% (-0.11 ± 0.02 Tg CH₄/year, $z=-$
 427 0.34) from 3.1 ± 0.1 (2000-2020) to 3.0 ± 0.1 Tg CH₄/year (2021-2025).
 428



429
 430 Figure 3. (a) Mean XGBoost model predicted wetland CH₄ emissions for 2021-2025. (b) Difference between
 431 mean emissions in 2021-2025 and 2000-2020. Reds indicate higher emissions in 2021-2025, blues lower. Insets
 432 indicate the zonal mean in 2° latitude bands.
 433

434
 435 We further assess regional wetland CH₄ emission changes in 2021-2025 compared with 2000-2020
 436 across 18 regions using XGBoost predictions (Figure 4) across 18 regions. Four regions show no
 437 significant change ($p \geq 0.05$ from two-sided paired significance test, labelled “ns”). Among the 14

438 regions with significant changes, nine regions show increases and five show decreases (Northern South
 439 America, Brazil, Southwest South America, Northern Africa, and Equatorial Africa). Notably, all
 440 regions with decreases are located in South America and Africa. The largest emission changes (2021-
 441 2025 relative to 2000-2020) occur in Brazil (-0.59 ± 0.09 Tg CH₄/year), Canada ($+0.46 \pm 0.05$ Tg
 442 CH₄/year), Southwest South America (-0.43 ± 0.07 Tg CH₄/year), Russia ($+0.31 \pm 0.03$ Tg CH₄/year),
 443 Southeast Asia ($+0.22 \pm 0.03$ Tg CH₄/year) and Equatorial Africa (-0.21 ± 0.04 Tg CH₄/year). The
 444 pronounced decline over Brazil possibly reflects the exceptional drought conditions in the Amazon
 445 basin during 2022-2024, including record-low river levels and anomalously warm, dry conditions
 446 (Espinoza et al., 2024). Such hydroclimatic drying reduces floodplain inundation and lowers water
 447 tables, which suppresses anaerobic conditions and thereby limits wetland methane production and
 448 emissions (Cui et al., 2024).
 449



450
 451 Figure 4. (a) Spatial distribution of the 18 regions; (b) Regional wetland CH₄ emission changes during 2021-2025
 452 relative to 2000-2020 across 18 regions using XGBoost modeled emissions. Bars show regional mean emission
 453 changes (annual mean in 2021-2025 minus that in 2000-2020). Error bars indicate 95% CI. A two-sided paired
 454 test was applied for each region to assess whether the 2021-2025 mean emission differs from the 2000-2020
 455 mean. Regions with $p \geq 0.05$ are labelled “ns”. Asterisks above bars denote significance: *** for $p < 0.001$, ** for
 456 $p < 0.01$, * for $p < 0.05$.
 457
 458

459 3.3 26-year emission trends and anomalies (2000-2025)

460 The emission predictions enable an assessment of long-term trends and interannual variability in
461 wetland emissions over 2000-2025. Annual emission time series show pronounced peaks in 2011 and
462 2016/2017 in the tropics, NH mid-latitudes, and globally (Figure S6). Monthly emission anomalies in
463 these regions (Figure S7) indicate that the annual maxima are driven by sustained positive anomalies
464 during Aug 2010-May 2011 and Aug 2016-Dec 2017, which coincide with La Niña conditions. This
465 alignment is consistent with prior studies reporting La Niña-driven wetting and expanded inundation
466 that enhance tropical wetland CH₄ emissions (Hodson et al., 2011; Lin et al., 2024; Murguia-Flores et
467 al., 2023; Zhang et al., 2020, 2018; Zhu et al., 2017), and with the broader evidence that CH₄ growth in
468 2020-2022 coincided with an unusual persistent La Niña event.

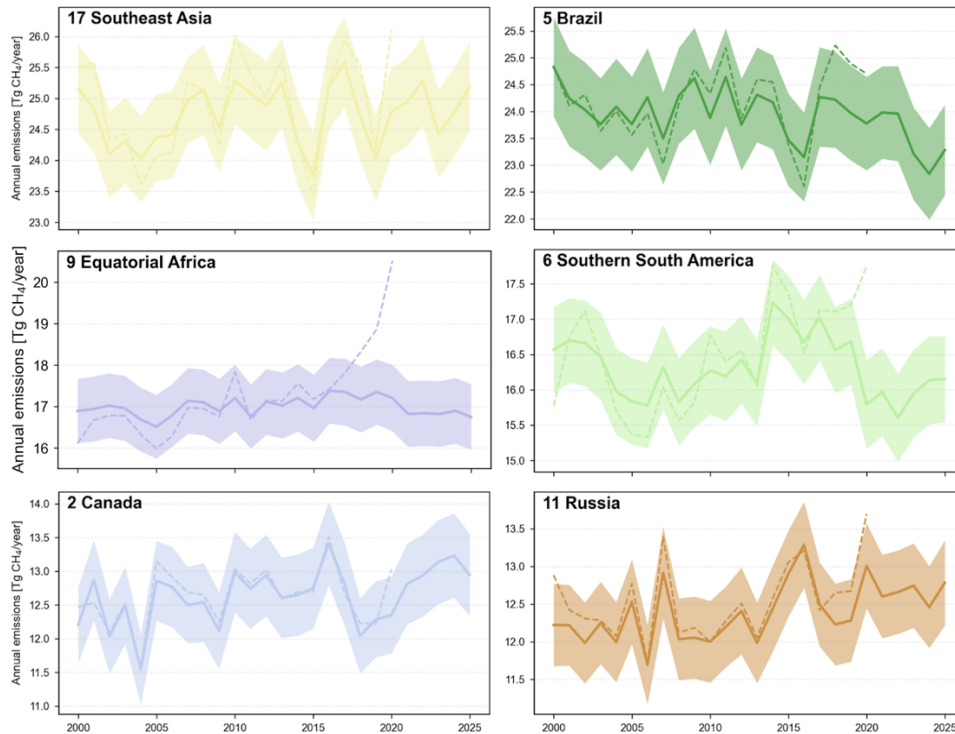
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470 In the 2020s, global and tropical annual emissions reach a minimum in 2023 (Figures S6, S7),
471 coincident with a strong El Niño event. This is supported by two recent studies reporting a sharp decline
472 in wetland emissions in South America (Ciais et al., 2026) and Amazonia (Quinn et al., 2025) in 2023
473 linked to El Niño-related drought. Our results further suggest that Brazil and Southern South America
474 are the key contributors, and both rank among the top six emitting regions globally (Figure 5). Annual
475 emissions in Brazil decline steadily across the study period, from 24.8 ± 0.9 Tg CH₄/year in 2000 to
476 23.3 ± 0.9 Tg CH₄/year in 2025, with the lowest emission in 2024 (22.8 ± 0.9 Tg CH₄/year) (Figure 5).
477 Emissions in Southern South America are relatively stable during 2000-2014, decrease markedly during
478 2014-2023 (from 17.2 ± 0.6 to 15.6 ± 0.6 Tg CH₄/year), and show a modest recovery in 2024/2025
479 (Figure 5).

480
481 The remaining four top emitting regions show contrasting behavior. Since 2019, Southeast Asia,
482 Canada and Russia show increasing emissions, whereas Equatorial Africa emissions decrease slightly
483 (Figure 5). Southeast Asian emissions fluctuate around 24.7 ± 0.5 Tg CH₄/year over 2000-2025, with a
484 minimum in 2015 (23.8 ± 0.7 Tg CH₄/year) and a maximum in 2017 (25.6 ± 0.7 Tg CH₄/year).
485 Emissions in Canada are comparatively stable during 2005-2015, while variability is larger before 2005
486 and after 2015: the amplitude reaches ~ 1.5 Tg CH₄/year in these periods, approximately double that
487 during 2005-2015 (~ 0.8 Tg CH₄/year). Equatorial African emissions remain 17.0 ± 0.2 Tg CH₄/year
488 during 2000-2025, despite two notable declines during 2002-2005 (-0.5 Tg CH₄/year) and 2019-2025 ($-$
489 0.6 Tg CH₄/year).

490
491 For regions showing increasing emissions in our XGBoost predictions (e.g., Southeast Asia, Canada,
492 and Russia), these changes are likely driven by wetter and/or warmer conditions that enhance
493 inundation, substrate supply, and CH₄ production. For instance, observed increases in Boreal-Arctic
494 wetland CH₄ emissions have been linked to warming and enhanced ecosystem productivity (Yuan et al.,
495 2024), while inversion-based studies highlight the role of La Niña-driven inundation in enhancing
496 emissions from Equatorial Asia (Qu et al., 2024; Lin et al., 2024; Feng et al., 2022). However, our
497 emulator predictions diverge from recent atmospheric inversion literature regarding African wetland
498 emissions. Multiple top-down studies indicate that enhanced emissions from tropical Africa contributed
499 substantially to the global CH₄ surge during 2020-2024 (Balasus et al., 2026; Qu et al., 2024). In

500 contrast, our climate-driven emulator predicts that emissions across Northern, Equatorial, and Southern
501 Africa remained relatively stable or slightly decreased during similar time periods (Figures 4, 5, S8).
502 We attribute this discrepancy to weaker performance and regional precipitation biases of ERA5 data
503 over parts of Africa in capturing complex African hydrological dynamics (Gebrechorkos et al., 2024).
504

505 Regional emission time series (Figures 5, S8) help explain the pronounced disagreement in 2019-2020
506 for global and tropical emissions, where predictions are substantially lower than the GMB estimates
507 (Figure S6). For 2019, the prediction mean emissions are 5.5 Tg CH₄/year below the GMB estimates for
508 global emissions (~3.5% underestimation), while emission changes in Equatorial Africa, Brazil, USA,
509 Southern Africa, and Southern South America jointly account for 4.3 Tg CH₄/year of this discrepancy.
510 In 2020, the modeled mean global emissions are 9.8 Tg CH₄/year lower than GMB (~6%
511 underestimation). The discrepancies in the tropics explain 9.2 Tg CH₄/year of the underestimation,
512 including major regions such as Eastern Africa, Southern South America, Southeast Asia, Brazil, and
513 Northern Africa. Notably, this disagreement in 2019-2020 is not evident in other latitude bands and
514 does not persist outside these years. We cautiously interpret this mismatch as the limitations of the
515 emulator and uncertainty in the underlying GMB ensemble. The climate predictors used in this study
516 can only indirectly capture ENSO (El Niño-Southern Oscillation) effects. Although the training period
517 includes several ENSO events, the unusual persistence of the 2020-2023 triple-dip La Niña may still be
518 underrepresented. In addition, most TD estimates from GMB use a similar prescribed OH (hydroxyl
519 radical) distribution, possibly causing OH-related uncertainty to be underrepresented, particularly
520 during anomalous periods such as COVID.
521



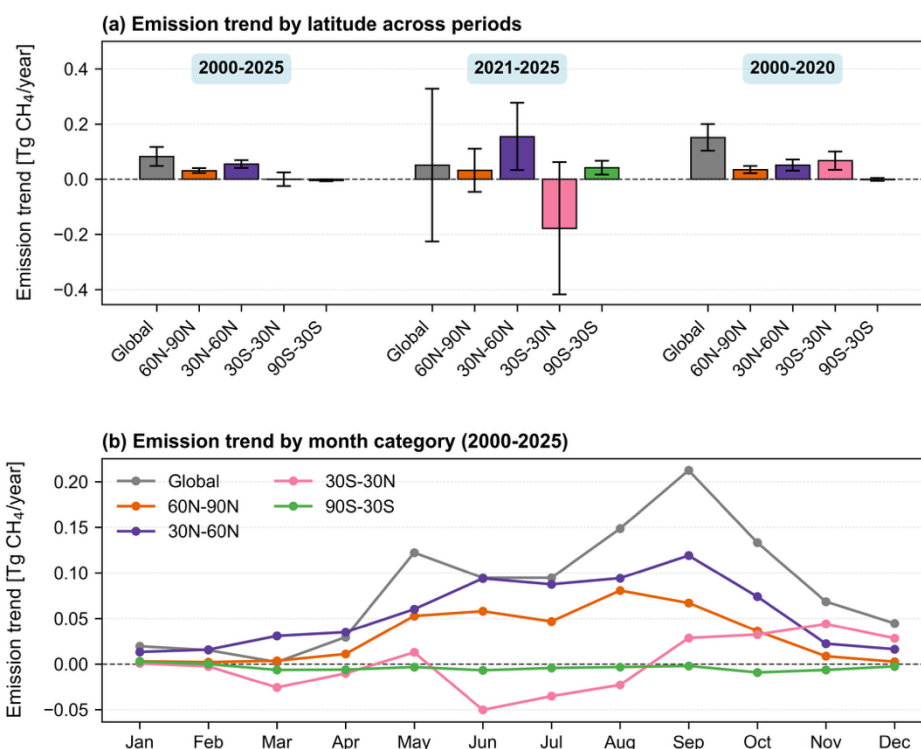
522 Figure 5. Annual wetland CH₄ emissions for the six top-emitting regions. Solid lines show model predictions of
 523 annual mean emissions, and shaded areas indicate 95% CI. Dashed lines show GMB emission estimates (mean).
 524 Panel titles indicate the region code and region name, which correspond to the region definitions shown in Figure
 525 4.
 526

527
 528 Finally, we examine trends in predicted monthly wetland CH₄ emissions across five latitude bands
 529 (Figure 6a) and 12 calendar-month categories (Figure 6b). Figure 6a compares these trends across three
 530 time periods (2000-2025, 2021-2025, and 2000-2020). Figure 6b shows the trends by 12 calendar-
 531 month categories for the 2000-2025 period. We summed predicted monthly emissions across all grid
 532 cells within each latitude band to form regional monthly time series, then computed monthly emission
 533 anomalies by subtracting the 2000-2025 climatological mean. We then estimated trends (Tg CH₄/year)
 534 using ordinary least squares linear regression of the anomaly time series. We calculated a z-score for
 535 each region and period by dividing the estimated trend by its standard error to assess the statistical
 536 significance of the trends.
 537

538 The global wetland CH₄ emission trend is three times lower in 2021-2025 compared with the 2000-2020
 539 period (0.05 ± 0.28 and 0.15 ± 0.05 Tg CH₄/year, respectively) (Figure 6a). However, the magnitude of
 540 changes in emission trends in NH mid-latitudes, the tropics and SH extratropics is greater in 2021-2025
 541 than 2000-2020 level. Emission trends in NH mid-latitudes increased three times from 0.05 ± 0.02 Tg
 542 CH₄/year ($z=4.94$) in 2000-2020 to 0.16 ± 0.12 Tg CH₄/year ($z=2.49$) during 2021-2025. The tropics
 543 exhibit a reversal of their emission trends from positive in 2000-2020 (0.07 ± 0.03 Tg CH₄/year,
 544 $z=3.95$) to negative in 2021-2025 (-0.18 ± 0.24 Tg CH₄/year, $z=-1.45$). Emission trends in SH

545 extratropics increased from nearly zero in 2000-2020 ($-8.74 \times 10^{-4} \pm 5.18 \times 10^{-3}$ Tg CH₄/year, $z=-0.33$) to
 546 higher than the global average in 2021-2025 (0.04 ± 0.02 Tg CH₄/year, $z=3.30$).
 547

548 Evaluating emission trends over individual months (Figure 6b), we find that the global trend peaks in
 549 late boreal summer, with the largest increases in August and September (0.15 ± 0.18 and 0.21 ± 0.13 Tg
 550 CH₄/year, respectively). The strongest seasonal intensification occurs in NH mid-latitudes during June-
 551 September ($0.09-0.12$ Tg CH₄/year), consistent with the dominant contribution from boreal growing-
 552 season emissions. For both NH mid- and high-latitudes, growth rates are systematically higher during
 553 May-October (growing season) than during November-April. In the tropics, monthly growth reaches its
 554 highest in Oct-Dec and lowest in Jun-Aug (negative growth). SH extratropics show near-zero growth
 555 throughout the year (slightly negative).
 556



557
 558 Figure 6. Regional and seasonal trends (Tg CH₄/year) in predicted wetland CH₄ emissions. (a) Trends across five
 559 latitude bands across three time periods (2000-2025, 2021-2025, and 2000-2020). (b) Trends for each calendar
 560 month over 2000-2025. Positive trends indicate an increase in CH₄ emissions, while negative values indicate a
 561 decrease. Error bars indicate 95% CI.
 562

563 3.4 Limitation

564 A limitation concerns source attribution in the TD estimates used for training. Although the TD
 565 inversions adopted BU wetland priors that exclude inland freshwater systems, the coarse spatial

566 resolution of TD inversions can still limit the separation of natural vegetated wetland emissions from
567 nearby open-water or inland-freshwater emissions in mixed floodplain environments. This uncertainty
568 cannot be quantified explicitly in the present framework because no independent source-resolved
569 posterior diagnostic is available. Future progress is likely to come from higher-resolution atmospheric
570 inverse frameworks. Two recent studies performed 25 km resolution atmospheric inversions using
571 satellite observations in regional and global applications and improved spatial attribution (East et al.,
572 2025; Hancock et al., 2025).

573
574 Another limitation is that our framework predicts wetland emissions from climate predictors only.
575 ENSO-related impacts can be captured only indirectly, while dynamic inundation and atmospheric
576 chemistry perturbations are outside the predictor space of the model.
577

578 **4 Conclusions**

579 This study presents a framework to extend global natural wetland CH₄ emissions from 2000 through
580 2025 by applying a machine-learning emulator to provide a lower-latency release of GMB estimates to
581 estimate wetland emissions and is intended for routine updates. The framework reproduces both spatial
582 patterns and interannual variability of the GMB estimates. The extended wetland CH₄ emission record
583 shows that the magnitude of emission changes in 2021-2025 is substantially larger than the long-term
584 trend over 2000-2020, even though the global mean emission change is relatively small.
585

586 A key emerging feature is accelerated emissions estimated in the NH mid-latitudes and high-latitudes.
587 Seasonal trend analyses indicate that these increases are concentrated in the warm season (May - Oct),
588 consistent with an amplification of the seasonal cycle and a strengthening of late-summer growth,
589 highlighting Northern mid- and high-latitude wetlands as regions where intensifying trends are most
590 evident in recent years.
591

592 In contrast, the tropics show a reversal in emission trend in the post-2020s, from an emission increase
593 during 2000-2020 to a decline in 2021-2025. Nevertheless, the tropics still dominate interannual
594 variability of global wetland emissions and can strongly influence the global CH₄ budget through
595 episodic hydroclimate extremes.
596

597 More broadly, this emulator framework may also apply to other low-latency Earth system applications
598 beyond wetland CH₄ emissions, particularly for environments and pollutants that can be constrained by
599 routinely updated predictor fields. At the same time, the present results point to several opportunities for
600 improvement, including better representation of dynamic inundation and improved separation of
601 wetland and inland-freshwater emissions.
602

603 Overall, this work proposes an operational, low-latency emulator framework that provides a scalable
604 pathway to track wetland CH₄ emissions in response to climate anomalies by leveraging routinely
605 updated Earth observation reanalysis data.
606
607
608

609 **Data Availability**

610 ERA5 monthly averaged data are available at Copernicus Climate Change Service Climate Data Store
611 at <https://cds.climate.copernicus.eu/datasets/reanalysis-era5-single-levels-monthly-means?tab=overview>
612 (DOI:10.24381/cds.f17050d7) (Accessed on 22-Jan-2026) (Copernicus Climate Change Service, 2023).
613 The global natural wetland methane emission dataset generated from this study is publicly available at
614 <https://doi.org/10.5281/zenodo.18870108> (Li et al., 2026a).
615

616 **Competing Interests**

617 At least one of the (co-)authors is a member of the editorial board of Earth System Science Data.
618

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628
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