



Title: ABCFlux v2: Arctic–boreal CO₂ and CH₄ monthly flux observations and ancillary information across terrestrial and freshwater ecosystems

Journal: Earth System Science Data

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290 Abstract

Measurements of surface-atmosphere carbon dioxide (CO₂) and methane (CH₄) fluxes have been relatively sparse across the Arctic tundra and boreal biomes, causing significant uncertainties in carbon budget estimates from the region. While the availability of Arctic-boreal carbon flux data has increased substantially over the past decade, the data have remained spread across different repositories, scientific articles, and unpublished sources, making it difficult to leverage. Here we present a new dataset of monthly Arctic-boreal carbon fluxes (ABCFlux v2) across terrestrial (wetlands and uplands) and freshwater (lakes and rivers) ecosystems compiled from previous syntheses including the Arctic-boreal CO₂ flux database (ABCFlux v1), the Boreal-Arctic Wetland and Lake Methane Dataset (BAWLD-CH₄), and the Global River Methane Database (GRiMeDB). In addition, we consider data from general-purpose (e.g., Zenodo) and flux network repositories, literature, and site principal investigators. The dataset includes surface-atmosphere CO₂ fluxes of gross primary production (GPP), ecosystem respiration (Reco), and net ecosystem exchange (NEE), alongside CH₄ fluxes. For aquatic ecosystems, we split CH₄ fluxes into diffusive and ebullitive flux pathways, and included potential emissions from transient storage in the water column (“storage fluxes”), alongside CO₂ and CH₄ concentrations dissolved in the surface water. Fluxes are measured through a variety of methods including chamber and eddy covariance techniques alongside bubble traps, ice-surveys, and concentration-based turbulence-driven modelling in aquatic ecosystems. The monthly flux data are reported together with supporting methodological and environmental metadata. The resulting ABCFlux v2 has 23,656 flux site-months, 8,182 concentration site-months, and 199 seasonal observations from 1,024 sites, and includes 55,560 reported fluxes (i.e. sum of GPP, Reco, NEE, and CH₄ fluxes) from the years 1984 to 2024. The majority of monthly observations occurred after 1999. Wetlands had the highest number of site-month observations (8,641), followed by boreal forest (6,981), lotic ecosystems (6,275), lentic ecosystems (3,725) and upland tundra (3,308). Measurements of CO₂ dominated the dataset across most ecosystem types (25,101) except for lentic ecosystems, where CH₄ flux site-months (3,024) were more frequent than CO₂ flux site-months (2,858). Overall, ABCFlux v2 includes 158% more site-months for terrestrial CO₂ flux data compared to ABCFlux v1. Integrating and updating BAWLD-CH₄ flux data from growing season averages to monthly fluxes resulted in 5,671 site-months of chamber CH₄ data compared to 762 site-years. This collaborative initiative, involving contributions from over 260 researchers, provides a comprehensive overview of the current state of the Arctic-boreal carbon flux network and its data, and serves as an important step in reducing uncertainties in Arctic-boreal carbon budgets and in enhancing our understanding of climate feedbacks. The data can be accessed at ORNL DAAC at <https://doi.org/10.3334/ORNLDAAC/2448> (Virkkala et al., 2025b).



1. Introduction

The Arctic-boreal region has historically been sparsely measured for carbon dioxide (CO₂) and methane (CH₄) fluxes (Baldocchi et al., 2018; Pallandt et al., 2022). This data sparsity, together with rapid warming, changes in hydrology, permafrost thaw, and other environmental shifts (Biskaborn et al., 2019; O'Neill et al., 2023; Rantanen et al., 2022; Webb et al., 2022), has created significant uncertainties in Arctic-boreal carbon budget estimates (Hugelius et al., 2024; Treat et al., 2024), hindering our capability to understand the fate of its large soil organic carbon stocks (Hugelius et al., 2014; Schuur et al., 2008, 2022). However, over the past decade, the availability of Arctic-boreal carbon flux data has increased substantially (Vogt et al., 2024), giving rise to new flux synthesis datasets (Kuhn et al., 2021; Virkkala et al., 2022) that have been widely used to improve our process-understanding (Kuhn et al., 2021), model intercomparisons (Tao et al., 2021; Treat et al., 2024), site-level trend assessments (See et al., 2024), and Arctic-boreal carbon budgets (Kuhn et al., 2025; Ramage et al., 2024; Virkkala et al., 2025a; Vonk et al., 2025; Yuan et al., 2024).

While significant progress has been made in Arctic-boreal carbon flux datasets, most existing syntheses do not include recently published flux data from 2020 onward, a period marked by rapid warming (Minobe et al., 2025) and increased disturbances - such as fires (Euskirchen et al., 2024; Kelly et al., 2024; Korkiakoski et al., 2023), thermokarst (Jorgenson et al., 2025), and vegetation shifts (Frost et al., 2025). Moreover, global flux repositories often fail to include Arctic-boreal-specific variables, such as permafrost, high-latitude vegetation types, or lake origin (e.g. glacial or thermokarst lake). Additionally, CO₂ and CH₄ fluxes from freshwater and terrestrial ecosystems have typically been studied and synthesized separately, hindering a holistic understanding of the Arctic-boreal carbon cycle, which can also lead to double counting of carbon fluxes (Casas-Ruiz et al., 2023; Kyzivat and Smith, 2023; Thornton et al., 2016). Finally, data remain scattered across repositories, scientific publications, and unpublished sources, making it difficult to understand how comprehensive and representative the current network of Arctic-boreal flux measurements is.

To address these research gaps, we compiled a dataset of Arctic-boreal CO₂ and CH₄ fluxes in terrestrial and freshwater ecosystems (ABCFlux v2) from flux repositories, data syntheses, literature, and data contributors, which are presented here. We built upon recent syntheses (Golub et al., 2023; Kuhn et al., 2021; Song et al., 2024; Stanley et al., 2022; Virkkala et al., 2022) and earlier Arctic-boreal terrestrial and freshwater CO₂ and CH₄ flux datasets (Belshe et al., 2013; McGuire et al., 2012; Natali et al., 2019; Olefeldt et al., 2013; Treat et al., 2018; Wik et al., 2016b), including ABCFlux v1 (Virkkala et al., 2022). The structure of ABCFlux v2 follows v1, which synthesized monthly terrestrial CO₂ fluxes. Compared to v1, ABCFlux v2 includes not only updated terrestrial CO₂ fluxes, but also terrestrial CH₄ fluxes. In addition, we expanded the dataset to freshwater ecosystems, including lentic (lakes, ponds, reservoirs, pools) and lotic (rivers and streams) waterbodies, and synthesized carbon fluxes and surface concentrations of dissolved CO₂ and CH₄. We also added several new variables to ABCFlux v2 from the BAWLD-



CH₄ and GRiMeDB database to include variables specific to freshwaters (see section 3 for details) (Kuhn et al., 2021; Stanley et al., 2023).

ABCFlux v2 comprises several measurement techniques that provide different measurement frequencies across multiple ecosystem scales. Eddy covariance is a common method for measuring temporal dynamics in terrestrial carbon fluxes on ecosystem scales and quantifying year-round net carbon emissions. In many cases, eddy covariance data are actively shared, processed and curated in global (Pastorello et al., 2020) and regional (Heiskanen et al., 2022; Novick et al., 2018; Ueyama et al., 2025) flux networks. However, not all Arctic-boreal sites are part of these networks, and overall, the Arctic-boreal region has a particularly low coverage of eddy covariance towers (Pallandt et al., 2022), especially for lakes (Eugster et al. 2022, Golub et al. 2023). Furthermore, eddy covariance, which aggregates fluxes over ecosystem scales (hundreds of meters), often cannot resolve issues regarding local-scale spatial heterogeneity in emission and uptake driven by small-scale variation in vegetation, hydrology, soil microclimate (Chen et al., 2012; Virkkala et al., 2024). Moreover, CH₄ flux estimates derived from eddy covariance generally do not distinguish between the multiple CH₄ emission pathways (diffusion, ebullition, plant-mediated transport; but see (Ueyama et al., 2023)), which are important to understanding processes controlling the total CH₄ fluxes (Bastviken et al., 2004; Kyzivat et al., 2022). Thus, relying solely on eddy covariance towers is insufficient for a comprehensive understanding of Arctic-boreal carbon fluxes and the ability to predict current and future emissions more accurately, emphasizing the value of other kinds of flux measurements including small-scale, ground-based techniques.

Small-scale, ground flux techniques most often consist of static or automated chamber measurements in terrestrial and freshwater ecosystems and concentration-based turbulence-driven modeling approaches in freshwaters (Kuhn et al., 2021; Stanley et al., 2023; Virkkala et al., 2021). Chamber techniques can assess fluxes across small footprints (ca. 0.3 to 1 m²), allowing for detailed assessments of environmental controls on fluxes (Kuhn et al., 2021). Flux gradient approaches, wherein gas samples are taken from the air and throughout the soil or snow profile to estimate net flux, have also been used in some terrestrial sites (Pirk et al., 2016). In freshwaters, diffusive fluxes can also be estimated from measurements of gas concentrations dissolved in the surface water and using turbulence-driven modelling approaches based on gas transfer velocities (Klaus and Vachon, 2020; Vachon and Prairie, 2013). Ebullitive fluxes can be derived from concentration bursts during chamber measurements (Bastviken et al., 2004), but are most commonly assessed using bubble traps, which can be coupled with ice-bubble surveys to reduce spatial uncertainties (Huttunen et al., 2001; Walter Anthony and Anthony, 2013; Wik et al., 2013). For freshwaters, ground-based measurement techniques are also used to capture storage fluxes. Storage fluxes refer to the sudden diffusive efflux triggered by lake turnover in spring and fall, or by gas that accumulates under the frozen layer of a lake and is emitted to the atmosphere when the ice melts in the spring (Jammot et al., 2015). Non-eddy covariance measurements of storage fluxes are measured from the difference between measured waterbody content of dissolved gas before and after the turnover or ice-off (Karlsson et al., 2013). While these micro-scale approaches advance the understanding of local processes and spatial variability in carbon fluxes, they are also accompanied by uncertainties due to potentially limited spatial and temporal representativeness, the disturbance that collars, floating bubble



traps, and chambers can cause on the ground or water surface (Welles et al., 2001), oxidation
410 in the water column prior to ice out (Pajala et al., 2023), the wide range of available methods to
determine gas transfer velocities (Hall and Ulseth, 2020; Klaus and Vachon, 2020; Raymond et
al., 2012), and the temporal representativeness of the manual sampling campaigns (Golub et
al., 2023; McGuire et al., 2012; Wik et al., 2016). Automated chambers and continuous
415 concentration measurements in terrestrial and aquatic ecosystems, provide more temporally
representative sampling relative to those from more limited manual sampling campaigns, but do
not solve for potential artifacts derived from ground disturbances. Overall, combining all fluxes
measured with these different techniques (Table 1) is an important benefit of ABCFlux v2
compared to other efforts focused on a single flux measurement technique or gas species.

In this community-driven effort, we integrated surface-atmosphere CO₂ and CH₄ fluxes into a
420 single, unified Arctic-boreal-specific dataset. Below, we provide a description of the dataset and
a summary of the flux network, and synthesize flux magnitudes across key land cover types,
spanning both terrestrial and freshwater ecosystems.

Table 1. A summary of the measurement techniques, carbon flux and concentration
observations, and key ecosystems included in ABCFlux v2. Terrestrial classes include dry and
425 moist tundra, bogs, fens, marshes, tundra wetlands, permafrost bogs, and boreal forest
ecosystems. Freshwater classes include lentic and lotic waterbodies. Storage flux refers to the
transient accumulative release of gasses during ice-out and water column mixing events.
Percentages represent the percent of site-months from the respective ecosystem (i.e. % of
terrestrial site-months or % of freshwater site-months). Percentages for the aquatic ecosystems
430 do not add up to 100 % due to overlaps, for example, where both diffusion fluxes and dissolved
concentrations were measured.

Measurement technique	Terrestrial	Freshwater
Eddy covariance	Eddy covariance (CO ₂ , CH ₄ flux) (64.1%)	Eddy covariance (CO ₂ , CH ₄ flux) (4.6 %)
Non-eddy covariance	Manual or automated chamber (CO ₂ , CH ₄ flux) (34.1%) Flux gradient approach (CO ₂ flux) (1.7%)	Diffusion based on chambers and turbulence-driven modeling derived from concentrations (CO ₂ , CH ₄ flux) (42.4 %) Ebullition based on bubble traps or chambers (CO ₂ , CH ₄ flux) (4.7 %) Storage flux: Water column survey (CO ₂ , CH ₄ flux) (3.6 %) <i>Note: Total CH₄ flux = CH₄ diffusion + CH₄ ebullition</i>
Dissolved gas concentration		Dissolved concentration at the water surface (CO ₂ , CH ₄) (87.0 %)



2. Data compilation and search

ABCFlux v2 focuses on the Arctic tundra and boreal biomes, as characterized in Dinerstein et al. (2017; Figure 1). It also includes some hemiboreal sites located within 500 km south of the boreal biome boundary, when data were available through public repositories or provided by data contributors (~2.5% of sites). These sites were included because hemiboreal ecosystems share key characteristics with boreal systems and may provide insight into potential trajectories of boreal ecosystems under changing climate (Berner and Goetz, 2022). The dataset compiles in situ measured CO₂ and CH₄ fluxes aggregated to monthly time periods (unit: g C m⁻² per month, i.e. g CO₂-C m⁻² per month for CO₂ fluxes and g CH₄-C m⁻² per month for CH₄ fluxes) from terrestrial and freshwater ecosystems, including boreal forests, wetlands, tundra, lentic and lotic waterbodies. In forested ecosystems, chamber CO₂ measurements were excluded as they typically do not represent the whole ecosystem fluxes (i.e. chamber measurements exclude trees). However, for CH₄, we accepted measurements of understory CH₄ fluxes as the CH₄ fluxes from trees are expected to be minimal (see Sect. 8.3 for challenges associated with this assumption). For freshwater ecosystems, we also included monthly average CO₂ and CH₄ concentrations dissolved in surface waters because this information helps to understand flux dynamics and can also be used to estimate fluxes (e.g. (Holgerson and Raymond, 2016)). We excluded flux data from experimental manipulation sites, with the exception for control sites within manipulations experiments. We included data from managed forests but excluded croplands, as croplands typically undergo intensive annual management (e.g., tillage, fertilization, and harvesting). In contrast, managed forests are generally managed on decadal timescales, which allows them to retain some functional characteristics of natural ecosystems, particularly in later stages of regrowth.

We used a monthly aggregation interval as it is a common and standard temporal frequency across many site-level, synthesis, and modeling studies, remote sensing products, and process models. However, there were some seasonally-aggregated data from previous syntheses and studies that we were not able to incorporate in a monthly format; these were kept in the dataset in seasonal format but they make up only <1% of the dataset. Monthly fluxes were primarily found derived by multiplying daily means (g C m⁻² d⁻¹) by the number of days in each month to calculate monthly cumulative fluxes (g C m⁻² month⁻¹), although methods varied based on available data and temporal resolution (see section 5.1).

The data compilation steps are detailed in a flow chart (Fig. 2). We compiled and harmonized data from syntheses, global and regional flux repositories and general data repositories (Table 2), publications, and direct submissions from data contributors. In cases where data for the same sites and periods were available from multiple sources, we prioritized user-contributed data over data extracted from repositories, syntheses, and publications. This prioritization was chosen due to the benefits associated with the expertise of data contributors with data processing at their site (e.g., gap-filling), and the inclusion of ancillary data. There are no



duplicate observations in the main dataset, i.e. only one flux estimate per flux type is given for each site-month, even though there may be multiple different data sources.

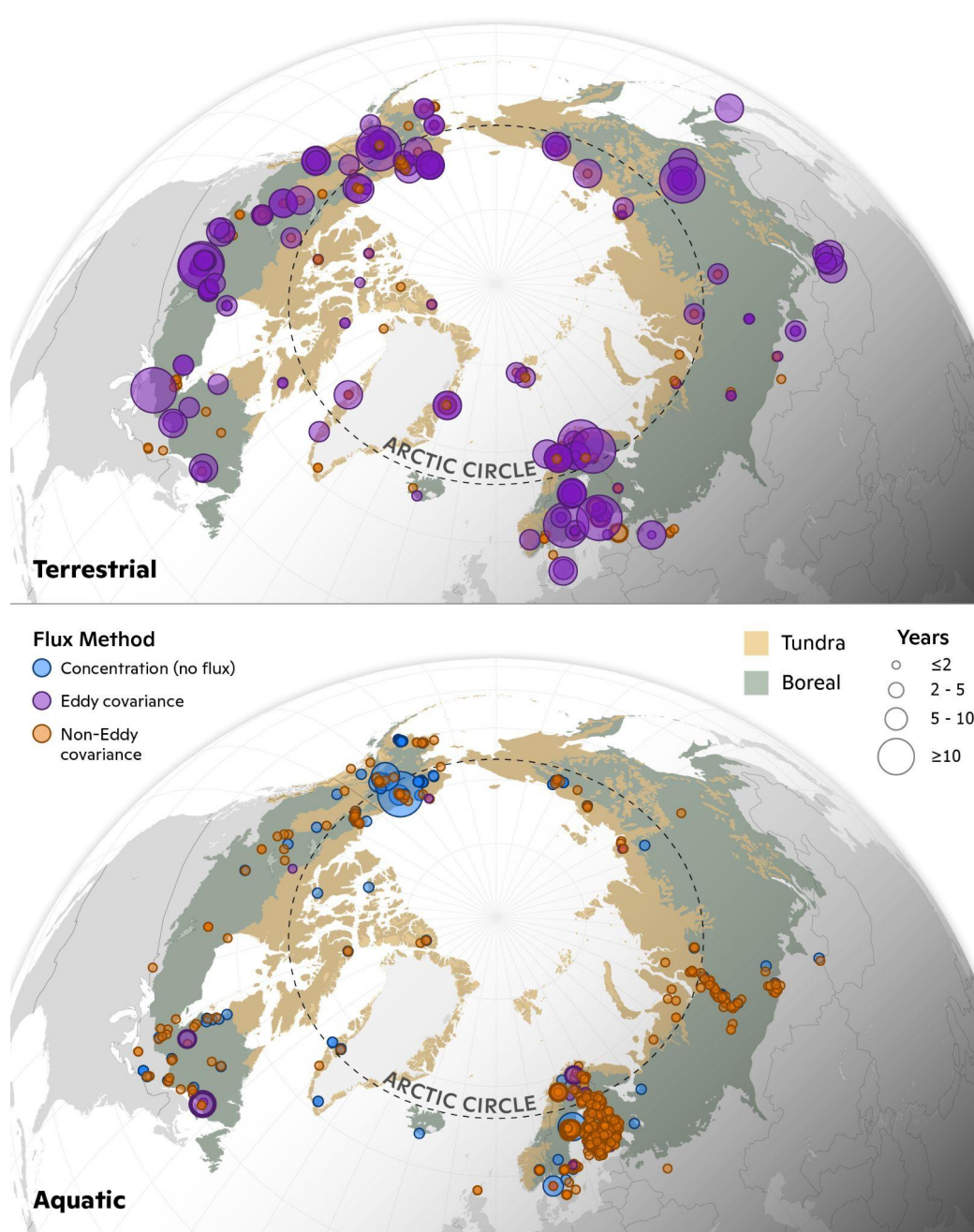




Fig. 1. Numbers of site-months for terrestrial CO₂ and CH₄ flux sites, and aquatic CO₂ and CH₄ flux sites across the Arctic-boreal region. The number of months represented by circles refers to total, not necessarily consecutive months. See Supplementary Fig. 7 for zoomed-in maps for the densely measured areas.

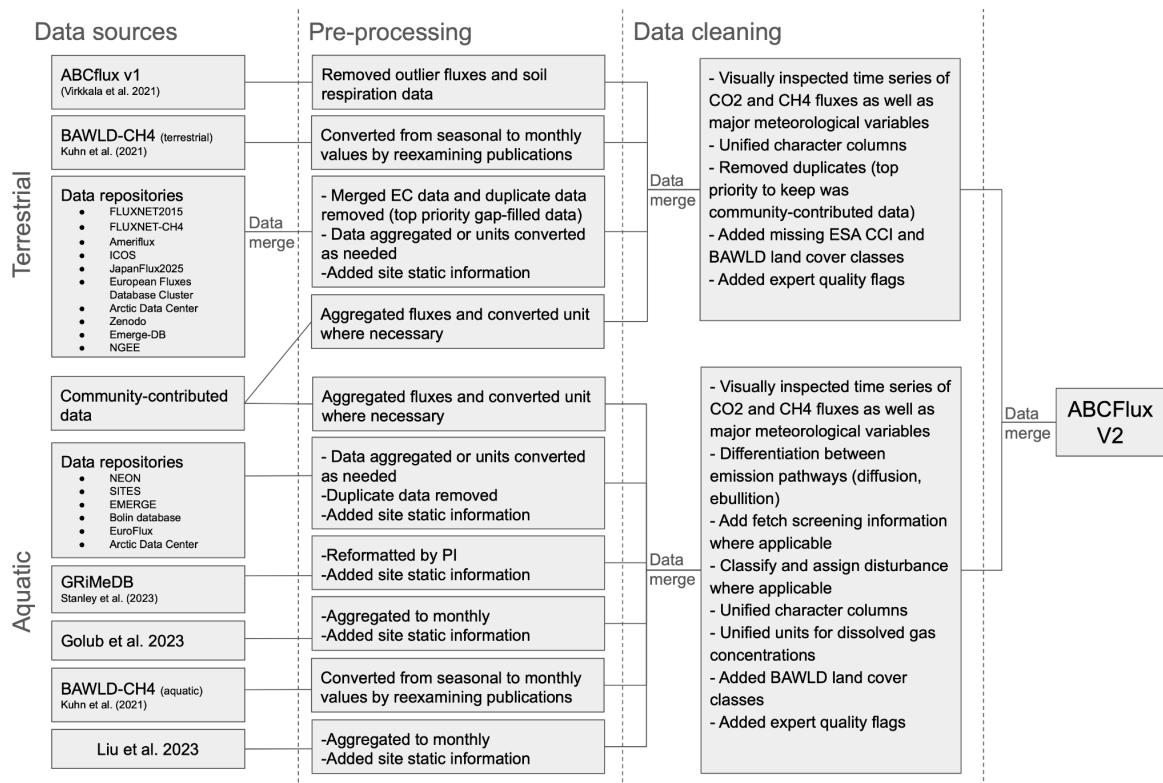


Fig. 2. Flowchart representing the main data compilation steps to produce ABCFlux v2.

2.1 Recent data syntheses

We incorporated data from various data syntheses into our dataset (Table 2). We integrated terrestrial CO₂ flux data included in ABCFluxv1 into our v2 dataset with some modifications to the original dataset. Notably, GPP values in v1 were reported as negative, whereas in v2 they are presented as positive to align with the convention used throughout this synthesis. We removed soil respiration data from forest floors as our focus here was on whole-ecosystem CO₂ fluxes. Some monthly chamber fluxes within the v1 dataset were spatial replicates (same coordinates and land cover) and for the purposes of v2, we aggregated these by taking a mean flux. We updated the soil moisture classification for several sites as well as several site names from v1 to be consistent with site names used in v2. Additionally, some data from v1 were replaced by more recent versions of the data found in flux repositories.



495 The terrestrial and aquatic CH₄ chamber flux data from the BAWLD-CH₄ dataset (Kuhn et al., 2021) were originally presented as average daily fluxes over the growing season (wetlands and uplands) and open-water season (lentic ecosystems) for each site, with the last year of data collection being 2019 (Table 2). For sites where data at monthly resolution were available in publications, we extracted monthly flux and supporting environmental data. If monthly flux or other ancillary data were not extractable from the literature, we reached out to the lead author for data contribution. For sites without monthly aggregated fluxes available, we present the fluxes as seasonal values and provide the start and end date of the measurement period.

500 Data in the Global River Methane Database (GRiMeDB, Table 2; (Stanley et al., 2023)) were aggregated to monthly resolution in ABCFlux v2. Fluxes in the global CO₂ lake and reservoir synthesis (Golub et al., 2023) were presented in half-hourly timesteps (Table 2), but were aggregated to monthly resolution. Dissolved gas concentrations in rivers from Liu et al. (2022) were converted to monthly averages. Overall, a significant portion of data (30%) incorporated in ABCFlux v2 originated from these terrestrial and aquatic data syntheses.

505 **Table 2.** Global and regional flux syntheses and repositories used in the data compilation of terrestrial and/or aquatic data. Information about the ecosystem, the name, reference or web page, flux method, gas species, temporal resolution, spatial extent and period of data coverage of the respective synthesis or repository are given.

Ecosphere	Type	Ecosystem	Name of repository or synthesis	Reference or web page (last access: August 4, 2025)	Flux method	Gas	Temporal resolution	Spatial extent	Data coverage
Both	Synthesis	Wetland, Lentic	BAWLD-CH ₄	Kuhn et al. (2021)	Mostly chamber	CH ₄	Seasonal	Arctic-boreal	1984-2020
Terrestrial	Synthesis	Terrestrial	ABCFlux v1	(Virkkala et al., 2022)	Eddy covariance, chamber	CO ₂	Monthly	Arctic-boreal	1989-2020
Terrestrial	Repository	Terrestrial	FLUXNET2015	(Pastorello et al., 2020); https://fluxnet.org/data/fluxnet2015-dataset/	Eddy covariance	CO ₂	Half-hourly to yearly	Global	1994-2014
Terrestrial	Repository	Terrestrial	FLUXNET-CH ₄	(Delwiche et al., 2021); https://fluxnet.org/data/fluxnet-ch4-community-product/	Eddy covariance	CH ₄	Half-hourly to daily	Global	2006-2018
Terrestrial	Repository	Terrestrial	Ameriflux	(Chu et al., 2023); https://ameriflux.lbl.gov/	Eddy covariance	CO ₂ , CH ₄	Half-hourly to yearly	Americas	1994-2021



Terrestrial	Repository	Terrestrial	ICOS	(Warm Winter 2020 Team and ICOS Ecosystem Thematic Centre, 2020); https://www.icos-cp.eu/	Eddy covariance	CO ₂ , CH ₄	Half-hourly to yearly	Europe	1996-2023
Terrestrial	Repository	Terrestrial	JapanFlux2024	Ueyama et al. (2025)	Eddy covariance	CO ₂ , CH ₄	Half-hourly to yearly	Japan and East Asia	1990-2023
Terrestrial	Repository	Terrestrial	European Fluxes Database Cluster	(Valentini, 2003); https://www.europe-fluxdata.eu/	Eddy covariance	CO ₂	Daily to monthly	Europe	1996-2008
Terrestrial	Repository	Terrestrial	Arctic Data Center	https://arcticdata.io/	Eddy covariance, chamber	CO ₂ , CH ₄	Varies	Global	Varies
Terrestrial	Repository	Terrestrial	Zenodo	https://zenodo.org/	Chamber	CO ₂ , CH ₄	Varies	Global	Varies
Terrestrial	Repository	Terrestrial	Next Generation Ecosystem Experiments	https://ngee-arctic.ornl.gov/	Chamber	CO ₂ , CH ₄	Varies	USA	Varies
Terrestrial	Repository	Terrestrial	EMERGE-DB	https://emerge-db.asc.ohio-state.edu/	Chamber	CO ₂ , CH ₄	Varies	Sweden	Varies
Aquatic	Synthesis	Lentic		Golub et al. (2023)	Eddy covariance	CO ₂	Half-hourly	Global	2005-2015
Aquatic	Synthesis	Lotic	GRiMeDB	Stanley et al. (2023)	Chamber, concentration	CO ₂ , CH ₄	Daily to seasonal	Global	1973-2021
Aquatic	Synthesis	Lotic		Liu et al. (2022)	Concentration	CO ₂	Daily to seasonal	Global	
Aquatic	Repository	Lentic, lotic	NEON	https://www.neonscience.org/	Concentration	CO ₂ , CH ₄	Half-hourly	US	2016-2022
Aquatic	Repository	Lentic	Bolin	https://bolin.su.se/data	Chamber, concentration, ebullition	CO ₂ , CH ₄	Varies	Global	Varies
Aquatic	Repository	Lentic, lotic	PANGAEA	https://doi.org/10.1594/PANGAEA.919986	Concentration	CH ₄	Varies	Global	Varies
Aquatic	Repository	Lentic, lotic	SITES	https://www.fieldsites.se/	Chamber, concentration	CO ₂	Varies	Global	Varies

2.2 Data repositories

510 We obtained a majority of terrestrial eddy covariance data from flux network repositories listed in Table 2. Because aquatic ecosystem eddy covariance data were sparsely available through



these flux networks, we relied on user-contributed data and those published in Golub et al., 2023. The repository data were downloaded between August 2023 and January 2024.

515 The terrestrial data that we downloaded from these repositories came in a variety of formats, though we gave preference to CO₂ flux data processed with the ONEFlux pipeline (Pastorello et al., 2020) and CH₄ flux data processed as part of the FLUXNET-CH₄ community product (Delwiche et al., 2021) when available since these data were gap-filled and quality-checked (i.e., FLUXNET2015, ICOS, and some of the Ameriflux datasets). This decision was made because of the strengths associated with the consistent data processing, quality-control, and recent updates. The ONEFlux pipeline produces datasets aggregated to different time resolutions (half-hourly, daily, monthly, and yearly) along with fluxes processed with various partitioning methods and friction velocity (USTAR) criteria, to remove data under low turbulence conditions (i.e., turbulence filters to correct biases). However, the ONEFlux pipeline does not perform footprint partitioning or fetch screening based on wind direction. Therefore, we assumed that any such filtering (e.g., for wind direction or land cover representativeness) was performed by the data provider of the site prior to ONEFlux processing, or that the reported fluxes represent the entire tower footprint. We used the datasets pre-aggregated to monthly mean fluxes (g C m⁻² d⁻¹) and opted for fluxes processed with a constant USTAR threshold and flux partitioning according to Reichstein et al. (2005) when available (for a justification, see 530 Virkkala et al., 2022). To calculate gap-fill percentages for each month, we used the half-hourly datasets produced by the ONEFlux pipeline and the quality flag associated with each flux (QC= 0 measured value).

For terrestrial eddy covariance data not processed with ONEFlux, we prioritized gap-filled data, though where it was not available, we accepted data that were not gap-filled (10% of all the 535 monthly eddy covariance data). The level of pre-processing (i.e. USTAR filtering, storage correction, etc) of this non-gap-filled data varied by data source but all data were quality checked prior to monthly aggregation. We justified this approach to increase the amount of data in this data-sparse region, and carefully assessed that the aggregated monthly fluxes were within a realistic range (within minimum and maximum monthly fluxes in similar environments).

540 We also searched terrestrial and aquatic data through several data repositories not focused solely on fluxes (see Table 2). To identify datasets of interest we used the same search words in the repositories as in the literature search (see Section 2.3). If the datasets identified from this search had associated publications, we reviewed the publications and extracted relevant information including that describing the environmental conditions of the study site.

545 2.3 Literature search

We conducted an exhaustive *Web of Science* search with search words (“carbon flux” or “carbon dioxide flux” or “methane flux” or “CH₄ flux” or “CO₂ flux” or “NEE” or “net ecosystem exchange”) and (“arctic” or “tundra” or “boreal”) up to December, 2023. For aquatic data, this search was extended up to August 2024. Based on the literature search, we added data from 550 three additional publications for terrestrial ecosystems focusing on chambers and 105 for aquatic ecosystems beyond those already included in BAWLD-CH₄, GRiMeDB, Liu et al., 2022



and Golub et al., 2023. The total number of new papers identified for terrestrial ecosystems was low because the recent BAWLD-CH₄ and ABCFluxv1 datasets already encompass studies up to 2018–2020, and more recent data were sourced from repositories or submitted directly by data contributors.

2.4 Community-contributed data

In order to capture data that are not easily extractable from literature nor found in repositories, we contacted ~180 researchers that were identified based on earlier reviews and syntheses, and our expert knowledge. This call for data began in April 2023, with final data submissions due in June 2024, though most contributions were given prior to January 2024. Out of the ~180 researchers contacted, we received 98 terrestrial and 43 aquatic datasets. These user-contributed flux data constitute 41% of observations in the overall dataset (46% of terrestrial, 32% of aquatic, 46% of CO₂, and 50% of CH₄). In addition to flux data, submissions from data contributors were more likely to include detailed site descriptions and data for ancillary variables that were not often available in repositories or papers. Submitted data from 23 new sites were previously unpublished (i.e., not published in scientific papers; 4% of site-months for terrestrial and 2.7% for aquatic) but had been processed using standard processing protocols or similar tools used at the site before.

3. Data columns

3.1 Summary of data columns

The ABCFlux v2 dataset is organized such that each row represents a unique combination of site and month. The data is grouped by site and arranged by time and all data is provided in a single file to facilitate use. All the columns in ABCFlux v2 are listed in Table 3, together with the percentage of data in each column, which has been subset by applicable data types (e.g. soil temperature applies only to terrestrial data). There are a total of 141 columns in ABCFlux v2 including 15 that contain flux data and 12 pertaining to measurement technique details (e.g. flux method details, partition method) and data quality (e.g. gap-fill percentage, number of chamber measurement days).

For eddy covariance measurements in terrestrial ecosystems, net ecosystem exchange of CO₂ (NEE) can be partitioned into gross primary productivity (GPP) and ecosystem respiration (Reco). In this dataset we include NEE as well as GPP and Reco, when available but we did not perform any data processing or flux partitioning, aside from unit conversions. Throughout this dataset, we report GPP and Reco as positive. NEE and CH₄ fluxes are reported with respect to the atmosphere, where positive values are a net source to the atmosphere from the ecosystem, and negative values are a net sink from the atmosphere to the ecosystem. For aquatic ecosystems, only the overall (non-partitioned) CO₂ flux is given because of the multiple origins of the CO₂ (Battin et al., 2023). The partitioning of GPP from Reco in aquatic systems is usually done with other methods (i.e. as commonly done with oxygen mass balances; (Staeher et al., 2010), but are not included in this dataset. For freshwater CH₄ fluxes derived with methods other than eddy covariance, we differentiate between diffusive and ebullitive emission pathways



and assume the sum of diffusion and ebullition to yield the total CH₄ flux. We also compiled CO₂ and CH₄ storage fluxes (as done for CH₄ by Kuhn et al., 2021 and Wik et al. 2016). Partitioning plant-mediated carbon fluxes from freshwaters is rare and not considered here.

In addition to carbon fluxes and dissolved gas concentrations, we gathered information
 595 describing environmental conditions of the site as well as general site characteristics. There are
 48 variables pertaining only to aquatic data (e.g. waterbody depth, water temperature) and 38
 tailored to terrestrial data (e.g. vegetation information, soil temperature). Fifty three of the
 columns represent static variables (e.g., biome, land cover), whereas the remaining columns
 vary monthly (e.g., fluxes, soil temperatures), seasonally (seasonal flux if monthly was not
 600 available), or annually (active layer depth). New columns were added to v2 compared to v1
 (Virkkala et al., 2022) to represent detailed descriptions of plant functional types (e.g.,
 evergreen and deciduous shrub coverage), deep soil temperatures (<10 cm), and permafrost
 thaw presence or absence in the top two meters. We also added new columns pertaining to CH₄
 fluxes that were not included in previous flux synthesis efforts, such as a categorical moisture
 605 class (wet-moist-dry) and BAWLD classes (Olefelt et al., 2021).

To differentiate between terrestrial and freshwater ecosystem classes, we followed the BAWLD
 classification system (Olefelt et al., 2021) and specified the classes in column “bawld_class”.
 BAWLD was specifically designed to separate key classes relevant for CH₄ cycling. There are 8
 terrestrial classes and 10 freshwater classes listed in the variable descriptions in Table 3, and
 610 the classes are more thoroughly discussed in their respective papers and metadata documents
 (Kuhn et al., 2021; Olefeldt et al., 2021). The column “bawld_class” was designed to
 differentiate between plot-level (sub-meter) variability in land cover types. Eddy covariance sites
 that had highly heterogeneous footprint with multiple BAWLD landcover classes were assigned
 the dominant BAWLD class for the tower footprint and should be interpreted with caution given
 615 the different landscape classification scales. When assigning lotic classes, we followed the
 BAWLD river size distinction (large rivers have a Strahler order >5, small rivers have a Strahler
 order ≤5). We first used the description provided by data contributors to determine if a small
 river was organic-poor or organic-rich. If the data contributors did not provide a site description,
 we deferred to the organic carbon geospatial data by Hugelius et al. (2020). We selected a 20%
 620 organic soil coverage threshold for this split based on a comparison with a small dataset
 containing site-level classifications of organic-poor and organic-rich soils, combined with our
 expert knowledge and cross-checked with the gridded percent cover data for rivers from
 BAWLD. However, since most lotic sites were ultimately classified using geospatial rather than
 site-level data, these classifications are inherently more uncertain. In some cases where
 625 classification was unclear, the site was left as ‘Unknown’.

In addition to the BAWLD classes, and to acknowledge classes relevant to CO₂ fluxes, we used
 a classification system from the European Space Agency (ESA) Climate Change Initiative (CCI)
 land cover product (ESA CCI 2016) which has been used in earlier syntheses (Virkkala et al.,
 2021). It differentiates between six different boreal forest/vegetation classes instead of the



630 broad Boreal forest class in BAWLD. There are in total 20 terrestrial classes and one general
aquatic class in ESA CCI, which are listed in the column descriptions in Table 3 and are more
thoroughly discussed in their respective metadata document (ESA CCI 2016). For the ESA CCI
class, we created two columns: “land_cover_plot” for the plot level and “land_cover_eco” for the
ecosystem level, to acknowledge the extent and scale of the land cover type associated with
635 each measurement. For example, a dry shrub-dominated plot at a palsamire received a shrub
class at the plot level, while the ecosystem-level class was water-logged, characterizing the
mostly wetland-dominated status of the ecosystem. For both the BAWLD and ESA CCI
columns, the categories were defined by data contributors or extracted from papers and
repositories and then unified by dataset developers through an expert assessment utilizing
640 additional columns (e.g., plant cover and vegetation description for BAWLD and ESA CCI
terrestrial classes, and lake size and sediment type for BAWLD aquatic classes). If no
information was available, the class was left as “NA”.



643 **Table 3.** Variable names and their description, and the percentage of data present. Percentages
644 were calculated based on the total relevant data for each variable. Subset refers to the data
645 category the respective variable applies to: eddy covariance (EC), and non-eddy covariance
646 (Non-EC) including all other measurement methods (chambers, concentration, etc.). Where the
647 subset remains blank, the variable applies to the whole dataset.
648



Variable category	Subset	Variable name	Unit	Description	Percent data present
Metadata		site_name		Site name as specified in data source. E.g. Hyytiälä	100.0
		site_reference		A more specific name used in data source. E.g. the name of the chamber plot (e.g. shrub) or the abbreviation used in eddy covariance data repositories (e.g. FI-Hyy).	100.0
		data_contributor_or_author		Data contributor(s) or primary author(s) associated with data set or publication	100.0
		email		Primary author email(s)	72.8
		extraction_source		Data source where data were extracted and compiled.	100.0
		citation		Citation for the data source: journal article, data citation, and/or other source (online repository link etc.). If the user contributed unpublished data, a journal article citation describing already published data can be added here.	90.8
		country		Country of the study site	100.0
		latitude	decimal degrees	Latitude of study site, as detailed as possible	100.0
		longitude	decimal degrees	Longitude of study site, as detailed as possible	100.0
		year	YYYY	Year in which data were recorded	99.8
		month	MM	Measurement month	99.2
Fluxes	Terrestrial	nee	g C m ⁻² month ⁻¹	Monthly cumulative net ecosystem exchange (NEE) for the entire measurement interval in g C in CO ₂ . Negative flux represents a net sink to the ecosystem.	78.4
	Terrestrial	gpp	g C m ⁻² month ⁻¹	Monthly cumulative gross primary productivity (GPP) for the entire measurement interval in g C in CO ₂ . Note: GPP is presented as positive (uptake) values.	60.9



	Terrestrial	reco	g C m-2 month-1	Monthly cumulative Ecosystem respiration (Reco) for the entire measurement interval in g C in CO ₂	68.3
	Aquatic	co2_flux	g C m-2 month-1	Monthly cumulative CO ₂ flux for the entire measurement interval. Refers to aquatic fluxes only where fluxes are commonly not partitioned between photosynthesis and respiration. Negative flux represents a net sink.	36.3
		ch4_flux_total	g C m-2 month-1	Monthly cumulative total CH ₄ flux for the entire measurement interval. Negative flux represents a net sink to the ecosystem.	27.5
	Aquatic-Non-EC	co2_flux_ebullition	g C m-2 month-1	Monthly cumulative ebullitive CO ₂ flux for the entire measurement interval	0.0
	Aquatic-Non-EC	ch4_flux_ebullition	g C m-2 month-1	Monthly cumulative ebullitive CH ₄ flux for the entire measurement interval	12.8
	Aquatic-Non-EC	ch4_flux_diffusion	g C m-2 month-1	Monthly cumulative diffusive CH ₄ flux for the entire measurement interval	93.6
	Aquatic-Non-EC	ch4_flux_storage	g C m-2 month-1	Monthly cumulative storage CH ₄ flux for the entire measurement interval. Storage flux is a diffusion flux but can be measured separately and is separated here because of its non-continuous rapid burst emission nature.	9.9
	Aquatic-Non-EC	co2_flux_storage	g C m-2 month-1	Monthly cumulative storage CO ₂ flux for the entire measurement interval. Storage flux is a diffusion flux but can be measured separately and is separated here because of its non-continuous rapid burst emission nature.	0.0
	Terrestrial	nee_seasonal	g C m-2 season-1	Cumulative NEE flux in a seasonal format (e.g. June 3rd to August 25th). This column is filled only if monthly data are not available. Negative flux represents a net sink.	0.4
	Aquatic	co2_flux_seasonal	g C m-2 season-1	Cumulative CO ₂ flux from freshwater systems in a seasonal format (e.g. June 3rd to August 25th). This column is filled only if monthly data are not available. Negative flux represents a net sink.	0.6



		ch4_flux_total_seasonal	g C m-2 season-1	Cumulative CH4 total flux in a seasonal format (e.g. June 3rd to August 25th). This column is filled only if monthly data are not available.	0.9
	Aquatic-Non-EC	ch4_flux_ebullition_seasonal	g C m-2 season-1	Cumulative CH4 flux from ebullition in a seasonal format (e.g. June 3rd to August 25th). This column is filled only if monthly data are not available.	0.5
	Aquatic-Non-EC	ch4_flux_diffusion_seasonal	g C m-2 season-1	Cumulative CH4 flux from diffusion in a seasonal format (e.g. June 3rd to August 25th). This column is filled only if monthly data are not available.	0.0
Meteorological		tair	°C	Monthly mean air temperature	53.7
		tair_height	m	Height of the air temperature measurement	23.3
		precip	mm	Monthly cumulative precipitation mm	35.3
		snow_depth	cm	Mean snow depth during the measurement interval	8.0
		ppfd	μmol m-2 s-1	Mean photosynthetically active radiation during measurement interval (in Photosynthetic Photon Flux Density, PPFD)	29.0
	Terrestrial	tsoil_surface	°C	Monthly mean surface soil temperature in ca. 0-10 cm depth	77.4
	Terrestrial	tsoil_surface_depth	cm	Depth of the surface soil temperature measurement	58.2
	Terrestrial	tsoil_deep	°C	Monthly mean deeper soil temperature at > 10 cm depth	25.3
	Terrestrial	tsoil_deep_depth	cm	Depth of the deeper soil temperature measurement	28.5
	Terrestrial	soil_moisture	VWC %	Monthly mean surface soil moisture in ca. 0-10 cm depth	45.2
	Terrestrial	moisture_depth	cm	Depth for the surface soil moisture measurement	29.9
	Terrestrial	thaw_depth	cm	Mean thaw depth during the measurement interval. Positive values represent depth below the soil surface.	8.3



	Terrestrial	alt	cm	Active layer thickness (maximum thaw depth). Thickness changes annually. Positive values represent depth below the soil surface.	14.5
	Terrestrial	water_table_depth	cm	Mean water table depth during the measurement interval; positive is below the surface, negative is above (inundated)	17.5
Measure ment details		flux_method	EC, Chamber, Snow diffusion, Ebullition, Concentration, Concentration (no flux)	Broad categories for how flux values were measured, may list more than one	100.0
		flux_method_detail		Details related to how flux values were measured. Specifies between manual vs. automated chambers, snow diffusion, open vs. closed-path eddy covariance, water sampling, ebullition trap, ice sampling	96.5
		flux_method_description		Details related to measurement method, e.g. chamber size/volume and deployment time, tubing length, use of bubble shield or not, height of the tower and wind measurements.	78.8
		gap_fill		Approach used to gap-fill the data	85.6
	Terrestrial	partition_method		Method used to partition NEE into GPP and RECO.	64.6
	EC	tower_corrections		Details related to processing corrections employed, including time, duration, and thresholds for u^* and heat corrections	75.8
	Chamber	diurnal_coverage	Day, Day and Night	Indicator whether data was collected during the day or during day and night	91.2
		instrumentation		Description of instrumentation used (e.g. type of greenhouse gas analyzer)	48.8
	EC- CO ₂	gap_fill_perc_co2	%	% of eddy covariance or automated chamber CO ₂ (nee, co2_flux) data that was gap-filled in the measurement interval (relative to standard measurement time step)	74.0



	EC-CH ₄	gap_fill_perc_ch4	%	% of eddy covariance or automated chamber CH ₄ data that was gap-filled in the measurement interval (relative to standard measurement time step)	30.9
	Non-EC-CO ₂	chamber_nr_measurement_days_co2		Number of days with chamber measurement of CO ₂ within the measurement interval (month)	42.9
	Non-EC-CH ₄	chamber_nr_measurement_days_ch4		Number of days with chamber measurements of CH ₄ within the measurement interval (month)	50.6
	Aquatic Non-EC-CO ₂	ebullition_nr_measurement_days_co2		Number of days with ebullition measurements of CO ₂ within the measurement interval (month)	1.5
	Aquatic Non-EC-CH ₄	ebullition_nr_measurement_days_ch4		Number of days with ebullition measurements of CH ₄ within the measurement interval (month)	2.7
	Terrestrial	nee_seasonal_interval	MM/DD/YYYY-MM/DD/YYYY	The period for the seasonal estimate (e.g. 06/04/2015-08/25/2015)	0.4
	Aquatic	co2_flux_seasonal_interval	MM/DD/YYYY-MM/DD/YYYY	The period for the seasonal estimate (e.g. 06/04-08/25)	0.5
		ch4_flux_seasonal_interval	MM/DD/YYYY-MM/DD/YYYY	The period for the seasonal estimate (e.g. 06/04-08/25)	1.1
	CO ₂	expert_flag_co2	0,1,2,3,4	0: no known issues, 1: terrestrial fluxes outside the 1st and 99th percentiles, aquatic lentic fluxes outside the 99th percentile, 2: terrestrial non-eddy covariance growing season measurements with 3 or less measurement days in the month and no modeling used to gap-fill, 3: terrestrial eddy covariance data with 3 or more consecutive months of 100% gap-filling, 4: site does not represent typical environmental conditions of the Arctic Boreal Zone	73.7
	Terrestrial-CO ₂	expert_flag_gpp	0,1	0: no known issues, 1: terrestrial fluxes outside the 1st and 99th percentiles	70.7



	Terrestrial- CO ₂	expert_flag_reco	0,1	0: no known issues, 1: terrestrial fluxes outside the 1st and 99th percentiles	79.3
	CH ₄	expert_flag_ch4	0,1,2,3,4	0: no known issues, 1: terrestrial fluxes greater than 30 g C m ⁻² month ⁻¹ , aquatic lentic fluxes outside the 99th percentiles, 2: terrestrial non-eddy covariance growing season measurements with 3 or less measurement days in the month and no modeling used to gap-fill, 3: terrestrial eddy covariance data with 3 or more consecutive months of 100% gap-filling, 4: site does not represent typical environmental conditions of the Arctic Boreal Zone	68.0
Site informatio n		biome	Tundra, Boreal, Temperate	Biome of the study site	100.0
		land_cover_eco	40=Mosaic natural vegetation (>50%) / cropland (<50%); 60=Tree cover, broadleaved deciduous; 70=Tree cover, needleleaved evergreen; 80=Tree cover, needleleaved deciduous; 90=Tree cover, mixed leaf type; 100=Mosaic tree & shrub (>50%) / herbaceous (<50%); 110=Mosaic herbaceous (>50%) / tree & shrub (<50%); 120=Shrubland; 121=Shrubland evergreen;	Dominant land cover class for the site following expert assignment and using class names of the ESA CCI land cover product. http://maps.elie.ucl.ac.be/CCI/viewer/download/ESACCI-LC-Ph2-PUGv2_2.0.pdf (section 9.1, global classification, page 81)	99.8



			122=Shrubland deciduous; 130=Grassland; 140=Lichens & mosses; 150=Sparse vegetation (<15%); 151=Sparse tree; 152=Sparse shrub; 153=Sparse herbaceous; 160=Tree cover, flooded fresh/brackish; 170=Tree cover, flooded saline; 180=Shrub/herbac eous flooded; 200=Bare areas; 210=Waterbodies		
		land_cover_plot	Same classes used as for "land_cover_eco" (see above)	Dominant land cover class for the site following expert assignment and using class names of the ESA CCI land cover product. http://maps.elie.ucl.ac.be/CCI/viewer/download/ESACCI-LC-Ph2-PUGv2_2.0.pdf (section 9.1, global classification, page 81)	99.8
		bawld_class	Bog, Fen, Marsh, Permafrost Bog, Wet Tundra, Dry Tundra, Moist Tundra, Boreal Forest, Rocklands, Large Lake, Midsize Glacial Lake, Small Glacial Lake, Midsize Peatland Lake, Small Peatland Lake, Midsize Yedoma Lake, Small Yedoma Lake, Large River, Small Organic-Rich River, Small Organic-Poor River	Dominant ecosystem class for the site following Boreal-Arctic Wetland-Lake Database (BAWLD) classes (Olefeldt et al. 2021). https://essd.copernicus.org/articles/13/5127/2021/	100.0



		veg_detail		Detailed vegetation description from data source/contributor, incl. dominant species and a description of vegetation changes at the site. In an aquatic site, this should have information on both the terrestrial and aquatic plants.	57.2
		permafrost	Yes, No	Reported presence or absence of permafrost	68.0
		permafrost_thaw	Yes, No	Permafrost thaw present	32.2
		thaw_category	Active layer detachment, Active layer thickening, Gradual thaw, Ice-wedge degradation, Palsa thaw, Retrogressive thaw slumps, River bank erosion, Subsidence, Talik formation, Thaw lake, Thaw pond, Thermokarst, Thermokarst mounds with talik	Broad category for types of thaw	14.8
		thaw_extent	0-33%, 34-66%, 67-100%	Spatial extent of thaw within the site (i.e. tower footprint, chamber/sampling plot, drainage basin for aquatic ecosystems)	16.3
		thaw_severity	Low: Change to the ecosystem that is temporary or does not have substantial impact on ecosystem identity or function, Moderate: Change to the ecosystem that is temporary but has a substantial impact on ecosystem function, High: Change to the ecosystem that is	Severity of dominant type of thaw following the classifications of ecosystem change presented in (Webb et al., 2025)	14.8



			lasting with consequences for ecosystem identity and function		
		thaw_dominant	Active layer detachment, Active layer thickening, Active layer thickening/Thaw Ponds, Gradual thaw, Palsa thaw, Retrogressive thaw slumps, River bank erosion, Subsidence, Thaw lake, Thermokarst, Thermokarst (pond), Thermokarst bog formation, and Upland thermokarst mounds with talik	Dominant type of thaw. If multiple categories were chosen in thaw_category, thaw_extent and thaw_severity primarily refer to the type of thaw listed here	14.4
		landform		Description of the geomorphological landforms associated with the site. E.g., polygonal features, palsas, cryoturbation, abrupt thaw features, drained lake basins.	14.0
		disturbance		Description of the recent disturbance history of the site or list "No" if there are no disturbances. Note that the disturbance might have been caused due to natural reasons (e.g. fire) and/or due to anthropogenic influences (e.g. drained peatland, harvested forest). If several disturbances have occurred during the recent decades, these can all be listed here but please list the last dominant disturbance first.	54.5
		disturbance_category	Altered hydrology, Animal herbivory, Artificial pond, Beavers, Drained lake, Drainage,	Broad categories for disturbances	55.5



			Erosion, Extreme weather, Fire, Forestry, Human paths, Insect herbivory, Land use change, None, Other, Peat mining, Reservoir, Roads, Seismic lines, Thaw, Tidal effects, Wastewater		
		disturb_dominant	Animal herbivory, Artificial pond, Beavers, Drained lake, Drainage, Erosion, Erosion/Thaw, Extreme weather, Fire, Forestry, Human paths, Insect herbivory, Land use change, None, Other, Reservoir, Roads, Seismic lines, Thaw, Wastewater	Dominant disturbance of the site. If multiple categories are listed in disturb_category, disturb_extent and disturb_severity primarily refer to the disturbance listed here	21.0
		disturb_year	Numeric variable (year), 0 = annual (e.g., annual grazing, annual nutrient additions)	Year of last dominant disturbance, 0-[year] indicates the disturbance is ongoing and began in the year given	21.9
		disturb_severity	Low: Change to the ecosystem that is temporary or does not have substantial impact on ecosystem identity or function, Moderate: Change to the ecosystem that is temporary but has a substantial impact on ecosystem function, High: Change to the ecosystem that is	Relative severity of last dominant disturbance following the classifications of ecosystem change presented in Webb et al. 2025	19.2



			lasting with consequences for ecosystem identity and function		
		disturb_extent	0-33%, 34-66%, 67-100%	Spatial extent of the dominant disturbance within the site (i.e. tower footprint, chamber/sampling plot, drainage basin for aquatic ecosystems)	14.8
		site_activity	Active, Non-active	The current flux measurement activity status of the site (i.e., measurements conducted each year).	72.4
Terrestrial variables	Terrestrial	dec_shrub	Absent, Present, Dominant	Are deciduous shrubs absent, present, or dominant? Examples of deciduous shrubs: Betula nana, Salix sp, Vaccinium uliginosum, Vaccinium myrtillus, Rubus chamaemorus	46.0
	Terrestrial	ev_shrub	Absent, Present, Dominant	Are evergreen shrubs absent, present, or dominant? Examples of evergreen shrubs: Empetrum sp, Cassiope sp, Loiseleuria sp, Vaccinium vitis-idaea, Rhododendron sp, Phyllodoce caerulea, Dryas octopetala	46.3
	Terrestrial	sedge	Absent, Present, Dominant	Are sedges absent, present, or dominant	45.1
	Terrestrial	non_sedge_herbaceous	Absent, Present, Dominant	Are grasses, rushes and forbs absent, present, or dominant	42.8
	Terrestrial	ev_needle_tree	Absent, Present, Dominant	Are evergreen needleleaf trees absent, present, or dominant	50.8
	Terrestrial	dec_needle_tree	Absent, Present, Dominant	Are deciduous needleleaf trees absent, present, or dominant	37.1
	Terrestrial	dec_broad_tree	Absent, Present, Dominant	Are deciduous broadleaf trees absent, present, or dominant	40.0
	Terrestrial	sphagnum_cover	Absent, Present, Dominant	Are Sphagnum mosses absent, present, or dominant	43.3
	Terrestrial	other_moss_cover	Absent, Present, Dominant	Are other mosses (non-Sphagnum) mosses absent, present, or dominant	36.8
	Terrestrial	canopy_height	m	Height of the vegetation canopy	23.5



			Wet = At least sometimes inundated or water table close to surface, Dry = well-drained, Moist = in between wet and dry		
	Terrestrial	soil_moisture_class		General descriptor of site moisture	53.6
	Terrestrial-forest	forest_age	YYYY	Forest age since last disturbance (anthropogenic/natural)	11.3
	Terrestrial	soil_depth	cm	Soil organic layer depth	21.3
	Terrestrial	soil_ph		Surface soil pH in ca. 0-10 cm depth	11.4
	Terrestrial	soil_perc_c	%	Surface soil C % in ca. 0-10 cm depth	14.0
	Terrestrial	soil_perc_n	%	Surface soil N % in ca. 0-10 cm depth	7.7
	Terrestrial	c_stock	kg C m-2	Soil organic carbon stock, ideally for the entire soil profile.	22.3
	Terrestrial	stock_depth	cm	Soil depth used in the stock calculation.	21.4
	Terrestrial	soil_type_detail		Soil type description	23.0
	Terrestrial	lai	m-2 m-2	Leaf area index	14.6
	Terrestrial	ndvi		Normalized difference vegetation index	4.0
Aquatic variables	Aquatic	waterbody_type	Lentic, Lotic	Type of waterbody: lentic (standing water) or lotic (flowing water)	99.5
	Aquatic	aquatic_site_sampling_location	Edge, Center, Both	The locations of the measurements within the waterbody	34.3
	Aquatic	water_body_trophic_status	Oligotrophic, Mesotrophic, Eutrophic	Trophic state classification describing the productivity of a lentic waterbody	38.1
	Aquatic	water_area	m2	Area of lentic waterbody	31.6
	Aquatic	water_depth	m	Mean, maximum or point-level depth of a lentic or lotic waterbody	34.5
	Aquatic	water_depth_location	Mean, Sampling location, Maximum	Description representing the location of the water depth measurement	5.7
	Aquatic	water_ph		Mean pH at water surface during the measurement interval	62.2
	Aquatic	water_n	mg l-1	Mean total nitrogen at water surface during the measurement interval	41.7



	Aquatic	water_p	mg l ⁻¹	Mean total phosphorus at water surface during the measurement interval	35.8
	Aquatic	water_doc	mg l ⁻¹	Mean dissolved organic carbon content at water surface during the measurement interval	62.0
	Aquatic	water_iceon	MM/DD/YYYY	Ice-on date	9.8
	Aquatic	water_iceoff	MM/DD/YYYY	Ice-off date	11.0
	Aquatic	water_temperature	°C	Mean surface water temperature during the measurement interval	76.8
	Aquatic	benthic_veg	Yes, No	Is benthic vegetation occurring at the site?	6.5
	Aquatic	emergent_veg		Details related to the emergent vegetation	5.3
	Aquatic	sediment	Minerogenic, Organic, Peat, Yedoma, Unspecified	Sediment type	23.2
	Aquatic	water_chlorophyll	mg l ⁻¹	Mean concentration of chlorophyll at water surface during the measurement interval	2.6
	Aquatic	water_do	mg l ⁻¹	Mean dissolved oxygen at water surface during the measurement interval	29.1
	Aquatic	water_co2	μmol l ⁻¹	Mean dissolved CO ₂ concentration at water surface during the measurement interval	77.5
	Aquatic	water_ch4	μmol l ⁻¹	Mean dissolved CH ₄ concentration at water surface during the measurement interval	75.8
	Aquatic	water_turbidity	FNU (Formazin Nephelometric Unit)	Mean turbidity of the water during the measurement interval	9.3
	Aquatic	water_conductivity	μS/cm	Mean electrical conductivity at water surface during the measurement interval	45.6
	Aquatic	k600	cm h ⁻¹	Gas transfer velocity normalized to a Schmidt number of 600	16.2
	Aquatic	k600_equation		Equation used to determine gas transfer velocity (k600)	17.5
	Aquatic	k600_method		Method used to determine gas transfer velocity (k600)	26.1
	Aquatic	stream_discharge	m ³ s ⁻¹	Stream discharge of lotic waterbody	18.1



	Aquatic	stream_velocity	m s ⁻¹	Stream velocity of lotic waterbody	2.1
	Aquatic	water_mixing_regime	E.g. Monomictic, Dimictic, Polymictic	Mixing regime of lentic waterbody to indicate the frequency of mixing throughout the year (once = monomictic, twice = dimictic, multiple times = polymictic)	8.3
	Aquatic	strahler_order		Strahler order of lotic waterbody to define stream size	26.3
	Aquatic EC	fetch_screening	Yes, No, Unknown	Indicator whether fetch screening was applied	99.2
	Aquatic EC	fetch_detail		Details about fetch screening	92.1
	Aquatic	air_co2	ppmv	Mole fraction of CO ₂ in the air above water surface during the measurement interval	18.5
	Aquatic	air_ch4	ppbv	Mole fraction of CH ₄ in the air above water surface during the measurement interval	13.9
	Aquatic Non-EC	water_d13ch4	permil	Ratio of stable carbon isotopes of CH ₄ at water surface during the measurement interval	6.0
	Aquatic Non-EC	water_d13co2	permil	Ratio of stable carbon isotopes of CO ₂ at water surface during the measurement interval	3.0
	Aquatic	isotopic_analysis_detail		Details regarding the isotopic analysis, e.g. determined from dissolved gas or ebullition	1.3
Policies and notes		data_usage	Tier1 = data are open and free for scientific and educational purposes, Tier2 = data producers must have opportunities to collaborate and consult with data users, Other (please specify)	Instruction of data usage	90.4
		data_version		Version number for data extracted from repositories or version number given by the data contributor based on their version tracking.	24.4



		notes		Additional relevant information	38.5
		id		Unique identifier given to each individual monthly entry at each site	100

649



650 3.2 Definition of site

To differentiate between measurement locations, we used two attributes: `site_name` and `site_reference`. The column “`site_name`” (e.g., Stordalen Mire) is considered a more general description of a site whereas “`site_reference`” (e.g., Stordalen Mire_Palsa Site_Chamber) is a more specific description of a plot/sub-site within a broader site and indicates the method of measurement. The distinction between `site_name` and `site_reference` is most evident among chamber studies where measurements may have been made across different types of vegetation or landscape characteristics within a single site (i.e., several `site_references` corresponding to a single “`site_name`”). We assigned a unique `site_reference` to a site, as long as it had a distinct land cover class, coordinates, or unique related environmental data. For eddy covariance tower measurements, the distinction between “`site_name`” and “`site_reference`” is less significant. Eddy covariance towers from the major flux repositories (e.g., FLUXNET, Ameriflux, ICOS) often have a FluxID assigned which was reflected in the `site_reference` of the data (e.g., Stordalen_SE-St1_tower). For user-contributed tower data that did not have a FluxID, `site_reference` is often the `site_name` along with a name specified by the data contributor with the addition of “_tower”. In instances where footprint analysis was applied to split tower data in addition to the ecosystem-level fluxes (sites Ranskalankorpi, Iskoras, Stordalen), the “`site_reference`” column specifies which ecosystem the flux comes from (e.g., Iskoras_NO-Isk_palsa_tower and Iskoras_NO-Isk_pond_tower). ABCFlux v2 comprises 1,024 individual site names and 5,121 individual site references. In order to not exaggerate the number of sites in this synthesis, we refer to the number of unique “`site_name`” unless otherwise specified.

Flux data for lentic waterbodies were aggregated to the waterbody level by averaging the observational data where several measurements were conducted within one waterbody. Therefore, spatial within-lake differences were not individually accounted for. However, the column “`aquatic_site_sampling_location`” provides information about the location within the waterbody where measurements were conducted, and differentiates between the edge and the center of the waterbody, or both if measurements were conducted across the waterbody which can have implications for total flux calculations. (Ray et al., 2023). It should be noted that the sampling location within the waterbody remained unknown (no location information was available) for 66% of the aquatic flux measurements, 22% were sampled from the center of the waterbody, 7% from both (center and edge), and 5% from the edge of the waterbody. Where measurements were taken within a single campaign along a large lotic waterbody over several kilometers, the river was divided into sections and flux data were aggregated for each section separately. This spatial aggregation was handled on a case-by-case basis and in close collaboration with data contributors. In instances where this spatial aggregation was applied, the “`site_name`” represents the river name (e.g., Teno) and “`site_reference`” reflects the river section (e.g., Teno_Karigasniemi_chamber, where Teno river is the name of the river, and Karigasniemi the measurement location).



690 4. Data quality and screening

We screened and cleaned data in ABCFlux v2 using expert judgement, informed by the gap-filled data percentage, quality flags and number of measurements days, if available. Our primary approach was to visually assess the time series of meteorological variables and fluxes for each site as well as the overall magnitudes in flux and supporting environmental data. In general, the quality control of all data was carried out in close cooperation with data providers.

4.1 Terrestrial fluxes

For repository data, we encountered 49 occurrences in eddy covariance site-level time series of CO₂ and CH₄ fluxes that had “flat lines” over several months of data (see Supplementary Fig. 1 panel A for an example), i.e., relatively constant flux values that did not vary by more than 3 g C m⁻² month⁻¹ over consecutive months. These “flat lines” often occurred at the beginning and end of time series and usually had a very high gap-fill percentage (a mean gap-fill percentage of 97%), indicating the value was based on very little measured data and was most often entirely gap-filled. We excluded flux data with “flat lines” if they were made up of three or more consecutive months with high gap-fill percentages (>75%).

Another issue that was identified during the quality check of the eddy covariance data submitted by data contributors, or extracted from data repositories, pertained to the winter months (Dec-Feb). During this period, NEE was occasionally found to be exactly zero with a gap-fill percentage of 100%. We excluded these zeroes from our dataset, as there should always be some variability due to measurement and data processing uncertainties, even during low-flux conditions. Moreover, previous studies (Kittler et al., 2017b; Natali et al., 2019; Watts et al., 2021) have shown that Arctic-boreal ecosystems can exhibit winter-season fluxes of significance, making the assumption of an exactly zero winter flux unlikely.

For 26% of CO₂ and 69% of CH₄ eddy covariance site-months (including data provided by data contributors and from flux repositories), there was no information about data quality and/or gap-filled data percentage. Consequently, we were not able to solely and systematically rely on these metadata for quality screening. Thus, we calculated the 1st and 99th percentiles for each combination of month, biome, and flux measurement method (EC and non-EC) and used them, together with the visual assessment of time series, to identify data that either strongly deviated from expected seasonal patterns or fell outside these percentile thresholds (Supplementary Table 2). We removed data if it was both outside of these percentiles and stood out visually (see Supplementary Fig. 1, panel C and D for examples). Primarily, we excluded months that were entirely gap-filled during winter and showed net CO₂ uptake beyond the 99th percentile. The visual inspection of time series also led to the removal of flux data from sites where one year in the dataset showed unrealistic patterns, such as potentially reversed signs (e.g., winter uptake and summer sources). These datasets were often downloaded from general data repositories like the Arctic Data Center and attempts to resolve the issues by contacting data contributors or reviewing relevant publications from the site were unsuccessful. At some sites, flux time series



730 followed a realistic seasonal pattern within the percentile thresholds (see Supplementary Table
 2) despite some of the monthly fluxes being entirely gap-filled (often during a few months of the
 winter season, or due to, e.g., one year of missing data in a longer time series). Due to the
 limited amount of data in the Arctic-boreal region, these fluxes were kept in the dataset, and this
 is noted in the gap-filled data percentage column. In total, we kept 26 sites that included 90-
 735 100% gap-filled flux data during the peak winter months (Dec-Feb) across all the measured
 years because the seasonal dynamics and magnitudes matched those from other years of data
 from the same site or similar ecosystem types (see e.g. Supplementary Fig. 1, panel B). Though
 we did remove months where the data repository had winter months that were 100% gap-filled
 and data contributors provided only growing season data and advised against including
 740 repository data.

We examined terrestrial non eddy covariance (i.e. chamber and snow pack diffusion) using the
 same approach as eddy covariance, accessing the 1st and 99th percentiles along with the
 visual inspection of each site time series. These fluxes were not removed, as the limited
 745 temporal coverage of these measurement methods made it difficult to interpret seasonal
 patterns from time series graphs. Additionally, converting these often temporally limited
 observations into monthly cumulative fluxes can yield values with considerable uncertainty;
 however, we retained these data given the overall scarcity of measurements in this region and
 instead rely on the quality flags described in Section 5.3 to guide data users.

750 For the supporting environmental data, we removed data with unchanging values across three
 or more months as they likely represented a seasonal average and not monthly data. We also
 removed soil temperature observations that were above 40°C and below -40°C as these were
 the approximate temperature ranges seen in ABCFlux v1 and BAWLD-CH4 and values outside
 755 of this range were assumed to be errors. The largest amount of cleaning was done for the water
 table depth to ensure that the sign of the data was aligned with the variable description for
 ABCFlux v2 (i.e., positive is below the soil surface, negative is above).

4.2 Freshwater fluxes

760 The footprints of eddy covariance towers over waterbodies often include surrounding non-
 aquatic ecosystems. Therefore, a fetch screening is commonly applied by data contributors of
 sites with mixed footprints (e.g. Lake Villasjön in Sweden; (Jammet et al., 2017) to separate flux
 contributions from aquatic and adjacent terrestrial ecosystems. The simplest approach to
 remove non-aquatic flux contributions is to apply a wind-directional fetch screening that
 excludes half-hourly fluxes from wind directions associated with land surfaces during data
 765 processing (also done in Golub et al., 2023). Apart from the wind-directional screening, more
 sophisticated approaches have been used (e.g. Bayesian modeling (Pirk et al., 2024)). In
 ABCFlux v2, we indicated whether fetch screening was applied and for which wind directions.
 Regardless of the type of approach used to filter out non-aquatic carbon fluxes for eddy
 covariance data, the number of data gaps tends to be larger than for terrestrial towers. Because
 770 of this, and the fact that most gap-filling approaches are tailored to terrestrial ecosystems, gap-
 filling for aquatic towers remains challenging. We included both gap-filled and non-gap-filled
 aquatic tower data and derived monthly cumulative fluxes based on the available data.



Within the process of quality screening, we also unified some variables such as the gas transfer velocity, which can be used to estimate diffusive fluxes based on dissolved gas concentrations and hydraulic properties. Various methods to derive gas transfer velocities can be used, which differ slightly for lentic and lotic ecosystems (Hall and Ulseth, 2020; Klaus and Vachon, 2020; Raymond et al., 2012). The gas transfer velocity can be expressed as a magnitude independent of gas and temperature when normalized to a Schmidt number (Sc) of 600 (k_{600} , i.e. normalized gas transfer velocity) for freshwater at 20 °C. The Schmidt number is defined as the ratio between kinematic viscosity and mass diffusivity but is often empirically determined, and quantifies the temperature-dependent molecular transport properties of each gas (Jähne et al., 1987). Where gas transfer velocities for a specific gas species (k_{gas}) were not normalized to a Schmidt number of 600, we converted them accordingly ($k_{600} = (600/Sc)^n \cdot k_{gas}$; (Cole and Caraco, 1998)), where n is determined by windspeed (Guérin et al., 2007). Furthermore, dissolved CO_2 and CH_4 concentrations (or partial pressures) were converted from a range of given units (ppm, ppb, μatm , mol/L, mmol/L, nmol/L, mmol/m³, mg/L) to $\mu mol/L$ following previous procedures from GRiMeDB (github code: <https://github.com/lukeloken/GlobalRiverMethane>).

5. Data usage notes

Despite extensive efforts in dataset cleaning, users of ABCFlux v2 should remain aware of certain considerations to avoid potential misinterpretation of the data.

5.1 Uncertainties related to gap-filling and flux partitioning

The approach used to gap-fill and estimate monthly cumulative fluxes varied within and across measurement methods. For eddy covariance data, the most common gap-filling technique was Marginal Distribution Sampling (MDS), used in 57% of site-months, following the ONEFlux pipeline processing approach (Pastorello et al., 2020). However, we also incorporated fluxes that were gap-filled using other methods including neural networks and non-linear regression. For terrestrial and aquatic eddy covariance data, the cumulative monthly flux was most commonly obtained by multiplying the gap-filled monthly mean flux rate given as per day ($g\ C\ m^{-2}\ day^{-1}$) by the number of days in the month. In instances where gap-filled data at terrestrial sites were not available, we multiplied the daily mean of the respective month by the days in that particular month, and indicated that no gap-filling was applied in the “gap_fill” column. With these non-gap-filled data, gaps covered 53% for NEE and 69% for CH_4 flux per month on average. For aquatic eddy covariance sites, half-hourly flux data adopted from Golub et al. (2023) had been gap-filled following (Pastorello et al., 2020) and were aggregated to monthly cumulative fluxes for ABCFlux v2, but data contributors also shared non-gap-filled data which we then aggregated to monthly cumulative fluxes and indicated that in the “gap_fill” column. As mentioned above, additional gaps in flux data occur where fetch screening was applied to heterogeneous tower footprints. For aquatic sites, the gap-fill percentage often reflects both general data gaps, such as those caused by sensor issues, maintenance, power outages, or poor turbulence conditions, and the additional exclusion of data due to fetch screening.



815 Consequently, gap-fill percentages between terrestrial and aquatic sites in ABCFlux v2 cannot be directly compared. Gap-fill percentages for aquatic data averaged 68% for CO₂ and 65% for CH₄.

820 The methods used to partition CO₂ fluxes into GPP and Reco at terrestrial eddy covariance sites varied, with the most common approach (66% of site-months) being that of (Lasslop et al., 2010; Reichstein et al., 2005)), which has been widely applied in global synthesis and upscaling studies (e.g., (Nelson et al., 2024)). In addition to the Reichstein et. al (2005) method, this dataset also includes fluxes based on other partitioning methods such as (Lasslop et al., 2010; Reichstein et al., 2005)) and (Runkle et al., 2013). Potential limitations and differences among partitioning methods in the Arctic-boreal context have been extensively discussed in Virkkala et al. (2022) (under “Uncertainties in eddy covariance flux partitioning”). In particular, nighttime partitioning (Reichstein et al., 2005) can introduce uncertainty in high-latitude regions where low-light nighttime conditions are limited during summer. However, when comparing multiple gap-filling and partitioning methods across sites, we found that the variability in annual GPP and Reco estimates was small (Desai et al., 2008; Keenan et al., 2019), lending confidence to the partitioned GPP and Reco estimates derived from the diverse methods used in this dataset.

830 Chamber and other non-eddy covariance flux measurements, although generally more temporally sporadic than eddy covariance, were often converted to monthly cumulative fluxes using a similar method as eddy covariance (i.e., measurements averaged and multiplied by days) or, in some cases, gap-filled with light- and temperature-response models; details related to the approach can be found in the “gap_fill” column. Similarly, meteorological data were often collected only during these sporadic measurements, and are thus not based on continuous meteorological measurements within a month. 20% and 32% of terrestrial non-eddy covariance NEE and CH₄ flux measurements, respectively, 68% and 51% of lentic chamber CO₂ and CH₄ flux measurements, and 14% and 16% of lotic chamber CO₂ and CH₄ flux measurements were derived from one single measurement day (however note that the majority of lotic data do not have this information, 80% of CO₂ and 76% of CH₄ data). Furthermore, these measurements were often conducted during daytime only (69% of terrestrial chamber measurements, 55% of lentic, and 90% of lotic flux measurements). The sporadic nature and lower data coverage of the non-eddy covariance data leads to uncertainty in monthly flux and meteorological data. The bias toward daytime measurements may lead to an overestimation of net CO₂ sinks in vegetated ecosystems due to less photosynthesis at night (Lai et al., 2012; Järveoja et al., 2020) and an overestimation of CH₄ emissions in lentic ecosystems due to calmer winds and cooler temperatures at night (López-Blanco et al., 2017; Sieczko et al., 2020; Voigt et al., 2023). Similarly, in lotic systems, measurements were typically conducted during the day, even though nighttime emissions may exceed daytime release, potentially leading to an underestimation of monthly CO₂ emissions (Attermeyer et al., 2021; Gómez-Gener et al., 2021). However, such biases were not clearly evident in the data, even when comparing monthly fluxes derived using simple averages versus light- and temperature-response models (Supplementary Fig. 2 and 3). Site-months based on daytime-only measurements, however, tended to show greater variability, with lower minimum and higher maximum fluxes compared to those including both day- and night-time data. The columns “diurnal_coverage”, “chamber_nr_measurement_days_co2”,



“chamber_nr_measurement_days_ch4” and “ebullition_nr_measurement_days” help the data users understand the temporal representativeness of the data.

5.2 Fluxes and periods captured by the dataset

860 Gas flux and dissolved gas concentration measurements were more abundant during the growing season (May-August) compared to the non-growing season (September-April). The average total number of observations per month was 4,038 during growing season months vs 1,467 during non-growing season months, and 50% of terrestrial data and 28% of aquatic data were collected in the non-growing season despite two thirds of the year being non-growing
 865 season. Non-growing season monthly fluxes were often more heavily gap-filled than those during the growing season (54% in the growing season vs. 68% in the non-growing season for CO₂, 61% vs. 63% for CH₄), which further contributes to higher uncertainties. This is problematic as the non-growing season, and in particular spring and autumn season CO₂ and CH₄ emissions in both terrestrial and freshwater ecosystems, are important for the annual carbon balance (Arndt et al., 2023; Lyu et al., 2024). In freshwater ecosystems, the CO₂ and
 870 CH₄ emissions during the spring ice-out period (“storage flux”) are known to contribute significantly to annual fluxes (e.g., 11 to 59% in subarctic lakes (Jammet et al., 2015; Juutinen et al., 2009a; Karlsson et al., 2013, 2024; Prėskienis et al., 2021)), but measurements during the spring period are limited. Assuming a spring ice-out period in May-June for lentic
 875 ecosystems and March-April for lotic ecosystems (following Song et al., 2024), 28% of the lentic site-months and only 6% of the lotic site-months captured this period. Furthermore, information on the timing of the ice-on and ice-off was rarely given for non-growing season measurements at aquatic sites. Therefore, annual CO₂ and CH₄ fluxes from aquatic ecosystems may be underestimated due to the lack of data during these influential seasonal periods. However, the
 880 extent of this underestimate is uncertain given the unknown role of CH₄ oxidation and diel variation in CO₂ consumption in the water prior to emissions (Pajala et al., 2023; Rudberg et al., 2021). At the same time, simple averaging of the dominating daytime summer fluxes to spring and autumn seasons may instead generate substantial overestimates for CH₄, while simultaneously underestimating CO₂ emissions in productive lakes where emissions can occur
 885 outside the growing season (Natchimuthu et al., 2016; Rudberg et al., 2021).

A substantial part (22%) of freshwater carbon fluxes synthesized in ABCFlux v2 were gained from samples taken exclusively in the center of waterbodies, excluding edges. This may be problematic as the spatial variability of carbon fluxes across depth zones in lentic ecosystems
 890 can be large and plays a significant role when estimating ecosystem-level emissions, particularly for CH₄ fluxes (Kuhn et al., 2023). Studies have shown that CO₂ emissions might be overestimated when only considering measurements from the center of lakes (Loken et al., 2019), whereas CH₄ emissions might be underestimated in some lakes (Juutinen et al., 2003), but not others (Schmiedeskamp et al., 2021). It is also important to note that the location of
 895 eddy covariance towers in freshwater ecosystems can contribute to under- or over-estimating fluxes depending on the location of localized emission hotspots (e.g. thermokarst features). Therefore, both under- and overestimations in aquatic cumulative ecosystem-level fluxes are possible, depending on the gas, sampling location, sampling frequency, and seasonal dynamics involved (Ray et al., 2023).



900 5.3 Quality flags based on expert assessment

To help data users assess data quality in the flux records, we included additional columns “expert_flag_co2”, “expert_flag_gpp”, “expert_flag_reco” and “expert_flag_ch4” (Table 4, Supplementary Fig. 4). These columns may aid in filtering out sites and/or observations that are atypical or highly uncertain and should be used with particular caution when scaling monthly
 905 fluxes to estimate budgets across larger domains. In other use cases, these observations may still be useful.

As a broad overview, we flagged CO₂ fluxes outside of the 1st and/or 99th percentiles (flag 1), uncertain fluxes due to sporadic non-eddy covariance measurements (flag 2), long periods of
 910 eddy covariance data that have been entirely gap-filled (flag 3), and sites that do not represent typical conditions across the Arctic-boreal region due to human induced changes (flag 4, Table 4). All other CO₂ fluxes were marked with a zero (i.e. representative, good-quality data).

In more detail, we used 1st and 99th percentiles for terrestrial CO₂ fluxes calculated separately
 915 for each combination of month, biome (tundra, boreal, temperate), and measurement method (EC, chamber, snow diffusion) to flag extremely high or low fluxes with 1. This procedure was applied independently to NEE (expert_flag_co2), GPP (expert_flag_gpp), and Reco (expert_flag_reco). This flag considered, for example, unusually low NEE values during peak winter months (i.e. fluxes from less than -25 g C m⁻² month⁻¹ between December and February),
 920 which represent unrealistically high winter net uptake values likely due to issues in eddy covariance data collection and gap-filling (Jentsch et al., 2021; Kittler et al., 2017). During the summer months (June-August), flag 1 captured eddy covariance NEE data below -140 g C m⁻² month⁻¹ and chamber NEE data below -425 g C m⁻² month⁻¹. For chamber data, these large negative NEE values were often based on single daytime measurements alone. Flag 2 marks
 925 terrestrial chamber CO₂ and CH₄ flux growing season (May-Aug) measurements with fewer than three observation days in a month, where monthly values were calculated using simple averages rather than models incorporating light and temperature. Such limited sampling and simplistic averaging can introduce bias by failing to capture environmental variability, especially during the growing season. The three-day threshold is based on the assumption that
 930 approximately one measurement per week is necessary to produce a more reliable cumulative estimate, as suggested by Virkkala et al. (2022). To account for long periods of missing and entirely gap-filled data, we introduced flag 3 which marks eddy covariance data with 3 or more consecutive months of 100% gap-filling. For flag 4, we flagged eddy covariance sites where there have been significant atypical changes to the landscape from humans such as sites where
 935 the organic soils were removed with a bulldozer (Euskirchen et al., 2017; Walter Anthony et al., 2024). While flag 1 was applied separately to NEE, GPP, and Reco, flags 2–4 were applied only to expert_flag_co2, as they pertain to measurement methods and site conditions rather than to the flux variables themselves.

940 For aquatic CO₂, fluxes from lentic ecosystems that exceeded the monthly 99th percentile (between 26.4 g C m⁻² month⁻¹ in January to 320.9 g C m⁻² month⁻¹ in June) were marked with flag 1. We refrained from flagging fluxes from lotic ecosystems since these may be affected, for



example, by higher turbulence and larger resulting gas exchange compared to stiller conditions in lentic ecosystems.

945

For terrestrial CH₄ fluxes, we applied flag 1 to fluxes higher than 30 g C m⁻² month⁻¹ as such high values are well beyond the range of previously published estimates (Kuhn et al., 2021). Flags 2–4 were applied to terrestrial CH₄ fluxes following the same criteria used for CO₂ fluxes.

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Similarly to terrestrial data, we flagged diffusive CH₄ fluxes for lentic ecosystems that were higher than 30 g C m⁻² month⁻¹ and were excluded from analyses in previous studies (such as BAWLD-CH₄) due to exceptionally high gas transfer velocities with flag 1. This flag was not applicable to aquatic eddy covariance sites since total CH₄ fluxes did not exceed the chosen threshold. Furthermore, we refrained from flagging high diffusive CH₄ fluxes for lotic ecosystems

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Flag 2 was not used for aquatic fluxes since 90% of the aquatic monthly cumulative CO₂ fluxes and 86% of CH₄ diffusion fluxes from non-eddy covariance methods were derived from 3 or less measurements per month. Furthermore, none of the aquatic eddy covariance sites showed 3 or more consecutive months of 100% gap-filled data (flag 3), and none of the aquatic sites experienced significant land use change which could be linked with extremely high fluxes (flag 4).

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Table 4: Quality flags assigned to the terrestrial and aquatic CO₂ and CH₄ fluxes based on expert assessment (columns “expert_flag_co2”, “expert_flag_gpp”, “expert_flag_reco” and “expert_flag_ch4”). Note that the data that the respective flags apply to may change between terrestrial and aquatic, as well as lentic and lotic fluxes, and between flux methods. Further detail is given in Sect. 5.3.

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CO ₂ Flag	Requirements	Applicable data
0	Terrestrial: NEE, GPP, Reco within the 1st and 99th percentiles Aquatic-lentic: CO ₂ fluxes within 99th percentile Aquatic-lotic: all CO ₂ fluxes	All terrestrial and aquatic
1	Terrestrial: NEE, GPP, Reco outside of 1st and 99th percentiles Aquatic-lentic: CO ₂ fluxes greater than 99th percentile	All terrestrial and aquatic (lentic)
2	Growing season NEE fluxes with 3 or less measurement days in a month AND no modeling used in gap-filling	May-August terrestrial non-eddy covariance
3	NEE fluxes with 3 or more consecutive months of 100% gap-filling	Terrestrial eddy covariance



4	Sites with significant land use changes from humans	Selected terrestrial sites
CH₄ Flag		
0	Less than 30 g C m ⁻² month ⁻¹	All terrestrial, aquatic: only lentic non-eddy covariance diffusion and eddy covariance total flux
1	Greater than 30 g C m ⁻² month ⁻¹	All terrestrial, Aquatic: only lentic non-eddy covariance diffusion
2	Growing season CH ₄ fluxes with 3 or less measurement days in a month AND no modeling used in gap-filling	May-August terrestrial non-eddy covariance
3	CH ₄ fluxes with 3 or more consecutive months of 100% gap-filling	Terrestrial eddy covariance
4	Sites with significant land use changes from humans	Selected terrestrial sites

970 6. Spatial and temporal distribution of the dataset

Throughout the following sections, we describe the spatial and temporal distribution of the dataset. We use the column “site_name” to identify unique sites, and group the flux measurement techniques following the three measurement categories presented in Table 1. The dissolved gas-category includes concentration data alone and is included in the site-month numbers presented Figures 3-6. The term “site-month” refers to monthly data (fluxes and concentrations) and excludes seasonal fluxes (199 observations in the overall dataset). The key ecosystem categories used in visualizations are described in Supplementary Text 1; the terrestrial tundra class characterizes non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

980 6.1. Number of site-months, unique flux values and sites

The ABCFlux v2 dataset comprises 28,930 site-months, where each row represents a month with one to several unique gas fluxes (e.g. one site may include both a CH₄ and GPP flux for a given month) for a given site. In total, the dataset includes 55,560 unique flux values spanning CO₂ and CH₄ fluxes and their surface-atmosphere transport pathways. This “unique flux” value reflects the sum of all non-NA flux entries in the “nee”, “gpp”, “reco”, “co2_flux”, “diffusion”, “ebullition”, “total_ch4”, “storage”, and “seasonal” columns (see Supplementary Table 1). For aquatic concentrations, there are 8,801 site-months and 15,668 unique aquatic concentration measurements. Throughout the following sections, we focus on site-months rather than unique flux values. The number of site-months per site varied from 1 month to 330 site-months (1 to



990 330 site-months at terrestrial sites, 1 to 149 site-months at aquatic sites), with an average of 15
 site-months (38 for terrestrial sites, 4 for aquatic sites); note that some sites had some sporadic
 months that were completely gap-filled in these estimates. We identified six large site clusters in
 the data (> 800 monthly terrestrial and aquatic observations within a 30 km² radius): Toolik Lake
 (USA), Abisko-Stordalen (Sweden), Hyytiälä-Siikanen (Finland), Degerö-Flakaliden (Sweden),
 995 Fairbanks (USA), and Utqiagvik (formerly Barrow; USA), see Supplementary Figure 11.

The dataset includes a total of 1,024 sites, comprising 337 terrestrial and 711 aquatic sites
 (Table 5). Most sites collected data primarily during the growing season (May–August), while
 115 sites operate year-round, the majority of which (112) are eddy covariance sites in terrestrial
 1000 ecosystems. ABCFlux v2 includes 16 terrestrial year-round CO₂ flux sites with >10 years of
 data, mainly located in the boreal biome (14 sites). The longest time series of CO₂ fluxes in
 ABCFlux v2 in the boreal biome were from Hyytiälä (27 years), Degerö (23 years),
 Fyodorovskoye (22 years), University of Alaska Fairbanks (21 years), and in the tundra biome
 sites Eight Mile Lake (14 years), Imnavait Creek Watershed Heath Tundra (14 years), and
 1005 Imnavait Creek Watershed Tussock Tundra (14 years). There were 30 terrestrial sites with both
 year-round CO₂ and CH₄ flux data (at least for some of the years; see Supplementary Table 3).
 For CH₄, there were 33 terrestrial sites with year-round data with the longest time series at Trail
 Valley Creek (9 years), Cherskii reference (8 years), and Scotty Creek Landscape (8 years). For
 the aquatic dataset, CO₂ fluxes were measured year-round at 7 eddy covariance and non-eddy
 1010 covariance sites (Bernard Lake, Eastmain Reservoir, Iskoras, Kuivajärvi, Romaine-2 reservoir,
 Vanajavesi, Simpevarp) and at three of them CH₄ fluxes were measured year-round (Bernard
 Lake, Iskoras, Romaine-2 reservoir). ABCFlux v2 also includes automatic chamber
 measurements from 23 sites. However, these sites have shorter time series and contribute 15%
 of monthly CO₂ chamber measurements and 6% of monthly CH₄ chamber measurements.
 1015



Table 5. Number of sites and site-months for the whole dataset separated by terrestrial (CO₂ and CH₄), aquatic (CO₂ and CH₄), CO₂ only and CH₄ for growing season (May-August) and non-growing season (September- April) months. The table also lists the number of year-round sites and year-round sites with 5+ years of data. Note that some sites are both terrestrial and aquatic. A year-round site was defined as having at least some data for all months (no 100% gap-filled months) for at least one full year in the time series; if no data on gap-filling percentage was provided then we assumed there was some data for all months. The sum of terrestrial and aquatic site counts exceeds the total number of sites because some sites have both terrestrial and aquatic measurements.

	Gas species	All		Growing season			Non-growing season		Year-round	
		Sites	Site-months	Sites	Site-months	Sites with 5+ years	Sites	Site-months	Sites	Sites with 5+ years
ABCFlux v2		1,024	29,062	947	16,328	122	564	12,500	115	51
Terrestrial		337	18,952	304	9,208	108	264	9,722	109	49
	CO ₂	255	16,257	238	7,330	104	220	8,927	105	49
	CH ₄	173	7,433	156	4,426	30	120	3,007	33	8
Aquatic		711	10,110	664	7,120	15	311	2,778	7	2
	CO ₂	627	8,791	605	6,253	11	291	2,494	7	2
	CH ₄	463	8,855	454	6,370	11	278	2,383	3	1
All CO ₂		875	25,101	829	13,583	114	502	11,421	111	51
All CH ₄		651	16,415	596	10,796	41	391	5,390	35	9

6.2 Regional coverage

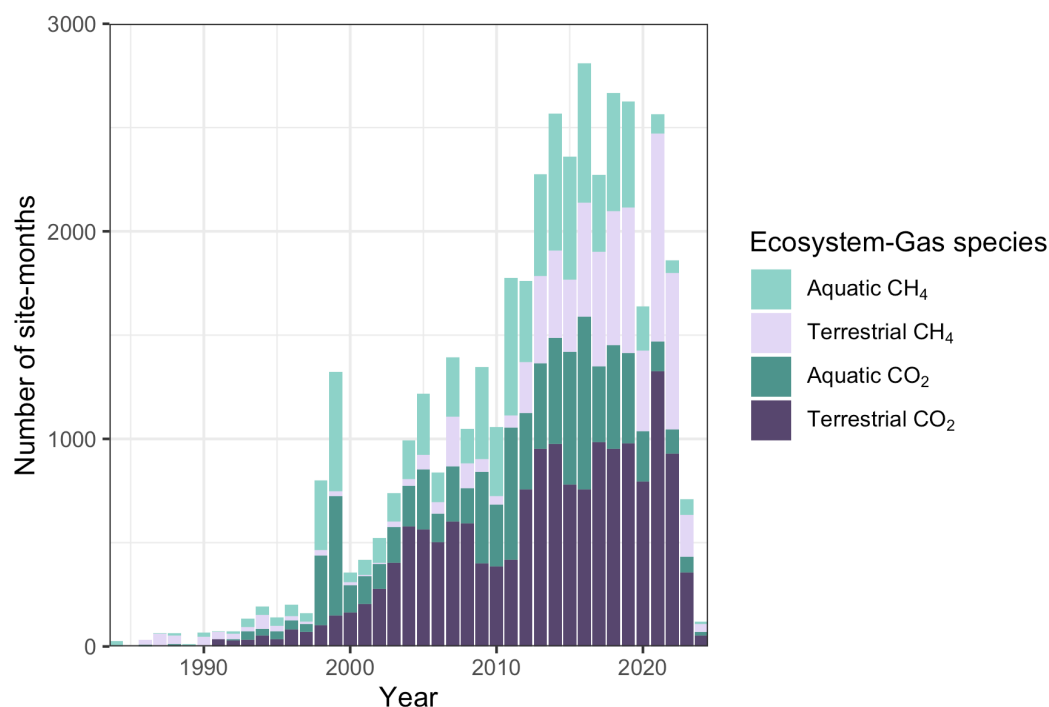
Terrestrial and aquatic sites are widely distributed across the Arctic-boreal domain (Fig. 1). However, the distribution of sites in ABCFlux v2 varies significantly across regions and key ecosystems in terms of the number of sites, site-months, and the gas species measured (Table 5, Supplementary Fig. 6). In terms of site-months in the full dataset, Alaska showed the highest coverage (24.5% of the dataset) with Sweden (20.9%), Canada (19.3%), and Finland (17.5%) following closely behind, and finally Russia (10.9%), Norway (3%), and Greenland (2.3%) (Supplementary Table 4). Among just the terrestrial data, Alaska, Finland, and Canada account for a majority of the site-months (23.6%, 20.6%, and 20.4% respectively), with Sweden only making up 14.8% of terrestrial observations. In contrast, measurements from Sweden made up the highest number (32%) of the aquatic site-months followed by Alaska (26.1%), Canada (17.8%), and Finland (11.6%). The majority of aquatic sites came from Finland (31.6%), and this



1040 dominance can be attributed to a few key aquatic studies that were incorporated in our dataset
(Juutinen et al., 2009b; Kortelainen et al., 2006). The density of sites in these countries is also
quite variable in terms of terrestrial and aquatic systems as well as the type of flux
(Supplementary Fig. 5).

6.3 Temporal and seasonal coverage

1045 Data in ABCFlux v2 span the years 1984 to 2024, with the majority of monthly observations
occurring after 1999 (92.5% in total, 95.1% terrestrial, 87.5% aquatic, Fig. 3). Years with the
largest amount of data are 2014-2019. The distribution of sites with more recent data is less
comprehensive (2022 as an example; Supplementary Fig. 9). This does not imply these sites
1050 discontinued data collection, but rather that the data were not made available at the time of this
synthesis. Regarding seasonal coverage, eddy covariance data are more evenly distributed
across the year compared to chamber and other non-eddy covariance measurement methods
(Supplementary Fig. 6). Terrestrial ecosystems show the highest data availability in July, with
1.5 to 3 times more data than other months, depending on the ecosystem (Fig. 4). In contrast,
1055 data coverage in aquatic ecosystems peaks in August, although overall coverage from June to
August remains similar. Aquatic data show a sharp decline during spring and autumn, whereas
terrestrial ecosystems experience a more gradual decrease outside the summer months.
Among terrestrial ecosystems, wetlands exhibit the strongest seasonal bias, with substantially
more observations in summer compared to other seasons.



1060 **Fig. 3. The number of CO₂ and CH₄ flux site-months across years. Numbers of**
observations are shown as stacked bars for each year. If a site-month measured both CO₂
 and CH₄, it is counted as separate site-months in the figure (1 CO₂ site-month
 and 1 CH₄ site-month). The distribution of sites with more recent data is less comprehensive (2022 as an
 example). This does not imply these sites discontinued data collection, but rather that the data
 1065 were not made available at the time of this synthesis

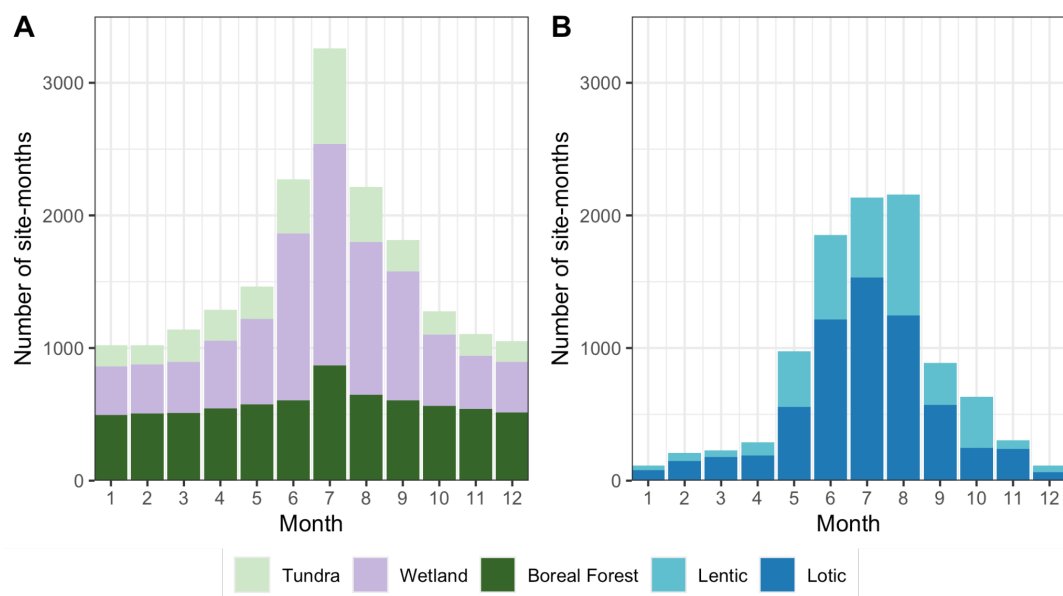


Fig. 4 The number of CO₂ and CH₄ monthly observations across months and key ecosystem types (A = terrestrial, B = aquatic). Aquatic concentration site-months are included in the figure. The terrestrial tundra class characterizes non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

6.4 Land cover type and disturbance coverage

Wetlands (including bogs, fens, permafrost bogs, tundra wetlands, marshes) had the highest number of flux observations (8,641 site-months), driven by the abundance of CO₂ flux measurements (Fig. 5). Boreal forests were the second most measured ecosystem type (6,981 site-months), while tundra systems were the least studied (3,308 site-months). CO₂ flux site-months dominated the dataset across most ecosystem types (24,048 site-months), except for lentic ecosystems, where CH₄ flux measurements (3,024 site-months) were more frequent than CO₂ flux measurements (2,858 site-months).

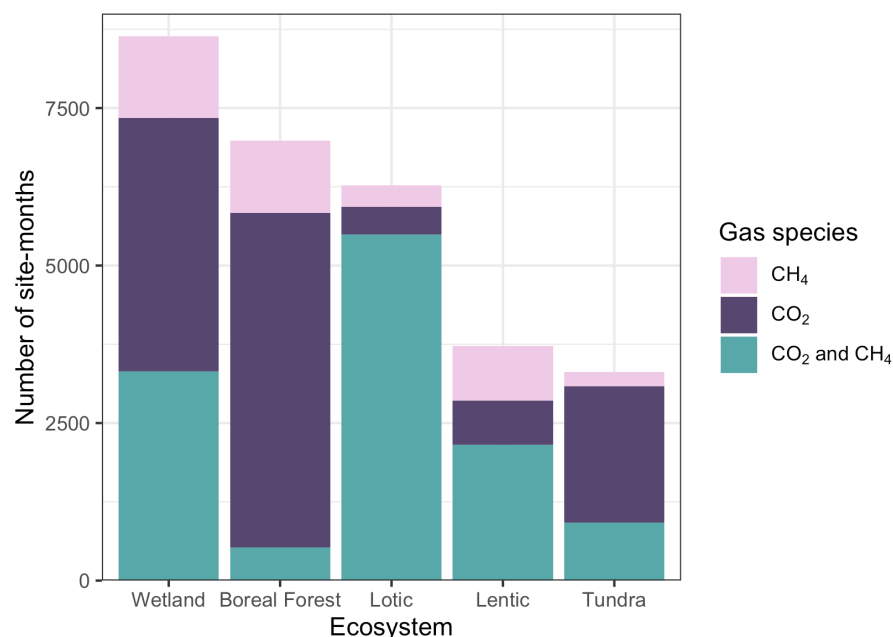


Fig. 5. The number of CO₂ and CH₄ flux and concentration site-months per key ecosystem types. Aquatic concentration site-months are included in the figure. The terrestrial tundra class characterizes non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

To better understand data coverage and calculate aerial extent across more detailed land and waterbody types, we used a combination of BAWLD and ESA CCI land cover data for terrestrial ecosystems (see Supplementary Text 1 for details) and a simplified version of BAWLD for aquatic ecosystems, focused on waterbody sizes. Overall, flux site-months were somewhat unevenly distributed across key land cover types relative to their areal extent (Fig. 6). Some classes were measured more in comparison to the area they cover (e.g., bogs, fens, permafrost bogs, wet tundra), while others showed more balanced representation (e.g., shrublands, graminoid ecosystems, evergreen needleleaf forests, mixed forests). In contrast, some classes were sparsely measured relative to their large areal extent (e.g., deciduous needleleaf forests, sparse vegetation and barren). For aquatic ecosystems, the most pronounced coverage biases were observed in the large lentic class, which was underrepresented relative to its total surface area, while small lotic ecosystems were disproportionately sampled compared to their small surface area extent. However, since flux data in our dataset are collected at the lake level, each site-month observation is weighted equally, regardless of lake size. This may not accurately reflect larger lakes, where a single observation can represent a much greater area. Similarly, this approach does not account for flux magnitudes or variability, which are variable across ecosystems. For example, lotic and wetland ecosystems often exhibit substantial CH₄ flux variability, necessitating more frequent observations for accurate representation compared to, for instance, barren ecosystems.

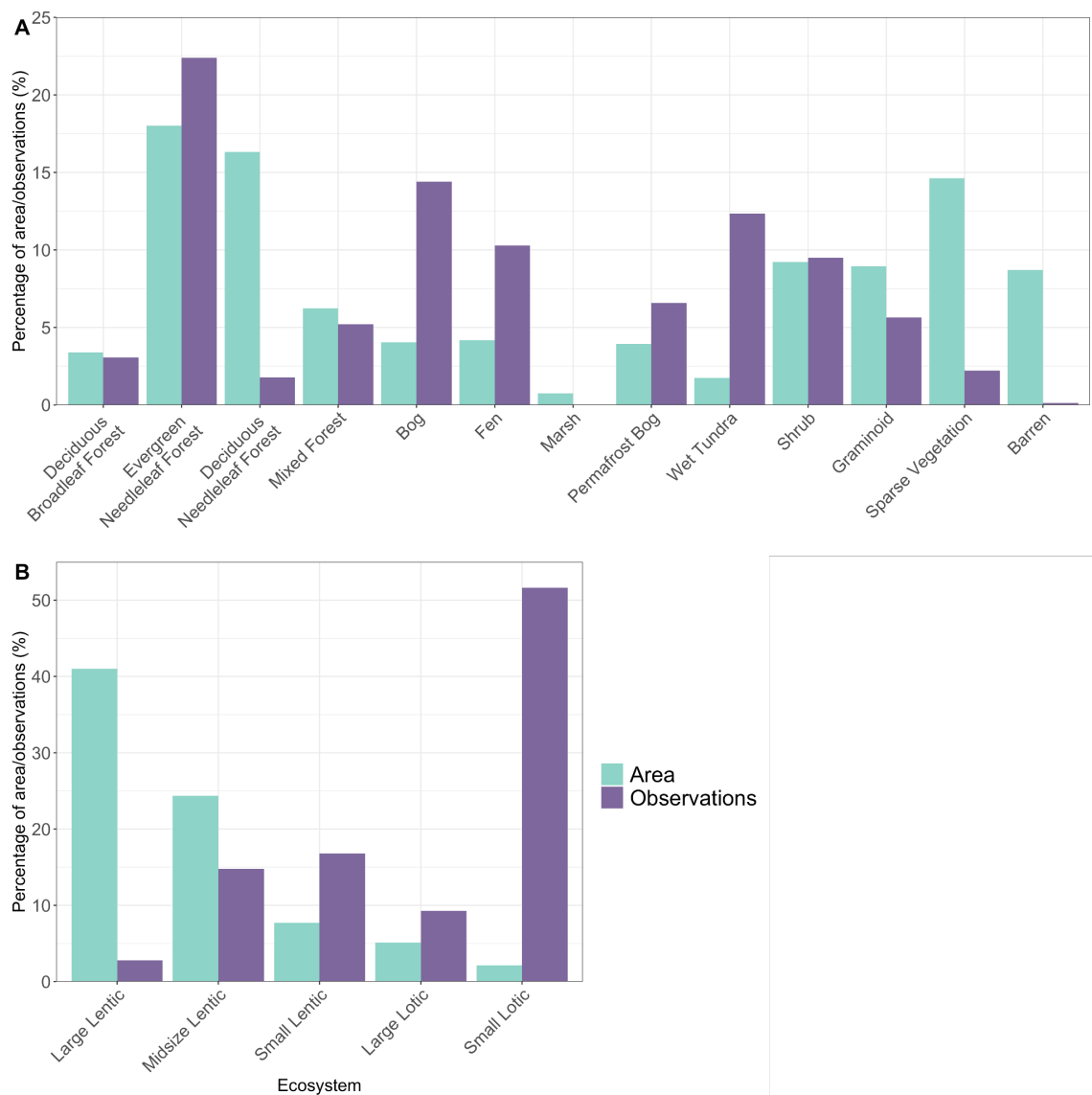


Fig. 6. The coverage of monthly flux observations and areas of key terrestrial (A) and aquatic (B) ecosystem types from the ESA land cover model. Aquatic concentration site-months are included in the figure. Some observations were left unclassified due to the lack of descriptive data and are not shown in the figure. There are 61 site-months of data from marshes, resulting in their proportion appearing nearly zero in the figure.



1110 There are 272 sites in ABCFlux v2 that reported a disturbance, constituting 11,363 site-months. These disturbances were broadly classified into 21 categories: altered hydrology, animal herbivory, artificial pond, beavers, drainage, drained lake, erosion, extreme weather, fire, forestry, human paths, insect herbivory, land use change, peat mining, reservoir, seismic lines, thaw, tidal effects, wastewater, and other. Among these categories, the most common
 1115 disturbance noted was “Thaw” (131 sites, 6,030 site-months), which can be attributed to the thaw category including various types of permafrost thaw. We further divide the thaw category into 13 categories: active layer detachment, active layer thickening, gradual thaw, ice-wedge degradation, palsa thaw, retrogressive thaw slumps, river bank erosion, subsidence, talik formation, thaw lake, thaw pond, thermokarst, thermokarst mounds with talik. Aside from thaw,
 1120 fire and forestry were the most frequent disturbances in terms of terrestrial sites (35 and 26 sites respectively). For aquatic sites, thaw was still the most common disturbance (70 sites) followed by wastewater, though the number of sites with this disturbance was much less (9 sites). The years that disturbances occurred varied from hundreds of years ago to more recent (2023) and on-going disturbances.

1125 6.5 Flux site data coverage comparison to earlier synthesis datasets

ABCFlux v2 includes 59% more sites and 158% more site-months for terrestrial CO₂ flux data compared to ABCFlux v1. ABCFlux v2 has substantially more sites with CO₂ data from the non-growing season (September–April; 212 vs. 141 in ABCFlux v1). ABCFlux v2 also has more recent data, with 2020–2024 making up 16% of the data; though the most recent years 2023–
 1130 2024 account for ~ 2% of site-months likely due to a delay in publication or processing of data.

While ABCFlux v2 has substantial overlap with the major international eddy covariance data repositories (FLUXNET2015, FLUXNET-CH₄, Ameriflux, and ICOS), ABCFlux v2 also incorporates a large number of additional sites and site-months from community-contributed
 1135 data (i.e., data provided directly by site PIs or researchers), enhancing both the spatial coverage and temporal range of the dataset. FLUXNET-2015 covers 26% of the terrestrial eddy covariance CO₂ site-months included in ABCFlux v2. However, many of these observations were also directly provided by data contributors or removed during data cleaning (see Sect. 3.1 for more details), reducing the net contribution of FLUXNET2015 to 18% of the terrestrial eddy
 1140 covariance CO₂ site-months. While other repositories such as Ameriflux (17% of terrestrial CO₂ site-months) and ICOS (10% of site-months) remain important data sources in the Arctic-boreal region, ultimately community-contributed data represent a substantial portion of the dataset (30% of site-months). This trend is even more pronounced in the terrestrial CH₄ eddy covariance data, where FLUXNET-CH₄ covers 28% of site-months in ABCFlux v2 but
 1145 contributes only 16% to the final dataset, while community-contributed data account for 58% of the observations.

Supporting data coverage remained relatively similar between the v1 and v2 datasets. For instance, soil organic carbon stock data were available for 16% of site-months in v1 compared
 1150 to 22% in v2, and soil moisture data for 35% in v1 versus 46% in v2. However, the v2



database shows improved coverage of disturbance information, increasing from 30% in v1 to 55% in v2.

1155 BAWLD-CH₄ terrestrial CH₄ flux data included 555 site-years (daily average over the growing season) of chamber data from 214 sites. Integrating and updating terrestrial CH₄ chamber data to monthly format in ABCFlux v2 resulted in 4520 site-months from 126 sites (using BAWLD-CH₄ “site” definition for comparison). BAWLD-CH₄ aquatic flux data included 396 site-years (daily average over the open-water season) of diffusive fluxes and 168 site-years of ebullitive fluxes from 391 and 151 sites, respectively. Integrating and updating aquatic flux data to monthly format in ABCFlux v2 resulted in 939-site-months from 286 sites for diffusion and 212
 1160 site-months from 27 sites for ebullition. The apparent loss of ebullition sites is because many of the original BAWLD-CH₄ sites include ebullitive flux data only at the seasonal timeframe and it was not possible to partition fluxes into monthly intervals. Seasonal ebullitive data were included in the “ch₄_flux_ebullition_seasonal” column instead.

1165 Of the lotic data, about 85% of site-months were transferred from the global river methane database (GRiMeDB), with the remaining 15% of data largely being submitted by data contributors or extracted from recent publications. Nine eddy covariance towers within the Arctic-boreal domain were included in the global analysis of lake and reservoir CO₂ fluxes (Golub et al., 2023). In comparison, ABCFlux v2 contains CO₂ flux data from 15 eddy covariance towers covering aquatic ecosystems, one of which was deployed on a Finnish river.
 1170 Otherwise, new tower sites were added from lakes in Canada and Scandinavia.

We also compared the number of sites in ABCFlux v2 with the recent ARctic greenhouse Gas Observation metadata version 1 (ARGO) (Vogt et al., 2024), where metadata across
 1175 observational platforms in the Arctic-boreal region were collected. Compared to eddy covariance site counts in ARGO, ABCFlux v2 includes approximately 40 fewer eddy covariance sites with carbon flux data from terrestrial and aquatic ecosystems. This discrepancy is due to data for those sites not being publicly available. In particular, ABCFlux v2 excludes recently established sites, whose data are not yet fully processed or shared, and older, short-term towers from the 2000s included in publications for which data could not be accessed. ABCFlux
 1180 v2 includes recent data since 2022 for only around half of the existing tower sites that are active according to ARGO, in most cases likely due to the delay in making data available. For non-eddy covariance sites, the site counts vary more strongly between ABCFlux v2 and ARGO due to differences in defining a ‘site’. When comparing unique site names between both datasets, ABCFlux v2 contains roughly three quarters of the sites from the observational platforms
 1185 Chamber and Dissolved in ARGO. In both datasets, the number of aquatic sites is larger than for terrestrial sites, whereas lentic and lotic sites are split evenly.

7. Flux synthesis

1190 Here we show a summary of monthly and annual CO₂ and CH₄ flux variability to provide a synthesis of flux magnitudes showing the data spread as well as highlight some uncertainties that the data user needs to be aware of. We used columns “nee” for terrestrial data and



“co2_flux” for aquatic data for the CO₂ balance estimate, and the “ch4_flux_total” column for the CH₄ flux for both ecosystems without differentiation between diffusion and ebullition. Monthly fluxes were summarized using the full dataset (i.e., not using any quality flags) and are reported in the main text for the key ecosystem classes as well as in the Supplementary material for the more detailed ecosystem classes (Supplementary Tables 5-10). We additionally summarize monthly concentrations of CH₄ and CO₂ for lentic and lotic ecosystems.

Total annual estimates were calculated by finding the monthly median fluxes per overarching class type (wetland, tundra, boreal forest, lentic lotic) across the entire dataset, and these were summed to obtain annual fluxes per each class. In this approach, median monthly fluxes from the entire year from terrestrial ecosystems were considered. For freshwater ecosystems, we summed emissions from May-October (6 months), and we assumed ice-emissions accounted for 17% of the total annual emissions (following Liu et al., 2022; Ramage et al., 2024). We calculated annual emissions using a subset of the data where all data with flags 1–4 were removed based on “expert_co2_flag” and “expert_ch4_flag” columns. The flags considered unrealistically high or low fluxes, uncertain fluxes due to sporadic chamber measurements, eddy covariance data with long periods of entirely gap-filled data, and sites that do not represent typical conditions across the Arctic-boreal region.

The monthly and annual CO₂ and CH₄ flux estimates from our dataset exhibit magnitudes and variability comparable to previous estimates across key ecosystems (Supplementary Figs. 12-14) and are briefly described below.

7.1 Average monthly fluxes and concentrations

Monthly NEE varied widely, ranging from -81.3 to 122.4 g C m⁻² month⁻¹ (2.5th–97.5th percentiles) and from -621.7 to 3850.0 g C m⁻² month⁻¹ (min–max) across the entire dataset (Fig. 7). The most extreme values were often driven by non-eddy covariance measurements with low data coverage per month. When subset by the five main ecosystems, monthly median NEE fluxes varied from -33.2 to 502.1 g C m⁻² month⁻¹ (Table 6). A few lotic sites in the boreal biome remained unfrozen in January and February and showed very high median fluxes (502.1 g C m⁻² month⁻¹); however it is generally assumed that most lotic, and also lentic, ecosystems are frozen and not actively releasing carbon this time of the year. Monthly median net emissions were highest in lotic ecosystems, followed by lentic ecosystems and October-April months in boreal ecosystems. Strongest net uptake values were found in boreal forests, followed by wetlands and tundra, whereas some aquatic net uptake up to -43.2 g C m⁻² month⁻¹ was observed as well (in total 24 lotic and 219 lentic site-months). Terrestrial ecosystems showed clear seasonal patterns in NEE, with uptake peaking in July (tundra and wetland) or June (boreal forests) and net emissions peaking in October-November (boreal forests and wetland) and September-October (tundra) (Table 6). Seasonal patterns in aquatic NEE were less clear, but net emissions were high throughout most of the year.



Monthly CH₄ flux varied from -0.1 to 6.2 (2.5th–97.5th percentiles) and from -9.3 to 145.4 (min–max) g C m⁻² month⁻¹. The extreme values were primarily associated with aquatic non-eddy covariance measurements, where diffusion and ebullition fluxes were summed to estimate total CH₄ flux. Ebullition fluxes in lentic ecosystems accounted on average for 53% (±33% standard deviation, 308 site-months) of the total CH₄ flux, and for 44% (±40%, 16 site-months) in lotic ecosystems. Six monthly total CH₄ fluxes exceeded expected ranges due to exceptionally high ebullition fluxes (>30 g C m⁻² month⁻¹). Monthly median CH₄ fluxes across key ecosystem types varied from -0.01 to 5.9 g C m⁻² month⁻¹ (Table 7). Monthly median CH₄ emissions were highest in lotic ecosystems, but numbers of site-months were small. Lentic ecosystems and wetlands followed (Fig. 8). Boreal forests were consistent small net CH₄ sinks (median) throughout the July–November period, while tundra remained neutral (0.0 g C m⁻² month⁻¹) or exhibited small net emissions, particularly in spring, autumn, and winter. 14% of tundra (non-wetland) and 16% of boreal forest site-months represented net CH₄ sinks. Emissions from wetlands, lentic and lotic ecosystems showed clear seasonal patterns with emissions peaking in July–August.

Table 6. Monthly median NEE (terrestrial) or CO₂ flux (aquatic) in g C m⁻² month⁻¹ across key ecosystems, with 25th–75th quantiles in parentheses. The terrestrial tundra class characterizes non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

Month	Boreal Forest	Tundra	Wetland	Lentic	Lotic
1	12.4 (6.5 to 19.8)	5.3 (1.5 to 9.7)	5.2 (2.5 to 8.6)	2.5 (0.6 to 7.3)	139.4 (59.7 to 386.3)
2	10.1 (4.7 to 15.8)	3.3 (0.3 to 8.4)	4.2 (1.7 to 6.6)	13.8 (2.7 to 22.8)	603.4 (547.8 to 659)
3	8.9 (3.1 to 15.6)	1.6 (0.3 to 6.1)	3.9 (1.3 to 7.2)	2.8 (1.1 to 10.4)	14.8 (6.0 to 57.2)
4	1.5 (-16.9 to 9.8)	1.6 (0.4 to 5.5)	2.4 (0.2 to 6.2)	6.9 (3.9 to 13.4)	50.6 (24.9 to 128.2)
5	-26.0 (-48.2 to -9.9)	4.0 (-0.3 to 9.5)	-2.3 (-12.2 to 2.7)	23.4 (6.9 to 52.8)	41.4 (17.4 to 90.2)
6	-48.0 (-71.9 to -26.1)	-13.8 (-33.3 to 2.8)	-15.4 (-29.8 to -3.2)	11.5 (4.8 to 26.1)	78.9 (35.5 to 195.9)
7	-34.1 (-59.8 to -14.7)	-26.6 (-53.3 to -4.4)	-32 (-54.6 to -12.9)	5.6 (1.8 to 14.9)	49.6 (17.9 to 161.3)
8	-11.8 (-33.6 to 7.1)	-11.9 (-27.5 to -1.3)	-17.5 (-33.3 to -8.0)	7.9 (2.1 to 19.2)	62.0 (24.1 to 167.6)



9	11.6 (-4.1 to 25.8)	12.2 (4.7 to 18.0)	1.3 (-5.6 to 9.8)	14.2 (4.9 to 31.0)	63.2 (25.5 to 164.6)
10	20.2 (12.3 to 33.8)	11.6 (7.5 to 17)	9.2 (5.4 to 14.8)	14.5 (5.3 to 24.2)	38.3 (15.0 to 75.3)
11	19.5 (12.1 to 29.7)	8.7 (4.4 to 13.7)	7.7 (4.9 to 11.8)	10.9 (7.4 to 14.8)	41.2 (14.7 to 122.3)
12	14.8 (7.6 to 24.4)	6.8 (2.5 to 11.6)	6.5 (3.7 to 10.4)	2.6 (-0.2 to 7.5)	14.8 (0 to 71.2)

1250

Table 7. Monthly median total CH₄ flux (g C m⁻² month⁻¹) across key ecosystems, with 25th–75th quantiles in parentheses. The terrestrial tundra class characterizes non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

Month	Boreal Forest	Tundra	Wetland	Lentic	Lotic
1	0 (-0.01 to 0.07)	0.01 (0.01 to 0.08)	0.1 (0 to 0.4)	0.1 (0 to 0.2)	0 (0 to 0)
2	0 (-0.01 to 0.08)	0.01 (0.01 to 0.05)	0.1 (0 to 0.2)	0.1 (0 to 0.2)	0 (0 to 0)
3	0 (0 to 0.05)	0 (0 to 0)	0.1 (0 to 0.3)	0.1 (0 to 0.2)	0.1 (0.1 to 0.2)
4	0 (-0.01 to 0)	0 (0 to 0)	0.1 (0 to 0.4)	0.1 (0 to 0.3)	0.1 (0 to 0.1)
5	0 (-0.01 to 0.17)	0.12 (0.02 to 0.28)	0.3 (0 to 0.9)	0.2 (0 to 0.9)	0.3 (0.1 to 3.2)
6	0 (-0.03 to 0.13)	0 (0 to 0.18)	0.4 (0.1 to 1.3)	0.4 (0.1 to 0.9)	1.1 (0.8 to 2.2)
7	-0.01 (-0.05 to 0.16)	0 (-0.04 to 0.09)	0.9 (0.2 to 2.5)	0.7 (0.2 to 1.3)	3.7 (0.3 to 17.6)



8	-0.01 (-0.04 to 0.1)	0 (-0.01 to 0.37)	1.2 (0.3 to 2.5)	0.7 (0.3 to 1.3)	3.1 (1.5 to 8.8)
9	0 (-0.03 to 0.19)	0.08 (0.01 to 0.71)	0.6 (0.1 to 1.5)	0.4 (0.2 to 0.7)	1.2 (0.5 to 1.7)
10	-0.01 (-0.03 to 0.06)	0.02 (0 to 0.14)	0.5 (0.2 to 0.8)	0.2 (0.1 to 0.3)	0.3 (0.2 to 1.1)
11	-0.01 (-0.03 to 0.06)	0.03 (-0.01 to 0.14)	0.3 (0.1 to 0.5)	0.2 (0 to 0.3)	0.1 (0.1 to 0.1)
12	0 (-0.02 to 0.04)	0.01(0 to 0.11)	0.2 (0.1 to 0.4)	0.2 (0 to 0.3)	0.1 (0 to 0.1)

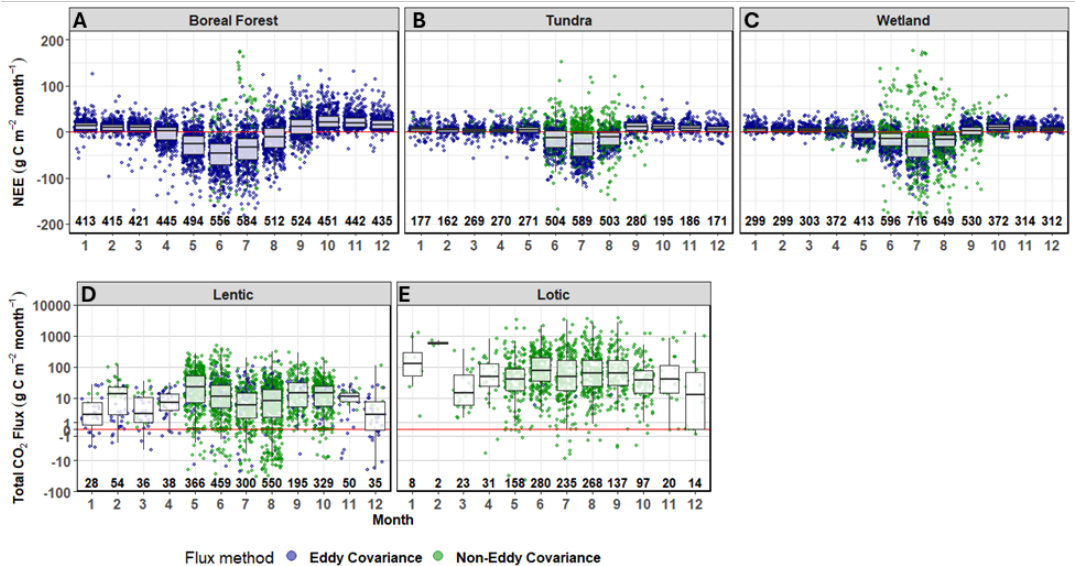


Fig. 7. Monthly NEE (terrestrial) and total CO₂ flux (freshwater) variability across key terrestrial and aquatic classes together with the number of site-months. Note that y axes for the lentic and lotic fluxes follow a pseudo-log scale and represent the CO₂ flux. The terrestrial tundra class characterizes non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

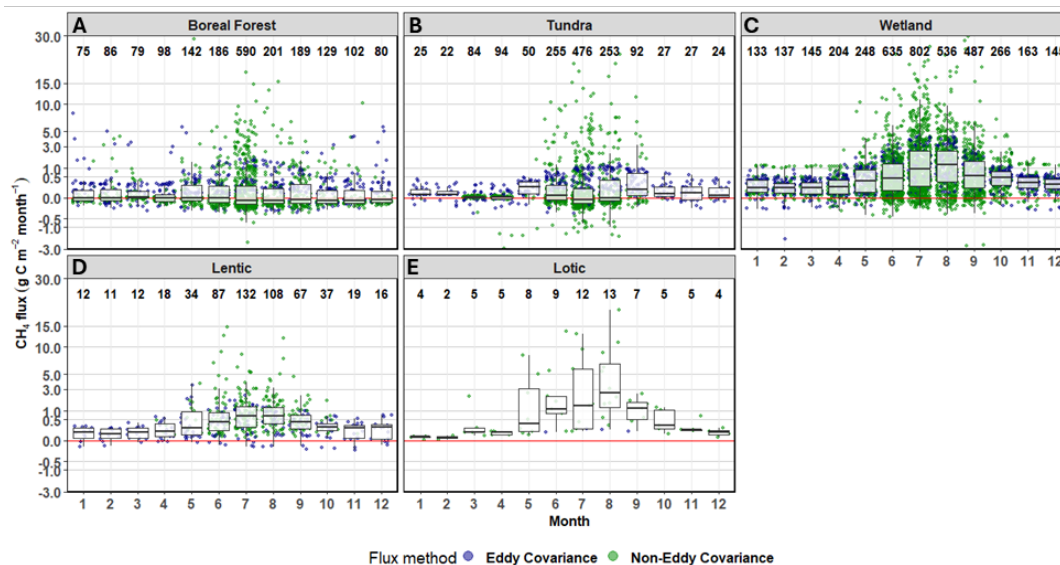


Fig. 8. Monthly CH_4 fluxes across key terrestrial and aquatic classes together with the number of site-months. Note that y axes follow a pseudo-log scale. The tundra classes characterize non-wetland ecosystems in the tundra biome (i.e., dry and moist tundra).

Monthly dissolved CO_2 concentrations varied from 12.6 to 5588 (2.5th–97.5th percentiles) and from 0 to 5480 (min–max) $\mu\text{mol L}^{-1}$. Monthly median CO_2 concentrations were generally highest in lotic ecosystems and followed a seasonal pattern of lower concentrations during the summer months (Fig. 9). Monthly CO_2 concentrations for lentic systems were highest in spring, but these months also had the smallest sampling sizes. Monthly dissolved CH_4 concentrations varied from 0 to 13.1 (2.5th–97.5th percentiles) with a maximum concentration of 530 $\mu\text{mol L}^{-1}$. Lotic ecosystems trended towards CH_4 higher concentrations and there were no clear monthly patterns for either freshwater ecosystem.

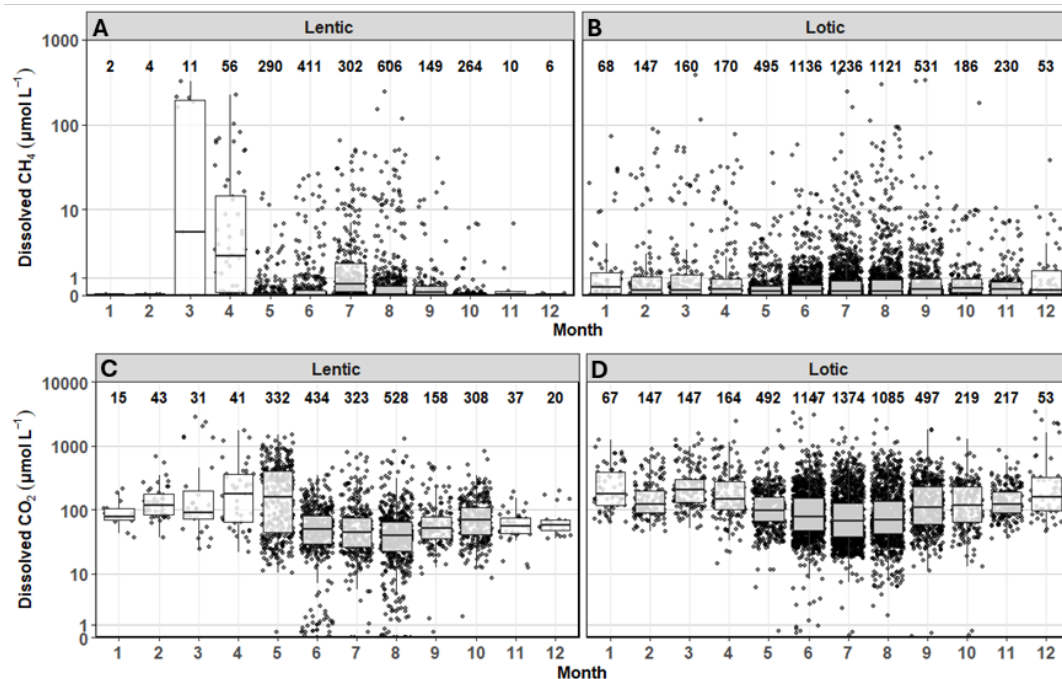


Figure 9. Monthly dissolved CH₄ and CO₂ concentrations across freshwater classes together with the number of site-months. Note that y axes follow a pseudo-log scale.

7.2 Average annual fluxes

Our annual NEE estimates showed a clear sink-to-source transition from net CO₂ sinks in wetlands (-34.0 g C m⁻² yr) forests (-22.2 g C m⁻² yr), and tundra (-10.7 g C m⁻² yr) to CO₂ sources in lentic (142 g C m⁻² yr) and especially lotic ecosystems (1180 g C m⁻² yr). However, the interquartile range for all three terrestrial classes included positive NEE estimates (Table X). Annual CH₄ fluxes demonstrated the largest median net CH₄ source for wetlands (5.6 g C m⁻² yr⁻¹) followed by lentic waterbodies (3.1 g C m⁻² yr⁻¹), lotic waterbodies (2.4 g C m⁻² yr⁻¹) and then tundra (0.6g C m⁻² yr⁻¹). Boreal ecosystems were near neutral (0.07 m⁻² yr⁻¹), with an interquartile range between -0.26 and 1.53 g C m⁻² yr⁻¹). It's important to note that these figures are not upscaled estimates and therefore may not accurately reflect the overall sink–source status of the Arctic–boreal region. For example, many freshwater systems exhibit strong CO₂ sink during the summer months (Fig. 8) and our approach using the median values for each ecosystem type likely masks this uptake signal and CH₄ uptake in upland ecosystems. Furthermore, we provide only a simplified estimate of the contribution of ice-out emissions to total annual estimates for freshwater ecosystems.

Table 8. Estimates of annual fluxes (g C m⁻² yr⁻¹) for each broadly defined ecosystem. Values represent the median annual emission estimate followed by first and third quartiles.



Ecosystem	Annual CH ₄ flux (g C m ⁻² yr ⁻¹)	Annual NEE flux (g C m ⁻² yr ⁻¹)
Wetland	5.6 (1.9, 11)	-34.0 (-111, 44.2)
Boreal Forest	0.071 (-0.26, 1.5)	-22.2 (-182, 117)
Tundra	0.62 (0.1, 3.6)	-10.7 (-96, 82.6)
Lentic	3.1 (0.46, 7.1)	142 (64.8, 245)
Lotic	2.4 (1.5, 4.0)	1180 (893, 2050)

8. Discussion

8.1 Future research directions

1305 While a detailed analysis of the underlying mechanisms of flux patterns is beyond the scope of
 this data description paper, our data compilation offers several new perspectives and research
 opportunities. For example, ABCFlux v2 opens up for the opportunity to further investigate the
 detailed characterization of land cover types, waterbody classifications, and disturbance history
 and with that provides valuable context often overlooked in recent syntheses, which have
 1310 typically employed coarse classifications (e.g., treating "boreal forest" as a single category, as
 seen in Virkkala et al., 2021 and Ramage et al., 2024 studies, or aquatic ecosystems split only
 across permafrost zones as in Song et al., 2024). With ABCFlux v2, it is possible to effectively
 detect temporal trends, including those in CH₄ fluxes. The inclusion of both CO₂ and CH₄ fluxes
 also allows for further analysis of their ratios under changing environmental conditions.
 1315 Moreover, the monthly format provides a clearer understanding of seasonal dynamics, offering
 an improvement over earlier studies that primarily focused on growing season or annual
 cumulative fluxes (Kuhn et al., 2021; Ramage et al., 2024; Virkkala et al., 2021).

ABCFlux v2 further allows for an improved understanding of some under-studied flux dynamics.
 1320 While summertime net CO₂ uptake was previously found at individual sites (Emmerton et al.,
 2016; Prėskienis et al., 2021), a broader-scale analysis of the underlying conditions and the
 extent of CO₂ uptake is worth undertaking. This points towards CO₂ and CH₄ sources playing a
 more complex role in regional carbon budgets than previously understood (Bogard et al., 2019;
 Tank et al., 2009). Lotic systems exhibit highly variable yet substantial per m² emission rates
 1325 that, according to our synthesis, appear higher than those of any other ecosystem type.
 However, the degree to which these elevated emissions occur across entire river networks
 remains uncertain, as does the influence of local (e.g., steep or shallow sections with high gas
 transfer velocities and emissions vs. flatter or deeper sections with lower emissions;
 (Natchimuthu et al., 2017) and circumpolar (e.g., climate gradients) drivers of variability.



1330 Importantly, the overall contribution of lotic systems to regional carbon budgets is constrained
 by their limited spatial extent—they occupy only about 0.5% of the Arctic–boreal domain
 (Olefeldt et al., 2021b). As a result, despite high per m² fluxes, total emissions from rivers may
 be smaller than those from more widespread ecosystems such as wetlands, which have lower
 per m² emissions but a much greater areal extent (Casas-Ruiz et al., 2023). Nevertheless,
 1335 ABCFlux v2 offers a valuable opportunity to address these aquatic knowledge gaps.

Our dataset also shows distinctive patterns in the seasonal dynamics of CH₄ exchange across
 different ecosystems. Particularly noteworthy is the contrast between non-wetland tundra and
 boreal forest systems: tundra areas function as net CH₄ sinks in ABCFlux v2 for only a single
 1340 month, while boreal forest ecosystems maintain CH₄ uptake throughout the summer period as
 well as autumn period (Fig. 8).

8.2 Remaining gaps

ABCFlux v2 demonstrates clear improvements in carbon flux data quantity over time but also
 1345 shows that some ecosystems (e.g., deciduous needleleaf forests, barren and sparsely
 vegetated ecosystems, large lentic ecosystems) or periods (non-growing season, and especially
 the challenging spring ice-off period in aquatic ecosystems) still remain poorly captured.
 Moreover, the studied sites are heavily clustered in a few regions in Alaska and Fennoscandia,
 therefore leaving significant spatial gaps in coverage in spite of the relatively high absolute
 1350 number of locations. In particular, despite Russia's large land mass (close to 60% of the
 domain), data from Russia only comprise 10.8% of the dataset in terms of site-months and
 15.3% of sites in ABCFlux v2 (with similar representativeness among terrestrial and aquatic
 data), making this region a critical data gap. At local scales, flux data across lentic depth zones
 and throughout entire lotic networks are critically needed to better estimate spatially
 1355 representative waterbody fluxes. Furthermore, most freshwater measurements were taken
 during daytime, leaving nighttime dynamics less constrained. To more reliably estimate the net
 ecosystem carbon balance, incorporating lateral fluxes into future studies would be a large
 benefit, especially in landscapes affected by permafrost thaw (Zolkos et al., 2022).

1360 Additionally, some specific flux mechanisms and environmental controls remain undersampled.
 This is particularly true for ebullition, which only makes up 12% of aquatic CH₄ fluxes in the
 dataset but can account for up to 90% of total aquatic CH₄ emissions (Walter et al., 2010; Kuhn
 et al., 2021). Moreover, plant-mediated CH₄ emissions in aquatic systems and CH₄ fluxes from
 terrestrial trees are not explicitly partitioned in the dataset; instead, they are included as part of
 1365 total fluxes measured primarily by eddy covariance. Improved partitioning of these flux pathways
 would enhance our understanding of the processes driving CH₄ dynamics and help to better
 constrain landscape-scale carbon budgets (Gauci et al., 2024; Iwata et al., 2018; Juutinen et al.,
 n.d.; Kankaala et al., 2005; Kyzivat et al., 2022). Important site-level environmental data such as
 thaw depth, soil carbon stocks, and comprehensive information on plant and microbial
 1370 communities are partly or entirely missing, yet would provide valuable insights into the
 processes governing fluxes. Likewise, more detailed disturbance metrics and a stronger



integration of disturbed sites into flux monitoring networks are needed to better capture the impacts of these changes on carbon fluxes and budgets.

1375 Although we excluded chamber measurements for CO₂ fluxes taken underneath the canopy of
 forests, as chambers do not capture the full ecosystem dynamics it is notable that these
 chamber measurements could be valuable end points and quality assessments when co-located
 with towers, providing a lower limit on respiration estimates. We did include below canopy
 chamber CH₄ flux measurements under the assumption that CH₄ fluxes from trees are
 1380 negligible. However, recent evidence suggests that trees may play a globally significant role in
 CH₄ uptake, even considering that CH₄ uptake by trees decreases with increasing latitude and
 approaches zero in low mean annual temperatures (Gauci et al., 2024; Sundqvist et al., 2012).
 Furthermore, some studies suggest that boreal trees emit CH₄ from tree stems (Klaus et al.,
 2024; Machacova et al., 2023; Vainio et al., 2022), suggesting more tree-based flux
 1385 measurements are needed and should be incorporated into future synthesis efforts.
 Furthermore, given the low data coverage of 2% for stable isotope CO₂ and CH₄ measurements
 among the aquatic data in ABCFlux v2, the lack of our understanding of emission pathways and
 sources becomes apparent. Therefore, increasing the number of observations across
 ecosystems would significantly improve the source attribution of emissions as well as our
 1390 process understanding.

While long-term, year-round terrestrial CO₂ flux sites have become more common, the need
 remains to expand the network to more aquatic and under-represented terrestrial sites where
 year-round measurements of both CO₂ and CH₄ fluxes are collected simultaneously. A larger
 1395 number of sites measuring these fluxes throughout the year would also inform and improve
 wintertime process understanding. Long-term CH₄ flux monitoring sites remain scarce, hindering
 our ability to detect temporal trends in CH₄ emissions—despite their potentially critical role in a
 changing climate (Turetsky et al., 2020). Thus, maintaining existing sites and setting up new
 permanent CH₄ flux sites is critical for accurate understanding on the changing carbon cycle in
 1400 Arctic-boreal regions. The availability of recent, “real-time”, data remains challenging as often
 post-processing times can delay the release of the data. Improving the turn-around time and
 associated pipelines from data collection to availability will improve future synthesis and
 modeling efforts.

1405 An additional notable data gap is the lack of information on nitrous oxide (N₂O), another
 significant GHG that is currently absent from the ABCFlux v2 dataset, which primarily focuses
 on CO₂ and CH₄. This omission reflects a broader data gap, despite growing evidence that N₂O
 emissions may become increasingly important for regional and global climate feedbacks as
 permafrost thaw accelerates (Voigt, Marushchak). As N₂O has a global warming potential over
 1410 300 times that of CO₂ over a 100-year timescale, even relatively small fluxes may significantly
 contribute to climate feedbacks. Incorporating N₂O measurements in future efforts would be
 critical to achieving a more comprehensive understanding of GHG dynamics in the Arctic-boreal
 domain.



8.3 The importance of regional networks

1415 While contributing data to global repositories remains highly encouraged, regional syntheses
 like this provide valuable insights into network status and development needs across the Arctic-
 boreal domain. In our case, it has helped in establishing a collaborative community that bridges
 terrestrial and aquatic researchers while integrating CO₂ and CH₄ flux data in a standardized
 format, creating a foundation for collectively improving Arctic-boreal carbon cycle
 1420 understanding. Looking ahead, we strongly advocate for maintaining a community-driven
 approach in future flux syntheses, similar to earlier ABCFlux initiatives (See et al., 2024;
 Virkkala et al., 2022, 2025a) and those supported by, for example, the Permafrost Carbon
 Network (Schuur et al., 2015, 2022). Including data contributors as co-authors not only ensures
 proper recognition of their contributions but also draws on their expertise in data interpretation.
 1425 Moreover, there is potential for efforts like ABCFlux to gradually evolve into more dynamic,
 continuously updated resources. While not yet realized, the aspiration to create a database
 capable of integrating recent flux measurements and associated metadata remains important for
 enabling iterative updates (e.g. <https://fluxnet.org/fluxnet-data-system/>) and the incorporation of
 new knowledge (see e.g., (Arctic Terrestrial Carbon Cycling, 2025)). Ultimately, such a
 1430 collaborative and adaptive approach enhances both the technical quality and scientific context
 of carbon flux syntheses, helping to advance our understanding of high-latitude carbon
 dynamics in a rapidly changing environment.

9. Data use guidelines

1435 ABCFlux v2 data is categorized into two usage tiers depending on the data source, as indicated
 in the "data_usage" column. Tier 1 data is open and free to use for scientific and educational
 purposes. In contrast, Tier 2 requires that data users give data producers the opportunity to
 collaborate and consult with them. However, it is recommended that all researchers reach out
 and collaborate with the dataset developers and relevant site data producers when using the
 1440 dataset as a core component of their analysis. This encourages a more informed and context-
 rich use of the data.
 If used, the dataset should be referenced by citing both this paper and the dataset (DOI:
 10.3334/ORNLDAAAC/2448).

10. Data availability

1445 The dataset associated with this publication can be found in ORNL DAAC at
<https://doi.org/10.3334/ORNLDAAAC/2448> (Virkkala et al., 2025b).

11. Conclusions

ABCFlux v2 provides the most comprehensive dataset of surface-atmosphere Arctic-boreal
 ecosystem CO₂ and CH₄ fluxes to date. It is particularly useful for machine learning or
 1450 regression modeling for drivers of fluxes, process model tuning, remote sensing-based



upscaling, and empirical studies aiming to understand carbon budgets and regional variability in flux magnitudes, as well as changes in fluxes through time.

Author contributions

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