



A Comprehensive Global Aquatic N₂O Emission Database (GANED): Unravelling N₂O Emission Patterns from Different Water Bodies

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Abstract. Nitrous oxide (N₂O) is not only one of the main potent greenhouse gases, but also currently the dominant ozone-depleting substance. The quantification of N₂O emissions from aquatic ecosystems, despite their global importance, is hindered by fragmented observations, inconsistent data reporting, and pronounced spatiotemporal variability. In this study, to improve accessibility, we introduce the Global Aquatic N₂O Emission Database (GANED; <https://doi.org/10.6073/pasta/4a086e49a4f308679b951293b380e7b9>, Nazir et al., 2026), a consolidated dataset comprising 5130 N₂O concentration records and 7386 N₂O flux measurements from 3002 sites across diverse aquatic systems including rivers, streams, lakes, reservoirs, ponds, estuaries, coastal waters, and open seas. The dataset integrates information on aquatic N₂O emission from 426 peer-reviewed publications across 8 continents, covering the period 1980-2023. While the number of observations has increased substantially since 2000, spatial coverage remains uneven, with significant gaps across Africa and parts of high-latitude regions, including Antarctica and South America. Our dataset revealed a highly skewed distribution of N₂O concentration and flux across aquatic ecosystems, with rivers and streams exhibiting the most significant variability and functioning as emission hotspots. Lakes and estuaries showed moderate variability and emission levels, whereas seas and coastal waters were characterized by consistently lower values. Pearson correlation coefficient revealed a strong positive relationship of N₂O fluxes with ammonium (NH₄⁺; R = 0.943, *p* < 0.001), nitrate (NO₃⁻; R = 0.691, *p* < 0.001), and nitrite (NO₂⁻; R = 0.807, *p* < 0.001). Significant negative correlations were found with dissolved oxygen (DO; R = -0.205, *p* < 0.05), dissolved organic carbon (DOC; R = -0.977, *p* < 0.05), and salinity (R = -0.636, *p* = 0.005), while non-significant associations were observed for water temperature, total nitrogen (TN), and total phosphorus (TP). The GANED dataset facilitates improved quantification of global aquatic N₂O inventories by providing comprehensive N₂O concentrations and fluxes in water bodies, as well as metadata describing sampling location, aquatic system type, and associated environmental parameters. The magnitude and patterns of N₂O emissions from water bodies provided by the GANED database are essential in defining how these aquatic ecosystems shape our climate, refining emission estimates, identifying drivers, and guiding mitigation strategies.



1 Introduction

Nitrogen (N) processes in terrestrial and aquatic ecosystems are primary source of atmospheric nitrous oxide (N_2O), a potent long-lived greenhouse gas (GHG) with a global warming potential 273 times higher than that of carbon dioxide (CO_2), and the leading anthropogenic driver of stratospheric ozone depletion (Marzadri et al., 2021; Tian et al., 2024; Ravishankara et al., 2009; Masson-Delmotte, 2024). Human activities, particularly agricultural intensification through large-scale synthetic fertilizer production via the Haber-Bosch process, have significantly perturbed the global N cycle. As a result, the atmospheric N_2O concentration has increased by ~25% above pre-industrial levels, with a current growth rate of 1.05 ppb yr^{-1} (IPCC, 2023; Water, 2019). Over the past four decades, global N_2O emissions have significantly increased by over 24%, rising from 270 ppb to 336 ppb (Christensen and Rousk, 2024), consequently contributing to 6.4% of total GHG radiative forcing over the period of 1750-2022 (IPCC, 2022; Gong et al., 2024). Recent estimates reveal consistent upward trends in global N_2O emissions, with both bottom-up assessments reporting an increase from $17.4 \text{ Tg N yr}^{-1}$ to $18.5 \text{ Tg N yr}^{-1}$, corresponding to an average increase rate of $0.043 \text{ Tg N yr}^{-2}$ (Yao et al., 2020), and top-down estimates indicating an increase from $15.4 \text{ Tg N yr}^{-1}$ to $17.0 \text{ Tg N yr}^{-1}$ over the period 1997-2020, reflecting a higher average rate of $0.085 \text{ Tg N yr}^{-2}$ (Tian et al., 2024). These divergent rates are primarily attributed to the continuous increase in N_2O emissions from both terrestrial and aquatic sources owing to increased food, feed, and energy demands (Davidson and Kanter, 2014).

Aquatic ecosystems contributed approximately 25-32% (5.3 Tg N yr^{-1}) of the total global N_2O emissions in 2020, second only to soils (Tian et al., 2020; Zheng et al., 2022; Zhang et al., 2024). These aquatic ecosystems, including inland waters (rivers, reservoirs, lakes, ponds, streams), estuaries, coastal zones, and oceans, represent significant yet highly uncertain sources of atmospheric N_2O (Wang et al., 2021a; Duvert et al., 2025). Inland waters emitted $\sim 319.6 \pm 58.2 \text{ Gg N yr}^{-1}$ of N_2O during the 2010s, representing an increase of about 207 Gg N yr^{-1} since the 1850s, with riverine systems contributing ~80% and lakes and reservoirs ~20% (Li et al., 2024). Consequently, N inputs from both agricultural and non-agricultural sources have made rivers, lakes, and reservoirs substantial contributors to the global N_2O budget (Zheng et al., 2022; Tian et al., 2024). Global riverine emissions are estimated at $30\text{-}35 \text{ Gg N}_2\text{O-N yr}^{-1}$, representing 0.16-0.19% of dissolved inorganic nitrogen (DIN) entering streams (Hu et al., 2016). Similarly, Zhang et al. (2026) reported that aquatic N_2O emissions are highest in the downstream reaches of reservoirs and in regions with dense populations and intensive agricultural activity, notably in eastern and southern Asia, southeastern North America, and Europe.

In aquatic ecosystems, N_2O production arises from complex microbial (nitrification, denitrification) and abiotic (chemodenitrification) pathways, modulated by nutrient supply, dissolved oxygen, water temperature, N, phosphorus (P), and pH (Tian et al., 2020; Wei et al., 2022; Pan et al., 2023). Aquatic N_2O emissions regulated by these key drivers vary widely by aquatic system types. Rivers exhibit increased fluxes driven by rapid water turnover and anthropogenic N loading, exacerbating eutrophication, warming, and pollution (Lu et al., 2023). Reservoirs and lakes are governed by stratification, with hypolimnetic anoxia favoring denitrification, triggering pulse N_2O releases (He et al., 2023). Similarly, estuaries function as hotspots, where high nutrient and organic matter inputs, coupled with hydrodynamic mixing, stimulate coupled nitrification and denitrification (e.g., Yangtze River estuary; Liu et al., 2025). Further, streams and ponds, despite their small size, have also been identified as



important sources of N₂O to the atmosphere due to high per-area fluxes from intense microbial activity and recurrent anoxia (Saunio et al., 2020; Bodmer et al., 2021). Oceans and coastal seas remain the largest contributors to global aquatic N₂O emissions, with oxygen-deficient zones disproportionately responsible for more than half of the total flux (Wang et al., 2025; Rees et al., 2022). Besides being predominantly sources of N₂O, some aquatic systems exhibit modest sink potentials. In undersaturated oceanic regions, atmospheric N₂O is absorbed and microbially consumed, sequestering 0.1-0.3 Tg N yr⁻¹ (De La Paz et al., 2025; Tian et al., 2024). Inland waters may also exhibit temporary sink function under precipitation-driven anoxia, reducing N₂O to N₂ (Liang et al., 2023). Although the net sink contribution of aquatic systems is minor compared with emissions, the sink processes of N₂O prior to emission could be important for unraveling spatiotemporal variabilities as well as for regional-global predictions (Zhang et al., 2024), considering the large potential for gross N₂O production in generally anaerobic aquatic environments (Wang et al., 2025).

Despite the importance of constraining N₂O budgets from aquatic ecosystems, substantial uncertainties persist due to a combination of spatial heterogeneity, data gaps, limited temporal coverage, environmental variability, and methodological inconsistencies. Higher emissions in regions with intensive anthropogenic activities, such as eastern and southern Asia, the southeastern United States, and parts of Europe, lead to significant spatial heterogeneity (Song et al., 2024; Wang et al., 2023). Tropical, arid, and high-latitude regions are poorly explored, limiting accurate upscaling to the global budget. Temporal variabilities, driven by rapid water turnover, or temporal events including eutrophication, warming, pollution, and precipitation-driven anoxia, are likely to contribute to pulse N₂O releases which are hard to measure (Zheng et al., 2022). Furthermore, production and emission processes are strongly influenced by local environmental factors, such as nutrient loading from fertilizers and manure, organic matter decomposition, agricultural runoff, salinity, and urbanization, leading to significant spatiotemporal variability (IPCC, 2023; Resplandy et al., 2024). Additionally, methodological differences introduce another major source of uncertainty, as different flux estimates frequently arise from the use of diverse measurement techniques. For example, variation in gas-transfer velocity (*k*) parameterizations and differences between chamber-based and concentration-gradient approaches can lead to flux discrepancies of 30-70%, depending on the methods and underlying assumptions (Hall Jr and Ulseth, 2020; Stanley et al., 2022). These differences highlight the difficulty of comparing flux data among aquatic types. For instance, fluxes are often much higher in some high-emission environments (e.g., estuaries) than in lower-emitting systems such as lakes and reservoirs. Collectively, these inconsistencies restrict the further consolidation of the global aquatic N₂O emission inventories, highlighting the need for improved monitoring networks, standardized methods, and better coverage of underrepresented areas to reduce uncertainties and refine global N₂O budgets.

The main objectives of this study are to: (1) construct a harmonized database for data acquisition and observation; (2) characterize the distributional patterns of N₂O concentrations and fluxes across various aquatic ecosystems; and (3) identify existing data gaps and propose future research directions to advance the understanding of N₂O dynamics in aquatic ecosystems. To achieve these objectives and to evaluate and predict N₂O emissions from aquatic ecosystems accurately, we developed a comprehensive database, the Global Aquatic N₂O Emission Database (GANED). This database compiles records from 3002 sites to investigate global distributional patterns of both N₂O concentrations and fluxes across aquatic ecosystems, linked with



100 associated biogeochemical variables. It also addresses discrepancies between concentration and flux observations, enabling
robust cross-system comparisons and improved prediction of aquatic N₂O emissions. By describing the dataset structure and
outline, we demonstrated its potential to reveal spatial, temporal, and biogeochemical controls on N₂O dynamics across diverse
aquatic systems. By integrating these attributes, GANED provides: (i) in-depth analyses of N₂O drivers, seasonal and spatial
variability; (ii) upscaling supports to improving national GHG inventories and global modeling frameworks; and (iii) robust
105 empirical evidence and theoretical basis for the quantitative assessment and management of aquatic N₂O emissions.

2 Methods

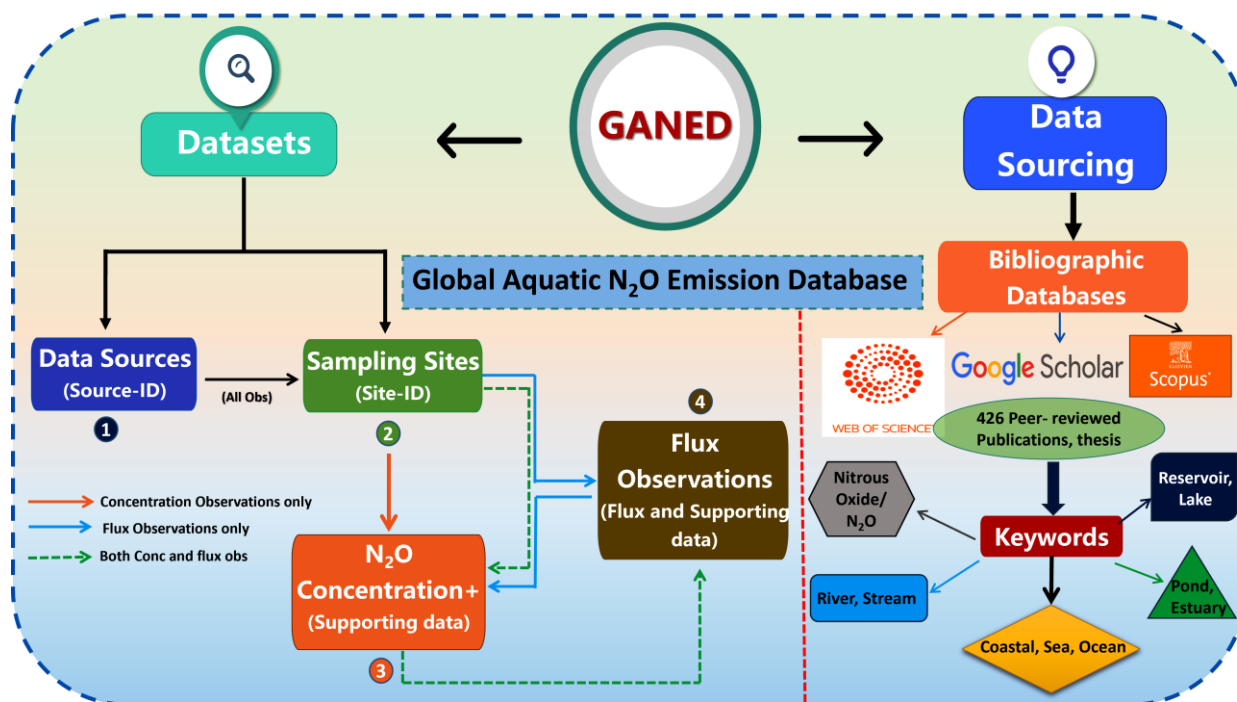
2.1 Literature search and data extraction

We conducted a comprehensive and systematic literature review of studies published in peer-reviewed journals between 1980
to 2023, following guidelines consistent with the Preferred Reporting Items for Systematic Reviews and Meta-Analyses
110 (PRISMA) framework, adapted for database synthesis (Moher et al., 2014). To synthesize GANED, original experimental data
were extracted directly from publications reporting N₂O concentrations and fluxes from aquatic systems, including rivers,
reservoirs, lakes, ponds, streams, estuaries, coastal areas, and seas (Fig. 1). GANED consolidates four interlinked datasets
encompassing (1) data sources, (2) sampling sites, (3) N₂O concentrations, and (4) N₂O fluxes, accompanied by associated
physicochemical parameters such as ammonium (NH₄⁺), nitrate (NO₃⁻), nitrite (NO₂⁻), dissolved oxygen (DO), dissolved
115 organic carbon (DOC), total nitrogen (TN), total phosphorus (TP), water temperature, salinity and pH, along with supporting
metadata such as geographic location, sampling dates, temporal information, climate data, and methodological details. These
datasets were connected through unique source identifiers, with each concentration and flux record assigned a distinct site
number to ensure consistency and traceability across data tables. Arrow lines illustrate the data flow and relational structure
among these datasets. Data compiled in GANED were sourced from peer-reviewed journal articles, master's theses, and
120 doctoral dissertations that reported quantitative N₂O measurements. Source material was obtained through systematic searches
of major bibliographic databases and repositories, including Web of Science, Scopus, and Google Scholar, using combinations
of the keywords including “nitrous oxide/N₂O”, “river”, “stream”, “reservoir,” “lake”, “pond”, “estuary”, “coastal”, “sea”, and
“ocean” as depicted in Fig. 1.

Inclusion criteria were rigorously applied to data reliability across studies. Specifically, (1) sources must report direct
125 measurements of N₂O concentrations or fluxes from aquatic ecosystems; (2) datasets must include geospatial coordinates or
clearly identifiable sampling site locations; (3) peer-reviewed journal articles, and dissertations/theses were prioritized, though
selected reports were incorporated if they provided novel data; (4) measurements must be accompanied by key environmental
variables, e.g., NH₄⁺, NO₃⁻, NO₂⁻, DO, DOC, TN, TP, water temperature, salinity and pH, to facilitate analysis of N₂O drivers.
Exclusion criteria were applied to eliminate studies that focused exclusively on terrestrial soils, atmospheric modeling without
130 empirical validation, or those lacking quantitative data, such as qualitative reviews. Following this screening process, a total
of 426 data sources met the inclusion requirements, comprising approximately 400 peer-reviewed journal articles and 26



academic theses or dissertations.



135 **Figure 1: Schematic overview of GANED and the relationship among its four meta-datasets.** The data workflow began with recording
 metadata for each source in the “Data Source”, where a unique ‘Source-ID’ was assigned. Site-specific information from each source
 was then entered into the “Sampling Sites”, with each site allocated a unique ‘Site-ID’ that was cross-referenced with its
 corresponding Source-ID. Both identifiers (Source-ID and Site-ID) were subsequently carried into the “Concentration and Flux”
 datasets to ensure complete traceability across all observations. N₂O observations were categorized based on site-date combinations,
 distinguishing between concentration data (orange) and flux data (blue) arrow lines. The “Concentration & Flux” datasets recorded
 140 Concentration records and flux measurements along with supporting data, respectively. Data sourcing for GANED carried through
 deep searching of the bibliographic databases (Web of Science, Google Scholar, Scopus) and related keywords.

2.2 Database Structure

The structure of GANED built upon established frameworks from analogous compilations in aquatic GHG research. Adapting
 this research approach, GANED was structured into four interrelated components, each designed to facilitate seamless data
 145 integration and analytical flexibility. These components were interconnected via unique identifiers for both data sources and
 sampling sites, allowing users to trace each observation to its origins while exploring spatiotemporal patterns and
 environmental correlates. The “Data Source” dataset stored metadata of each publication or report, uniquely identified by a
 ‘Source-ID’. The “Sampling Sites” dataset recorded information for each observational site, each associated with its
 corresponding data source through the same ‘Source-ID’. These identifiers were used to associate site-level attributes with the
 150 “N₂O Concentration” and “N₂O Flux” datasets, ensuring consistent referencing and integrity. This structured linkage enabled
 robust integration, filtering, and querying across datasets, thereby supporting advanced analyses of spatial and temporal N₂O



variability. The database also included flags for data aggregation (spatial and temporal) and observation types, allowing users to filter the dataset for methodological consistency. Thus, this modular design not only aligned with successful precedents in aquatic GHG databases but also enhanced usability for hypothesis testing and upscaling exercises specific to N₂O dynamics.

155 The data workflow for N₂O observations is further demonstrated in Fig. 2.

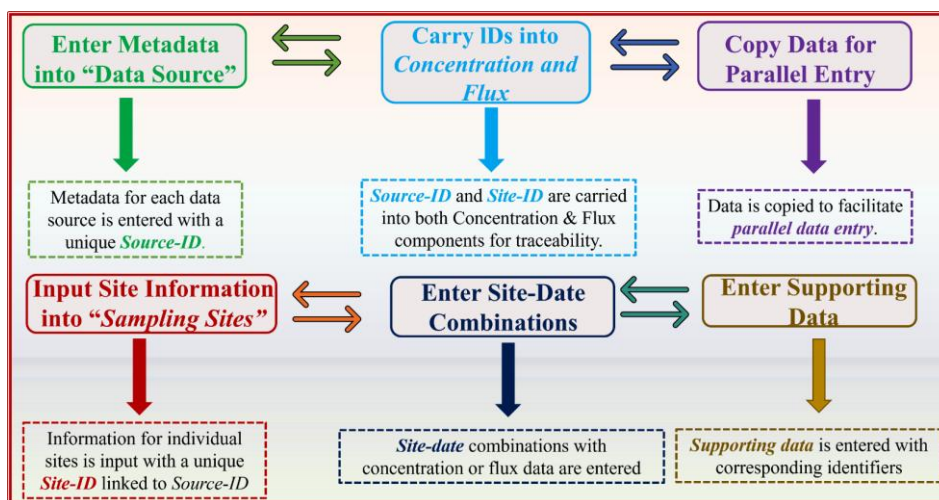


Figure 2: Workflow for entering data into the GANED. A structured, step-by-step protocol for inputting metadata, site-specific information, and ancillary concentration and flux datasets into the system, ensuring rigorous data integrity and streamlined processing through standardized workflows.

160 2.2.1 Data Source

The “Data Source” component comprised of all scientific literature used in constructing GANED. Each N₂O data source was identified by a unique ‘Source-ID’ along with corresponding bibliographic details, including title (Title), author(s) (Author), publication type (Type), data source name (SRC-Name), publication year (Pub-year), and digital object identifiers (DOI). All column titles for “Data Source” are provided in Table A1. In several instances, the compiled datasets were supplemented with additional information, including related physicochemical parameters sourced from external databases or obtained directly from the original author(s) to clarify incomplete or ambiguous records. To resolve uncertainties (e.g., unit inconsistencies) or to acquire missing/more detailed information (e.g., site- or date-specific concentrations or fluxes), we contacted the corresponding authors as well to verify or provide supplementary data. The source dataset represented the integration of previously unpublished data provided by these authors, along with a detailed account of all modifications and corrections.

170 2.2.2 Sampling Sites

Site-specific information was extracted from tables, figures, and text descriptions provided by the authors within the source publications using Web-Plot-Digitizer (<https://automeris.io/WebPlotDigitizer/>). This process yielded geospatial and descriptive attributes for the 3002 unique sampling sites, including information on country, continent, aquatic type, site name, channel



type, latitude, longitude, depth (m), surface area (km²), elevation (m), and status (e.g., concentration-only, flux-only, or both).
175 The “Sampling Sites” dataset summarized the basic attributes of all locations where N₂O measurements were collected. Each site is assigned a unique identifier ‘Site-ID’ and a corresponding site name (typically taken directly from the original data source/publication). Sites were also cross-referenced with the “Data Source” dataset through the shared ‘Source-ID’. A detailed description of all columns included in the “Sampling Site” dataset is provided in (Table A2).

The spatial extent associated with each ‘Site-ID’ varied among data sources. In some cases, sites represented (1) discrete
180 sampling points, (2) geomorphically distinct study reaches, or (3) large channel sections, drainage networks, or other defined geographic units. For sites containing multiple sampling points, results were generally averaged to represent mean conditions. Most of the sites had geographic coordinates, though in cases where coordinates were missing, approximate locations were derived from maps or figures in the original publications using Google Earth (Google, 2025, <https://earth.google.com/>). All coordinates were verified and corrected as needed to minimize spatial errors. Gas sampling at closely located sites (often
185 referred to as ‘high-density sites’) can complicate site delineation and data interpretation. To address this, we merged data from points with minimal differences in latitude and longitude, treating them as a single representative site. For that, firstly, latitude and longitude precision were standardized to three decimal places to ensure spatial consistency. Secondly, in cases where samples were collected from multiple points or depths within a single channel cross-section, data were averaged to produce one representative record (especially for the Pearl River basin and the streams in Guangdong Province, China). However, some
190 ambiguity remained when distinguishing between repeated observations at the same site and distinct nearby locations (typically 10-50 m apart). In such instances, sites with overlapping coordinates and sampling on different days were treated as a single site based on careful evaluation.

Many studies on aquatic N₂O dynamics aimed to determine whether, and to what extent, specific features such as upstream reservoirs, point-source discharges, thermokarst formations, or gas extraction activities affect N₂O concentrations and fluxes,
195 typically with the presumption of an overall acceleration. Similarly, investigations focusing on N₂O hotspots (e.g., springs, floodplains, backwaters, ditches, or canals) might introduce a bias toward higher emission estimates (Stanley et al., 2022). Nevertheless, we still included these data points in GANED, allowing us to examine the outcomes of anthropogenic influence on aquatic N₂O emission. Additionally, to enhance interpretability, we devised a system of channel codes to be used when identifying intended site types. The description of channel codes has been demonstrated in Table 1. Standardizing these
200 classifications posed significant challenges, often requiring subjective judgment in the context of N₂O emissions from aquatic systems. For example: (1), the influence of upstream dam (code US) on downstream sites (code DS) varied with distance (Wang et al., 2021b); (2) symmetric linear canals in agricultural and urban areas had unclear origins, with uncertainty over whether natural streams were channelized (code CH) or new ditches/canals (codes DIT, CAN) were created; in this case unspecified straight channels were coded as CH (3) wetland springs (WTS) were frequently difficult to locate due to their
205 diffuse boundaries; and (4) floodplain channels (FP) posed complexities in distinguishing riverine from wetland systems; therefore, the FP code was applied to lentic habitats such as backwaters or connected floodplain lakes. Due to these ambiguities, we recommend that these channel-type codes be interpreted with caution in conducting comparative or spatial analyses of



aquatic N₂O emissions.

210 **Table 1: Codes indicating site- or channel-specific attributes, as well as the presence of conditions potentially influencing N₂O concentrations or fluxes. Imposition of codes was based on information provided in the data source and/or direct observation of the site. Codes are recorded in the Channel-type field of the component “Sampling Sites” dataset.**

Code	Description
CAN	A canal or artificial channel having hardened channel boundaries.
CH	Channelized (A channel, in which all or most of its straight-line segments are of the same width and variations of channel direction are usually angled features instead of curves).
CD	Channel in a river delta
US	Upstream: direction of water toward the source or origin of a river or stream.
DS	Downstream: Site or region farther away from the source, usually at a lower elevation.
DIT	Ditch, normally agricultural drainage, an unhardened channel
FP	Flood plain site in a water body linked to the main channel that is lentic or is defined as a floodplain lake or a backwater.
GT	The end or “toe” of a glacier, where the glacier’s forward movement (advancing) or melting (retreating) stops.
IMP	The existence of more than one and typically small impoundment in the vicinity of a site (e.g., several rivers in Europe).
PI	permafrost that influences the local environment, particularly in regions adjacent to thermokarst outflows.
PSD	At once (< 1 km) downstream of any point source discharge.
SP	A spring channel is a natural or artificial pathway that directs water from a spring to other areas, such as rivers, streams, or reservoirs.
TH	The location that naturally receives thermogenic N ₂ O or by anthropogenic activities.
WTS	Wetland stream; the location is in a wetland or directly downstream of the outlet of a wetland.

2.2.3 Concentration and flux dataset

215 GANED was integrated with two core datasets, “N₂O Concentration” and “N₂O Flux”, extracted from diverse global aquatic systems across various temporal and spatial scales. Both datasets were linked to the “Data Sources” and “Sampling Sites” via ‘Source-ID’ and ‘Site-ID’. The datasets were further structured to include associated GHG measurements (CO₂, CH₄), as well as physical and chemical parameters to facilitate comprehensive analysis of N₂O dynamics (see Tables A3 and A4 for a complete list of columns and their descriptions). Most concentration and flux data were directly extracted from the tables of peer-reviewed literature, unpublished field measurements, and existing repositories provided by the authors. In some instances,



values were obtained from figures, and data were digitized and standardized to ensure consistency within the database.
220 Temporal precision varied across sites; some records specified exact days and months, while others reported only the month
or year. To standardize the temporal resolution, all data were adjusted to a monthly level. For records lacking specific start
dates, two additional columns “Sampling-Year” and “Sampling-Month” were introduced to indicate the month and year of
data collection. Temporal aggregations were recorded in the “Aggregated-Time” field, which was marked as “No” when the
start date (Date-start) and end dates (Date-end) were identical. Spatially aggregated sampling points were indicated by the
225 “Aggregated-Space” field and marked as “Yes” when the mean value of the aggregated points was used to represent the data.

Measuring both dissolved N₂O concentration and flux from aquatic systems involved a combination of sampling
techniques, gas analyses, and, in some cases, mathematical modeling. Among the recorded data sources, 52% of the total
observations used a common approach of combined measurements of dissolved gas concentration (conc) and the gas exchange
coefficient (*k*) to estimate N₂O emissions. Direct gas sampling followed by analysis via gas chromatography or mass
230 spectrometry was also commonly employed. Additionally, chamber-based flux measurements, comprising 48% of all
observations, were also widely used, such as floating and static chambers. In the chamber method, the flux rate was calculated
based on the change in N₂O concentration within the chamber over time according to the Eq. (1):

$$FN_2O = \frac{V_{\text{chamber}} \cdot \Delta CN_2O}{A_{\text{surface}} \cdot \Delta t} \quad (1)$$

235 Where FN_2O is the N₂O flux (μmol⁻²day⁻¹), V_{chamber} is the chamber volume (L), ΔCN_2O is the observed change in N₂O
concentration inside the chamber (μmolL⁻¹), A_{surface} is the chamber surface area (m²), and Δt is the duration of the chamber
deployment (in days).

N₂O flux via gas transfer velocity was typically calculated using the Eq. (2) (Henry’s Law Method):

$$Flux = k \times \Delta C / H \quad (2)$$

240 Where Flux represents N₂O flux, expressed in (mmol m⁻² day⁻¹). The gas transfer velocity *k* (m d⁻¹) strongly depends on water
temperature and hydrodynamic conditions and was calculated using the widely accepted general equation Eq. (3) for N₂O
(Wanninkhof, 2014):

$$k_{N_2O} = k_{600} \times (Sc_t / 600)^{-n} \quad (3)$$

Where k_{600} is the normalized gas transfer velocity, and Sc_t is the Schmidt number of N₂O at in situ temperature, and ‘n’ is the
245 Schmidt number exponent (0.5 cm h⁻¹ for lakes, reservoirs, rivers and streams). Under hydrodynamic conditions, such as wind
speed for standing waters (lakes, ponds, reservoirs) and velocity/channel slope for running waters (coastal/open oceans), this
 k_{600} was further determined as using the following Eq. (4):

$$k_{600} = 0.251 \times U_{10}^2 \times (Sc_t/600)^n \quad (4)$$

Where 0.251 is an empirically determined constant used in the equation to relate wind speed to gas transfer velocity, and U_{10}
250 is the wind speed at 10 meters above the water surface (in ms⁻¹). In eq (2), ΔC represents the concentration gradient, and *H*
denotes Henry’s law constant. The concentration gradient was calculated as the difference between the N₂O concentration in



the water and the N₂O equilibrium concentration with the atmosphere, as, $\Delta C = (C_{\text{water}} - C_{\text{eq}})$.
This study compiled fluxes from diverse aquatic systems reported in the recorded publications (Data-source), and k values were adopted directly from the original publications whenever provided. When recalculations were necessary, k was assigned using system-specific ranges established in recent syntheses, with temperature normalization performed using the Schmidt number dependency with an exponent of -0.5.

Some studies additionally employed stable isotope labeling techniques to trace the sources and cycling of N₂O pathways, offering unique insights into its biogeochemical behavior and environmental impact, despite their relatively limited application. The integration of these approaches across GANED enhanced measurement accuracy and comprehensiveness, facilitating a deeper understanding of N₂O dynamics across diverse aquatic ecosystems.

2.3 Data standardization and quality assurance

Several steps were undertaken to assess the reliability and quality of the data compiled within GANED. This process included evaluating the accuracy of digitized data, followed by a series of additional verification procedures. Each data entry was cross-checked by a co-author other than the original contributor, when necessary. These checks encompassed validation of site details, variable units, and gas data entries, either through spot checks or full reviews, depending on the dataset size and whether the data were manually entered or imported from external files. After data entry, all gas and physicochemical variables were standardized to consistent units (Tables A3 and A4). Specifically, all concentrations were converted to ($\mu\text{mol L}^{-1}$) and fluxes to ($\text{mmol m}^{-2} \text{day}^{-1}$). Unit conversions employed standard molar volumes and gas constants; measurements originally reported in nmol L^{-1} , ppm or ppb were converted using Henry's law where appropriate. After converting all values to standard units, variables were visualized to detect irregularities, and any extreme values were cross-checked with the original data source. Unresolved records were excluded from the dataset, and missing data points for ancillary measurements or secondary metadata fields were explicitly denoted with 'NA' (Not Applicable/Not Available).

The quality control involved several steps: (1) cross-verification of units and values against original publications; (2) detection of irregularities detection using statistical thresholds (values >3 standard deviations from the mean per aquatic type were excluded, affecting $<1\%$ of entries); (3) handling of short data (having a single observation) in dataset files by noting limitations to ensure robustness in aggregated analyses; (4) assessment of measurement uncertainty, incorporation reported error ranges where available, for example, chamber-based flux measurements.

2.4 Uncertainty, plausibility assessment, and guidance for reuse

The GANED database integrates measurements compiled from 426 peer-reviewed studies across diverse aquatic environments, sampling designs, analytical methods, and reporting formats. Consequently, uncertainty in GANED arises from multiple sources, including heterogeneity in original measurement approaches, variation in temporal sampling frequency, incomplete reporting of ancillary variables, digitization of graphical information where tabulated values were unavailable, and differences in study-specific quality-control procedures. In addition, concentration measurements and direct flux observations are not always interchangeable, particularly in hydrologically dynamic systems, and this methodological heterogeneity should be



285 considered when comparing records across aquatic types.

To reduce these uncertainties, we applied harmonized quality-assurance procedures during database construction, including standardized unit conversion, metadata harmonization, cross-checking of extracted values, removal of clearly implausible outliers, and explicit structuring of the database into source, site, concentration, and flux components.

A formal validation against an independent global reference product is currently not possible due to the lack of a globally
290 consistent benchmark dataset for aquatic N₂O observations across all major water bodies. Instead, following ESSD principles, we provide a plausibility assessment based on internal consistency checks, cross-system comparisons, methodological evaluation, and explicit identification of underrepresented regions, ecosystems, and time periods.

Users should interpret GANED as a harmonized dataset product useful for comparison, meta-analysis, and hypothesis generation, while being aware of its representativeness limits. Notably, the spatial concentration records in Asia and North
295 America, lack of long-term repeated observations, and the limited coverage of some transitional or intermittently connected aquatic systems may constrain global extrapolation and should be considered in downstream analyses.

2.5 Statistical Analysis

All data processing and analyses were conducted in R statistical software (version 4.5.1, R Core Team, 2025) using packages such as ‘dplyr’ (version 4.2; Wimberly (2023) and ‘data.table’ (Dowle and Srinivasan, 2023) for data manipulation, the “sf”
300 package (version 4.2; Mishra et al., 2024) for spatial data handling, ‘stats’ for statistical analyses, and ‘ggplot2’ and ‘patchwork’ (version 4.2; Kabacoff; Thomer and Rayburn, 2024) for visualizations. Descriptive statistics (means, medians, standard deviations) were used to summarize the database, highlighting site distributions by country/continent and aquatic system type. Grouped means facilitated comparison of N₂O concentrations and fluxes across ecosystem categories. Temporal trends were assessed by aggregating observations into annual averages, followed by linear regression to identify changes from 1980 to
305 2023. To explore environmental drivers, correlation coefficients were calculated between N₂O fluxes and associated physiochemical variables, using pairwise complete observations to handle missing values. Statistical significance was denoted as *** for $p < 0.001$, ** for $p < 0.01$, and * for $p < 0.05$ for multiple comparisons.

3 Results

3.1 Introduction to the GANED

310 GANED comprised 5130 N₂O concentration records and 7386 N₂O flux records from 3002 unique sites spanning 1980-2023 (Fig. 3a). Notably, more than 85% of the concentration records and 80% of the flux measurements were published since 2015, with rivers and streams accounting for half of all observations (Fig. 3b). GANED revealed significant variation in N₂O emissions across different water body types, highlighting the growing significance of aquatic contribution to the global N₂O budget. As illustrated in Fig. 3b, rivers contribute the largest share (41.04%) of the total emissions, followed by lakes (13.02%),
315 reservoirs (14.36%), and streams (8.43%), whereas seas and coastal areas contributed the least (2.63% and 1.53%, respectively)



despite their large surface area, contribute minimally to overall aquatic N₂O emissions. Long-term observations were rare, with only 2% (62/3002 sites; n = 62) having more than 10 concentration observations (total=1629 observational counts), and 4% (130/3002 sites; n = 130) having more than 10 flux records (total=3807 counts). Among these sites, observation counts ranged from 13 to >100 per site. The longest concentration record spans 11 years (Priscu et al., 2008), whereas the longest flux record spanned 19 years (De Wilde and De Bie, 2000).

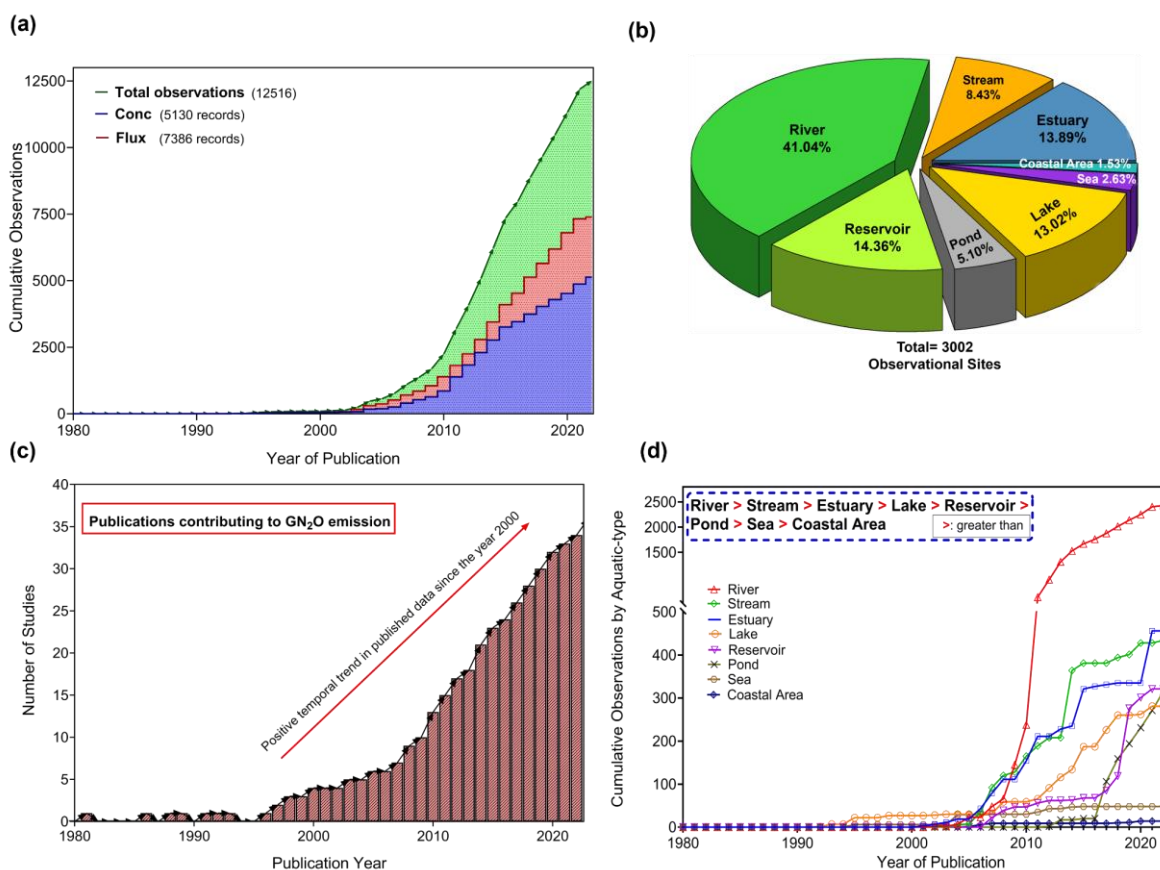
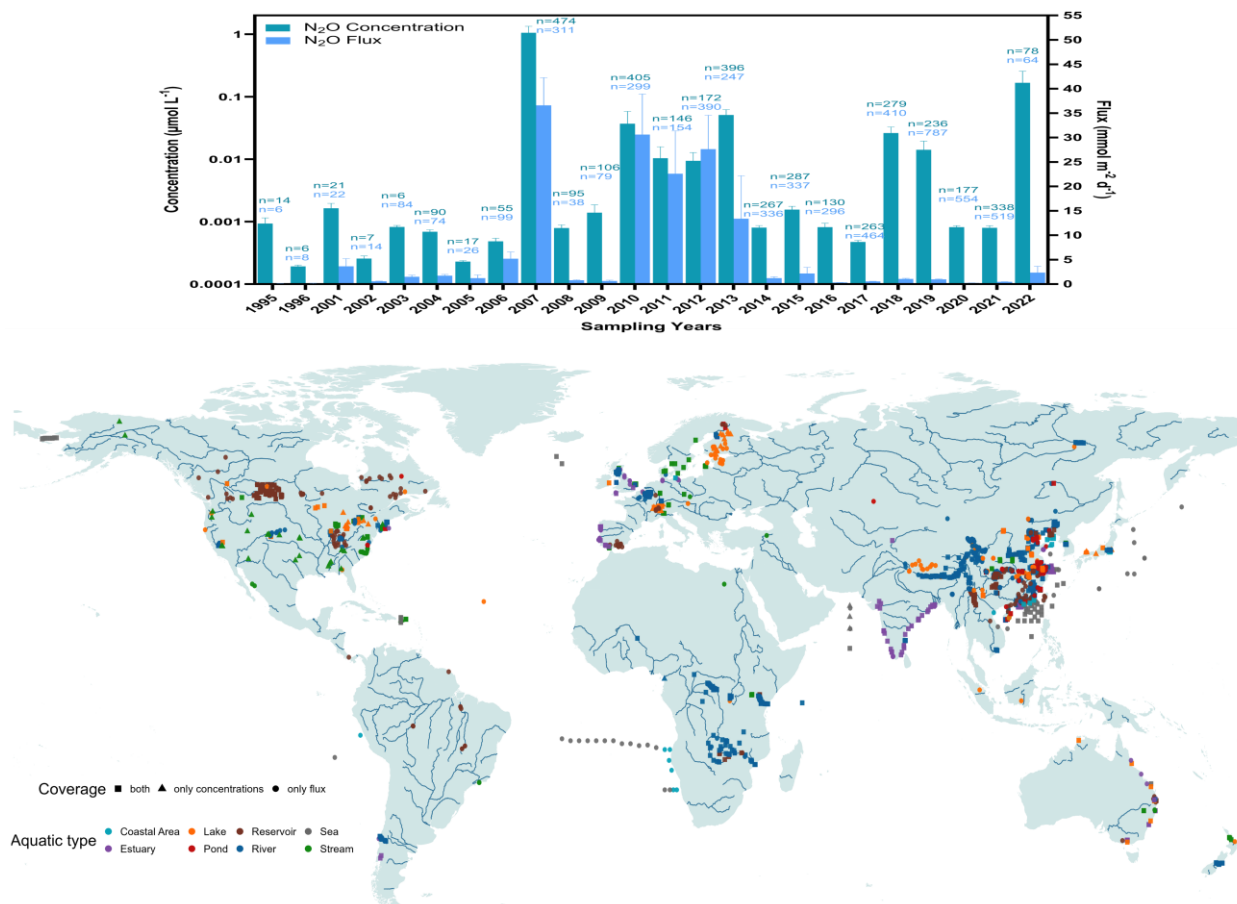


Figure 3: Global N₂O emissions (concentration and flux) from aquatic bodies over time, (a) cumulative observations of N₂O concentration and flux data, (b) proportion of water bodies contributions to net N₂O emission, (c) publications contributing to global N₂O emission, (d) the number of cumulative observations showing emission trends across different aquatic types over time.

Spatially, the majority of N₂O observations were concentrated in Asia (60%), followed by North America, which accounted for 15% of total N₂O observational sites. Europe, Oceania, and Africa had moderate but meaningful contributions compared to Asia and North America. In contrast, areas with extensive inland and coastal water bodies, such as Antarctica, Central America, and South America, were sparsely represented or entirely lacking observations, highlighting persistent geographical biases in global N₂O monitoring (Fig. 4). The heterogeneous distribution of observations underscored both the global relevance and uneven coverage of aquatic N₂O studies, offering important insights into spatial variability of concentrations and fluxes



across diverse water bodies. Further, number of observations for all variables by continent are shown in (Table S5).



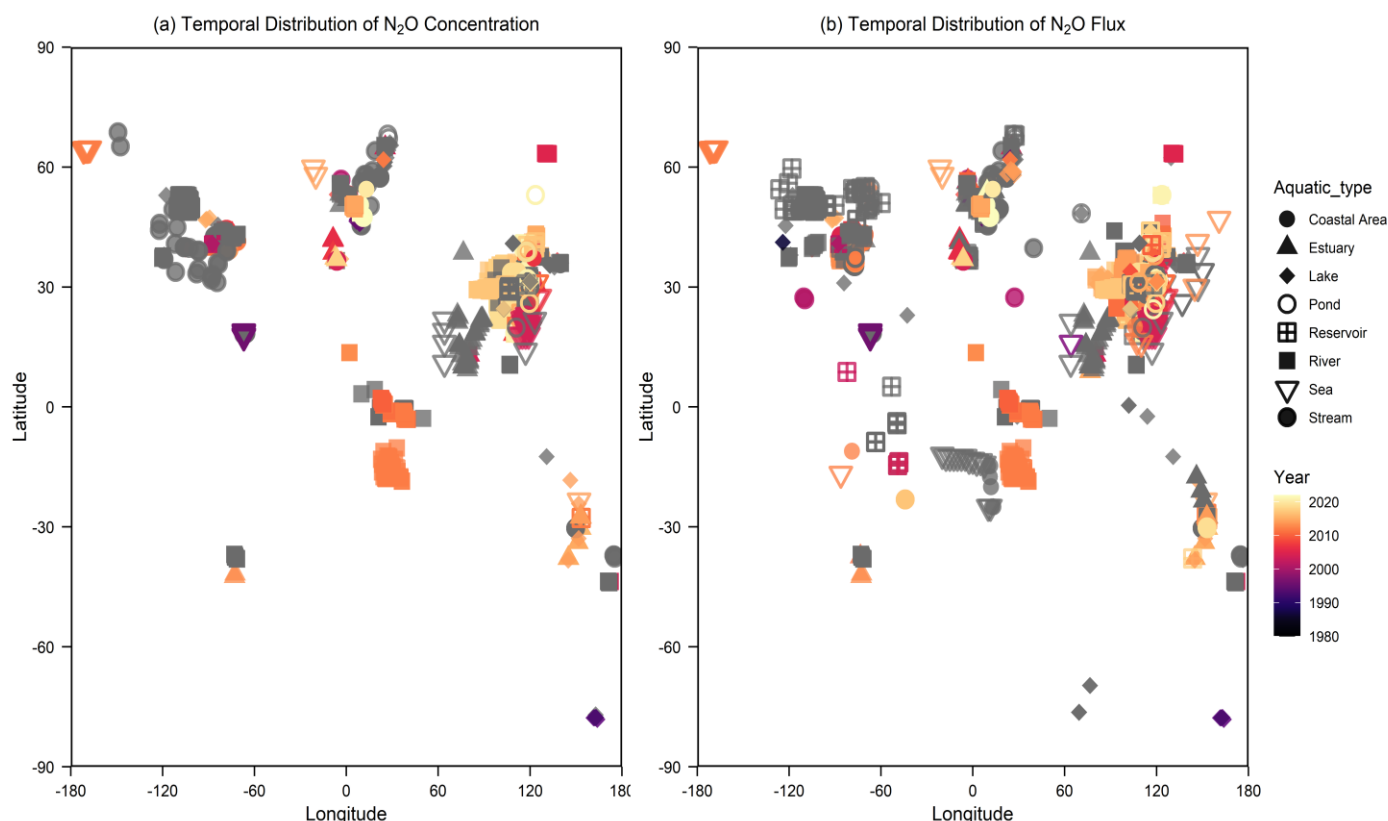
335 **Figure 4: Global distribution of N₂O observations in the database. Different symbols on the map represent observational sites where N₂O emissions were observed, categorized by the type of aquatic system (e.g., River, Stream, Lake, Estuary, Reservoir, Pond, Sea, and Coastal Area). The color coding identifies the type of aquatic type, while the shape and color of each symbol indicate whether the measurements focus on N₂O concentration, flux data or both.**

The temporal distribution of N₂O concentration (a) and flux (b) measurements across various aquatic types has been illustrated in Fig. 5. The observational data revealed a clear latitudinal gradient, with an increasing number of studies over time. Prior to 2000, concentration records were largely restricted to the Northern Hemisphere (30°-60°N), mainly across North America, Europe, and eastern Asia, reflecting the early focus on temperate mid-latitudes (Fig. 5a). After 2000, observations expanded across the Southern Hemisphere and tropics, achieving near-global coverage by 2010-2020, with notably accelerated records in South America, sub-Saharan Africa, Southeast Asia, and Australia. Recent measurements (2010-2020; yellow/red symbols) from tropical estuaries and rivers (10°S-30°N), notably in Southeast Asia and the Amazon Basin, frequently reported the highest N₂O concentrations.

The corresponding flux data followed similar temporal trends in sampling distribution consistent with the concentration data



(Fig. 5b). Early fluxes (1980-2000) were concentrated in mid-latitude regions, whereas more recent flux observations (2010-2020) were frequently reported within tropical zones (15°S-15°N), particularly in estuaries, rivers, and reservoirs across Southeast Asia, eastern China, India, and South America. Flux observations were also evident in Southern Hemisphere subtropical regions. Overall, rivers and streams consistently exhibited high fluxes across latitudes, with the most pronounced emissions occurring in tropical and subtropical systems, emphasizing their critical role as persistent aquatic N₂O sources.

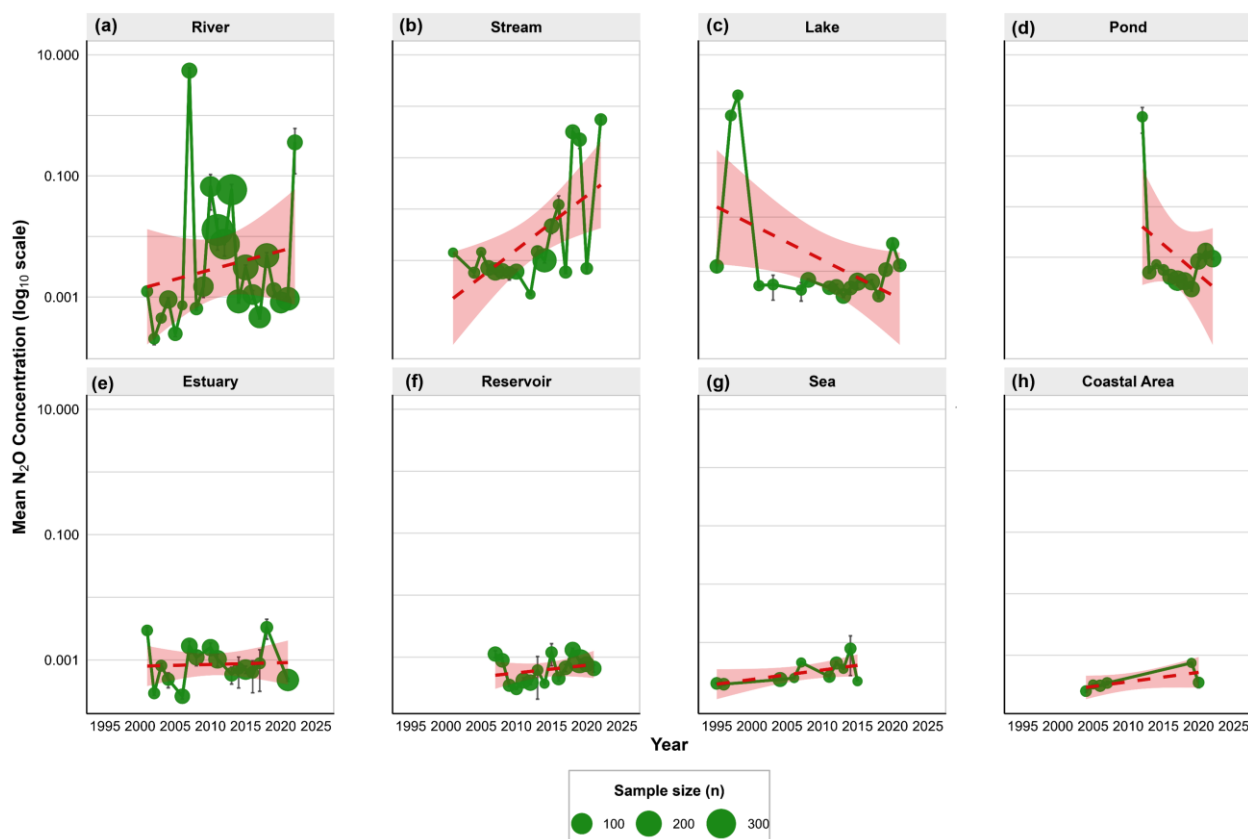


355 **Figure 5: Temporal distribution of N₂O emissions across different aquatic types over the years, (a) temporal distribution of N₂O concentration (left), with data points color-coded by years, ranging from 1980 (purple) to 2020 (yellow/red), and (b) temporal distribution of N₂O flux (right) across longitude and latitude coordinates over same period of time.**

Temporally, long-term trends in mean N₂O concentrations across aquatic types varied substantially (Fig. 6). Rivers and streams generally exhibited the highest concentrations and greatest interannual variability. Rivers showed extreme fluctuations, with mean concentrations occasionally exceeding 1.0 (on the log₁₀ scale), though no definitive long-term linear trend was evident due to high variability. Streams, however, displayed a distinct increasing trend in N₂O concentration, particularly notable from 2010 onward. Lakes and ponds exhibited a sharp decrease following a peak around 2011. Transitional and marine systems, such as estuaries, reservoirs, seas, and coastal areas, were characterized by the lowest overall N₂O concentrations, generally remaining stable or showing only marginal increasing trends over the decades.

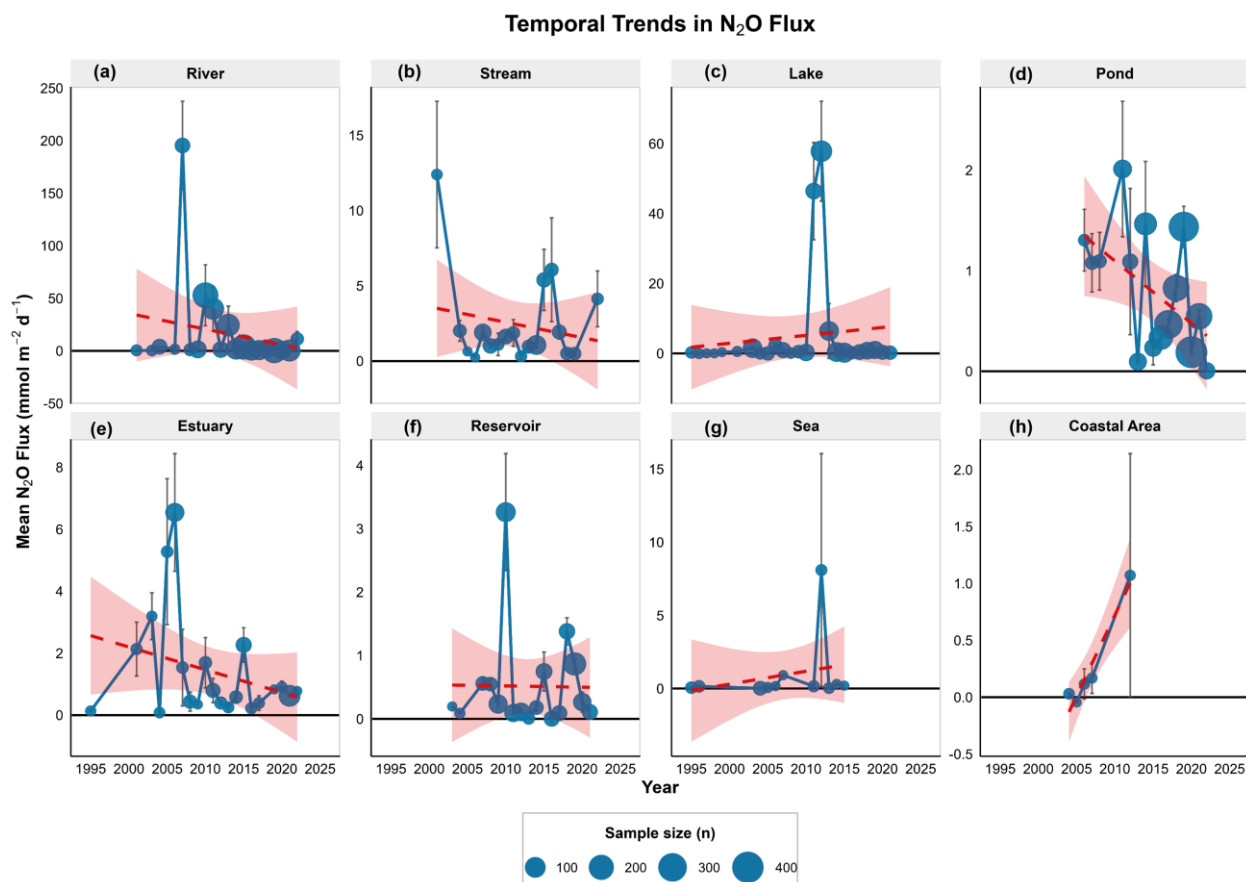


Temporal Trends in N₂O Concentration



365 **Figure 6: Temporal trends in mean dissolved N₂O concentration (log₁₀ scale) across aquatic types: (a) River, (b) Stream, (c) Lake, (d) Pond, (e) Estuary, (f) Reservoir, (g) Sea, and (h) Coastal Area. Green dots represent annual means concentration with circle size proportional to sample size (n= 100, 200, 300); pink shading and red line show the loess-smoothed trend with 95% confidence interval. Vertical error bars indicate standard error. Subplots are arranged in two rows with a shared y-axis per row to facilitate comparison.**

Temporal trends in mean N₂O fluxes were highly dependent on aquatic type and characterized by significant temporal
370 heterogeneity and episodic high-emission events (Fig. 7). Rivers dominated emission magnitudes, driven by extreme episodic
peaks, ~200 mmol m⁻² d⁻¹. Despite these peaks, the overall trend line for rivers suggested a slight decrease or stabilization in
recent years. Streams, lakes, ponds, and estuaries generally exhibited decreasing trends in mean N₂O flux over the observed
period, despite the increasing concentrations noted in streams in Fig. 6. Lakes and reservoirs were characterized by generally
low baseline fluxes punctuated by singular, intense emission years, e.g., the Lake peak in 2011. Seas generally reported low
375 and stable flux rates. Coastal areas showed a marked increasing trend in emissions, although data availability for this specific
category was limited to the period prior to 2015.



380 **Figure 7: Temporal trends in mean N₂O flux (mmol m⁻² d⁻¹) across aquatic types: (a) River, (b) Stream, (c) Lake, (d) Pond, (e) Estuary, (f) Reservoir, (g) Sea, and (h) Coastal Area. Blue dots represent annual mean fluxes; pink shading and red line show the loess-smoothed trend with 95% confidence interval. Dot size is proportional to the sample size in each year (100-400). Vertical error bars indicate standard error.**

3.2 Distribution patterns of N₂O concentrations and fluxes

385 The distributions of observed N₂O concentrations and fluxes in GANED were strongly positively skewed, with the majority of the data being collected predominantly occurring in the Northern Hemisphere (Fig. 8). Most concentration measurements ranged between 10⁻⁴ and 10⁻³ μmol L⁻¹ (Fig. 8a), with rivers and streams contributing the largest share of observations within this interval. Isolated but exceptionally high concentrations (> 400 μmol L⁻¹) were recorded, reflecting the presence of intense local emission hotspots (Sun et al., 2009). N₂O fluxes displayed a similar pattern, spanning from -9.3 to 164 mmol m⁻² d⁻¹, with ~500 observations falling between 10⁻² and 10⁰ mmol m⁻² d⁻¹ (Fig. 8b). Approximately 13% of reported fluxes were zero, below detection limits, or negative, emphasizing the inherent variability and uncertainty associated with flux measurements across diverse aquatic environments.

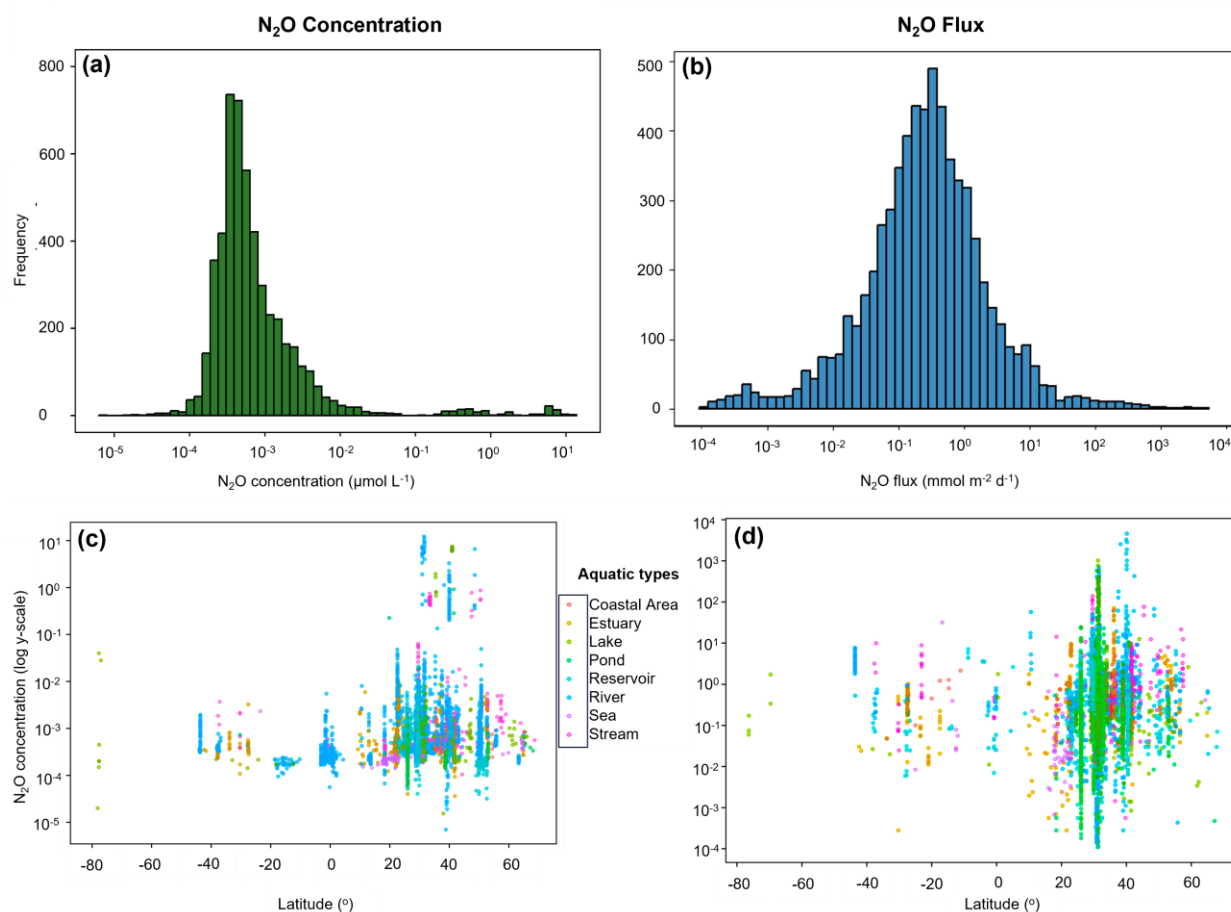
390 Latitudinal patterns provided additional insight into the distribution of concentration and flux observations across climatic zones and revealed substantial spatial heterogeneity and sampling bias (Fig. 8c,d). Data coverage was heavily concentrated in



the Northern Hemisphere, particularly within subtropical and temperate regions between 20°N and 60°N. This latitudinal band not only contained the majority of sampling efforts but also exhibits the highest emission magnitudes. Within this dominant
395 Northern Hemisphere region, the temperate zone (~40°N-60°N) showed dense clustering of data, primarily from rivers, lakes, and reservoirs, with concentrations generally centered around $10^{-2} \mu\text{mol L}^{-1}$. However, the subtropical zone (20°N-40°N) displayed the highest variability. Rivers, streams, lakes, and reservoirs in these regions not only accounted for the majority of available observations but also produced the most substantial emissions, highlighting their role as major global hotspot of aquatic N_2O emission.

400 In contrast, the tropical zone (from the equator to ~23.5° N/S) and the entire Southern Hemisphere were characterized by sparse observational data coverage. Although localized elevated concentrations occurred in tropical rivers and reservoirs near the equator (0°), their frequency and peak magnitudes were substantially lower than those recorded in the northern subtropics. High-latitude systems, including boreal and polar regions beyond 60°N and much of the area south of 40°S, were similarly underrepresented across all aquatic ecosystem types. Overall, rivers, streams, and lakes in the northern subtropical and
405 temperate bands continued to dominate current knowledge of high aquatic N_2O emissions, while pronounced data gaps persist for tropical and Southern Hemisphere regions.

More than 80 % of all measurements originated from high-income and upper-middle-income economies (World Bank classification, 2024), with the top five contributing countries, China ~38 %, United States ~15 %, Germany ~8 %, India ~5 %, and United Kingdom ~4 %, accounting for ~70 % of the entire dataset. China alone contributes the majority of extreme hotspots
410 (Concentration: $>50 \mu\text{mol L}^{-1}$ and flux: $>100 \text{mmol m}^{-2} \text{d}^{-1}$), reflecting intense sampling effort in the heavily fertilized Yangtze, Yellow River, and Hai River basins and southeastern coastal provinces. The United States showed a similar pattern in the Mississippi River basin and Midwestern agricultural states. In contrast, low-income and lower-middle-income countries together contributed <6 % of records, of which these observations were largely restricted to a small number of major river systems such as the Congo, Amazon, and Ganges-Brahmaputra.



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Figure 8: Distribution of N₂O concentrations and fluxes across different aquatic types, derived from the GANED. (a,b) represent frequency histograms of N₂O concentrations (µmol L⁻¹) and fluxes (mmol m⁻² d⁻¹), respectively. (c,d) display the latitudinal distribution of N₂O concentrations (left) and fluxes (right) on a log scale. Different colors of observations denote the distinct aquatic type.

420 3.3 Statistical overview of N₂O concentration and flux by aquatic types

Statistical summaries of N₂O concentrations and fluxes in the GANED dataset by aquatic ecosystem types revealed marked differences in central tendencies and extreme values (Table S2). Lakes exhibited the highest mean dissolved N₂O concentration, 0.406 µmol L⁻¹, followed by rivers, 0.077 µmol L⁻¹, and streams 0.029 µmol L⁻¹. Rivers also recorded the highest maximum concentration with 12.23 µmol L⁻¹. In contrast, ponds, estuaries, coastal waters, and seas generally showed low concentrations, with median values frequently at or near zero and maximum values consistently below 0.011 µmol L⁻¹ (Table S3).

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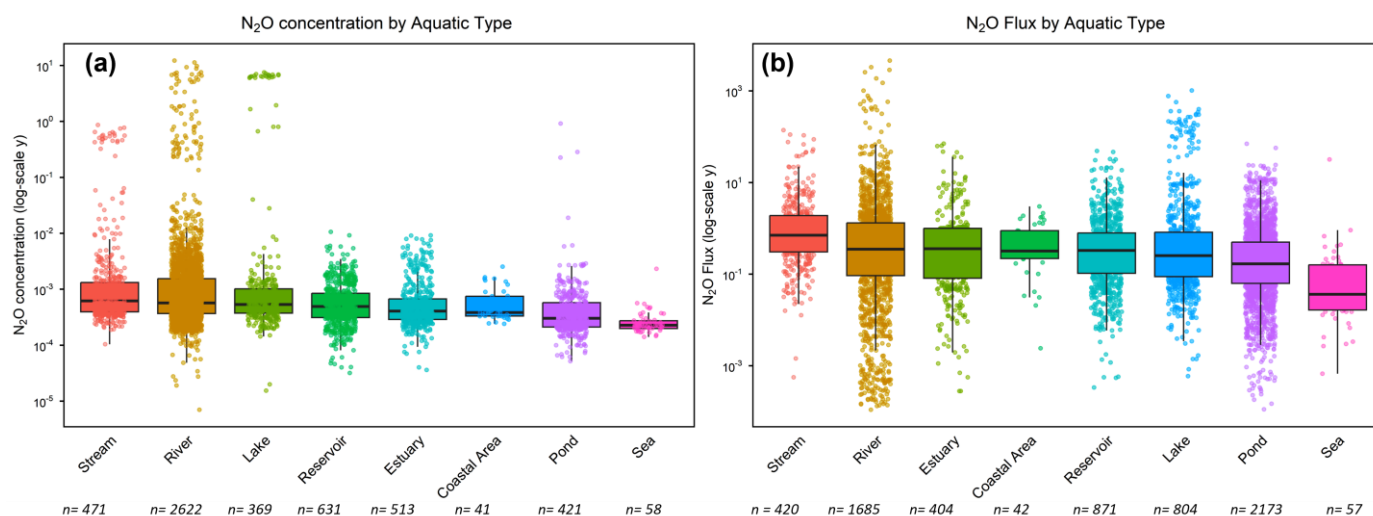
For N₂O fluxes, rivers displayed both the highest mean flux, 15.45 mmol m⁻² d⁻¹ followed by lakes, 10.34 mmol m⁻² d⁻¹, and streams 3.75 mmol m⁻² d⁻¹, respectively. Estuaries, reservoirs, coastal waters, and seas exhibited substantially lower mean fluxes (all < 1 mmol m⁻² d⁻¹) and accounted for the majority of zero or negative values, indicating these systems often functioned near equilibrium or as weak N₂O sinks (Table S4). Furthermore, summary statistics for concurrently measured



430 GHGs (CO₂ and CH₄) concentrations and fluxes are provided in Table S1 and Fig. S1.

Boxplot analysis of N₂O concentrations and fluxes across different aquatic types in the GANED dataset highlighted significant variability and distinct emission profiles, with sample sizes denoted below each aquatic category (Fig. 9). For N₂O concentrations (Fig. 9a), rivers exhibited the widest interquartile range (IQR) and the highest median value on a logarithmic scale, reflecting their role as primary emission hotspots. Streams and lakes displayed moderate medians and IQRs, while ponds, 435 estuaries, reservoirs, coastal areas, and seas exhibited progressively lower medians. Seas had the lowest median concentration, consistent with their limited number of observations (n = 58). The pronounced scattering of outliers, particularly in rivers and streams, further highlighted the influence of localized nutrient enrichment and key biogeochemical processes such as nitrification and denitrification.

Correspondingly, for N₂O fluxes, rivers and ponds emerged as prominent emission hotspots, exhibiting the greatest magnitude and variability in fluxes, with broader outliers accounting for 1685 and 2173 observations, respectively (Fig. 9b). Lakes and 440 estuaries exhibited intermediate flux levels, whereas seas and coastal waters consistently displayed the lowest N₂O flux values.



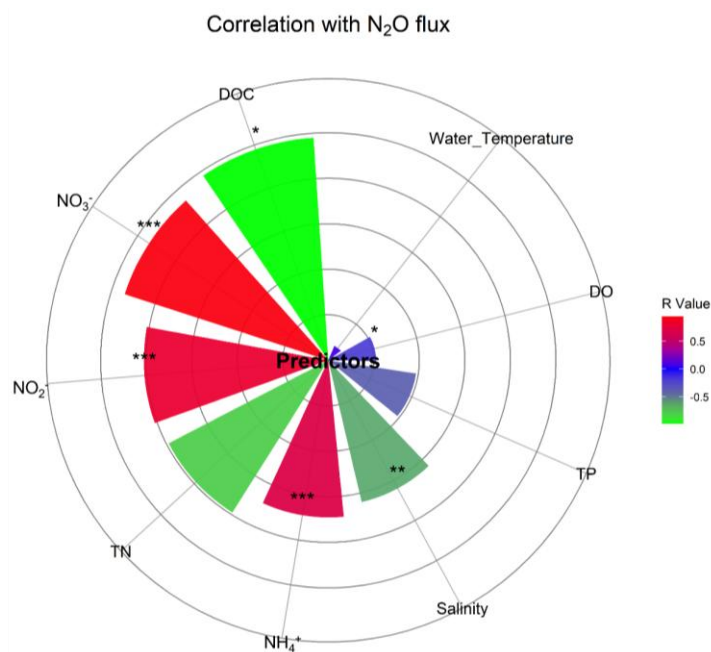
445 **Figure 9: N₂O concentration and flux across aquatic types.** Data was derived from the GANED, comprising 5,126 concentration records and 6456 flux measurements. Boxplots representing the median (central line), interquartile range (box), and data range (whiskers) of distribution of dissolved N₂O concentration (left), and N₂O flux (right) across different aquatic types. Missing or NA values were omitted.

3.4 N₂O flux dynamics and environmental predictors

The relationships between N₂O fluxes and key environmental drivers were examined using correlation analyses, revealing significant associations that highlighted the underlying biogeochemical controls on N₂O emissions from aquatic bodies (Fig. 450 10). NH₄⁺ exhibited a strong positive correlation with N₂O fluxes ($R^2 = 0.889$, $p < 0.001$), emphasizing the role of reduced N availability in driving N₂O emissions, potentially via nitrification or coupled nitrifier denitrification. NO₃⁻ also showed a



455 significant positive relationship ($R^2 = 0.477, p < 0.001$), underscoring the contribution of oxidized inorganic N as a substrate for N_2O production. DOC displayed a moderate positive correlation, consistent with its role in providing electron donors for heterotrophic processes such as denitrification. In aquatic systems, the DOC: NO_3^- ratio is frequently used to indicate the preferred pathway of NO_3^- reduction, with higher ratios favoring complete or incomplete denitrification, which can produce N_2O as an intermediate over assimilatory uptake or dissimilatory reduction to ammonium (DNRA). The positive association observed here with both DOC and inorganic N species (NH_4^+ , NO_3^- , NO_2^-) suggested that denitrification likely contributed substantially to N_2O fluxes in these systems, particularly under conditions of sufficient organic carbon availability. In contrast, TP had a weak negative correlation ($R^2 = 0.148, p = 0.614$), indicating a limited influence on N_2O fluxes. DO displayed a negative correlation ($R^2 = 0.042, p = 0.048$), suggesting that lower oxygen levels and cooler temperatures may favor denitrification processes. Similarly, water temperature exhibited an inverse relationship with N_2O fluxes ($R^2 = 0.469, p = 0.012$). Salinity showed a moderate correlation ($R^2 = 0.404, p = 0.005$; $R^2 = 0.063, p = 0.206$), whereas TN had a negligible non-significant effect ($R^2 = 0.63, p = 0.206$). Additional quantitative results for all analyzed predictors are provided in (Table 2) and summary statistics in (Table S6).



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Figure 10: Radial correlation plot depicting the relationships between N_2O fluxes and key environmental predictors. The correlation strengths (R^2 values) are shown with color gradients (red to green indicating positive to negative associations) at significance levels (* $p < 0.001$, ** $p < 0.01$, * $p < 0.05$).**

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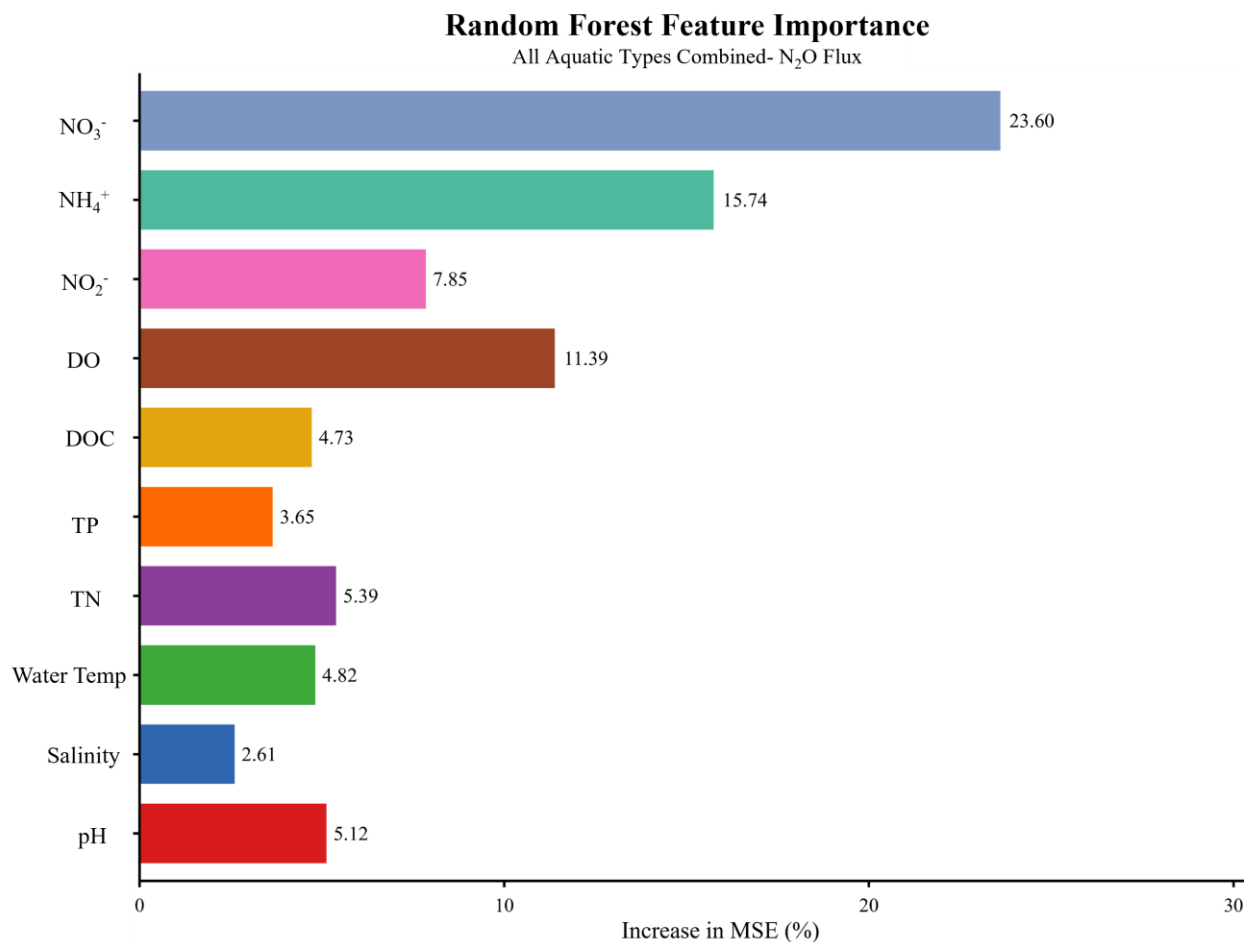
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Table 2: Linear regression models based on the parable portion of the data (approximately 190 rows with non-NA N₂O-Flux values after filtering). The R value is the Pearson correlation coefficient, R² obtained as R² = (R-value)², and significance is based on the p-value (* for p < 0.001, ** for p < 0.01, * for p < 0.05).**

Predictors	Slope	Intercept	R Value	R ² Value	P Value	Std. Error	Significance
NH ₄ ⁺	3.318	-0.160	0.943	0.889	5.11×10 ⁻¹¹	0.262	***
NO ₂ ⁻	16.179	-0.677	0.807	0.651	5.65×10 ⁻⁶	2.647	***
NO ₃ ⁻	3.738	-1.674	0.691	0.477	3.68×10 ⁻⁶	0.874	***
Salinity	-0.181	5.024	-0.636	0.404	0.005	0.055	**
DO	-0.131	1.379	-0.205	0.042	0.048	0.065	*
DOC	-0.261	1.131	-0.977	0.955	0.023	0.040	*
Water-Temperature	0.012	0.297	0.070	0.004	0.469	0.017	Ns
TP	-0.002	0.301	-0.386	0.148	0.614	0.003	Ns
TN	-0.413	1.136	-0.794	0.630	0.206	0.224	Ns

Note: pH (excluded) had insufficient non-NA pairwise data for regression in the recorded N₂O flux observations.

480 Furthermore, using a machine learning (ML) approach, the predictive performance of the random forest (RF) model of ML for estimating N₂O flux across all aquatic types combined in GANED was analyzed and complemented by feature importance analysis. The model was trained on a dataset with an 80% training and 20% testing split, and performance was evaluated using
 485 root mean square error (RMSE), mean absolute error (MAE), and the coefficient of determination (R²), with RF demonstrating robust predictive capability despite the inherent variability in global N₂O flux data. RMSE = 84.41, MAE = 6.96, and R² = 0.13 values for RF indicated the explained variance in the N₂O flux data. The permutation feature importance results from the RF model, expressed as percentage increase in mean squared error (%IncMSE) when each feature was randomly permuted (Fig. 11), identified NO₃⁻ as the dominant predictor with a value of 23.60%, followed by NH₄⁺, 15.74%; DO, 11.39%, and NO₂⁻, 7.85%. The remaining environmental predictors showed lower but notable contributions: TN, 5.39%; pH, 5.12%; water temperature, 4.82%; DOC, 4.73%; TP, 3.65%, and salinity, 2.61%.



490 **Figure 11: Feature importance of environmental predictors in the RF model for N₂O flux. The bar chart shows the increase percentage in Mean Squared Error (IncMSE%) for each environmental predictor, with a higher percentage (%) indicating greater predictive importance for N₂O flux.**

To further unravel the controls on N₂O emissions across different water bodies in GANED, we conducted individual RF models for each aquatic type using only samples with complete data for all predictors (Fig. S2). Model performance varied considerably due to differences in sample size, data variability, and inherent emission ranges (Table 3). Lakes showed the highest R² = 0.2424, followed by estuaries R² = 0.2250, and ponds R² = 0.1817, respectively. Marine systems, seas and coastal areas, exhibited negative R² values, indicating poor predictive skill, likely due to very low sample sizes, n=70 and n=49, respectively, and limited variability in emissions.

Permutation-based feature importance, expressed as IncMSE%, revealed marked heterogeneity in the dominant controls on N₂O flux across aquatic types. In rivers, NO₃⁻: 32.97% and NH₄⁺: 21.60%, together accounted for >50% of predictive importance, with DO and water temperature contributing substantially, ~15% each. Streams were primarily governed by water temperature = 20.22%, and DO = 19.69%, followed by TP and NO₃⁻. Ponds, the most data-rich lentic system, showed NH₄⁺

500



505 =16.0% as the leading predictor, closely followed by NO_2^- , temperature, and TN = ~12% each. In reservoirs, salinity emerged as the top predictor, accounting for 16.83% of IncMSE %, ahead of NO_3^- =14.58% and pH, 11.2%. Estuaries exhibited strong influence of DO, 17.62%; NO_3^- , 15.38%; water temperature, 14.54%; and NH_4^+ , 14.32%. Lakes displayed the most balanced importance distribution, with pH, 5.86%, and TN , 5.21%, marginally leading and no single predictor exceeding 6%. Marine systems (sea, $n = 70$; coastal area, $n = 49$) yielded unreliable importance rankings due to insufficient sample size and low emission variability.

Table 3: Random Forest (RF) model performance and permutation feature importance (% increase in MSE) for predicting N_2O flux by aquatic type in GANED. Types ordered by number of complete cases. Feature importance values rounded to one decimal place

Aquatic type	n	R^2	Top predictors Feature Importance (% IncMSE)
River	1830	0.120	NO_3^- (32.97%), NH_4^+ (21.60%), DO (14.74%), Temp. (14.22%)
Stream	431	0.090	Temp. (20.22%), DO (19.69%), TP (15.09%), NO_3^- (14.97%)
Pond	2476	0.182	NH_4^+ (16.0%), NO_2^- (12.23%), Temp. (12.08%), TN (11.67%)
Reservoir	1119	0.170	Sal. (16.83%), NO_3^- (14.58%), pH (11.2%), TN (11.06%)
Estuary	438	0.225	DO (17.62%), NO_3^- (15.38%), Temp. (14.54%), NH_4^+ (14.32%)
Lake	973	0.242	pH (5.86%), TN (5.21%), TP (3.46%), NH_4^+ (3.44%)
Sea	70	-0.017	Limited reliability (low n)
Coastal area	49	-0.099	Limited reliability (low n)

510 4 Discussion

The magnitude and heterogeneity of N_2O emissions from global aquatic bodies remain challenging to quantify, largely due to the fragmented nature of existing observations and pronounced spatiotemporal inconsistency of available flux measurements (Song et al., 2024). The GANED database represents the most comprehensive observation-based synthesis of aquatic N_2O concentrations and fluxes, enabling a mechanistic interpretation of emission patterns across different water-body types rather than treating aquatic environments as a uniform source. As a global compilation, its main contribution is to make previously fragmented N_2O concentration and flux observations more accessible, comparable, and reusable through standardized structure, metadata linkage, and quality assurance. By integrating 7,386 flux measurements and 5,130 concentration records, this study elucidates that N_2O emission dynamics are strongly system-specific and primarily controlled by substrate availability. We also discuss the main features of the compiled dataset, its heterogeneity, its current coverage gaps, and the implications for future reuse and synthesis.

4.1 Aquatic ecosystems as heterogeneous N_2O observation domains



The GANED highlights that aquatic N₂O observations are distributed across strongly contrasting aquatic settings, including flowing inland waters, standing waters, estuarine systems, and marine environments. The dataset shows substantial variation in the magnitude and spread of reported concentrations and fluxes among these aquatic types, indicating that global aquatic N₂O observations should not be treated as a uniform data class. For data users, this heterogeneity is an important feature of the compilation because it supports cross-system comparison while also indicating that aquatic type, sampling context, and methodological differences should be considered explicitly in synthesis, model evaluation, and comparative analyses. The highly right-skewed distributions of both N₂O concentrations and fluxes (Fig. 8) constitute a central outcome of GANED, confirming that rivers, streams, ponds, and certain lakes exhibit higher and more variable per-area N₂O fluxes. In contrast, open seas are characterized by marginal fluxes close to zero, and by a substantial occurrence of negative fluxes (approximately 13% of observations), highlighting their episodic role as N₂O sinks (Ansari et al., 2024). The elevated fluxes observed in nutrient-enriched rivers and streams can be attributed to distinct aquatic process controls, including direct N loading (Tonina et al., 2021) and rapid microbial turnover in hyporheic zones and sediments (Yan et al., 2024). These processes often produce per-unit-area N₂O emissions that exceed those of many terrestrial soils, though driven by fundamentally different biogeochemical mechanisms. Importantly, recent evidence has indicated that low-order streams and headwaters often contribute disproportionately to N₂O emissions relative to larger rivers, a pattern linked to their high surface-area-to-volume ratios, shorter water residence times, and stronger coupling between sediments and the overlying water column (Marzadri et al., 2021; Hall Jr and Ulseth, 2020).

Our statistical analysis supports a predominantly substrate-driven control mechanism for aquatic N₂O emissions. In rivers and streams, NO₃⁻ and NH₄⁺ together account for more than 50% of RF predictive importance, indicating the co-occurrence of nitrification in aerobic sediment surfaces, denitrification within hypoxic microsites, and nitrifier-denitrification under elevated NH₄⁺ conditions (Lu et al., 2023). This mechanistic framework explains the wide distribution of riverine fluxes, particularly at sites influenced by point-source effluent or agricultural drainage, rather than implying that rivers act as persistently high N₂O emitters. GANED further reveals intermediate yet episodic flux behavior in lakes and estuaries, driven by denitrification, transient re-oxygenation pulses, and incomplete reduction of N₂O during turnover events (He et al., 2023; Liu et al., 2025). Ponds exhibit disproportionately high per-area emissions due to high surface-area-to-volume ratios, organic-rich sediments, and frequent anoxia conditions coupled with redox oscillation. In the RF model, emissions in ponds were primarily regulated by NH₄⁺, NO₂⁻, and temperature, consistent with intense nitrifier-denitrification at shallow benthic interfaces (Yang et al., 2020). In contrast, seas and coastal waters display low per-area and less variable fluxes, as N₂O production is distributed over large water volumes and oxygen-deficient zones occupy only limited spatial extents (Wang et al., 2025). The comparatively low predictive power of the RF model for these water bodies primarily reflects both limited sample sizes and intrinsically low flux variance, underscoring the need for improved database synthesis to better constrain marine N₂O emission potential, particularly within data-poor oxygen-deficient zones (Rees et al., 2022).

4.2 Spatiotemporal coverage, data biases, and remaining gaps



555 The database shows that the number of published aquatic N₂O observations has increased substantially since 2000 (Fig. 3), but
the spatial and temporal distribution of records remains uneven. Understanding the spatial and temporal variability in N₂O
concentrations and fluxes across aquatic bodies is essential for improving global emission models. Given the high global
warming potential of N₂O, robust quantification of emissions from diverse aquatic bodies, from rivers and lakes to open seas
and coastal waters, is critical for reliable climate assessments. Over the past decade, research on aquatic N₂O emissions has
560 undergone significant spatial expansion. A large fraction of sites is concentrated in Asia and North America, whereas Africa,
South America, and high-latitude regions remain comparatively underrepresented (Stanley et al., 2022). Temporal continuity
is also limited, with relatively few sites containing long-term repeated observations. Although these gaps do not diminish the
intrinsic value of GANED as a synthesis product, they establish critical constraints on representativeness that must be
considered for regional comparison, global extrapolation, or gap-filling.

565 GANED highlights both the global significance of emission hotspots and the potential underestimation associated with limited
spatiotemporal sampling. Our analysis reveals pronounced spatial bias in existing N₂O observations, with more than 60% of
observational sites located in Asia, predominantly reflecting concentrated research efforts in East Asia. This is consistent with
the regions having high anthropogenic N input and dense hydrological monitoring networks (Tian et al., 2024; Wang et al.,
2023). Conversely, large areas of Africa, South America, and high-latitude regions remain sparsely monitored (Fig. 4), creating
570 substantial blind spots in global upscaling efforts. Such spatial inequity mirrors gaps identified in other global GHG databases
(Rosentreter et al., 2021) and accentuates the risk of bias towards temperate and subtropical aquatic systems, while
underrepresenting tropical and polar aquatic fluxes, which may exhibit distinct biogeochemical dynamics for N₂O production
(Lauerwald et al., 2025).

The spatial distributions of N₂O emission, as described in Sect. 3.1, display strong latitudinal clustering within subtropical to
575 temperate regions of the Northern Hemisphere (20°N-60°N), reflecting strong sampling biases and sparse observational
coverage in the tropics and the Southern Hemisphere. This uneven coverage persists despite the high potential for elevated
emissions in warm, frequently anoxic aquatic environments characteristic of these underrepresented regions (Flecker et al.,
2022). Such geographic skewness is further reinforced by socioeconomic disparities in monitoring capacity and research
investment. Over 80% of available measurements originate from high- and upper-middle-income economies, with China,
580 contributing approximately 38%, and the United States, around 15%, dominating observed emission hotspots in intensively
fertilized river basins such as the Yangtze and Mississippi rivers, respectively. In contrast, lower-income regions account for
less than 6% of observations, restricting insights into large, comparatively understudied systems including the Congo and
Amazon basins.

585 Additionally, clear urban biases are also evident in recorded datasets, predominantly for rivers and streams. Many high-
emission records are derived from urbanized catchments, often from polluted rivers in China and Europe (Wang et al., 2022;
Li et al., 2022), where aquatic bodies receive substantial inputs of treated wastewater effluents. For instance, rivers in rapidly
urbanizing regions of East Asia exhibit N₂O fluxes that are significantly higher than those observed in natural forested or
mountain catchments. Across economic contexts, observations from high-income countries such as the United States and



Germany tend to cluster within agricultural-urban interfaces, with mean fluxes on the order of $\sim 0.1\text{--}0.15$ mmol m⁻² d⁻¹. In contrast, hotspots in upper-middle-income economies, particularly China, frequently exceed 100 mmol m⁻² d⁻¹ in densely populated basins, indicating that population-driven river modification and nutrient loading can amplify N₂O emissions more strongly than in predominantly rural, low-income settings. Furthermore, the Southern Hemisphere remains notably underrepresented, especially at latitudes south of the equator, for both concentration and flux measurements, highlighting substantial knowledge gaps in high-latitude and southern tropical aquatic ecosystems (Kong et al., 2017; Zhu et al., 2025).

Although spatial coverage of N₂O measurements has improved considerably over the past decade, recent syntheses consistently highlight persistent limitations in temporal resolution. The majority of published datasets remain based on discrete sampling campaigns, often yielding only one or a small number of observations per site, which constrains the ability to capture short-term variability and episodic emission events (Cui et al., 2024; Que et al., 2023). In GANED, the temporal dimension of data availability reveals an exponential increase in observations since the early 2000s (Fig. 3c,d), reflecting growing scientific recognition of the role of aquatic bodies in the global N₂O budget.

4.3 Biogeochemical drivers and environmental controls

The primary objective of compiling GANED was to consolidate global N₂O data, providing a foundation for more comprehensive analyses in future research. To achieve this goal, the database integrates detailed observations that allow systematic exploration of the relationships between N₂O dynamics and potential environmental drivers. This integration facilitates a more robust evaluation of the factors controlling N₂O fluxes, and is expected to enhance the accuracy and reliability of assessments of N₂O distributions from regional to global scales, while improving quantification of aquatic emissions to the atmosphere. In this study, we examined a subset of key drivers, including NH₄⁺, NO₃⁻, NO₂⁻, DO, DOC, TN, TP, water temperature, and salinity, for their correlations with N₂O flux records (Fig. 10). The inclusion of these drivers increases the utility of the dataset for comparative assessment and for evaluating patterns in relation to aquatic system type and sampling context. Correlation analyses in GANED identified that inorganic N species were the dominant positive drivers of N₂O emissions (NH₄⁺: R² = 0.943, $p < 0.001$; NO₃⁻: R² = 0.691, $p < 0.001$; NO₂⁻: R² = 0.807, $p < 0.001$), whereas DO (R² = -0.205, $p = 0.048$) and water temperature (R² = 0.07, $p = 0.46$) exhibited negative or weak associations, highlighting denitrification as a key pathway (Xu et al., 2025). These strong correlations with inorganic N species further align with the understanding that N₂O is primarily produced as an intermediate or byproduct of microbial nitrification and denitrification processes (Amado and Roland, 2017). Further, moderate positive correlations with DOC suggest that carbon availability enhances heterotrophic processes, with lower DOC:NO₃⁻ ratios (carbon-limited conditions) likely promoting incomplete denitrification and N₂O production, whereas high DOC:NO₃⁻ ratios (carbon-rich conditions) may favor DNRA or assimilation in nutrient-enriched systems. The relationship between TP and N₂O was non-significant and exhibited substantial data variability, suggesting that TP exerts minimal direct control over N₂O dynamics. Although, TP is a well-documented driver of eutrophication (Dodds and Smith, 2016), its potential indirect influence on N₂O production through algal bloom warrants further investigation (Wang et al., 2023). Additionally, the observed correlation between DO and N₂O flux indicate that oxygen availability partially regulates



N₂O production, with hypoxic conditions in warmer waters promoting high N₂O production. This pattern is consistent with observations from oxygen-deficient zones, which are estimated to contribute over half of global oceanic N₂O emissions (Wang et al., 2025).

625 Environmental drivers of N₂O fluxes vary across aquatic system types, reflecting distinct biogeochemical mechanisms. Within GANED, rivers and streams (176 of 436 study sources) exhibit high NH₄⁺ and NO₃⁻ concentrations derived from runoff and sewage inputs, which drive nitrification in oxic zones and incomplete denitrification in hypoxic reaches, further intensified by flow turbulence. Moreover, urban river showed amplified fluxes under low DO conditions. Lakes and reservoirs (94 sources) which are characterized by high mean nutrient concentrations, experience stratification that induces hypolimnetic anoxia, exacerbating denitrification. Reservoirs, in particular, exhibit pulsed N₂O emissions modulated by temperature and nutrient
630 gradients (Leon-Palmero et al., 2025). Estuaries (57 sources), despite lower means fluxes, exhibit strong influences of salinity (R² = 0.636, *p* = 0.005) and hydrodynamic mixing, which promote coupled nitrification-denitrification at interfaces (Maxey et al., 2024), with nutrient pulses from upstream further increasing variability. Ponds and aquaculture (40 sources; intermediate fluxes) are affected by intensive management practices that elevate NH₄⁺ loading and stimulate nitrification, while anoxic
635 sediments drive denitrification (Wang et al., 2023), though aeration can suppress N₂O production. In seas and coastal areas, dilution and equilibrium generally suppress fluxes, although upwelling zones with low DO can create localized N₂O hotspots via denitrification (Lachkar et al., 2024). Across this GANED database's diverse aquatic types, nutrient inputs emerge as a universal driver of emissions, with urban pollution amplifying fluxes in inland waters. In contrast, less-impacted natural waters are primarily sensitive to DO and DOC (Yoon et al., 2023), highlighting the interplay between nutrient availability and redox
640 conditions in regulating N₂O production.

Furthermore, a ML approach using the RF model revealed that inorganic N substrates, particularly NO₃⁻ and NH₄⁺, consistently rank among the top predictors of N₂O fluxes in freshwater and estuarine waters (Fig. 11), supporting substrate availability as a primary global driver (Liu et al., 2025). Collectively, these N species account for nearly 40% of total predictive importance. In addition, other environmental drivers, including DO, water temperature, and salinity, exert stronger influence in specific
645 aquatic types, reflecting the heterogeneity of N₂O emission mechanisms. Among these, the weak correlation between water temperature and N₂O flux conceals pronounced system-specific variability. Temperature exerts divergent effects across aquatic systems, depending on the dominant N₂O production pathways, substrate availability, and physical structure. In stratified systems, such as lakes and reservoirs, temperature primarily influences N₂O emissions indirectly by promoting thermal stratification, which induces hypolimnetic anoxia and decouples zones of N₂O production from surface emission interfaces (Li et al., 2024; Wu et al., 2024). In contrast, in well-mixed shallow systems, including rivers, streams, and shallow ponds, temperature directly stimulates microbial metabolic activity associated with both nitrification and denitrification, resulting in a positive relationship between temperature and N₂O emissions (Wang et al., 2023). Moreover, DO emerges as a critical
650 secondary regulator (Fig. 10), reflecting the sensitivity of N₂O production to redox conditions, whereas salinity plays a relatively minor role at the global scale (Li et al., 2019). The rankings of these predictors align with the central role of
655 nitrification and denitrification as the dominant microbial pathways of N₂O production in aquatic environments (Yao et al.,



2020; Kuypers et al., 2018). The pronounced importance of NO_3^- reflects its function as the terminal electron acceptor in denitrification and its accumulation under oxic and sub-oxic conditions, which favor incomplete reduction to N_2O rather than N_2 (Benckiser et al., 2015; Hassan et al., 2022). Similarly, high NH_4^+ feature importance indicates substantial contribution from nitrifier denitrification and hydroxylamine oxidation, processes that dominate in environments with elevated NH_4^+ loading (Zhang et al., 2023). Importantly, these outcomes are derived directly from empirical patterns within GANED rather than from model assumptions or literature-derived interpretation, providing an evidence-based foundation for targeted future process studies.

4.4 Dataset heterogeneity and implications for interpretation

A notable feature of GANED is the strong heterogeneity in reported N_2O concentrations and fluxes across sites, aquatic types, and observation methods. This heterogeneity reflects both natural environmental variation and differences in sampling design, observation frequency, and measurement approach across the source literature. Rather than attempting to remove this variability, the database preserves it in a harmonized structure so that users can evaluate patterns, filter records, and define subsets appropriate to their applications. The skewed distributions within GANED, illustrated by approximately 13% of zero or negative fluxes, combined with pronounced spatial heterogeneity, including Northern Hemisphere and economic biases, indicate that current global N_2O estimates likely remain incompletely represented, which should be acknowledged in broad-scale interpretations derived from the database. Although the sparse data from these regions suggest lower peak fluxes, they may nonetheless harbor significant emission hotspots under warming scenarios. This heterogeneity observed in GANED should be regarded as an inherent characteristic of global aquatic N_2O observations, rather than a limitation of the database. Consistent with the patterns discussed in Sect. 4.1-4.3, this variability reflects real differences in the processes governing N_2O production and emission across diverse aquatic system types and environmental contexts. Rivers and streams, which account for roughly 80% of high-emission records, highlight the potential for targeted mitigation strategies, such as riparian buffer implementation, to reduce denitrification-driven N_2O production (Slate et al., 2024; Wang et al., 2025).

In contrast, lakes and reservoirs exhibit broader but more episodic variability, pointing to a dominant role of internally regulated processes. Here, stratification and the transient accumulation of reactive N can trigger short-lived emission peaks rather than persistently elevated fluxes (Shen et al., 2025). Estuaries represent intermediate emissions, where the interplay between freshwater and marine inputs, along with organic-rich sediments, modulates the balance between N_2O production and consumption. This dynamic is evident in our RF-model analyses, which identify salinity, DO, and NO_3^- as key predictors in these systems (Fig. 11 and Fig. S2). Seas and coastal waters exhibit comparatively low median N_2O fluxes in the current dataset; however, the broad variability and limited sampling of oxygen-deficient or eutrophic boundary zones suggest that these patterns likely reflect observational gaps rather than consistently low emissions.

Overall, the heterogeneity observed in GANED is structured and interpretable. It illustrates how key drivers identified in this study, particularly inorganic N availability and redox conditions, manifest differently across hydrological contexts, providing an empirical foundation for assessing the ability of models to reproduce these cross-system contrasts in N_2O emission dynamics.



690 Furthermore, comparisons with concurrent GHG datasets (CO₂ and CH₄; Table S1; Fig. S1) reveal co-emission patterns, supporting the development of integrated aquatic GHG budgets, in which N₂O accounts for roughly 7-10% of total aquatic emissions.

4.5 Limitations, gaps, and future data needs

695 Despite broad coverage, GANED inherits several limitations from its underlying observational sources. Some aquatic environments such as intermittent or transitional systems remain weakly represented, and some pathways or ancillary variables are insufficiently documented in the available literature. The datasets exhibit pronounced spatial biases, with approximately 60% of sites located in Asia and an additional 15% in North America. This imbalance highlights the need for expanded monitoring in underrepresented regions to reduce uncertainties in global flux estimates. Long-term records were scarce in the dataset, with only 2-4% of sites having more than 10 observations, limiting the ability to resolve interannual trends. Similarly, the scarcity of long-term time series (>10 years) at most sites limits our ability to discern climate-driven trends, interannual variability, and the effects of extreme events on N₂O emissions. Currently, no continuous records exist for N₂O fluxes, and most available concentration data are limited to a small number of geographically clustered sites.

700 Furthermore, investigations in arid drainage systems remain particularly limited, more so than anticipated based on their relatively small river surface area. This specific gap likely reflects the dominant research focus on streams and rivers as major, continuous contributors to the global atmospheric N₂O budget (Tikkasalo et al., 2025), leading to the assumption that arid systems contribute relatively little. In such systems, dry-wet alternations and transient inundation events can substantially influence N₂O dynamics by shifting redox conditions, altering nutrient availability, and stimulating microbial processes (De Klein et al., 2017). For instance, ephemeral streams, floodplains, and reservoirs in arid and semi-arid regions may remain inactive during dry phases but emit significant N₂O pulses upon rewetting (Webb et al., 2023; Stringer et al., 2021). Thus, what may initially appear as a “gap from arid zones” is more accurately a lack of representation of intermittent and transitional aquatic systems, which likely contribute disproportionately to regional N₂O budgets (Upadhyay et al., 2023). One of the most persistent gaps in current flux data collection is the limited quantification of plant-mediated emissions. Although such fluxes can account for a substantial proportion of total N₂O release in wetlands and shallow lakes (Bodmer et al., 2021), their contribution to other aquatic systems remains largely unquantified. In our GANED database, plant-mediated fluxes were excluded, as only two studies were found to explicitly report this pathway in streams (Sanders et al., 2007; Wilcock and Sorrell, 715 2008). These limitations identify priorities for future data collection and database expansion rather than shortcomings unique to GANED itself.

720 Future efforts should prioritize the development of automated sampling technologies, such as field-deployable online GHG monitoring systems for dissolved N₂O (Piatka et al., 2024), to improve temporal resolution and reduce observational gaps. Integrating these high-frequency measurements with machine-learning approaches will further enhance gap-filling and support predictive modeling across diverse aquatic systems. Additionally, coupling in situ observations with remote-sensing products and global hydrological models will enable more robust upscaling of flux estimates and help narrow uncertainties in the global



725 aquatic N₂O budget. By making GANED openly accessible, we aim to facilitate continuous community-driven updates that strengthen flux estimates and support forthcoming IPCC AR7 assessments (IPCC, 2023). Collectively, these advances will help inform targeted mitigation strategies, ranging from wetland restoration to precision nutrient management, to reduce aquatic N₂O emissions and contribute to achieving the 1.5 °C climate targets.

5 Code and data availability

The scripts used for data harmonization, quality-control processing, statistical analysis, and figure generation are archived in Environmental Data Initiative (EDI) and are available at: <https://doi.org/10.6073/pasta/4a086e49a4f308679b951293b380e7b9> (Nazir et al., 2026). The archived materials include the R scripts used for unit conversion, data cleaning, statistical summaries, Random Forest analysis, and manuscript figure production. Dataset sheets and code archives are documented in repository (edi.2267.1) to support reproducibility.

6 Conclusion

735 The Global Aquatic N₂O Emission Database (GANED) represents the most comprehensive synthesis of empirical N₂O observations, encompassing 5,130 dissolved concentration records and 7,386 flux measurements from diverse aquatic types, including rivers, streams, lakes, reservoirs, ponds, estuaries, seas, and coastal areas, spanning 1980-2023. By harmonizing data from 426 peer-reviewed sources and explicitly linking N₂O dynamics to co-measured biogeochemical drivers, GANED reveals pronounced heterogeneity in emission patterns across aquatic bodies. GANED improves access to previously fragmented information and supports more transparent comparison across aquatic system types. Rivers and streams emerge as dominant emission hotspots, exhibiting the highest and most variable fluxes, primarily driven by elevated inorganic N substrates, such as NH₄⁺, NO₃⁻, and NO₂⁻. In contrast, lakes, reservoirs, estuaries, and ponds show intermediate levels, and open seas and coastal waters consistently display lower emissions, occasionally functioning as weak N₂O sinks. Random forest analyses further confirm substrate availability as the primary global control on aquatic N₂O production, with redox conditions such as DO exerting secondary but system-specific influences. These patterns highlight the central role of microbial nitrification, denitrification, and nitrifier-denitrification pathways in regulating aquatic N₂O emissions.

745 GANED advances N₂O research by providing the first global dataset that systematically distinguishes emission mechanisms across aquatic system types through interlinked biogeochemical metadata. Critically, it demonstrates that aquatic N₂O emissions are primarily substrate-controlled rather than climate-controlled, and reveals systematic biases from concentration-only measurements that may underestimate fluxes in dynamic systems like streams. Despite substantial progress in observational coverage since 2000, GANED highlights persistent spatial biases, such as concentrated observations (~60% of sites) in Asia, and critical data gaps in Africa, South America, high-latitude regions, and low-income economies. These economic disparities result in well-documented emissions in high-income countries (China, the United States) while substantial



755 gaps persist in low- and middle-income regions. These imbalances, together with the scarcity of long-term time series and the underrepresentation of intermittently connected or episodic systems, likely contribute to considerable uncertainty in current global aquatic N₂O budgets. Existing estimates suggest that inland waters, estuaries, and oceans together account for approximately 25-32% ($\approx 5.3 \text{ Tg N yr}^{-1}$) of total anthropogenic N₂O emissions, yet the pronounced economic and latitudinal skew in GANED implies that contributions from tropical and Southern Hemisphere systems may be systematically underestimated.

760 GANED marks a pivotal shift from documenting “where” emissions occur to understanding “why” they differ across systems, enabling targeted mitigation interventions. Moreover, GANED provides an essential empirical foundation for refining national GHG inventories, improving process-based and machine-learning models, and identifying targeted mitigation opportunities. By elucidating mechanistic rather than purely spatial controls on emissions, the database shifts the paradigm from treating aquatic systems as uniform “hotspots” toward recognizing substrate-driven heterogeneity amenable to management. In particular, the results highlight the potential effectiveness of reducing N loading in rivers, streams, and urban-impacted waters. Making GANED reusable synthesis product for comparative analysis, intercomparison, and future refinement of aquatic N₂O 765 assessments will facilitate community-driven expansion, integration with emerging high-frequency and remote-sensing products, and more robust constraints on aquatic N₂O contributions in future IPCC assessments.



7 Appendices

770 Appendix A: GANED tables and variables

Table A1: Column titles and description of their content for the GANED “Data Source” component.

Column title	Description
Title	Title of data source
Author	Last name of the leading author.
Type	Type of data source either its Journal article, master’s thesis or doctoral dissertation
SRC-Name	Name of the data source outlet, such as a journal, data repository. For studies with both published papers and associated datasets, the journal name is listed here.
Pub-year	Year of publication, data release, or acquisition of an unpublished dataset.
Source-ID	Unique identifier assigned to the data source.
Paper-DOI	DOI or hyperlink for the journal article or other publication reporting the N ₂ O data.
Additional-data	“Marked “Yes” if additional data were obtained directly from the author for any field.



Table A2: Column titles and content description for the GANED “Sampling Sites”.

Column title	Definition
Source-ID	Unique identifier assigned to each data source in the ‘Data Source’ dataset.
Site-ID	Unique identifier for each sampling site.
Continent	Asia, Africa, North America, South America, Antarctica, Europe, Australia
Country	Countries categorized by their continents/sub-continent
Aquatic-type	Water bodies based on their ecological characteristics (lakes, rivers, ponds, estuary, reservoirs, coastal area, sea)
Site-Name	Name assigned to the site.
Region-Name	Name of the area from the data source taken based on the specified region.
Latitude	Geographic latitude (decimal degrees) in the WGS84 ensemble (EPSG:4326).
Longitude	Geographic longitude (decimal degrees) in the WGS84 ensemble (EPSG:4326).
Elevation-m	Reported elevation of the site (meters above sea level).
Slope-m-per-m	Reported channel slope (m m^{-1}).
Status	N_2O Concentration, flux or records containing both
Channel-type	Codes describing specific channel attributes or site conditions; categories and definitions are provided in Table 1.
Comments	Additional information or clarifications regarding the site or data source (if any).



775 **Table A3: Column titles and definitions for the GANED “Concentration dataset”.**

Column title	Definition
Source-ID	Unique identifier for each publication in the ‘Data Source’ component.
Site-ID	Unique identifier for each sampling site in the ‘Sampling Sites’.
Site-Name	Unique name of the sampling site from the “Sites” component.
Conc-Name	Identifier for the sampling event at a site. Matches Conc-Name when concentration measurements exist for the same site-date combination.
Sampling-year	Corresponding year of observation recorded.
Sampling-month	Monthly precision of observation data of targeted site.
Date-start	First date of sample collection.
Date-end	Last date of sample collection; identical to Date-start if data are not aggregated over time.
Aggregated-Space	Indicates whether values are averaged across more than one site (“yes” or “no”).
Aggregated-Time	Indicates whether values are averaged across more than one sampling date (“yes” or “no”).
Observation type	sources providing only gas concentration marked as ‘Conc’ and for Conc+flux marked as ‘Both’.
N ₂ O concentration	Concentration records of N ₂ O (μmol L ⁻¹) from sampling sites.
CO ₂ concentration	CO ₂ concentration (μmol L ⁻¹) measured concurrently with N ₂ O.
CH ₄ Concentration	CH ₄ concentration (μmol L ⁻¹) measured concurrently with N ₂ O.



Table A4: Column titles and definitions for the GANED “Flux dataset”.

Column Title	Definition
Source-ID	Unique paper identifier from the data source
Site-ID	Unique site identifier from the sampling sites
Site-Name	Unique site name from the sampling sites
Flux-Name	Unique name for the sampling event at the site; same as Conc-Name in the concentration table if both concentration and flux data for the same site-date combination are available
Date-start	First sampling date
Date-end	Last sampling date; this is the same date as Date-start if data are not aggregated over time.
Aggregated-Space	Yes or no; “yes” if N ₂ O data entered are averages from >1 site
Aggregated-Time	Yes or no; “yes” if N ₂ O data entered are averages from >1 date
Flux-Method	Methodological category used to measure diffusive gas flux. Categories (with brief explanations in italics) are the following. – <i>chamber (unspecified) – unspecified response</i> use of an unspecified type of chamber (static or free-floating) and pattern of change gas concentration – <i>floating chamber – unspecified response</i> chamber unrestrained and able to float downstream during flux measurement – <i>floating chamber – linear response</i> – <i>conc+k</i> diffusive flux calculated using the equation $\text{flux} = k (C_w - C_{eq})$, where k is gas exchange coefficient, C_w =N ₂ O concentration measured in water, and C_{eq} =N ₂ O concentration in water in equilibrium with the atmosphere
N ₂ O-Flux	Total measured N ₂ O flux (mmolm ² d ⁻¹)
CO ₂ _Flux	measured total CO ₂ flux (mmolm ² d ⁻¹)
CH ₄ -Flux	measured total CH ₄ flux (mmolm ² d ⁻¹)
Total-N ₂ O-Flux-SD	Standard deviation of the mean total N ₂ O flux
Total_N ₂ O_Flux_Median	Median measured total N ₂ O flux (mmolm ² d ⁻¹)
Comments	Any additional relevant information regarding data entered in this row
Flux-unit	Common unit for all diffusive N ₂ O flux data (mmol m ⁻² day ⁻¹)
DO-mgL	Dissolved oxygen (mg L ⁻¹).
pH	pH of the water sample.
NO ₃	Nitrate or nitrite + nitrate concentration (μmol L ⁻¹) measured concurrently with N ₂ O.
NH ₄	Ammonium concentration (μmol L ⁻¹) measured concurrently with N ₂ O.
TN	Total nitrogen (TN) measured concurrently with N ₂ O.
TP	Total phosphorus (TP) measured concurrently with N ₂ O.
DOC	Dissolved organic carbon (DOC) measured concurrently with N ₂ O.
WaterTemp-degC	Water temperature (_C) measured concurrently with N ₂ O.



8 Supplement.

The supplement related to this article is available online at: <https://doi.org/XXXXX/essd.-supplement>.

780 9 Author Contribution

Muhammad Junaid Nazir and Longfei Yu conceptualized the study and writing the original draft. Jing Wei investigated, supervised and review the study. Yunjie Zhang, Jing Zou and Wenping Yuan conducted the analysis, data curation and prepared figures. All authors contributed to the investigation, visualization, data acquisition, data entry, data checking, writing, reviewing and editing the original draft.

785 10 Competing interests

The authors declare that they have no conflict of interest.

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