

Australia's terrestrial industrial footprint and ecological intactness

Ruben Venegas-Li¹, Scott Atkinson^{1,2}, Milton A U de Andrade Junior¹, Rachel Fletcher³, Peter Owen⁴, Lucia Morales-Barquero¹, Bora Aska¹, Miguel Arias-Patino⁵, Hedley S. Grantham^{6,7}, Hugh Possingham¹, Oscar Venter⁵, Michelle Ward^{1,8}, James E.M. Watson¹

5 ¹ Centre for Biodiversity and Conservation Science, The School of the Environment, University of Queensland, St Lucia, 4072, Australia

² United Nations Development Program, New York, NY, United States

³ The Wilderness Society, GIS, South Brisbane, Queensland, Australia

⁴ The Wilderness Society (South Australia) Inc., 111 Franklin Street, Adelaide, SA 5000, Australia

10 ⁵ Natural Resources and Environmental Studies Institute, University of Northern British Columbia, Prince George, British Columbia V2N 4Z9, Canada

⁶ Center for Ecosystem Science, School of Biological, Earth and Environmental Sciences, University of New South Wales, Sydney, New South Wales, Australia

⁷ Bush Heritage Australia, Melbourne, Victoria, Australia

15 ⁸ School of Environment and Science, Griffith University, Nathan, QLD 4111, Australia

Correspondence to: Ruben Venegas Li (r.venegas@uq.edu.au)

20 **Abstract.**

Australia's unique biodiversity faces significant threats from anthropogenic activities that drive habitat destruction and degradation. This study presents the first comprehensive national-scale cumulative pressure map for terrestrial Australia since the 1980s, providing key insights into human disturbance of the landscape. We developed a Human Industrial Footprint (HIF) index map at a 100 m spatial resolution, incorporating 16 nationally relevant pressure layers, which offers a contemporary
25 assessment of cumulative pressures on Australia's landscapes. The HIF was used to derive an Ecological Intactness Index (EII), which accounts for habitat loss, quality, and fragmentation, to provide an estimate of an area's structural intactness. A technical validation comparing visually scored pressures in 1397 stratified random samples using high-resolution satellite images revealed a strong agreement with the mapped pressure values (the HIF). We also conducted an uncertainty (sensitivity) analysis by adjusting individual pressure scores by up to $\pm 50\%$ across 100,000 simulations, which showed a moderate impact
30 on cumulative pressure scores, confirming the robustness of our approach. We believe both the HIF and EII datasets can be valuable tools for guiding conservation efforts, such as informing protected area expansion, ecosystem restoration priorities, and biodiversity offset strategies. By offering a detailed assessment of cumulative pressures and ecological integrity, this study addresses a critical knowledge gap, and can support evidence-based decision-making for Australia's biodiversity conservation and sustainable development objectives. The HIF, EII, and scaled pressure layers are available at [10.5281/zenodo.15833395](https://doi.org/10.5281/zenodo.15833395)
35 (Venegas-Li et al., 2025).

Introduction

Australia is globally recognized as one of the most biodiverse countries on Earth, hosting an array of species and ecosystems found nowhere else (Chapman, 2009). However, since European colonization, industrial activities such as agriculture, forestry, and urbanization have caused widespread habitat destruction, fragmentation, and pollution of the natural environment. As a result, during the last 200 years, one-third of native vegetation has been lost (Bradshaw, 2012; Kingsford et al., 2009; Ward et al., 2019). Over 2,100 species and 100 ecological communities are now legislated as threatened with extinction in the near term, and 103 species have become extinct (Commonwealth of Australia, 2025). The threatening processes driving these declines remain largely unabated (Kearney et al., 2023; Legge et al., 2023; Woinarski et al., 2015), and understanding how their extent and intensity interact with natural systems is key to informing conservation policy (Kearney et al., 2019).

The field of cumulative pressure mapping, in which data on multiple pressures are integrated under a spatial model (maps), has become a widely used approach to estimate human pressures on the environment (Watson et al., 2023b). Here, we use the term “pressure” to denote human activities with the potential to harm nature (Borja et al., 2006; Martins et al., 2012), broadly corresponding to “direct threats” or “stressors” in the IUCN Threat and Stress Classification Scheme (Salafsky et al., 2008, 2025). Such pressure maps are increasingly used as proxies for human influence on ecological state and condition, particularly within pressure-state-response frameworks used to guide adaptive planning and management (Watson and Venter, 2019). The conceptual foundations of cumulative pressure maps emerged in the 1980s (Lesslie and Taylor, 1983, 1985; McCloskey and Spalding, 1989), but the discipline has expanded rapidly over the past two decades, with advances in Earth observation and geographic information systems (Watson et al., 2023b; Watson and Venter, 2019). The Human Footprint of Sanderson and colleagues (2002) is arguably one of the most influential early global assessments of humanity’s influence on the terrestrial planet, and mapped at a 1 km resolution, provided a framework to quantify anthropogenic influence across nine major pressures. This framework has since been refined and adapted to incorporate additional pressures (Kennedy et al., 2019; Venter et al., 2016a), regional contexts (González-Abraham et al., 2015; Hirsh-Pearson et al., 2022; Martinuzzi et al., 2021; Theobald, 2013; Woolmer et al., 2008), and alternative models for aggregating pressures (Halpern et al., 2008; Theobald, 2013), while recent efforts have achieved spatial resolutions of 100–300 m and annual updates (Gassert et al., 2023; Mu et al., 2022; Theobald et al., 2025). Comparable methods have also been applied in marine systems to quantify the extent and intensity of human use of the oceans (Ban et al., 2010; Halpern et al., 2008, 2015; Micheli et al., 2013).

Cumulative pressure maps are understood to represent potential human influence rather than the realised ecological state or condition of natural systems (Theobald et al., 2025; Venter et al., 2016b). Nonetheless, they have become foundational datasets for ecological research, conservation planning, and environmental reporting, where higher pressures correspond to degraded or lower ecological integrity areas, and lower pressures to areas closer to their natural state. For example, these maps have been used to evaluate relationships between human pressures and species extinction risk (Di Marco et al., 2018; Ramírez-Delgado et al., 2022; Torres-Romero et al., 2025), analyse changes in global mammal distributions (Tucker et al., 2021), population level changes in great apes' behaviour and densities (Kühl et al., 2019; Ordaz-Németh et al., 2021), as well as model

70 the spread of infectious diseases (Skinner et al., 2023). Moreover, cumulative pressure maps have been used in major environmental assessments, including the IPBES Global Assessment (Purvis et al., 2019), the Intergovernmental Panel on Climate Change (IPCC) reports (Masson-Delmotte et al., 2018), and the latest Global Biodiversity Outlook (Hirsch et al., 2020), where they have directly informed indicators of human impact and ecosystem condition.

Pressure maps have been used as surrogates for ecological intactness. However, intactness (often used as a synonym for areas
75 of high integrity) describes the degree to which systems retain their natural composition, structure, and function (Nicholson et al., 2021). Pressure maps may therefore not fully capture intactness, as they do not account for the spatial configuration and habitat-quality context surrounding each pixel (Theobald et al., 2025). To overcome this, Beyer and colleagues (2020) developed a metric to estimate ecological intactness, which integrates relative habitat quality with the degree of fragmentation, using cumulative pressure maps as the base layer. This approach provides a spatially explicit measure of the structural
80 dimension of integrity, complementing cumulative pressure maps that represent direct human influence. Developing intactness metric datasets is particularly important in the context of the global conservation agenda (Mendez Angarita et al., 2025) because targets have been set for retaining ecological intactness in the Kunming-Montreal Global Biodiversity Framework (GBF) (CBD, 2022), to which Australia is a signatory and has made commitments to. Specifically, the ecosystem component of the GBF's Goal A aims to ensure "the integrity, connectivity, and resilience of all ecosystems are maintained, enhanced, or
85 restored, substantially increasing the area of natural ecosystems by 2050". This is to be achieved through activities including protection and restoration (Targets 1-3) (CBD, 2022).

In Australia, Lesslie and colleagues carried out pioneering work in the 1980s to create the first pressure map at a national scale (Lesslie et al., 1988; Lesslie and Taylor, 1983, 1985). However, no similar efforts have been carried out subsequently for the country, making the available national data highly dated. Global efforts have mapped pressures across Australia, but some of
90 these use a limited set of globally available datasets to represent pressures (Gassert et al., 2023; Mu et al., 2022; Sanderson et al., 2002; Williams et al., 2020) and therefore miss nation-specific critical pressures (Hirsh-Pearson et al., 2022). Others, such as Theobald et al. (2020, 2025), incorporate additional pressures and finer spatial resolution, which likely improves their representation of human influence across many landscapes (Arias-Patino et al., 2024). However, these models may remain constrained by the need to use globally consistent datasets, which might fail to represent features such as rural roads, farm
95 dams, and small-scale mining that are sometimes better mapped within national boundaries. National-scale assessments can overcome some of the limitations of global models by integrating detailed, locally curated datasets derived from bottom-up data collection and long-term government monitoring programs (González-Abraham et al., 2015; Hirsh-Pearson et al., 2022; Martinuzzi et al., 2021; Theobald, 2013; Woolmer et al., 2008). Such datasets are subject to national quality standards and aligned with official land-use reporting frameworks, thereby improving policy relevance (Martinuzzi et al., 2021; Scott and
100 Rajabifard, 2017).

This study aims to produce two complementary national datasets representing cumulative industrial pressures and intactness in Australia around 2022 (based on the median year of the input datasets): the Australian Human Industrial Footprint (HIF) and the Ecological Intactness Index (EII). The HIF is a cumulative pressure map at 100 m spatial resolution that integrates 16

105 nationally significant pressures, while the EII translates these pressures into an ecologically interpretable measure of structural
intactness, following the metric proposed by Beyer and colleagues (2020). Together, the HIF and the EII offer critical tools
for guiding conservation and restoration efforts, aligning with Australia's commitments under the Global Biodiversity
Framework, including those targeting highly intact ecosystems (Target 1), where to undertake restoration (T2) where the
important areas to protect are (T3), and government objectives and policies such as the Threatened Species Action Plan
(Commonwealth of Australia, 2022).

110 **2 Methods**

2.1 Overview of the Human Industrial Index Mapping Method

We adapted the Human Footprint Index methodological approach (Sanderson et al., 2002) to create a cumulative pressure map
for Australia, incorporating best practices from studies that have refined this method globally and regionally over the past two
decades (e.g.; Arias-Patino et al., 2024; Gassert et al., 2023; Hirsh-Pearson et al., 2022; Watson et al., 2023a; Woolmer et al.,
115 2008). Each mapped pressure aligns with one or more IUCN-CMP threat classes (Salafsky et al., 2025, see Table S1), ensuring
conceptual consistency with global biodiversity reporting frameworks (e.g., GBF). This alignment helps guide the selection of
pressures to include, clarifies relationships between them, and informs the choice of appropriate datasets and schemes.

We selected pressures based on (i) national data availability at suitable spatial and thematic resolution, (ii) relevance to natural
systems as identified in previous footprint frameworks. Consistent with earlier human footprint assessments, we focused on
120 observable industrial pressures, those directly resulting from human activities such as infrastructure, extraction, and land use.
Pressures such as pollution, invasive species, climate change, and fire were excluded because suitable national spatial data
were unavailable. Climate change and changed fire regimes present additional challenges in distinguishing natural from
human-induced events (Bowman et al., 2020; Theobald et al., 2025). Nevertheless, we acknowledge that changes in fire
regimes increasingly threaten Australian biodiversity (Doherty et al., 2024; Ward et al., 2020). The framework we use remains
125 flexible, allowing future integration of new pressures (such as changed fire regimes) as suitable datasets become available.

We identified 16 human pressures relevant to Australia with available spatial data (Table 1): 1) intensive land uses, 2)
buildings, 3) mining and quarrying, 4) human population density, 5) croplands, 6) pasturelands, 7) forestry plantations, 8)
reservoirs and dams, 9) farm dams, 10) roads, 11) railways, 12) hiking trails, 13) oil pipelines, 14) gas pipelines, 15) energy
transmission lines, and 16) navigable waterways. We assigned a score between 0 and 10 to each pressure, with each pressure's
130 score relative to other pressures (Fig. 1 and Table S2 in the Supplement). For all the pressures, scores were assigned according
to their "direct" disturbance on the area they overlap. For pressures 8-16, we also assigned a score to adjacent areas to reflect
indirect disturbances, such as edge effects, habitat fragmentation, and more cryptic forms of disturbance, such as potential
access for humans or invasive species to areas previously inaccessible. We largely followed past human footprint studies to
assign pressure scores to ensure comparability with these (Hirsh-Pearson et al., 2022; Sanderson et al., 2002; Venter et al.,
135 2016b; Woolmer et al., 2008). Several pressure scores were refined through discussions within our author team, which includes

researchers with extensive experience in Australian ecosystems and land-use and pressure mapping, to better reflect national conditions and data characteristics. While this approach provided context-specific refinements, future national applications could further strengthen the scoring scheme through structured decision-science methods (e.g., via Delphi methods).

140 Following previous human footprint maps, we defined intensive land uses, mining, cropland, and pasturelands as mutually exclusive pressures, as each represents a complete replacement of natural land cover, unlikely to co-occur with the others within the 100 m pixels used for mapping. Moreover, reservoirs and farm dams are also mutually exclusive, as are paved, unpaved roads, and hiking trails. All other pressure overlaps were allowed to overlap, mainly because they represent additional disturbances that can co-occur spatially in the 100 m pixels used in our analyses (e.g., roads, pipelines, or population density). To avoid spatial overlap among incompatible land uses, we applied a hierarchical rule in which overlaps were resolved
145 by retaining the pressure with the highest intensity score (0–10), for example: built environments \geq mining > cropland > pastureland. This ensured that only one of these mutually exclusive land-use pressures contributed to each pixel’s cumulative score, while allowing other co-occurring pressures to be summed cumulatively.

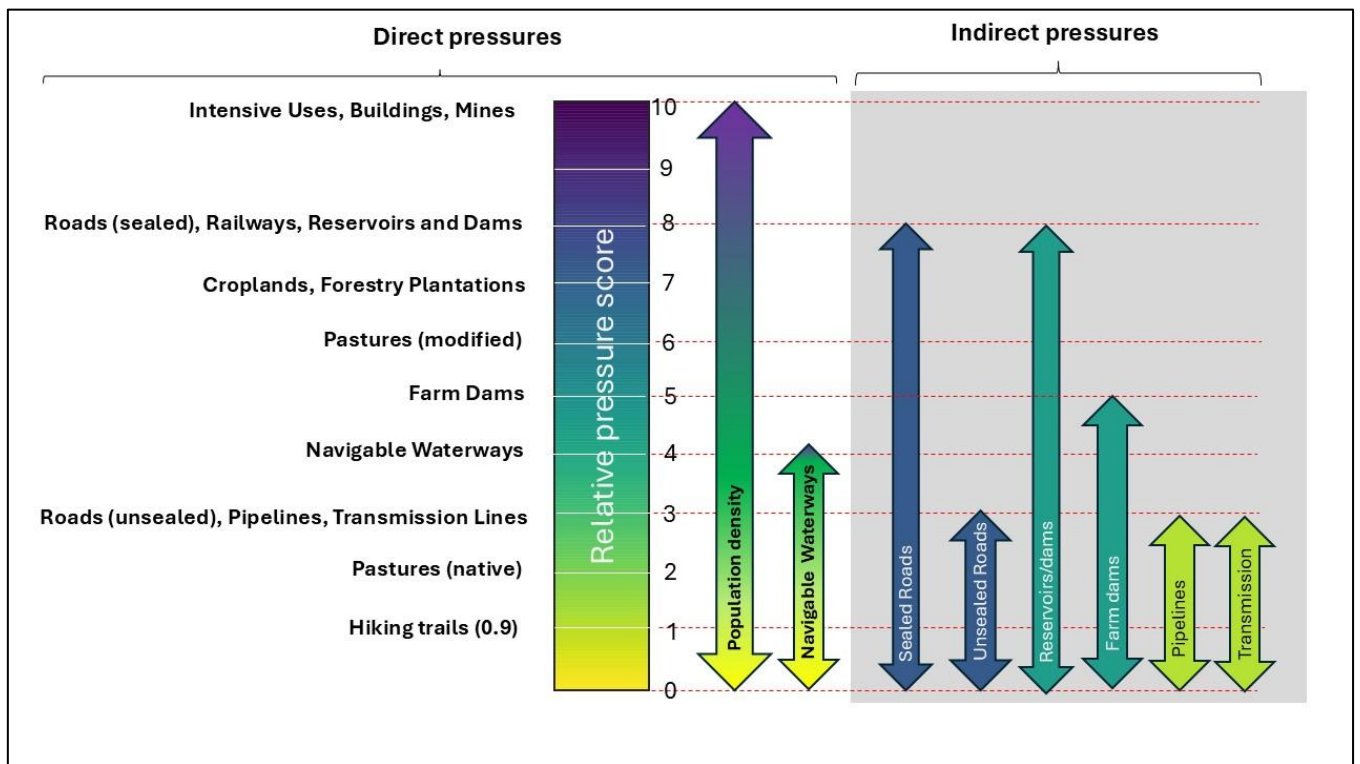
After creating the weighted pressure layers, we summed them up to generate the terrestrial Industrial Footprint map. The analysis was conducted at a 100 m spatial resolution using the Australian Albers Equal Area projection (EPSG:3577). This
150 resolution represented a balance between the 50 m resolution of the primary land use data and the overall accuracy of the dataset. We used a binary rule to rasterise (using `gdal_rasterize`, with the `-at` option) polygon and line features to 100 m raster cells, where any pixel intersecting a feature was assigned the corresponding pressure value. All individual layers were processed using GRASS GIS (GRASS Development Team, 2024) or Google Earth Engine (Gorelick et al., 2017). The uncertainty analysis was conducted in Python 3.6 (Van Rossum and Drake Jr, 1995), and graphics were developed using the
155 R Package `ggplot` (Wickham, 2016).

Table 1 The pressures included in mapping the Australian Industrial Footprint, and the specific data layers utilised in the mapping process.

Data layer	Date	Scale/ Resolution/ Positional accuracy	Data Source
Intensive land uses	2017**	50 m *	Catchment Scale Land Use of Australia – Update December 2023 version 2 (ABARES 2024)
Buildings	2018	Not specified	Australian Building Footprints (Microsoft 2022)
Mining/ Quarrying	2023	50 m *	Catchment Scale Land Use of Australia – Update December 2023 version 2 (ABARES 2024), updated with data distributed by states and territories (Table S3)
Population density	2023	100 m	WorldPop High-Resolution Dataset: Australia, 2024. (WorldPop 2024)
Croplands	2017**	50 m *	Catchment Scale Land Use of Australia – Update December 2023 version 2 (ABARES 2024)
Pasturelands	2017**	50 m *	Catchment Scale Land Use of Australia – Update December 2023 version 2 (ABARES 2024)
Forestry (plantations)	2023	50 m *	Australia Forests 2023 (ABARES 2023), complemented with Catchment Scale Land Use of Australia – Update December 2023 v. 2 (ABARES 2024)
Reservoirs/ dams	2025	50 m *	National Hydrological Database of Geoscience Australia (Cresswell et al. 2015), created in 2015 and updated in 2025.
Farm dams	2021	nd	Malerba et al. (2021), and references therein.
Roads	2024	± 2 m urban areas. ± 10 m in rural areas	National Roads (Geoscape Australia, 2024) and Open Street Maps (OpenStreetMap contributors, 2024)
Railways	2021	Accuracy ± 20m	Foundation Rail Infrastructure (Geoscience Australia 2021)
Hiking Trails	2024	Not specified	National Roads (Geoscape Australia 2024), and Open Street Maps (2024)
Oil Pipelines	2022	Not specified	Oil Pipelines of Australia (Geoscience Australia 2022)
Gas Pipelines	2022	Not specified	Gas Pipelines of Australia (Geoscience Australia 2022)
Transmission lines	2021	Not specified.	Electricity Transmission Lines (Geoscience Australia 2021) Digitized using satellite images with a resolution of 0.15- 2.5 m
Navigable waterways	2016	300 m	Generated for this study following the methods in Venter et al. (2016)

160 **Median year of all pixels mapped within the pressure layer, as the date of mapping varies between 2008 and 2023.

*A seamless raster was created by combining land use vector data showing a single dominant land use for each location based on the management objective of the land manager. The scale of mapping varies between 1:50,000 and 1:250,000.



165 **Figure 1** Direct and indirect scores assigned to each of the 16 pressures used to estimate the Australian Industrial Footprint. We set
 indirect effects to extend 5 km from roads, 500 m from reservoirs and dams, and 2.75 km from pipelines and transmission lines.

2.2 Mapped pressures

In this subsection, we outline the rationale for including each of the 16 pressures used in this analysis, the scores assigned to them, and a short description of the data used to represent each pressure. We took advantage of the 2023 update of the
 170 Catchment Scale Land Use of Australia (ABARES, 2024), henceforth called CLUMP 2023, to represent several pressures. CLUMP 2023 has a high level of thematic detail (its tertiary classification has 189 classes), uses a standard national classification (Australian Land Use and Management, ALUM Version 8) developed under the Australian Collaborative Land Use and Management Program (ACLUMP), and is updated regularly (e.g., previous updates include 2015, 2017, 2018, and 2020). Although distributed as a 50 m raster, CLUMP 23 is derived from detailed vector datasets produced by each State and
 175 Territory government, through a bottom-up mapping process that combines aerial and satellite images, cadastral and tenure information, zoning data, expert input, and field validation. The vector datasets span various dates (2008–2023) and mapping scales (1:5,000 to 1:250,000) (Fig S2 in the Supplement), reflecting differences in land-use intensity and regional update cycles. The national raster was generated by the Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES) in ArcGIS, so that each 50 m cell represents a single dominant land use based on the management objective of the
 180 land manager. We resampled this dataset to 100 m resolution using a majority filter.

While CLUMP 2023 integrates data of varying scale, age, and jurisdictional origin, it provides a robust, nationally standardised and accepted representation of land use with strong alignment to official State and Commonwealth reporting. Older and coarser data correspond with arid and semi-arid regions where land use changes are less frequent. Therefore, we supplemented these data with more recent, higher-resolution datasets to improve accuracy and currency in these regions. For example, South
185 Australia's arid zones, which rely on older, coarser mapping, are dominated by large expanses of native pasturelands, and changes in pressures can be captured using current and finer data on buildings, farm dams, mines, roads, and other linear infrastructure. A notable modification from previous applications of the HIF methodology is the exclusion of nightlight data as a proxy for infrastructure in rural areas or working landscapes like mine sites (Venter et al., 2016b). We opted not to use nightlights because equivalent or higher-quality national datasets already map these features directly (e.g., CLUM 2023,
190 Geoscience Australia's pipelines and building footprint layers).

Although some of the input datasets differ slightly in their reference years, most represent conditions between 2017 and 2023 and together provide an up-to-date national view of cumulative industrial pressures. These small temporal differences reflect the different update cycles of national and state agencies, but because most pressures (such as infrastructure, land use, and mining) are long-lasting features, they are unlikely to meaningfully affect the overall spatial patterns of pressure across
195 Australia.

2.2.1 Intensive land uses

This category includes pressures from land uses typically linked with infrastructure and human settlements, such as residential, industrial, intensive horticulture and animal production (e.g., glasshouses, piggeries), and the infrastructure supporting services and utilities. Lands affected by these pressures are often heavily modified and constructed, making it unlikely they will revert
200 to a natural state. These areas experience significant disruption of natural processes, leading to habitat loss and the exclusion of wildlife and ecosystem services (Venter et al. 2016). Therefore, we assigned intensive land uses a score of 10. These pressures were mapped using the CLUMP 2023 dataset, which aggregates land use data from 47 tertiary classes (Table S4) into this category. This category broadly aligns with the "Built-up" pressure identified in other human industrial footprint analyses (Sanderson et al. 2002; Venter et al. 2016; Hirsh-Pearson et al. 2022).

2.2.2 Buildings

Buildings remove natural habitat under the footprint of the construction site and are often associated with habitat clearing in areas surrounding the buildings. Here, we assigned a score of 10 to any pixel overlapping a building, which assumes some modification around the building in pixels not fully occupied by its footprint. This pressure was used as a proxy for human settlement and industrial activities outside urban areas that are potentially not captured through coarser mapping by CLUMP
210 2023. Building data was obtained from Microsoft (2018), which reports a false positive ratio of ~1% in 1000 randomly sampled buildings. We conducted our own visual inspection of 1000 buildings and observed false positives, especially in arid regions, where big rocks were misidentified as buildings. To reduce these errors, we limited our analysis to buildings located within

200 meters of roads or mining areas, as these are typically associated with built structures. We acknowledge that this filtering may produce a slight underestimation of pressures in remote areas where small buildings exist, but roads are unmapped.

215 **2.2.3 Mining and quarrying**

Multiple activities associated with mining, from exploration to post-closure, will negatively affect biodiversity and ecosystem services (Boldy et al., 2021; Sonter et al., 2018). Mining reshapes the landscape, alters waterways and wetlands, increases erosion, and causes pollution from noise, dust, and emissions (Haddaway et al. 2019, and references therein). Due to these multiple environmental impacts, we assigned a score of 10 to the direct pressures from mining. To create the mining pressure
220 layer, we updated the mining land use class from the CLUMP 2023 dataset with state-level datasets (see Table S3 for data sources). Note that while most mining data occupies the affected area, some polygons represented the mining tenement, which, while a limitation, captures activities such as human habitation and road construction.

2.2.4 Human population density

Environmental degradation in a particular area is often associated with proximity to human populations due to activities such
225 as recreation, hunting, logging, and the introduction of non-native species. Following Venter et al. (2016), we converted a human population density layer into a pressure layer with scores between 0 and 10. Locations with more than 1,000 people per km² were assigned a score of 10, assuming population density reaches saturation at this level. For areas with densities below 1,000 people/km², we scaled the pressure score logarithmically using the formula: Pressure Score = 3.333 * log (population density +1). We used the WorldPop dataset (WorldPop, 2018), which provides population density estimates at a
230 100 m resolution for its most recent update (2020).

2.2.5 Croplands

Croplands are often completely converted ecosystems and are subject to high levels of pesticide and fertilizer use and destructive slash-and-burn techniques, and as a consequence, have become the main driver of biodiversity decline and the degradation of the natural landscape (Green et al., 2005; Maxwell et al., 2016). Following Venter and colleagues (2016), we
235 assigned croplands a pressure score of 7, as some native species can still utilize croplands (Grass et al., 2019), unlike in most built environments. We obtained cropland data from the CLUMP 2023 dataset using 56 tertiary classes associated with these activities (Table S5). We acknowledge that different crops might exert different pressure intensity on the environment, and while our study maintains consistency with what has been done in previous HIF studies, future work could explore modifying cropland pressures based on intensification levels and biochemical conditions.

240 **2.2.6 Pasturelands**

Grazing impacts ecosystems through the creation of fence networks, soil compaction, trampling, and intensive browsing of native vegetation, the spread of invasive species, and altered fire regimes (Kauffman and Krueger, 1984). Domestic herbivores

also have multi-trophic effects on plant and animal biodiversity, contributing to biodiversity loss (Filazzola et al. 2020). In this study, we diverged from previous human-industrial footprint analyses (Venter et al., 2016b), which assigned a uniform score of 4 to pasturelands. Instead, we made a distinction between modified and native pasturelands under production, a classification provided in the CLUMP 2023 dataset.

Modified pasturelands, characterized by 50% or more dominant exotic species and irrigation practices (ABARES, 2016), were assigned a score of 6 due to significant vegetation modification and frequent livestock grazing (see Table S6 for CLUMP 2023 tertiary classes). Native pasturelands, which have undergone minimal or no deliberate modification, were assigned a score of 2. This lower score was assigned to native pasturelands to be conservative, as there might be a great similarity between pasturelands in arid zones not often grazed and areas not classified as grazing lands. Data on livestock stocking density and grazing intensity are largely unavailable, making further differentiation within this land use impossible. However, these areas are still associated with varying levels of fencing, soil compaction, and browsing by farmed animals, unpaved roads, and altered fire regimes, which have an associated impact on their native ecological communities (Tulloch et al., 2023).

Spatially explicit national data (e.g., from the Australian Bureau of Statistics) are lacking, and livestock distribution in large arid regions are highly variable through time. Global datasets such as the Gridded Livestock of the World (GLW4; FAO, 2022) and Annual Global Gridded Livestock Mapping, 1961–2021 (Du et al., 2025) were not included due to their coarse resolution (5-10 km), their somewhat outdated baseline, and inability to capture local or seasonal livestock movements. We therefore relied on CLUM 23 based to represent grazing pressure, and we acknowledge this limitation in the Discussion.

2.2.7 Forestry plantations

Australia's plantation forests covered 1.96 million hectares in 2016, mainly comprising exotic pines (softwood) and *Eucalyptus* (hardwood) (ABARES, 2018). Plantation forests remove habitat for species, including tree cavities, and can alter paths of travel and fire regimes (Bradstock et al., 2002; Brockerhoff et al., 2008). Given that these plantations are (typically) monocultures, we assigned a pressure score of 7, akin to croplands. The forestry pressure layer was created by merging the CLUMP 2023 dataset, using only the plantation forest classification, with plantation forests from the Australia Forests 2023 dataset (ABARES, 2023). These layers were merged as we argue they complement each other. The CLUMP 23 dataset does not include some plantations observed in the Australia Forest 2023, while the Australia Forest 2023 dataset does not include plantations that have been recently clear-cut and are presently bare land, but that will most likely be replanted. We did not account for pressures from forestry undertaken in native forests, as spatially explicit records of activities in these areas are not consistently mapped across the continent.

2.2.8 Reservoirs and dams

Dams and reservoirs inundate the land, altering their hydrology and often converting terrestrial ecosystems into aquatic ones, causing habitat loss for many terrestrial and freshwater species, as well as altering local ecosystems (Barnett and Adams, 2020; Poff and Hart, 2002). Dams can also disrupt sediment transportation and fish migration, change water quality, and increase the

275 risk of invasive species (Bunn and Arthington, 2002; Johnson et al., 2008; Liermann et al., 2012; Syvitski et al., 2005). Given
this, we assigned reservoirs (including the dam wall area where present) a pressure score of 8. While we focused on the direct
pressure from the inundated footprint, we also established a 500 m buffer, where the score is assigned using an exponential
decaying function. Furthermore, while dams also exert pressures downstream, we do not consider this in the current analysis,
as data for Australia is unavailable. Data for these pressures were obtained from the National Hydrological Database of
280 Geoscience Australia (Cresswell et al., 2015), which was created in 2015 and updated in 2025.

2.2.9 Farm dams

Farm dams (sometimes called ‘agricultural ponds’) in the Australian agricultural landscape are ubiquitous; Malerba and
colleagues (2021) found over 1.765 million dams across the country, covering an area of 4,678 km² and storing more than 20
times the amount of water in Sydney Harbour. These human-made features catch and store water for livestock, irrigation, crop
285 spraying, firefighting, and other domestic purposes. But, while often small in scale, farm dams can significantly affect
biodiversity and biogeochemical cycles (Liddicoat et al., 2022; Woolmer et al., 2008). They directly modify the environment,
accumulate pollution from run-off, and can produce greenhouse gases.

We assigned a score of 5 to farm dams, which is extended to 500 m from the dam to account for changes to environmental
processes. The buffer assumes a conservative distance of concentrated grazing, trampling, and degradation associated with
290 access to water points by livestock which can extend up to 3 km (Cowley et al., 2015; Materne et al., 2025; Washington-Allen
et al., 2004); 65% of farm dams overlapped pasturelands. It also assumes potential spillage during intense rainfall which creates
tides of water and, in some cases, flooding when multiple dams are present (Kazarovsky, 1996; Pisaniello and Tingey-Holyoak,
2017). We do not consider downstream pressures due to data availability, and because many farm dams are constructed off-
stream (Section 2.2.8). We obtained farm dam data from Malerba and colleagues (2021), who compiled it from different State
295 sources.

2.2.10 Roads

Roads exert numerous direct and indirect pressures on terrestrial and aquatic ecosystems, including habitat loss, fragmentation,
mortality from construction, roadkill, animal behavior change, alteration of the physical and chemical environment, spread of
invasive species, and increased use of areas by humans (Rytwinski and Fahrig, 2015; Trombulak and Frissell, 2000). This
300 linear infrastructure has direct and indirect pressures on the environment, which are accounted for in the pressure scoring. We
adapted the scoring systems for roads from Venter and colleagues (2016) to assign these direct and indirect pressure scores
and differentiate between sealed and unsealed roads, as we recognize that many roads in regional areas are rarely used. A score
of 8 was assigned to a 0.3 km buffer from sealed roads and a decreasing pressure from 7.9 to 0.25 outward up to 5 km from
the road. Unsealed roads were assigned a score of 3, including a 0.3 km buffer from the road and a decreasing pressure from
305 2.9 to 0.25 outward up to 5 km from the road. The 5 km maximum distance follows Arias-Patiño et al. (2025), based on
observed road impacts on mammals; for instance, in Australia McCall et al. (2010) reported that sugar glider survival was

~70% lower within 5 km of major roads. This distance also serves as a proxy for accessibility and is consistent with recent human-footprint mapping efforts (Hirsh-Pearson et al., 2022), while remaining more conservative than early global maps (Venter et al., 2016b; Williams et al., 2020). However, we acknowledge that this is a major simplification, since off-road accessibility varies with topography, land tenure, land cover, and hydrology. Future pressure mapping in Australia could explore the use of friction-surface and travel-time models (Nelson et al., 2019; Weiss et al., 2020).

To create the roads layer, we merged the National Roads dataset (Geoscape Australia, 2024) and the Open Street Map (OpenStreetMap contributors, 2024) data. Integrating two road datasets introduces the possibility of ‘double-counting’ when the same road is represented in both datasets, due to the features not spatially aligning exactly within the two datasets; however, because we are rastering linear features at a 100m pixel size, small errors in alignment are most likely removed. Larger spatial alignment errors of the same feature, where found, may be representative of the real-world nature of unsealed roads and tracks in rural and outback areas in Australia, where road users may seasonally take different paths due to high water levels or other factors. Full SQL attribute and data tag queries used for each road type (sealed, unsealed, track, patch) and for each input dataset, can be found in Table S7-S8.

320 **2.2.11 Railways**

Like roads, railways are linear infrastructures that directly remove habitat, resulting also in fragmentation that can produce edge effects (Fuentes-Montemayor et al., 2009). However, railways are less conducive to providing access to natural environments than roads, given that passengers would usually only disembark at rail stations. The direct pressure of railways was assigned a pressure score of 8, with a 50 m buffer on either side of the railway. We used data from the Foundation Railway of Australia (Geoscience Australia, 2021), which includes open, closed, and other tracks. We removed features with a dismantled, proposed, or disused status from this dataset.

2.2.12 Hiking trails

We also included the disturbance from trails, as these are often pathways for human access-related pressures (hunting, invasive weeds, etc.) into remote and protected areas. We assigned a low direct pressure score of 0.9 to trails to acknowledge their presence while remaining conservative in estimating human pressures in remote areas. This conservative weighting reflects both the variable quality and completeness of trail data across jurisdictions and expert input from wilderness mapping practitioners, who note that limited or infrequently used trails may not substantially compromise wilderness character. To create the trails layer, we merged the National Roads dataset (Geoscape Australia, 2024) and the Open Street Map (OpenStreetMap contributors, 2024) data, using the query described in Table S7-S8.

335

2.2.13 Transmission lines, and oil and gas pipelines

Pipelines and transmission corridors represent important linear infrastructure that exerts multiple pressures on natural systems. Their construction and maintenance lead to habitat loss and fragmentation of natural habitats and can facilitate the spread of invasive species (Benítez-López et al., 2010). Moreover, pipeline leaks and spills can pollute the soil and water and contribute to greenhouse gas emissions (Brandt et al., 2014). Transmission lines can affect the mortality of flying animals (Bevanger, 1998) and increase the risk of wildfires (Keeley and Syphard, 2018). We believe these linear features represent pressures in a similar way to unsealed roads (and often have unsealed service roads alongside them). Therefore, we set a direct pressure score of 3 with a buffer of 300 m around this type of infrastructure and an indirect score decaying from 3 to 0.25 outwards to 2.5 km from the infrastructure (Arias-Patino et al., 2024).

2.2.14 Navigable waterways

Navigable waterways - in the form of navigable rivers, lakes, and marine coastlines - facilitate human accessibility to the natural environment in a way analogous to roads. We created the navigable pressure layer by applying the methods described in Venter et al. (2016), on a 100 m resolution. Areas directly alongside navigable waterways have a pressure of 4, which decreased exponentially outwards 15 km (Venter et al. 2016).

2.3 Technical validation and uncertainty (sensitivity) analysis of the Human Industrial Footprint

2.3.1 Validation and accuracy assessment

We followed the methods outlined in Arias et al. (2024) and Venter et al. (2016) to evaluate the agreement between the HIF map and the pressures observed in situ. To this end, a single assessor visually identified industrial pressures observed through very high-resolution satellite images (<1 m, typically 0.3–0.5 m in urban and coastal areas) within 1,397 randomly stratified sample plots. The satellite images were obtained from web map services such as Google Maps (Google, 2024.), Bing Maps (Microsoft, 2024.), and Basemap ArcGIS (ESRI, 2024.), and they all corresponded to the years 2020-2023.

We defined five strata corresponding to pressure classes (Table 2), allocating a number of samples to each, following Olofson et al. (2013, 2014), distributing them according to the area and expected error for each stratum. This strategy aims to prevent oversampling large strata like low-pressure areas and minimize the standard error for small regions like high-pressure areas, which could result in an overestimation of the accuracy. Because the stratified random sampling was based on HIF value ranges, the validation sample is not entirely independent of the resulting map. However, this approach ensured that the full gradient of cumulative pressures was represented, particularly in low- and high-pressure areas that occupy small proportions of Australia's land area. Each sample location was selected randomly within strata, and validation relied on independent, very high-resolution satellite imagery, providing an objective assessment of accuracy.

The sample distribution was as follows: no pressure= 477, very low pressure = 30, low pressure = 565, moderate pressure= 205, high pressure = 120. Each sample plot consisted of a 100 m window (matching the analysis' spatial resolution) and five

surrounding buffers at 300, 500, 1000, 2750, and 5000 m to aid in recording both direct and indirect pressures (Arias et al. 2024). Scores were assigned as per Table S9. Indirect pressures were recorded based on the nearest observed feature and its area of influence, with scores assigned using the mean value of the two closest buffers. The sum of all observed pressure scores represented each plot's assumed actual state of in-situ pressures, while corresponding HIF values were extracted from the map for comparison.

Both HIF and validation scores were normalized to a 0-1 scale using the min-max normalization formula. We used the maximum value observed in our map rather than the theoretical maximum, as there is no location where all pressure factors overlap simultaneously, following Venter et al. (2016). To quantify the level of agreement between the HIF and validation scores, we utilized the root mean square error (RMSE) (Chai and Draxler, 2014). The RMSE expresses the average error in the units of the variable of interest, tending to penalize large errors; a lower RMSE indicates higher agreement between the HIF and the validation scores. We also calculated the percentage of validation samples with agreement between the HIF and validation scores, considering the HIF to match the validation score if they were within 20% of each other on the 0-1 scale (Venter et al., 2016b). This $\pm 20\%$ tolerance provides a complementary measure of continuous-scale agreement, offering an intuitive indication of how close predicted and observed values are before categorization. It does not replace the categorical accuracy metrics outlined below, but adds context to the RMSE and R^2 results by highlighting overall precision and bias trends. Finally, we also conducted an accuracy assessment of the classified five-level HIF map to provide a complementary perspective focused on the reliability of the pressure classes used for spatial interpretation and conservation planning. Using the same validation samples, we estimated the user's, producer's, and overall accuracy metrics using the proportion of area, as implemented in the 'mapaccuracy' R package (Costa 2024). Agreement was determined by whether each sample plot fell within the same pressure class in both the HIF and the visual interpretation. The accuracy assessment also allowed us to produce error-adjusted estimates of the area for each class (Olofsson et al., 2013), supporting future applications that rely on categorical representations of the HIF. This accuracy was done, as numerous research studies have applied thresholds to cumulative pressure maps to categorize the level of human influence and inform conservation interventions at different scales. For example, it has been used to identify the last of the wild (Sanderson et al., 2002), the most globally intact areas (Watson et al., 2016; Williams et al., 2020), wilderness areas and vegetation condition assessments in Australia (Lesslie et al., 1988; Lesslie and Taylor, 1985; Thackway and Lesslie, 2008), and to assess the extinction risk to species (Di Marco et al., 2018).

2.3.2 Uncertainty analysis.

To understand the degree of uncertainty in our results, associated with the scores assigned to the different pressures, we followed Arias et al. (2024) and randomly adjusted intensity scores in the validation samples, by up to $\pm 50\%$ using the bootstrap technique. We chose to adjust pressure scores by up to $\pm 50\%$, a wider range than the $\pm 30\%$ used by Arias-Patiño et al. (2024), in order to test the robustness of our cumulative pressure scores under a wider range of values. This approach allowed us to evaluate whether the model remains stable even when pressure intensity values are varied well beyond the expected range of expert-derived variability. Each simulation involved selecting a random factor between 0.1% and 50%, which was then applied

400 to each pressure layer. Specifically, we multiplied this factor by the original pressure intensity (PI) value for each layer and randomly added or subtracted the result from the validation sample. We adjusted the pressure intensity (PI) by layer (s) as follows.

$$\text{Modified } PI_s = PI_s \pm (PI_s \times n); \quad n = \text{random number between } 0.01 - 0.5 \quad (1)$$

405

Using the modified scores, we calculated the simulated cumulative pressure value for each validation plot containing mapped values and then assessed the error by comparing it to the original mapped values. This simulation was repeated 100,000 times to ensure statistical robustness. Finally, we generated an uncertainty map by interpolating the standard deviation of the error using the inverse distance weighting (IDW) technique.

410 **2.3.3 Comparison with Global Human Footprint datasets**

To assess the added value of the fine-scale national Human Industrial Footprint (HIF), we carried out a visual comparison with global Human Footprint datasets available at 1 km for 2013 (Williams et al., 2020) and at 100 m resolution for 2020 (Gassert et al., 2023). These comparisons were used to qualitatively evaluate how well the HIF captures the spatial patterns of cumulative pressures relative to global assessments. Because the datasets differ in resolution, input data, set of mapped
415 pressures, and some assumptions, interpreting a direct quantitative comparison is of limited value.

Table 2. The HIF was classified into five descriptive categories representing diverse levels of pressure.

Category	Corresponding HIF score	Description
No pressure	< 1	Areas free or almost free of measurable or indirect pressures from linear infrastructures and human population.
Very low pressure	≥ 1 and < 2	Areas where the score arises primarily from indirect pressures from linear infrastructure or areas with a human population density of at least 1 person/km ² .
Low pressure	≥ 2 and < 6	Areas with scores primarily due to grazing in native pastures, isolated linear infrastructure with lower pressure scores (pipelines, electricity transmission lines, unpaved roads).
Moderate pressure	≥ 6 and < 10	Areas representing modified agricultural landscapes and a relatively lower level of cumulative pressures (e.g., grazing in native pastures, unpaved roads, and some level of human density).
High pressure	≥ 10	Areas highly modified by humans, such as mines, urban areas, and other settlements, and a higher level of cumulative pressures (e.g., overlap of intensive farming and roads).

420 2.4 The ecological intactness index

When pressure maps such as the HIF are used as proxies of ecological condition or intactness, it is done based on thresholds applied to individual pixels with a certain pressure value (Watson et al., 2016; Williams et al., 2020). However, while the HIF incorporates some indirect pressures that spread out to a buffer from the direct pressure, the value of each pixel does not explicitly account for the spatial configuration and habitat-quality context surrounding that pixel. For example, a narrow strip of native vegetation between agricultural fields could appear to have no pressure because there is no indirect pressure for cropland in the HIF; however, such a strip is impacted by significant edge effects and unmapped human presence, indicating it is somewhat degraded and not intact as the HIF would indicate. Here, we overcome this by calculating an intactness metric (Beyer et al., 2020) for Australia, which is sensitive to changes in habitat area, quality, and fragmentation (and therefore captures the structural component of ecological integrity (Nicholson et al., 2021), which is well known to influence the ability of an area to support a diversity of species (Betts et al., 2017; Fischer and Lindenmayer, 2007; Hanski et al., 2013).

For each 100 m cell, intactness was estimated as a function of the spatial configuration and quality of surrounding habitat, with contributions from neighbouring cells declining with distance. The metric is parameterized to decrease monotonically with increasing fragmentation, reflecting both the number and separation of habitat patches, and to scale with total habitat area and quality. The HIF served as the base layer, normalized to a 0–1 scale and inverted so that higher values represent greater habitat quality (i.e., lower pressure). Intactness values were computed within a circular moving window of 5 km radius using the

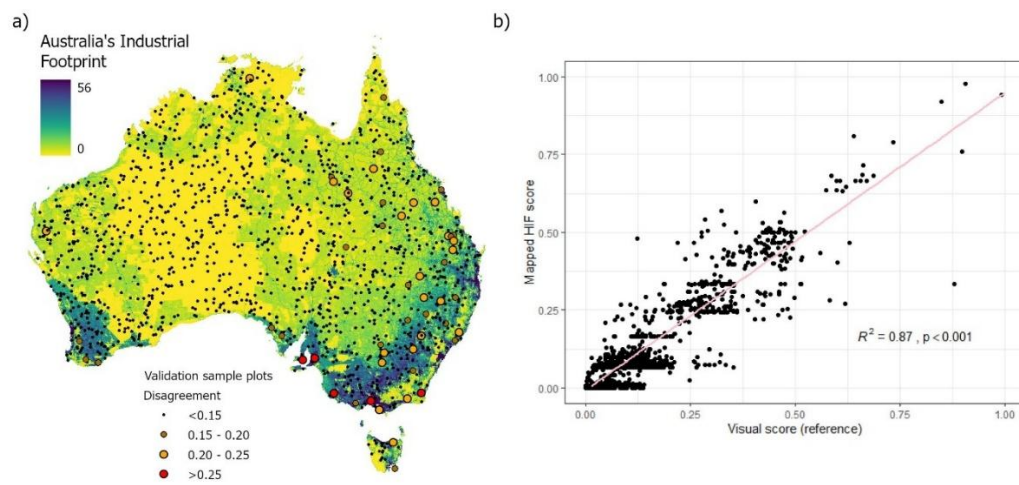
kernel function described by Beyer et al. (2020), which integrates both patch size and isolation effects. The resulting EII represents the relative degree to which each location retains the spatial configuration and quality characteristic of structurally intact ecosystems.

3. Results

440 3.1 Human Industrial Footprint map and pressures overview

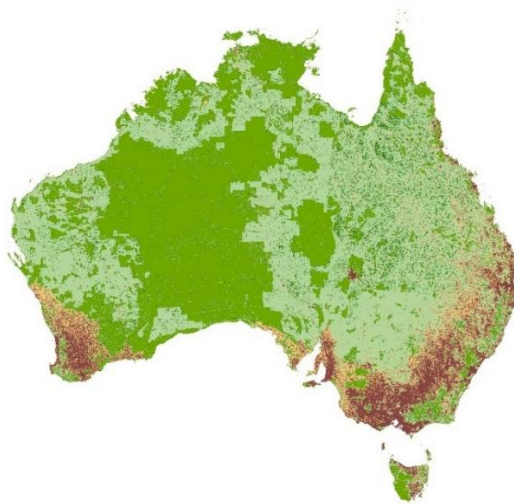
The Australian Human Industrial Footprint Index (Fig 2a) map covers an area of 7,692,047 km² and has a spatial resolution of 100 m. The scores range between 0 (areas with no mapped pressures) and 56.5 (densely populated built-up urban regions), with a mean score of 3.05 ± 4.18 , and a median of 2.25. The HIF scores are highly skewed to low values, as seen in the classified map (Fig 3, Fig S3), which shows that more than one-third of the Australian landscape (32%) is free or almost free (score <1) of the 16 pressures included in this analysis, and another 2.9% experiences very low pressures (i.e., scores of <2) (Fig 3). Another 47.5% of the Australian landscape has a low industrial pressure footprint (HIF value ≥ 2 and < 6). These low-pressure areas are primarily pastoral leases that operate without extensive introduction of non-native pastures. However, this analysis does not account for stocking intensity, and we acknowledge that the pressure in some of these areas might be underestimated. Finally, 14.2% of the Australian landscape presents more considerable industrial pressures (scores ≥ 6), with 450 5.6% of the land being under moderate pressure (scores between 6 and 10) and a further 8.5% experiencing high industrial pressure (scores ≥ 10).

The most prevalent pressures in Australia, in terms of areal coverage, are roads and grazing in native pastures, which exert some level of pressure across approximately 66% and 46% of the nation, respectively (Table 3).



455 **Figure 2. a) Australia's human industrial footprint map on land, showing the location of the validation sample plots. The larger points are those plots where the HIF and the visual score differed by more than 20% on a normalized (0-1) scale. b) Relationship between the reference score (visual score) and the score obtained through the HIF for Australia.**

Legend	Pressure class	Mapped area Km ²	CI 95% (Km ²)
	No pressure	2,429,491	93,862
	Very Low	223,451	66,900
	Low	3,903,901	102,525
	Moderate	255,383	48,853
	High	879,820	57,266



460 **Figure 3. Australian Human Industrial Footprint map categorized into five industrial pressure classes, from no pressure to very high pressure (see Table 2). The table shows the error-adjusted area, and 95% confidence intervals estimated for each class.**

Table 3. The percentage of the mapped area that each pressure covers across the Australian landscape.

Pressure	Land (%) under each form of pressure
Intensive land uses	0.4
Buildings	0.4
Mining	0.1
Population Density ¹	1.2
Croplands	5.7
Forestry (Plantations)	0.3
Modified pastures	2.8
Native pastures	47.0
Roads ²	66.1
Railways	0.1
Farm dams	7.0
Reservoirs/Dams	0.2
Pipelines (oil and gas)	2.2
Transmission lines	1.2
Navigable waterways	8.6
Hiking Trails	0.1

465 ¹A population density score of 0.1 (out of 10), corresponding to a population density of ~0.1 persons/km² was used as a threshold to define the presence of this pressure in a pixel. If the threshold is not applied, scores as low as 0.001 will be quantified, and the total area (%) of land affected by this pressure will increase to 95%.

²Paved and unpaved roads were aggregated to calculate this statistic.

3.2 Technical validation

The technical validation results indicate a strong agreement between the Human Industrial Footprint scores and those obtained through visual interpretation. A strong relationship ($R^2= 0.87$) exists between the human industrial footprint and the validation scores (Fig 2b). The RMSE for the 1397 validation plots was 0.059 on the normalized 0-1 scale, indicating an average deviation of approximately 6%. In addition to these continuous measures, 98% of validation plots showed agreement within $\pm 20\%$ between the HIF and the observed scores, providing an intuitive indication of how closely the mapped values match on-the-ground conditions. Only 27 plots fell outside this range (Fig 2a), five higher and 22 lower than the observed score, suggesting the HIF tends to underestimate pressure. Even when we consider a stricter threshold of 15% for agreement, we still obtained a 96.2% match between the HIF and the visual scores.

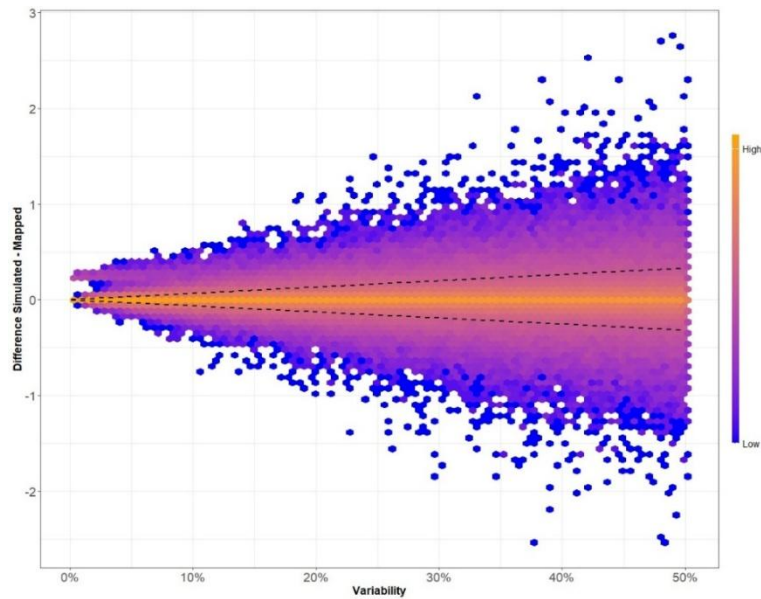
The overall accuracy of the classified map is 85.0%, where most errors arise from the HIF underestimating the in-situ pressure observed during the visual inspection of high-resolution satellite images (Table 4). These suggest that the HIF can be considered a conservative estimate of human pressures on the environment. Moreover, the confusion matrix shows that the very low-pressure class has both low producer's and user's accuracy, indicating the difficulty of detecting low-impact activities that can occur in highly intact landscapes.

Overall, the combination of continuous and categorical validation metrics demonstrates that the HIF map achieves practical reliability for mapping the cumulative industrial pressures at a national scale that were considered here.

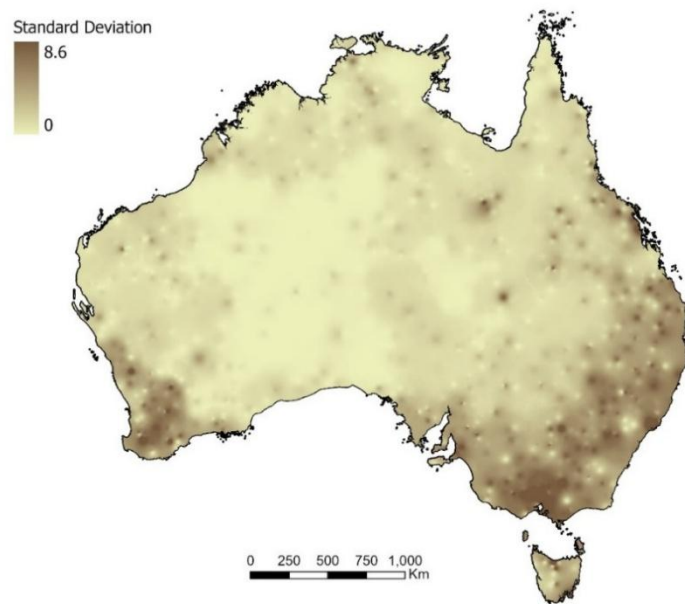
3.2 Uncertainty in pressure values

Changes in the pressure scores had only a limited impact on the cumulative pressure values. Adjusting the scores, either increasing or decreasing them by up to 50%, resulted in a moderate difference (error) between simulated and mapped values (mean = 0.002 ± 1.129). The maximum and minimum errors observed were 2.745 and -2.57, respectively, representing slightly above one-quarter of the full pressure scale.

Across the 100,000 simulations, nearly 90% of validation plots with mapped features (Figure 4) exhibited errors within a narrow range, between -0.15 and 0.16 (corresponding to the 5th and 95th percentiles). As expected, larger adjustment factors led to increased variability; however, even at the maximum adjustment level of 50%, 73% of plots still displayed relatively small errors (ranging from -2.5 to 2.5). The uncertainty map (Fig. 5) shows that areas where multiple pressures converge, particularly densely populated regions near major cities, are more vulnerable to uncertainties in pressure values. Higher uncertainties in areas with high HIF values likely reflect the accumulation of positional and classification errors from overlapping layers and the fine-grained heterogeneity typical of developed landscapes.



500 **Figure 4. Density plot depicting the difference between simulated value and mapped value (y-axis) relative to the percentage of variability of pressure scores (x-axis). The color scale represents the number of plots that include this transition, with orange indicating a high number of plots and blue indicating a few plots (legend is log-scaled). Black dashed lines represent the 5th and 95th percentiles of the distribution of the difference. Plots with no mapped features were excluded from the analysis.**



505 **Figure 5. Spatial distribution of the uncertainty of pressure scores across Australia when these are increased or decreased by 50%. Darker tones represent areas with high standard deviation of the mean cumulative pressure value.**

3.3 Comparison with global footprint datasets

A visual comparison between the national HIF and the global Human Footprint maps at 1 km and 100 m resolutions reveals broadly similar patterns of cumulative pressures across the continent. However, clear mismatches are evident even at a coarse scale. For instance, large parts of inland Australia appear pressure-free in the global maps, yet these areas coincide with native pasturelands captured in the Australian analysis (Fig. 6a). At finer scales, differences become more apparent, arising both from the coarser resolution of the 1 km dataset and from the inclusion of additional nationally curated pressures in the HIF. Examples include Kangaroo Island (Fig. 6b) and the city of Townsville and its surroundings (Fig. 6c), where the national HIF captures unpaved roads, forestry, and pasturelands that are absent in the global products. The Australian HIF also shows finer detail in cumulative pressures within urban centres and peri-urban areas, where features such as farm dams, reservoirs, and unpaved roads are more accurately represented.

520

Table 4 Error matrix showing the performance of a thematic map of five pressure classes obtained from the classification of the HIF against the pressure class obtained from the reference data (visual scores), using sample counts (for an error matrix estimated by proportions of area see Table S10). Accuracy measures are presented with a 95% confidence interval, and the overall accuracy was 85%.

525

Pressure classes from HIF	Reference data					Total	Producer's accuracy	Weights
	No pressure	Very Low	Low	Moderate	High			
No pressure	408	21	49	1		479	97.66 ± 1.23	0.362
Very Low	9	11	8	1		29	27.4 ± 11.98	0.021
Low	1	6	526	4	10	547	90.04 ± 1.94	0.475
Moderate			18	71	71	160	75.18 ± 10.98	0.056
High			3	7	172	182	70.58 ± 4.31	0.085
Total	418	38	604	84	253	1397		
Users accuracy	85.18 ± 3.19	37.93 ± 17.97	96.16 ± 1.61	44.38 ± 7.72	94.51 ± 3.31			

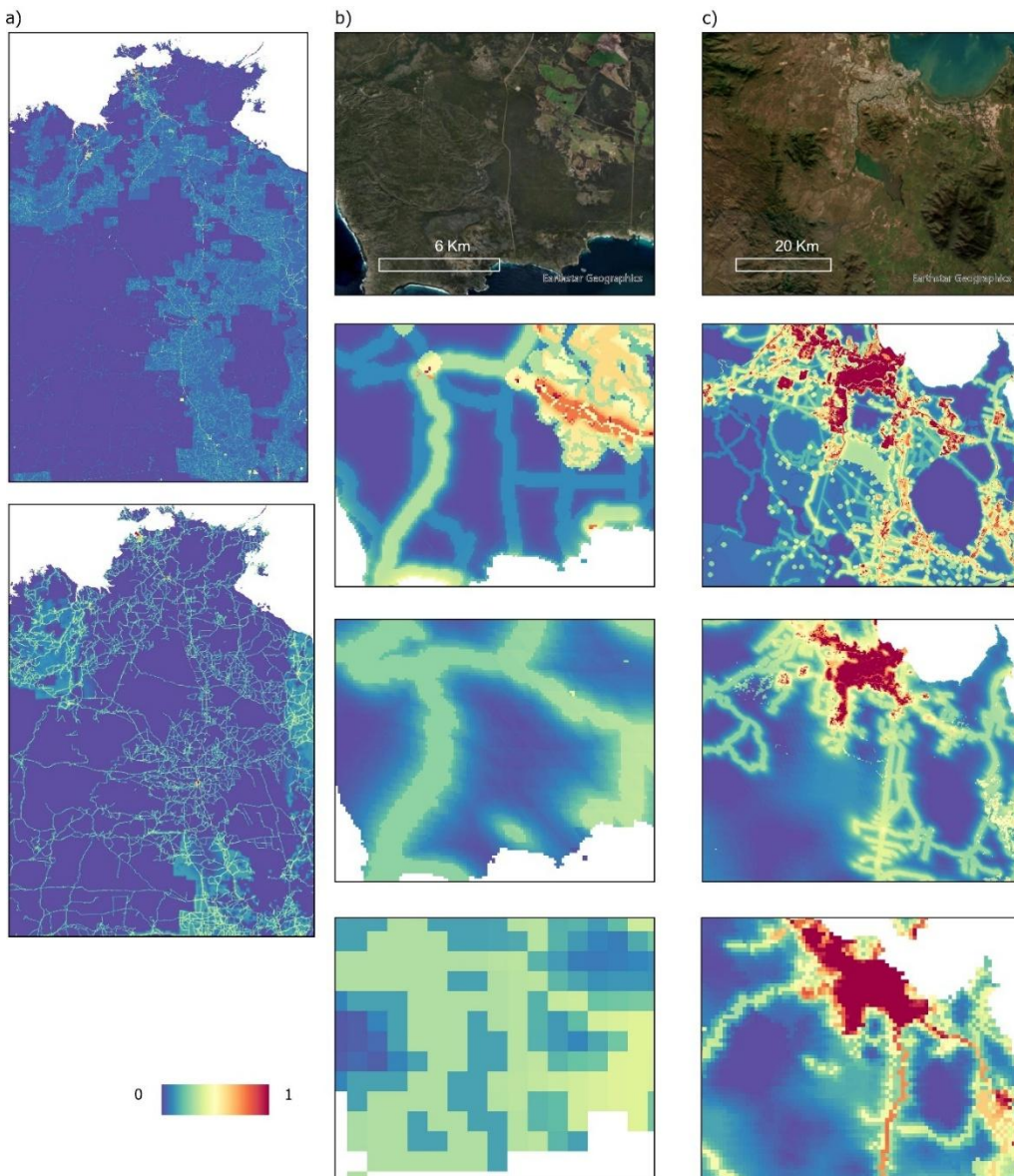
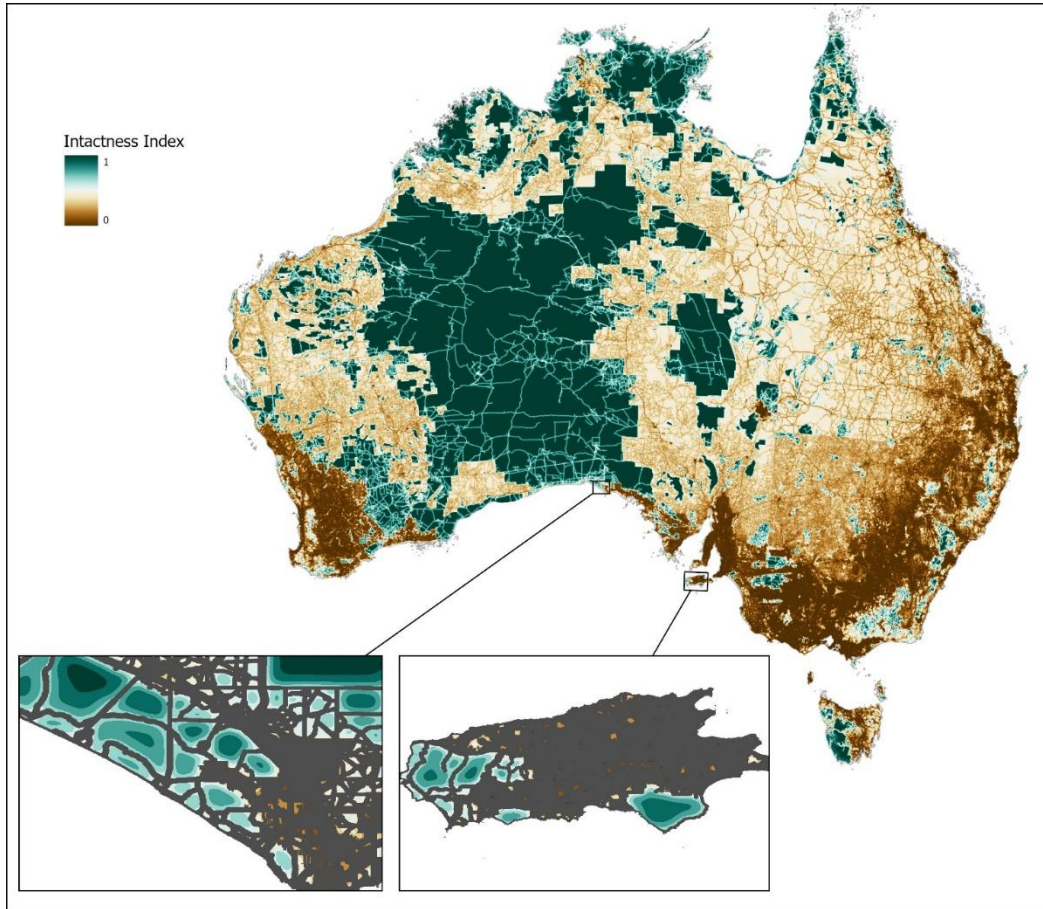


Figure 6. Comparison between the Australian Human Industrial Footprint (HIF) and global Human Footprint datasets. (a) Continental-scale comparison showing extensive areas where the Australian HIF (top) maps pressures from grazing in native pasturelands that are not visible in the global products (bottom). (b) Regional example for Kangaroo Island, South Australia, illustrating improved representation of pressures such as unpaved roads, forestry, and pasturelands in the national HIF relative to global datasets. (c) Urban and peri-urban examples for Townsville, Queensland, showing that the national HIF captures finer-scale gradients of disturbance and additional features such as reservoirs, farm dams, and unpaved roads. Values range from 0 (low pressure) to 1 (high pressure). From top to bottom in panels (b) and (c) are satellite imagery of the area of interest, followed by the Australian HIF, the global 100 m, and the global 1 km datasets, respectively.

3.5 Ecological Intactness Index

The Ecological Intactness Index map for Australia (Fig. 7) covers the same area as the HIF and was calculated using the same spatial resolution of 100 m. The mean intactness value obtained through this map for Australia is 0.52 ± 0.32 (on a scale of 0 to 1, with 1 representing high intactness). Approximately 60.5% of the country has an EII value of < 0.5 , with 9.4% of the landscape experiencing the most severe levels of degradation (EII < 0.1). The inset maps show how the EII value varies in areas with no mapped pressures, depending on the spatial configuration and habitat-quality context surrounding each pixel, to provide a surrogate of structural intactness.



545 **Figure 7. Ecological intactness index map for Australia calculated based on Beyer et al. (2020) using the Australian Industrial Footprint as an input habitat condition layer. The two inset maps show that, despite some areas not having observable industrial pressures (areas with mapped pressures are in grey), their intactness values depend on the habitat quality (HIF value) of surrounding pixels.**

4 Discussion

550 We present two complementary national datasets: the Human Industrial Footprint (HIF), the first national cumulative pressure map since the National Wilderness Inventory of the 1980s and 1990s (Lesslie and Taylor, 1985), and the complementary Ecological Intactness Index (EII), a map to represent ecological structural intactness across the nation. Both extend the established Human Footprint framework (Sanderson et al., 2002; Venter et al., 2016) using nationally curated and thematically detailed inputs, harmonised under a consistent scoring and validation approach at a 100 m spatial resolution.

555 The HIF represents a cumulative model of industrial pressures rather than direct ecological conditions. High HIF values indicate areas with greater concentration or intensity of human activities, often associated with degraded states of natural systems; low values indicate areas with fewer detectable or less intense pressures, associated with intact states. As in previous studies, HIF values should be interpreted ordinally, not linearly (Watson et al., 2016; Williams et al., 2020). For example, a pixel with a value of 20 does not imply that it experiences double the pressure compared to a pixel with a value of 10; rather,
560 it can be assumed to be experiencing a higher level of cumulative disturbance. The classified HIF map provides an intuitive framework for communication and comparison across regions, facilitating policy and planning, and has been used for many diverse conservation applications, including determining species' risk of extinction and the degree of human influence on protected areas (Allan et al., 2022; Jones et al., 2018; Di Marco et al., 2018; Torres-Romero et al., 2025).

The EII complements the HIF by translating mapped pressures into an estimate of ecological intactness, capturing
565 fragmentation and degradation, without reference to specific pressures. It has been used to guide protected area expansion, restoration planning, and application of the mitigation hierarchy (Mappin et al., 2019; Venegas-Li et al., 2024), and, as an indicator to the GBF, will help nations measure progress toward global conservation goals related to integrity.

We do note that cumulative impact maps remain imperfect representations of the complex human-environment interactions (Halpern and Fujita, 2013). They are constrained by the quality and completeness of the underlying datasets and by the
570 assumptions made to integrate them. All these are fully documented here to ensure transparency and reproducibility, and we provide all data and code to openly encourage refinement and improvements as new information becomes available.

5 Limitations and Uncertainty

While comprehensive, the product we present here is subject to several limitations. Some are inherent in cumulative pressure mapping (Watson et al., 2023b); others reflect the data used and assumptions in this analysis. Our analysis does not account
575 for all pressures, either because we lacked suitable data to represent them, or because we focused on those that are observable and related to access by humans. Excluded pressures include individual oil and gas wells, invasive species, disease, pollution, climate change, changes in groundwater regimes, and fire-regime shifts.

Consequently, some areas mapped as intact could be severely affected by unmapped disturbances. Similarly, the ecological response to equivalent pressures likely varies among ecosystems, meaning HIF values indicate potential rather than realized
580 impacts. This being said, the HIF has been shown to be an excellent proxy for assessing species extinction and ecological

degradation (see discussion in Watson et al. 2023b). We also note that rasterising vector data at 100 m resolution introduces a degree of generalisation, particularly for narrow linear features such as roads and pipelines, which may occupy only a fraction of a pixel. This can slightly overestimate their direct footprint, although we make the assumption that, in most cases, the ecological influence of such infrastructure extends beyond this area.

585 In the methods section, we outlined the specific limitations of each data set and the assumptions we made, but we discuss some in more detail here. Roads and grazing in native pasturelands are the two most prevalent pressures in the Australian landscape, but remain imperfectly mapped. Although we merged the Geoscape National Roads and OpenStreetMaps data, a close inspection using high-resolution satellite images shows that many unpaved and private roads are missing, and the completeness of the data varies between states and territories. For example, rural roads in New South Wales appear to be better mapped than
590 in the adjoining areas in Queensland and South Australia, likely underestimating pressures in those regions. Grazing intensity in native pastures is poorly documented in Australia; we therefore applied a conservative pressure score of 2, which might underestimate the degree of pressure in some areas.

One key concern with additive methods for mapping cumulative pressure maps is the use of expert judgment to assign pressure intensity scores to each spatial layer (Halpern and Fujita, 2013; Korpinen and Andersen, 2016). To minimize subjectivity, we
595 adopted scoring approaches from established, peer-reviewed global and national studies (Arias-Patino et al., 2024; Hirsh-Pearson et al., 2022; Venter et al., 2016b; Woolmer et al., 2008), and apply them consistently via open scripts. To further address concerns about subjectivity and its influence on the final cumulative pressure scores, we conducted a comprehensive uncertainty analysis. Following the methodology of Arias-Patiño et al. (2024), we tested robustness by varying all scores by up to $\pm 50\%$ across 100,000 simulations; nearly 90% of validation plots showed errors within a narrow range, indicating low
600 sensitivity to reasonable variations in pressure score inputs. Future refinements could employ structured expert elicitation or data-driven calibration to strengthen scoring schemes.

6 Data and code availability

The datasets generated from this work are available at Zenodo <https://zenodo.org/records/15833395> (Venegas-Li et al., 2025). It is provided in a standard raster format (tif). The code to create the individual pressure layers and the human footprint are
605 available through the same repository.

7 Conclusions

The Human Industrial Footprint (HIF) and Ecological Intactness Index (EII) developed in this study provide a contemporary assessment of cumulative pressures accross Australia's landscapes and a proxy for ecological degradation at 100 m resolution. These datasets offer valuable insights for understanding human impacts on biodiversity and ecosystem intactness and degree
610 of degradation, addressing a long-standing gap in national-scale pressure mapping. By incorporating 16 nationally relevant

pressure layers, the HIF provides a more accurate and context-specific representation of industrial influences than global-scale analyses, improving our ability to guide conservation and land-use planning. Both layers should be of interest to all those involved in biodiversity management when considering Australia's Strategy for Nature 2024-2030 (Commonwealth of Australia, 2024) and Nature Positive Plan (DCCEEW, 2022), as well as its global commitments to the Kunming-Montreal Global Biodiversity Framework with respect to Targets 1-4 especially (CBD, 2022). By identifying areas of high intactness and those under significant industrial pressure, these datasets can inform protected area expansion, ecosystem restoration priorities, and biodiversity offset strategies. Beyond conservation policy, the HIF and EII have applications in environmental impact assessments, regional land-use planning, and climate adaptation and mitigation strategies. Their integration into national and subnational decision-making processes can help halt further biodiversity loss, improve connectivity between protected areas, and support sustainable development objectives.

Author Contributions

RVL, SA, RF, PO, and JW conceptualized the study, with contributions to the development of the methods from all other co-authors. RVL, BA, and SA prepared the data. SA implemented the HIF and EII mapping. MAUAJ, RVL, LM, and MAP contributed to the technical validation and uncertainty analysis. Funding was secured by PO and RF. The original draft was prepared by RVL, JW, and SA, with review and editing contributions from all other co-authors. JW was the leader of the project.

Competing Interests

The authors declare that they have no conflict of interest.

Acknowledgements

This research was funded by The Wilderness Society, and we appreciate the many conversations with staff around these products.

References

- ABARES: The Australian Land Use and Management Classification Version 8, 2016.
- ABARES: Australia's State of the Forests Report 2018, Canberra, 2018.
- 635 ABARES: Forests of Australia, 2023.
- ABARES: Catchment Scale Land Use of Australia – Update December 2023 version 2 [Dataset], Australian Bureau of Agricultural and Resource Economics and Sciences, Canberra, 2024.

- Allan, J. R., Possingham, H. P., Atkinson, S. C., Waldron, A., Marco, M. Di, Butchart, S. H. M., Adams, V. M., Kissling, W. D., Worsdell, T., Sandbrook, C., Gibbon, G., Kumar, K., Mehta, P., Maron, M., Williams, B. A., Jones, K. R., Wintle, B. A., Reside, A. E., and Watson, J. E. M.: The minimum land area requiring conservation attention to safeguard biodiversity, *Science* (1979), 376, 1094–1101, <https://doi.org/10.1126/SCIENCE.ABL9127>, 2022.
- Arias-Patino, M., Johnson, C. J., Schuster, R., Wheate, R. D., and Venter, O.: Accuracy, uncertainty, and biases in cumulative pressure mapping, *Ecol Indic*, 166, 112407, <https://doi.org/https://doi.org/10.1016/j.ecolind.2024.112407>, 2024.
- Ban, N. C., Alidina, H. M., and Ardron, J. A.: Cumulative impact mapping: Advances, relevance and limitations to marine management and conservation, using Canada’s Pacific waters as a case study, *Mar Policy*, 34, 876–886, <https://doi.org/10.1016/j.marpol.2010.01.010>, 2010.
- Barnett, Z. C. and Adams, S. B.: Review of Dam Effects on Native and Invasive Crayfishes Illustrates Complex Choices for Conservation Planning, <https://doi.org/10.3389/fevo.2020.621723>, 2020.
- Benítez-López, A., Alkemade, R., and Verweij, P. A.: The impacts of roads and other infrastructure on mammal and bird populations: A meta-analysis, <https://doi.org/10.1016/j.biocon.2010.02.009>, 2010.
- Betts, M. G., Wolf, C., Ripple, W. J., Phalan, B., Millers, K. A., Duarte, A., Butchart, S. H. M., and Levi, T.: Global forest loss disproportionately erodes biodiversity in intact landscapes, *Nature*, 547, 441–444, <https://doi.org/10.1038/nature23285>, 2017.
- Bevanger, K.: Biological and conservation aspects of bird mortality caused by electricity power lines: A review, [https://doi.org/10.1016/S0006-3207\(97\)00176-6](https://doi.org/10.1016/S0006-3207(97)00176-6), 1998.
- Beyer, H. L., Venter, O., Grantham, H. S., and Watson, J. E. M.: Substantial losses in ecoregion intactness highlight urgency of globally coordinated action, *Conserv Lett*, 13, e12692, <https://doi.org/10.1111/CONL.12692>, 2020.
- Boldy, R., Santini, T., Annandale, M., Erskine, P. D., and Sonter, L. J.: Understanding the impacts of mining on ecosystem services through a systematic review, *Extr Ind Soc*, 8, 457–466, <https://doi.org/10.1016/J.EXIS.2020.12.005>, 2021.
- Borja, Á., Galparsoro, I., Solaun, O., Muxika, I., Tello, E. M., Uriarte, A., and Valencia, V.: The European Water Framework Directive and the DPSIR, a methodological approach to assess the risk of failing to achieve good ecological status, *Estuar Coast Shelf Sci*, 66, <https://doi.org/10.1016/j.ecss.2005.07.021>, 2006.
- Bowman, D. M. J. S., Kolden, C. A., Abatzoglou, J. T., Johnston, F. H., van der Werf, G. R., and Flannigan, M.: Vegetation fires in the Anthropocene, <https://doi.org/10.1038/s43017-020-0085-3>, 2020.
- Bradshaw, C. J. A.: Little left to lose: Deforestation and forest degradation in Australia since European colonization, *Journal of Plant Ecology*, 5, <https://doi.org/10.1093/jpe/rtr038>, 2012.
- Bradstock, R. A., Williams, J. E., and Gill, A. M.: *Flammable Australia: The fire regimes and biodiversity of a continent.*, Cambridge University Press, 2002.
- Brandt, A. R., Heath, G. A., Kort, E. A., O’Sullivan, F., Pétron, G., Jordaan, S. M., Tans, P., Wilcox, J., Gopstein, A. M., Arent, D., Wofsy, S., Brown, N. J., Bradley, R., Stucky, G. D., Eardley, D., and Harriss, R.: Methane leaks from North American natural gas systems, <https://doi.org/10.1126/science.1247045>, 2014.

- Brockerhoff, E. G., Jactel, H., Parrotta, J. A., Quine, C. P., and Sayer, J.: Plantation forests and biodiversity: Oxymoron or opportunity?, *Biodivers Conserv*, 17, <https://doi.org/10.1007/s10531-008-9380-x>, 2008.
- Bunn, S. E. and Arthington, A. H.: Basic principles and ecological consequences of altered flow regimes for aquatic biodiversity, <https://doi.org/10.1007/s00267-002-2737-0>, 2002.
- 675 CBD: Kunming-Montreal Global Biodiversity Framework, Montreal: Convention on Biological Diversity, 14 pp., 2022.
- Chai, T. and Draxler, R. R.: Root mean square error (RMSE) or mean absolute error (MAE)? -Arguments against avoiding RMSE in the literature, *Geosci Model Dev*, 7, <https://doi.org/10.5194/gmd-7-1247-2014>, 2014.
- Chapman, A. D.: Numbers of Living Species in Australia and the World (2nd ed.). Report for the Australian Biodiversity Information Services, Department of the Environment, Water, Heritage and the Arts., 2009.
- 680 Commonwealth of Australia: Threatened species action plan 2022-2023, Canberra, 2022.
- Commonwealth of Australia: Australia's Strategy for Nature 2024–2030, 2024.
- Commonwealth of Australia: Species Profile and Threats Database. 1 [accessed 4 February 2025, 2025].
- Cowley, R. A., Jenner, D., and Walsh, D.: What distance from water should we use to estimate paddock carrying capacity?, in: PROCEEDINGS OF THE AUSTRALIAN RANGELAND SOCIETY BIENNIAL CONFERENCE, 2015.
- 685 DCCEEW: Nature Positive Plan: better for the environment, better for business, Canberra, 2022.
- Doherty, T. S., Macdonald, K. J., Nimmo, D. G., Santos, J. L., and Geary, W. L.: Shifting fire regimes cause continent-wide transformation of threatened species habitat, *Proceedings of the National Academy of Sciences*, 121, e2316417121, <https://doi.org/10.1073/pnas.2316417121>, 2024.
- 690 Du, Z., Yu, L., Zhao, Y., Li, X., Liu, X., Li, X., Hao, P., Chen, Z., Ma, X., and Wang, H.: Annual global gridded livestock mapping from 1961 to 2021, <https://doi.org/10.5194/essd-2025-175>, 9 April 2025.
- Fischer, J. and Lindenmayer, D. B.: Landscape modification and habitat fragmentation: A synthesis, <https://doi.org/10.1111/j.1466-8238.2007.00287.x>, 2007.
- Fuentes-Montemayor, E., Cuarón, A. D., Vázquez-Domínguez, E., Benítez-Malvido, J., Valenzuela-Galván, D., and Andresen, E.: Living on the edge: Roads and edge effects on small mammal populations, *Journal of Animal Ecology*, 78, <https://doi.org/10.1111/j.1365-2656.2009.01551.x>, 2009.
- 695 Gassert, F., Venter, O., Watson, J. E. M., Brumby, S. P., Mazzariello, J. C., Atkinson, S. C., and Hyde, S.: An operational approach to near real time global high resolution mapping of the terrestrial Human Footprint, *Frontiers in Remote Sensing*, 4, <https://doi.org/10.3389/frsen.2023.1130896>, 2023.
- 700 Geoscape Australia: National Roads dataset Update February 2024, January 2024.
- Geoscience Australia: Foundation Rail Infrastructure [Data set], <https://ecat.ga.gov.au/geonetwork/home/api/records/459c6d44-58fa-458d-824d-37cc33ee398e>, 2021.
- González-Abraham, C., Ezcurra, E., Garcillán, P. P., Ortega-Rubio, A., Kolb, M., and Creel, J. E. B.: The human footprint in Mexico: Physical geography and historical legacies, *PLoS One*, 10, <https://doi.org/10.1371/journal.pone.0121203>, 2015.

- 705 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., and Moore, R.: Google Earth Engine: Planetary-scale geospatial analysis for everyone, *Remote Sens Environ*, 202, <https://doi.org/10.1016/j.rse.2017.06.031>, 2017.
- GRASS Development Team: Geographic Resources Analysis Support System (GRASS) Software, Version 8.4. , <https://grass.osgeo.org>, 2024.
- Grass, I., Loos, J., Baensch, S., Batáry, P., Librán-Embid, F., Ficiciyan, A., Klaus, F., Riechers, M., Rosa, J., Tiede, J., Udy,
710 K., Westphal, C., Wurz, A., and Tschamtko, T.: Land-sharing/-sparing connectivity landscapes for ecosystem services and biodiversity conservation, *People and Nature*, 1, 262–272, <https://doi.org/https://doi.org/10.1002/pan3.21>, 2019.
- Green, R. E., Cornell, S. J., Scharlemann, J. P. W., and Balmford, A.: Farming and the fate of wild nature, *Science* (1979), 307, 550–555, https://doi.org/10.1126/SCIENCE.1106049/SUPPL_FILE/GREEN.SOM.PDF, 2005.
- Haddaway, N. R., Cooke, S. J., Lesser, P., Macura, B., Nilsson, A. E., Taylor, J. J., and Raito, K.: Evidence of the impacts of
715 metal mining and the effectiveness of mining mitigation measures on social-ecological systems in Arctic and boreal regions: A systematic map protocol, *Environ Evid*, 8, <https://doi.org/10.1186/s13750-019-0152-8>, 2019.
- Halpern, B. S. and Fujita, R.: Assumptions , challenges , and future directions in cumulative impact analysis, *EcoSphere*, 4, 1–11, <https://doi.org/10.1890/ES13-00181.1>, 2013.
- Halpern, B. S., Walbridge, S., Selkoe, K. a, Kappel, C. V, Micheli, F., D’Agrosa, C., Bruno, J. F., Casey, K. S., Ebert, C., Fox,
720 H. E., Fujita, R., Heinemann, D., Lenihan, H. S., Madin, E. M. P., Perry, M. T., Selig, E. R., Spalding, M., Steneck, R., and Watson, R.: A global map of human impact on marine ecosystems., *Science*, 319, 948–52, <https://doi.org/10.1126/science.1149345>, 2008.
- Halpern, B. S., Frazier, M., Potapenko, J., Casey, K. S., Koenig, K., Longo, C., Lowndes, J. S., Rockwood, R. C., Selig, E. R., Selkoe, K. A., and Walbridge, S.: Spatial and temporal changes in cumulative human impacts on the world’s ocean, *Nat
725 Commun*, 6, 7615, 2015.
- Hanski, I., Zurita, G. A., Bellocq, M. I., and Rybicki, J.: Species-fragmented area relationship, *Proc Natl Acad Sci U S A*, 110, <https://doi.org/10.1073/pnas.1311491110>, 2013.
- Hirsch, T., Mooney, K., Cooper, D., and Maruma Mrema, E.: Global biodiversity outlook 5, in: Secretariat of the Convention on Biological Diversity, 2020.
- 730 Hirsh-Pearson, K., Johnson, C. J., Schuster, R., Wheate, R. D., and Venter, O.: Canada’s human footprint reveals large intact areas juxtaposed against areas under immense anthropogenic pressure, *Facets*, 7, 398–419, https://doi.org/10.1139/FACETS-2021-0063/SUPPL_FILE/FACETS-2021-0063_SUPPLEMENT4.PDF, 2022.
- Johnson, P. T. J., Olden, J. D., and Vander Zanden, M. J.: Dam invaders: Impoundments facilitate biological invasions into freshwaters, *Front Ecol Environ*, 6, <https://doi.org/10.1890/070156>, 2008.
- 735 Jones, K. R., Venter, O., Fuller, R. A., Allan, J. R., Maxwell, S. L., Negret, P. J., and Watson, J. E. M.: One-third of global protected land is under intense human pressure, *Science* (1979), 360, 788–791, <https://doi.org/10.1126/science.aap9565>, 2018.
- Kauffman, J. B. and Krueger, W. C.: Livestock Impacts on Riparian Ecosystems and Streamside Management Implications... A Review, *Journal of Range Management*, 37, <https://doi.org/10.2307/3899631>, 1984.

- Kazarovsky, P.: Farm dams do not have impact on large floods or do they?, in: Hydrology and Water Resources Symposium 1996: Water and the Environment; , 1996.
- 740 Kearney, S. G., Carwardine, J., Reside, A. E., Fisher, D. O., Maron, M., Doherty, T. S., Legge, S., Silcock, J., Woinarski, J. C. Z., Garnett, S. T., Wintle, B. A., and Watson, J. E. M.: The threats to Australia's imperilled species and implications for a national conservation response, <https://doi.org/10.1071/PC18024>, 2019.
- Kearney, S. G., Watson, J. E. M., Reside, A. E., Fisher, D. O., Maron, M., Doherty, T. S., Legge, S. M., Woinarski, J. C. Z., 745 Garnett, S. T., Wintle, B. A., Ritchie, E. G., Driscoll, D. A., Lindenmayer, D., Adams, V. M., Ward, M. S., and Carwardine, J.: Threat-abatement framework confirms habitat retention and invasive species management are critical to conserve Australia's threatened species, *Biol Conserv*, 277, <https://doi.org/10.1016/j.biocon.2022.109833>, 2023.
- Keeley, J. E. and Syphard, A. D.: Historical patterns of wildfire ignition sources in California ecosystems, *Int J Wildland Fire*, 27, <https://doi.org/10.1071/WF18026>, 2018.
- 750 Kennedy, C. M., Oakleaf, J. R., Theobald, D. M., Baruch-Mordo, S., and Kiesecker, J.: Managing the middle: A shift in conservation priorities based on the global human modification gradient, *Glob Chang Biol*, 25, 811–826, <https://doi.org/10.1111/GCB.14549>, 2019.
- Kingsford, R. T., Watson, J. E. M., Lundquist, C. J., Venter, O., Hughes, L., Johnston, E. L., Atherton, J., Gawel, M., Keith, D. A., MacKey, B. G., Morley, C., Possingham, H. P., Raynor, B., Recher, H. F., and Wilson, K. A.: Major conservation policy 755 issues for biodiversity in oecania, *Conservation Biology*, 23, <https://doi.org/10.1111/j.1523-1739.2009.01287.x>, 2009.
- Korpinen, S. and Andersen, J. H.: A global review of cumulative pressure and impact assessments in marine environments, <https://doi.org/10.3389/fmars.2016.00153>, 2016.
- Kühl, H. S., Boesch, C., Kulik, L., Haas, F., Arandjelovic, M., Dieguez, P., Bocksberger, G., McElreath, M. B., Agbor, A., Angedakin, S., Ayimisin, E. A., Bailey, E., Barubiyo, D., Bessone, M., Brazzola, G., Chancellor, R., Cohen, H., Coupland, C., 760 Danquah, E., Deschner, T., Dowd, D., Dunn, A., Egbe, V. E., Eshuis, H., Goedmakers, A., Granjon, A. C., Head, J., Hedwig, D., Hermans, V., Imong, I., Jeffery, K. J., Jones, S., Junker, J., Kadam, P., Kambere, M., Kambi, M., Kienast, I., Kujirakwinja, D., Langergraber, K. E., Lapuente, J., Larson, B., Lee, K., Leinert, V., Llana, M., Maretti, G., Marrocoli, S., Martin, R., Mbi, T. J., Meier, A. C., Morgan, B., Morgan, D., Mulindahabi, F., Murai, M., Neil, E., Niyigaba, P., Ormsby, L. J., Orume, R., Pacheco, L., Piel, A., Preece, J., Regnaut, S., Rundus, A., Sanz, C., Van Schijndel, J., Sommer, V., Stewart, F., Tagg, N., 765 Vendas, E., Vergnes, V., Welsh, A., Wessling, E. G., Willie, J., Wittig, R. M., Yuh, Y. G., Yurkiw, K., Zuberbühler, K., and Kalan, A. K.: Human impact erodes chimpanzee behavioral diversity, *Science* (1979), 363, <https://doi.org/10.1126/science.aau4532>, 2019.
- Legge, S., Rumpff, L., Garnett, S. T., and Woinarski, J. C. Z.: Loss of terrestrial biodiversity in Australia: Magnitude, causation, and response, <https://doi.org/10.1126/science.adg7870>, 2023.
- 770 Lesslie, R. G. and Taylor, S. G.: Wilderness in South Australia : an inventory of the state's relatively high quality wilderness areas, edited by: Taylor, S. G., Geography, U. of Adelaide. D. of, and Studies, U. of Adelaide. C. for E., Centre for Environmental Studies, University of Adelaide, Adelaide [S. Aust.], 1983.

- Lesslie, R. G. and Taylor, S. G.: The wilderness continuum concept and its implications for Australian wilderness preservation policy, *Biol Conserv*, 32, 309–333, [https://doi.org/https://doi.org/10.1016/0006-3207\(85\)90021-7](https://doi.org/https://doi.org/10.1016/0006-3207(85)90021-7), 1985.
- 775 Lesslie, R. G., Mackey, B. G., and Preece, K. M.: A Computer-based Method of Wilderness Evaluation, *Environ Conserv*, 15, <https://doi.org/10.1017/S0376892900029362>, 1988.
- Liddicoat, C., Ciganovic, P., Sindjic, M., and Wright, M.: Farm Dams. A guide to siting, design, construction and management on Eyre Peninsula, <https://cdn.environment.sa.gov.au/landscape/docs/ep/Farm-Dams-Guide.pdf>, 2022.
- Liermann, C. R., Nilsson, C., Robertson, J., and Ng, R. Y.: Implications of dam obstruction for global freshwater fish diversity, *Bioscience*, 62, <https://doi.org/10.1525/bio.2012.62.6.5>, 2012.
- 780 Malerba, M. E., Wright, N., and Macreadie, P. I.: A continental-scale assessment of density, size, distribution and historical trends of farm dams using deep learning convolutional neural networks, *Remote Sens (Basel)*, 13, <https://doi.org/10.3390/rs13020319>, 2021.
- Mappin, B., Chauvenet, A. L. M., Adams, V. M., Di Marco, M., Beyer, H. L., Venter, O., Halpern, B. S., Possingham, H. P., 785 and Watson, J. E. M.: Restoration priorities to achieve the global protected area target, <https://doi.org/10.1111/conl.12646>, 2019.
- Di Marco, M., Venter, O., Possingham, H. P., and Watson, J. E. M.: Changes in human footprint drive changes in species extinction risk, *Nature Communications* 2018 9:1, 9, 1–9, <https://doi.org/10.1038/s41467-018-07049-5>, 2018.
- Martins, J. H., Camanho, A. S., and Gaspar, M. B.: A review of the application of driving forces - Pressure - State - Impact - 790 Response framework to fisheries management, <https://doi.org/10.1016/j.ocecoaman.2012.07.029>, 2012.
- Martinuzzi, S., Radeloff, V. C., Pastur, G. M., Rosas, Y. M., Lizarraga, L., Politi, N., Rivera, L., Herrera, A. H., Silveira, E. M. O., Olah, A., and Pidgeon, A. M.: Informing forest conservation planning with detailed human footprint data for Argentina, *Glob Ecol Conserv*, 31, <https://doi.org/10.1016/j.gecco.2021.e01787>, 2021.
- Masson-Delmotte, V., Zhai, P., Pörtner, H.-O., Roberts, D., Skea, J., Shukla, P. R., Pirani, A., Moufouma-Okia, W., Péan, C., 795 and Pidcock, R.: Global warming of 1.5 C, An IPCC Special Report on the impacts of global warming of, 1, 43–50, 2018.
- Materne, C., Taber, L., Cowley, R., and Wirf, B.: Cattle walk further than 3 km from water in central Australia, but only if they have to!, in: PROCEEDINGS OF THE 12TH INTERNATIONAL RANGELAND CONGRESS, 2025.
- Maxwell, S. L., Fuller, R. A., Brooks, T. M., and Watson, J. E. M.: Biodiversity: The ravages of guns, nets and bulldozers, *Nature*, 536, 143–145, <https://doi.org/10.1038/536143a>, 2016.
- 800 McCall, S. C., McCarthy, M. A., van der Ree, R., Harper, M. J., Cesarini, S., and Soanes, K.: Evidence that a highway reduces apparent survival rates of squirrel gliders, *Ecology and Society*, 15, <https://doi.org/10.5751/ES-03445-150327>, 2010.
- McCloskey, J. M. and Spalding, H.: A reconnaissance-level inventory of the amount of wilderness remaining in the world, *Ambio*, 18, 1989.
- Mendez Angarita, V. Y., Larsen, P. B., Marcolin, L., and Di Marco, M.: Reconciling Different Forms of Ecological Integrity 805 to Aid the Kunming-Montreal Global Biodiversity Framework, *Conserv Lett*, 18, e13088, <https://doi.org/https://doi.org/10.1111/conl.13088>, 2025.

- Micheli, F., Halpern, B. S., Walbridge, S., Ciriaco, S., Ferretti, F., Frascchetti, S., Lewison, R., Nykjaer, L., and Rosenberg, A. A.: Cumulative human impacts on Mediterranean and Black Sea marine ecosystems: assessing current pressures and opportunities., *PLoS One*, 8, e79889, <https://doi.org/10.1371/journal.pone.0079889>, 2013.
- 810 Mu, H., Li, X., Wen, Y., Huang, J., Du, P., Su, W., Miao, S., and Geng, M.: A global record of annual terrestrial Human Footprint dataset from 2000 to 2018, *Sci Data*, 9, <https://doi.org/10.1038/s41597-022-01284-8>, 2022.
- Nelson, A., Weiss, D. J., van Etten, J., Cattaneo, A., McMenomy, T. S., and Koo, J.: A suite of global accessibility indicators, *Sci Data*, 6, <https://doi.org/10.1038/s41597-019-0265-5>, 2019.
- Nicholson, E., Watermeyer, K. E., Rowland, J. A., Sato, C. F., Stevenson, S. L., Andrade, A., Brooks, T. M., Burgess, N. D.,
815 Cheng, S.-T., Grantham, H. S., Hill, S. L., Keith, D. A., Maron, M., Metzke, D., Murray, N. J., Nelson, C. R., Obura, D., Plumptre, A., Skowno, A. L., and Watson, J. E. M.: Scientific foundations for an ecosystem goal, milestones and indicators for the post-2020 global biodiversity framework, *Nature Ecology & Evolution* 2021, 1–12, <https://doi.org/10.1038/s41559-021-01538-5>, 2021.
- Olofsson, P., Foody, G. M., Stehman, S. V., and Woodcock, C. E.: Making better use of accuracy data in land change studies:
820 Estimating accuracy and area and quantifying uncertainty using stratified estimation, *Remote Sens Environ*, 129, <https://doi.org/10.1016/j.rse.2012.10.031>, 2013.
- Olofsson, P., Foody, G. M., Herold, M., Stehman, S. V., Woodcock, C. E., and Wulder, M. A.: Good practices for estimating area and assessing accuracy of land change, <https://doi.org/10.1016/j.rse.2014.02.015>, 2014.
- OpenStreetMap contributors: OpenStreetMap roads dataset. Retrieved 20/11/2024, 2024.
- 825 Ordaz-Németh, I., Sop, T., Amarasekaran, B., Bachmann, M., Boesch, C., Brncic, T., Caillaud, D., Campbell, G., Carvalho, J., Chancellor, R., Davenport, T. R. B., Dowd, D., Eno-Nku, M., Ganas-Swaray, J., Granier, N., Greengrass, E., Heinicke, S., Herbinger, I., Inkamba-Nkulu, C., Iyenguet, F., Junker, J., Bobo, K. S., Lushimba, A., Maisels, F., Malanda, G. A. F., McCarthy, M. S., Motsaba, P., Moustgaard, J., Murai, M., Ndokoue, B., Nixon, S., Nseme, R. A., Nzoo, Z., Pintea, L., Plumptre, A. J., Roy, J., Rundus, A., Sanderson, J., Sereckx, A., Strindberg, S., Tweh, C., Vanleeuwe, H., Vosper, A., Waltert,
830 M., Williamson, E. A., Wilson, M., Mundry, R., and Köhl, H. S.: Range-wide indicators of African great ape density distribution, *Am J Primatol*, 83, e23338, <https://doi.org/https://doi.org/10.1002/ajp.23338>, 2021.
- Pisaniello, J. D. and Tingey-Holyoak, J. L.: How to achieve best-practice integrated land-use and safety assurance policy for hazardous on-farm water storage, *Land use policy*, 62, <https://doi.org/10.1016/j.landusepol.2016.12.019>, 2017.
- Poff, N. L. and Hart, D. D.: How dams vary and why it matters for the emerging science of dam removal, *Bioscience*, 52,
835 [https://doi.org/10.1641/0006-3568\(2002\)052\[0659:HDVAWI\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2002)052[0659:HDVAWI]2.0.CO;2), 2002.
- Purvis, A., Molnár, Z., Obura, D., Ichii, K., Willis, K., Chettri, N., Dulloo, M., Hendry, A., Gabrielyan, B., and Gutt, J.: Status and trends-nature, 2019.
- Ramírez-Delgado, J. P., Di Marco, M., Watson, J. E. M., Johnson, C. J., Rondinini, C., Corredor Llano, X., Arias, M., and Venter, O.: Matrix condition mediates the effects of habitat fragmentation on species extinction risk, *Nature Communications*
840 2022 13:1, 13, 1–10, <https://doi.org/10.1038/s41467-022-28270-3>, 2022.

- Van Rossum, G. and Drake Jr, F. L.: Python reference manual, Centrum voor Wiskunde en Informatica Amsterdam, 1995.
- Rytwinski, T. and Fahrig, L.: The Impacts of Roads and Traffic on Terrestrial Animal Populations, in: Handbook of Road Ecology, <https://doi.org/10.1002/9781118568170.ch28>, 2015.
- 845 Salafsky, N., Salzer, D., Stattersfield, A. J., Hilton-Taylor, C., Neugarten, R., Butchart, S. H. M., Collen, B., Cox, N., Master, L. L., O'Connor, S., and Wilkie, D.: A standard lexicon for biodiversity conservation: Unified classifications of threats and actions, *Conservation Biology*, 22, <https://doi.org/10.1111/j.1523-1739.2008.00937.x>, 2008.
- 850 Salafsky, N., Relton, C., Young, B. E., Lamarre, P., Böhm, M., Chénier, M., Cochrane, E., Dionne, M., He, K. K., Hilton-Taylor, C., Latrémouille, C., Morrison, J., Raymond, C. V., Seddon, M., and Suresh, V.: Classification of direct threats to the conservation of ecosystems and species 4.0, *Conservation Biology*, 39, e14434, <https://doi.org/https://doi.org/10.1111/cobi.14434>, 2025.
- Sanderson, E. W., Jaiteh, M., Levy, M. A., Redford, K. H., Wannebo, A. V., and Woolmer, G.: The Human Footprint and the Last of the Wild: The human footprint is a global map of human influence on the land surface, which suggests that human beings are stewards of nature, whether we like it or not, *Bioscience*, 52, 891–904, [https://doi.org/10.1641/0006-3568\(2002\)052\[0891:THFATL\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2002)052[0891:THFATL]2.0.CO;2), 2002.
- 855 Scott, G. and Rajabifard, A.: Sustainable development and geospatial information: a strategic framework for integrating a global policy agenda into national geospatial capabilities, *Geo-Spatial Information Science*, 20, <https://doi.org/10.1080/10095020.2017.1325594>, 2017.
- Skinner, E. B., Glidden, C. K., MacDonald, A. J., and Mordecai, E. A.: Human footprint is associated with shifts in the assemblages of major vector-borne diseases, *Nat Sustain*, 6, <https://doi.org/10.1038/s41893-023-01080-1>, 2023.
- 860 Sontter, L. J., Ali, S. H., and Watson, J. E. M.: Mining and biodiversity: key issues and research needs in conservation science, *Proceedings of the Royal Society B*, 285, <https://doi.org/10.1098/RSPB.2018.1926>, 2018.
- Syvitski, J. P. M., Vörösmarty, C. J., Kettner, A. J., and Green, P.: Impact of humans on the flux of terrestrial sediment to the global coastal ocean, *Science (1979)*, 308, <https://doi.org/10.1126/science.1109454>, 2005.
- 865 Thackway, R. and Lesslie, R.: Describing and mapping human-induced vegetation change in the Australian landscape, *Environ Manage*, 42, 572–590, 2008.
- Theobald, D. M.: A general model to quantify ecological integrity for landscape assessments and US application, *Landsc Ecol*, 28, <https://doi.org/10.1007/s10980-013-9941-6>, 2013.
- Theobald, D. M., Kennedy, C., Chen, B., Oakleaf, J., Baruch-Mordo, S., and Kiesecker, J.: Earth transformed: Detailed mapping of global human modification from 1990 to 2017, *Earth Syst Sci Data*, 12, <https://doi.org/10.5194/essd-12-1953-2020>, 2020.
- 870 Theobald, D. M., Oakleaf, J. R., Moncrieff, G., Voigt, M., Kiesecker, J., and Kennedy, C. M.: Global extent and change in human modification of terrestrial ecosystems from 1990 to 2022, *Sci Data*, 12, 606, <https://doi.org/10.1038/s41597-025-04892-2>, 2025.

- Torres-Romero, E. J., Eppley, T. M., Ripple, W. J., Newsome, T. M., Krofel, M., Carter, N. H., Ordiz, A., de Oliveira, T. G.,
875 Selva, N., and Penteriani, V.: Global scale assessment of the human-induced extinction crisis of terrestrial carnivores, *Sci Adv*,
11, eadq2853, <https://doi.org/10.1126/sciadv.adq2853>, 2025.
- Trombulak, S. C. and Frissell, C. A.: Review of ecological effects of roads on terrestrial and aquatic communities, *Conservation
Biology*, 14, 18–30, <https://doi.org/10.1046/J.1523-1739.2000.99084.X>, 2000.
- Tucker, M. A., Santini, L., Carbone, C., and Mueller, T.: Mammal population densities at a global scale are higher in human-
880 modified areas, *Ecography*, 44, <https://doi.org/10.1111/ecog.05126>, 2021.
- Tulloch, A. I. T., Healy, A., Silcock, J., Wardle, G. M., Dickman, C. R., Frank, A. S. K., Aubault, H., Barton, K., and
Greenville, A. C.: Long-term livestock exclusion increases plant richness and reproductive capacity in arid woodlands,
Ecological Applications, 33, e2909, <https://doi.org/10.1002/EAP.2909>, 2023.
- Venegas-Li, R., Grantham, H. S., Rainey, H., Diment, A., Tizard, R., and Watson, J. E. M.: An operational methodology to
885 identify Critical Ecosystem Areas to help nations achieve the Kunming–Montreal Global Biodiversity Framework, *Conserv
Lett*, 17, <https://doi.org/10.1111/conl.13037>, 2024.
- Venegas-Li, R., Atkinson, S. C., Aurelio Uba de Andrade Junior, M., Fletcher, R., Owen, P., Morales Barquero, L., Aska, B.,
Grantham, H., Possingham, H., Venter, O., Ward, M., and Watson, J.: Australia’s terrestrial industrial footprint and ecological
intactness [Dataset], <https://doi.org/10.5281/ZENODO.14999051>, 2025.
- 890 Venter, O., Sanderson, E. W., Magrath, A., Allan, J. R., Beher, J., Jones, K. R., Possingham, H. P., Laurance, W. F., Wood,
P., Fekete, B. M., Levy, M. A., and Watson, J. E. M.: Global terrestrial Human Footprint maps for 1993 and 2009, *Sci Data*,
3, <https://doi.org/10.1038/sdata.2016.67>, 2016a.
- Venter, O., Sanderson, E. W., Magrath, A., Allan, J. R., Beher, J., Jones, K. R., Possingham, H. P., Laurance, W. F., Wood,
P., Fekete, B. M., Levy, M. A., and Watson, J. E. M.: Sixteen years of change in the global terrestrial human footprint and
895 implications for biodiversity conservation, *Nat Commun*, 7, 12558, 2016b.
- Ward, M., Tulloch, A. I. T., Radford, J. Q., Williams, B. A., Reside, A. E., Macdonald, S. L., Mayfield, H. J., Maron, M.,
Possingham, H. P., Vine, S. J., O’Connor, J. L., Massingham, E. J., Greenville, A. C., Woinarski, J. C. Z., Garnett, S. T.,
Lintermans, M., Scheele, B. C., Carwardine, J., Nimmo, D. G., Lindenmayer, D. B., Kooyman, R. M., Simmonds, J. S., Sonter,
L. J., and Watson, J. E. M.: Impact of 2019–2020 mega-fires on Australian fauna habitat, *Nat Ecol Evol*, 4,
900 <https://doi.org/10.1038/s41559-020-1251-1>, 2020.
- Ward, M. S., Simmonds, J. S., Reside, A. E., Watson, J. E. M., Rhodes, J. R., Possingham, H. P., Trezise, J., Fletcher, R., File,
L., and Taylor, M.: Lots of loss with little scrutiny: The attrition of habitat critical for threatened species in Australia, *Conserv
Sci Pract*, 1, <https://doi.org/10.1111/csp2.117>, 2019.
- Washington-Allen, R. A., Van Niel, T. G., Ramsey, R. D., and West, N. E.: Remote sensing-based piosphere analysis, *GIsci
905 Remote Sens*, 41, <https://doi.org/10.2747/1548-1603.41.2.136>, 2004.
- Watson, J., Venegas-Li, R., and Atkinson, S.: South Australian Human Industrial Footprint and Intactness Assessment, DEW
Technical Report 2023/82, 2023a.

- Watson, J. E. M. and Venter, O.: Mapping the Continuum of Humanity's Footprint on Land, *One Earth*, 1, 175–180, <https://doi.org/10.1016/J.ONEEAR.2019.09.004>, 2019.
- 910 Watson, J. E. M., Shanahan, D. F., Di Marco, M., Allan, J., Laurance, W. F., Sanderson, E. W., Mackey, B., and Venter, O.: Catastrophic Declines in Wilderness Areas Undermine Global Environment Targets, *Current Biology*, 26, 2929–2934, <https://doi.org/10.1016/j.cub.2016.08.049>, 2016.
- Watson, J. E. M., Ellis, E. C., Pillay, R., Williams, B. A., and Venter, O.: Mapping Industrial Influences on Earth's Ecology, <https://doi.org/10.1146/annurev-environ-112420-013640>, 2023b.
- 915 Weiss, D. J., Nelson, A., Vargas-Ruiz, C. A., Gligorić, K., Bavadekar, S., Gabrilovich, E., Bertozzi-Villa, A., Rozier, J., Gibson, H. S., Shekel, T., Kamath, C., Lieber, A., Schulman, K., Shao, Y., Qarkaxhija, V., Nandi, A. K., Keddie, S. H., Rumisha, S., Amratia, P., Arambepola, R., Chestnutt, E. G., Millar, J. J., Symons, T. L., Cameron, E., Battle, K. E., Bhatt, S., and Gething, P. W.: Global maps of travel time to healthcare facilities, *Nat Med*, 26, <https://doi.org/10.1038/s41591-020-1059-1>, 2020.
- 920 Wickham, H.: *Ggplot2: Elegant graphics for data analysis*, 2016.
- Williams, B. A., Venter, O., Allan, J. R., Atkinson, S. C., Rehbein, J. A., Ward, M., Di Marco, M., Grantham, H. S., Ervin, J., Goetz, S. J., Hansen, A. J., Jantz, P., Pillay, R., Rodríguez-Buriticá, S., Supples, C., Virnig, A. L. S., and Watson, J. E. M.: Change in Terrestrial Human Footprint Drives Continued Loss of Intact Ecosystems, *One Earth*, 3, 371–382, <https://doi.org/https://doi.org/10.1016/j.oneear.2020.08.009>, 2020.
- 925 Woinarski, J. C. Z., Burbidge, A. A., and Harrison, P. L.: Ongoing unraveling of a continental fauna: Decline and extinction of Australian mammals since European settlement, <https://doi.org/10.1073/pnas.1417301112>, 2015.
- Woolmer, G., Trombulak, S. C., Ray, J. C., Doran, P. J., Anderson, M. G., Baldwin, R. F., Morgan, A., and Sanderson, E. W.: Rescaling the Human Footprint: A tool for conservation planning at an ecoregional scale, *Landsc Urban Plan*, 87, 42–53, <https://doi.org/10.1016/J.LANDURBPLAN.2008.04.005>, 2008.
- 930 WorldPop: Global High Resolution Population Denominators Project - Funded by The Bill and Melinda Gates Foundation (OPP1134076). , 2018.