

# Mapping key soil micronutrients across the Tibetan Plateau

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**Abstract.** Soil micronutrient supply sustains critical ecological functions but exhibit poorly quantified distribution patterns in high-altitude ecosystems. This study bridges this knowledge gap through a large-scale investigation across the Tibetan Plateau, a cold-arid region where cryogenic weathering, aridity, and suppressed pedogenesis interact to govern micronutrient cycling. We assembled a plateau-wide dataset from 526 sites with triplicate surface soils (0-10 cm) per site (n = 1,660). Four micronutrients (Fe, Mn, Zn, and V) were measured by laboratory X-ray fluorescence (XRF) for all samples and calibrated and validated against inductively coupled plasma mass spectrometry (ICP-MS) for a subset of samples. Four micronutrients were paired with multi-source predictors (climate, vegetation, soil covariates, topography, grazing disturbance, and weathering proxy). Elemental contents span broad ranges, with site-level summaries (mean  $\pm$  SD, mg·kg<sup>-1</sup>) of Fe 22,864.30  $\pm$  7,589.01, Mn 576.74  $\pm$  206.44, Zn 27.24  $\pm$  8.55, and V 56.99  $\pm$  19.33. Random Forest model was employed to quantify controls and generate high-resolution spatial maps. Key results reveal that pronounced regional heterogeneity is driven primarily by weathering intensity with secondary modulation from climate and topography covariates. Element-specific spatial patterns were observed, with Fe enrichment in southeastern/southern plateau, Mn gradients increasing southeastward. Zn hotspots in central-eastern and western marginal zones, and V exhibits a moderate spatial gradient, with higher contents in the southeastern Tibetan Plateau and relatively lower values in the northwest. We provide 1-km maps of all four micronutrients together with pixel-wise uncertainty layers to support benchmarking of process-based micronutrient cycling models and to inform sustainable ecosystem management under climate change. Predictions for major elements are robust, whereas trace-level elements, particularly Zn, exhibit comparatively higher methodological uncertainty despite calibration. Accordingly, users should interpret absolute Zn contents with caution and refer to the accompanying uncertainty diagnostics when applications require high-precision estimates. The dataset is openly available at TPDC (<https://doi.org/10.11888/Terre.tpdc.303242>; Huo et al., 2025).

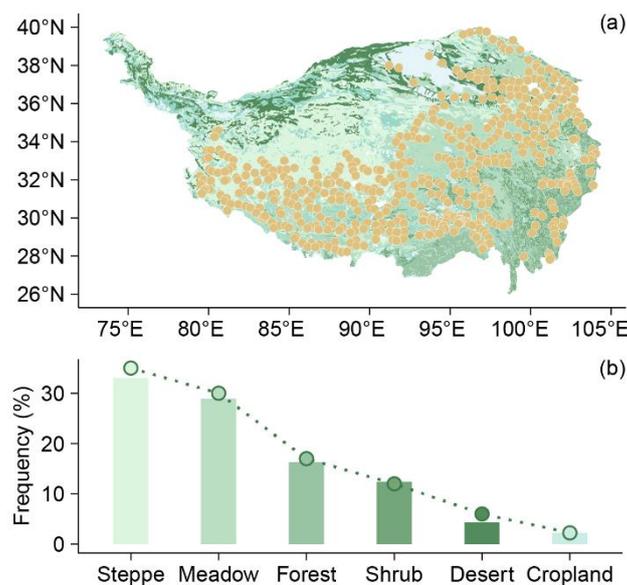
## 28 **1 Introduction**

29 As essential yet trace-level components of living systems, micronutrients (e.g., Fe, Mn, Zn, and V) sustain fundamental  
30 ecological processes, including photosynthesis (Fe, Mn; Fischer et al., 2015; Schmidt et al., 2020), respiration (Fe; Dallman,  
31 1986), enzymatic/redox functions (Zn, V; Hänsch et al., 2009), and biological nitrogen fixation (V; O'Hara, 2001). Crucially,  
32 micronutrient gradients in soils propagate through trophic chains, directly influencing human nutrition and health;  
33 deficiencies exacerbate global malnutrition burdens (Fageria et al., 2002; White et al., 2005). Despite their pivotal role in  
34 ecosystem stability and food security (Presteele et al., 2016; Stehfest et al., 2019), critical knowledge gaps persist regarding  
35 the distribution patterns and drivers of soil micronutrients from regional to global scales.

36 The Tibetan Plateau represents a uniquely important yet underrepresented region for addressing these knowledge gaps. As  
37 the high-altitude landmass and the most extensive alpine permafrost region at low to mid-latitude (Yao et al., 2012), the  
38 Tibetan Plateau supports vast areas of cold- and nutrient-limited ecosystems, where plant productivity, microbial activity, and  
39 biogeochemical cycling are particularly sensitive to micronutrient availability (Han et al., 2022; Tian et al., 2019). In such  
40 environments, micronutrients may act as co-limiting or even primary limiting factors alongside nitrogen and phosphorus, yet  
41 their spatial distributions and environmental controls remain poorly quantified (Dong et al., 2023; Pan et al., 2024).  
42 Moreover, the Tibetan Plateau is experiencing rapid climate change characterized by accelerated warming and increasing  
43 moisture availability, which is expected to fundamentally alter soil weathering regimes, redox conditions, and  
44 mineral-organic interactions (Yu et al., 2024; Cheng et al., 2024). Without robust baseline assessments of soil micronutrients,  
45 it remains difficult to evaluate how ongoing climate change may reshape nutrient limitation patterns and ecosystem  
46 functioning across high-altitude regions.

47 Soil micronutrient supply originates from coupled physicochemical weathering and biological mediation, critically regulated  
48 by local climate and topography (Ochoa-Hueso et al., 2020; Hartmann et al., 2023). In cold-arid high-altitude regions,  
49 particularly the Tibetan Plateau, extreme environmental interactions uniquely govern micronutrient cycling. Cryogenic  
50 processes such as glacial erosion and freeze-thaw cycles, accelerate physical bedrock weathering to mobilize lithogenic  
51 micronutrient reservoirs, while aridity concurrently constrains chemical weathering and elemental release (Mu et al., 2020;  
52 Mu et al., 2016). Low temperatures suppress biological turnover and synergize with aridity to compromise pedogenesis  
53 through clay deficits and diminished mineral reactive sites, thereby reducing elemental retention capacity (Dijkstra et al.,  
54 2004). The combined effects of these opposing processes jointly determine the overall abundance and spatial heterogeneity  
55 of soil micronutrients across alpine soils, likely differing from the patterns of soil micronutrients observed in temperate and  
56 tropical regions.

57 Despite the ecological significance of soil micronutrient, our understanding of their spatial distribution and controlling  
58 factors across the Tibetan Plateau remains limited. Previous studies have been largely confined to localized transects or  
59 site-specific investigations (e.g., the Heihe River Basin and the Tibetan Plateau Highway), offering limited spatial coverage  
60 and representativeness, and providing insufficient insight into how climate, vegetation, topography and lithology jointly  
61 regulate micronutrient patterns at the regional scale (Zhang et al., 2012; Guan et al., 2017; Bu et al., 2016). To address these  
62 knowledge gaps, we conducted a large-scale field investigation across the Tibetan Plateau, establishing 526 sampling sites  
63 distributed across representative temperature and moisture gradients (**Fig. 1**). The sampling design encompassed the plateau's  
64 dominant vegetation types and lithological classes. Using this dataset, we analyzed distribution patterns and key controlling  
65 factors for four essential micronutrients (Fe, Mn, Zn, V). We then applied a Random Forest algorithm to generate  
66 high-resolution spatial distribution maps of these micronutrients, representing the first comprehensive quantification at this  
67 scale and resolution.



68  
69 **Figure 1.** Sampling strategy and ecosystem representativeness. (a) Spatial distribution of sampling sites superimposed on China's  
70 1:1,000,000 vegetation map. (b) Areal proportions of ecosystem types (bars) versus sampling point frequency distribution (dots) across  
71 corresponding ecosystems. Similar bar and dot heights indicate that the sampling is proportionally representative.

## 72 2 Methods

### 73 2.1 Field survey and soil micronutrients analysis

74 We analyzed 1,660 topsoil samples collected from 526 locations during a 2019-2021 growing season (July-August) field  
75 survey across the Tibetan Plateau (79-105°E, 27-40°N; **Fig. 1a**). Sampling sites represent major plateau ecosystems: forests,  
76 shrubs, steppes, meadows, deserts, and croplands (**Table 1**). Our field sampling followed a proportional, area-weighted

77 design, whereby the number of sampling sites allocated to each ecosystem type was approximately proportional to its areal  
78 extent across the Tibetan Plateau. This design aims to provide a plateau-wide, regionally representative characterization of  
79 surface-soil micronutrients across dominant ecosystem types. The sites span broad environmental gradients, ranging from  
80 759 to 5565 m in elevation,  $-7.83$  to  $18.46$  °C in mean annual temperature (MAT), and 23 to 898 mm in mean annual  
81 precipitation (MAP), effectively capturing the plateau's topographic and climatic variability. Site was selected using  
82 standardized criteria: maintaining relative homogeneity in species composition, community structure, and habitat conditions,  
83 and avoiding proximity to roads or areas with frequent human activity. At each site, we established a  $10 \times 10$  m<sup>2</sup> quadrat and  
84 collected triplicate soil samples along the diagonal at three positions (0 m, 7.1 m, and 14.1 m; 0-10cm depth), corresponding  
85 to the start, midpoint, and end of the quadrat diagonal. This design allows sampling across the spatial heterogeneity within  
86 each plot while maintaining a standardized sampling framework. This depth was chosen because it represents the  
87 biologically active surface horizon most relevant to plant uptake and microbially mediated cycling, and is widely used in  
88 regional soil inventories for comparability across ecosystems (Du et al., 2017; Cardon et al., 2013). This consistent protocol  
89 ensures cross-site comparability while capturing the variability of surface layer that most strongly interacts with vegetation  
90 and climate. Geographic coordinates, elevation, community type, and species composition (Cheng et al., 2022) were  
91 systematically documented.

92 To ensure the reliability of soil micronutrient measurements, we adopted a two-step analytical strategy. First, all 1,660  
93 topsoil samples collected from 526 sampling locations were air-dried, sieved (2 mm), and analyzed in the laboratory using a  
94 third-generation X-ray fluorescence spectrometer (XRF). All micronutrient data reported in this study are based on these  
95 laboratory XRF measurements. Second, to independently validate the XRF-derived concentrations, a subset of 218 samples  
96 was randomly selected and re-analyzed using inductively coupled plasma-mass spectrometry (ICP-MS), a widely accepted  
97 reference method for elemental analysis (Simon, 2005). Among the examined elements, Fe ( $R^2 = 0.81$ ), Mn ( $R^2 = 0.67$ ), Zn  
98 ( $R^2 = 0.49$ ), and V ( $R^2 = 0.77$ ) exhibit clear and statistically significant linear relationships with ICP measurements,  
99 demonstrating moderate to strong correlation (**Fig. S1**). The increased dispersion observed at low contents reflects known  
100 limitations of XRF near detection thresholds, rather than method failure. Moreover, our analytical strategy follows  
101 established practice in large-scale geochemical mapping, where XRF is commonly used for high-throughput determination  
102 of total/major-element composition, while ICP-based methods are commonly employed in parallel for complementary  
103 measurements (e.g., extracted fractions, waters, or elements requiring lower detection limits) and, where appropriate, for  
104 inter-method comparison and validation. In addition to GEMAS and the NGSAs, similar XRF-based workflows have been  
105 implemented in other regional/national baseline programmes (e.g., BGS G-BASE in the UK, the FOREGS Geochemical

106 Atlas of Europe, and the Global Geochemical Baselines framework). These precedents support the robustness and suitability  
 107 of using XRF with targeted ICP verification for plateau-scale geochemical datasets (Table S1).

108 **Table 1.** Ecosystem classification and sampling coverage on the Tibetan Plateau.

Biome	Characteristics and dominant plant species	elevation ranges	climate conditions (temperature/precipitation)	No. of samples	No. of locations
Steppe	Alpine steppes, dominated by cold-adapted herbaceous species such as <i>Stipa purpurea</i> , features sparse vegetation adapted to cold-arid conditions.	1961-5151 m	-7.1-8.0 °C; 34.32-785.61 mm	569	180
Meadow	Alpine meadows feature dense, low-stature vegetation sustained by year-round low temperatures, high humidity, and water-retentive soils. These ecosystems thrive on gentle slopes and valley floors at higher elevations, hosting relatively diverse flora with characteristic dominance of sedges including <i>Kobresia pygmaea</i> and <i>K. humilis</i> .	2661-5565 m	-7.8-9.2 °C; 158.42-848.70 mm	499	154
Forest	Forests on the Tibetan Plateau concentrate primarily in the southeastern region, dominated by high-altitude cold-temperate coniferous forests. These humid-adapted ecosystems feature <i>fir (Abies)</i> and <i>spruce (Picea)</i> species as characteristic components.	1453-4237 m	-0.6-16.5 °C; 400.34-898.48 mm	265	87
Shrub	Tibetan shrublands primarily occur in arid and alpine zones, characterized by low-growing, drought-tolerant dwarf shrubs such as <i>Lonicera (honeysuckle)</i> and <i>Rhododendron</i> species adapted to nutrient-poor soils and extreme climatic conditions.	2169-5022 m	-5.2-11.8 °C; 301.40-868.63 mm	190	64
Desert	Alpine deserts occur in extremely arid, cold regions and exhibit extremely sparse vegetation dominated by <i>arid-tolerant dwarf shrubs</i> and <i>herbs</i> .	2108-5158 m	-7.1-5.6 °C; 22.64-443.43 mm	101	29
Cropland	Cropland, concentrated in river valleys and basin floors and is dominated by highland barley ( <i>Hordeum vulgare</i> var. <i>nudum</i> , “qingke”), with spring wheat ( <i>Triticum aestivum</i> ), rapeseed ( <i>Brassica napus</i> ), and potato ( <i>Solanum tuberosum</i> ) commonly cultivated; vegetation cover is strongly seasonal and often bare after harvest.	759-4360 m	0.5-18.4 °C; 112.53-783.06 mm	36	12

109 Biomes are grouped by diagnostic characteristics; dominant plant species are italicized. Elevation range gives the minimum-maximum  
 110 elevation (m) of sampling sites within each biome. Climate conditions report site-level ranges of mean annual temperature (°C) and mean  
 111 annual precipitation (mm). No. of samples is the number of soil samples analyzed (0-10 cm, three replicates per site), and No. of locations  
 112 is the number of unique sampling sites.

## 113 2.2 Soil Properties

114 Soil samples were sifted through 2 mm sieve, discarding visible stones and extracted roots. Soil pH was measured using the  
 115 potentiometric method, and soil texture analysis, quantifying clay, silt, and sand content fractions, was determined using a  
 116 laser diffraction particle size analyzer (Mastersizer 2000, Malvern, UK). Soil organic carbon (SOC) content was quantified  
 117 via the potassium dichromate oxidation method (Walkley-Black) with external heating. The chemical index of alteration

118 (CIA) was calculated using the molar proportions of  $\text{Al}_2\text{O}_3$ ,  $\text{CaO}^*$ ,  $\text{Na}_2\text{O}$ , and  $\text{K}_2\text{O}$  according to the formula:  $\text{CIA} = [\text{Al}_2\text{O}_3 /$   
119  $(\text{Al}_2\text{O}_3 + \text{CaO}^* + \text{Na}_2\text{O} + \text{K}_2\text{O})] \times 100$  (Fedo et al., 1995). The major-element oxides used to calculate CIA ( $\text{Al}_2\text{O}_3$ ,  $\text{CaO}$ ,  
120  $\text{Na}_2\text{O}$  and  $\text{K}_2\text{O}$ ) were obtained from laboratory XRF measurements, which provide a well-established basis for CIA-based  
121 evaluation of relative chemical weathering gradients, and data quality was ensured through standard QA/QC including  
122 replicate analyses and certified reference materials. CIA reflects the relative loss of mobile base cations (Ca, Na, K)  
123 compared with the enrichment of immobile Al during weathering (McLennan, 1993). Higher CIA values indicate stronger  
124 chemical weathering and more advanced soil development, whereas lower values suggest weaker weathering and limited  
125 leaching of base cations (Nesbitt et al., 1982).

### 126 **2.3 Environmental variables**

127 We considered geographic, climatic, biological, and edaphic drivers. Field measurements provided location (longitude,  
128 latitude), slope, aspect and vegetation type. For the spatial prediction of micronutrient distributions, however, the gridded  
129 predictor variables (e.g., elevation, slope, aspect, vegetation cover) were derived from online datasets in order to upscale  
130 site-level observations to the Plateau scale. Slope and aspect data came from the National Tibetan Plateau Data Center  
131 (<https://data.tpc.ac.cn>). The digital elevation model (DEM) data were collected from the Resource and Environment  
132 Science and Data Center (<https://www.resdc.cn/>). Climate variables, mean annual temperature (MAT) and mean annual  
133 precipitation (MAP) were downloaded from the Climate Data Store (<https://cds.climate.copernicus.eu/#!/home>). The Aridity  
134 Index (AI), calculated as mean annual precipitation/mean annual reference evapotranspiration, was obtained from the Global  
135 Aridity Index dataset (Trabucco et al., 2018), where higher values indicate greater humidity. Vegetation types (Forest, Shrub,  
136 Meadow, Steppe, Desert, Cropland) followed the 1:1,000,000 China Vegetation Map classification (Hou, 2019). The  
137 normalized difference vegetation index (NDVI) data were obtained from an Earthdata Search (<https://search.earthdata.nasa.gov/search>). The net primary productivity (NPP) data were obtained from the study by Chen et al. (2023) and  
138 were calculated using the CASA model (Potter et al., 1993). The grazing activity data were obtained from statistical  
139 yearbooks. Based on the lithological data published by Dijkshoorn et al. in 2018, the rock types on the Tibetan Plateau were  
140 classified into acidic igneous rock (IA), acidic metamorphic rock (MA), clastic sedimentary rock (SC), carbonate rock (SO),  
141 aeolian facies rock (UE), and fluvial facies rock (UF).

### 143 **2.4 Relative importance analysis and soil micronutrient mapping**

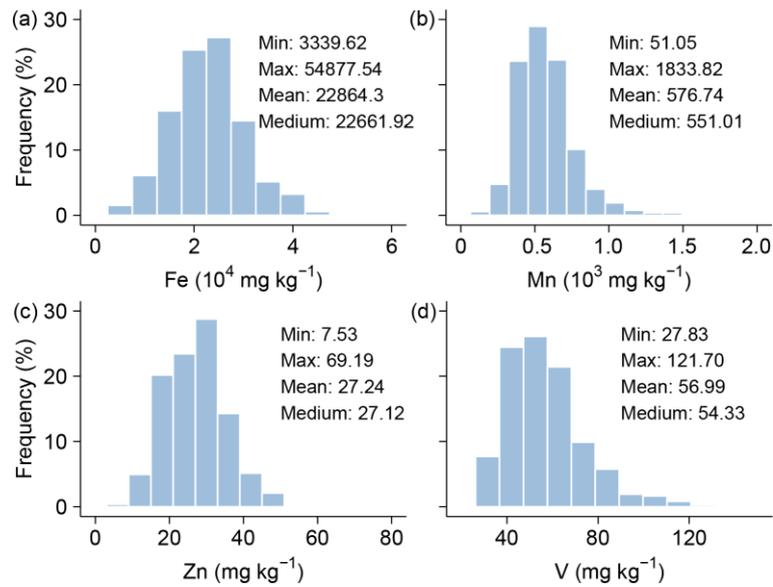
144 Soil micronutrient measurements were preprocessed to detect and remove outliers exceeding the mean  $\pm 3$  standard  
145 deviations. After data screening, the retained sample sizes are: Fe 1654 (from 1660), Mn 1630 (from 1646), Zn 1655 (from

146 1661), and V 857 (from 867). To evaluate the relative importance of predictors in explaining soil micronutrient variability  
147 across the Tibetan Plateau, we applied the random forest permutation importance (%IncMSE) to independently  
148 cross-validate the driver rankings. For spatial prediction, we developed four area-wide random forest models (each  
149 comprising 500 trees) targeting Fe, Mn, Zn, and V contents. The models were trained using a suite of environmental  
150 predictors, including topographic features (DEM, slope, aspect), climate variables (MAT, MAP, AI), vegetation indices  
151 (NDVI, NPP), soil covariates (texture, SOC, pH, CIA), and grazing intensity. Random forest was selected for its ability to  
152 model complex, nonlinear relationships and interactions among diverse types of predictors. Model hyperparameters were  
153 optimized using grid search combined with tenfold cross-validation. To assess model generalizability, we examined the  
154 extent to which the predictor parameter space in the validation set overlapped with that of the original training data. Model  
155 performance was evaluated by comparing predicted versus observed values using scatterplots (predicted on the x-axis,  
156 observed on the y-axis) following the method of Piñeiro et al. (2008), with models achieving strong predictive performance  
157 ( $R^2 = 0.6-0.7$ ). The uncertainty layer is calculated as the inter-pixel variance among tree predictions. All spatial predictions  
158 were produced on a uniform 1 km grid ( $0.01^\circ$ ) across the Tibetan Plateau to maintain consistent spatial support for both  
159 predictions and associated uncertainty estimates. All statistical analyses were conducted using R version 3.4.4.

## 160 **3 Results**

### 161 **3.1 Soil micronutrients content**

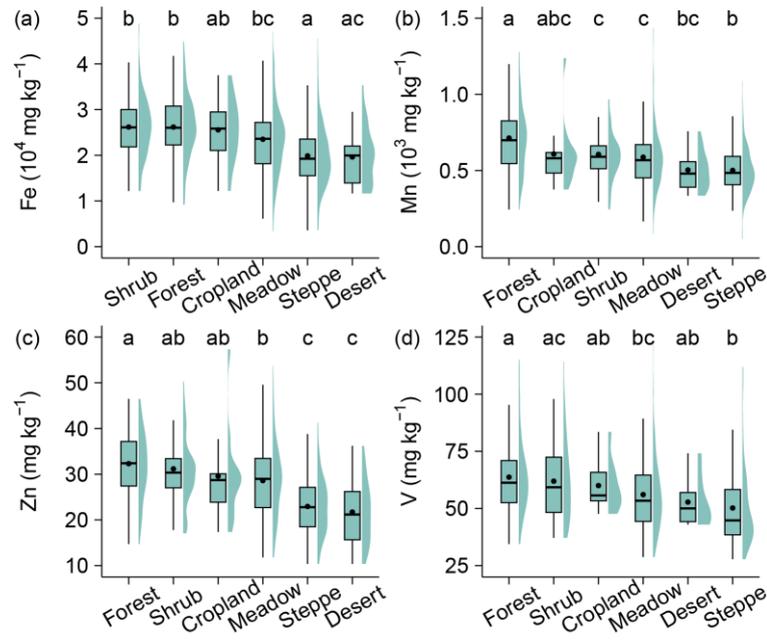
162 Across all sites, soil micronutrient contents varied widely, with mean value of  $22,864.30 \pm 7,589.01$  for Fe (mean  $\pm$  SD,  
163  $\text{mg}\cdot\text{kg}^{-1}$ ),  $576.74 \pm 206.44$  for Mn,  $27.24 \pm 8.55$  for Zn, and  $56.99 \pm 19.33$  for V. Coefficients of variation ( $\text{CV} = \text{SD}/\text{mean}$ )  
164 were 33% for Fe, 36% for Mn, 31% for Zn, and 34% for V. Collectively, Fe and Mn dominate in absolute abundance,  
165 whereas Zn and V occur at tens of  $\text{mg}\cdot\text{kg}^{-1}$ , indicating heterogeneous but orderly micronutrient levels across the Plateau (**Fig.**  
166 **2**).



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168 **Figure 2.** Frequency distributions of soil micronutrients (Fe, Mn, Zn, V) across the Tibetan Plateau.

169 **3.2 Soil micronutrients across vegetation types**

170 Soil micronutrient content (Fe, Mn, Zn, and V) varied significantly among different vegetation types (**Fig. 3**). Fe contents  
 171 were highest in shrub and forest ecosystems, with mean values of  $26,264.11 \text{ mg}\cdot\text{kg}^{-1}$  and  $26,090.66 \text{ mg}\cdot\text{kg}^{-1}$ , respectively,  
 172 significantly exceeding values observed in desert ( $19,762.66 \text{ mg}\cdot\text{kg}^{-1}$ ) and steppe ecosystems ( $19,852.37 \text{ mg}\cdot\text{kg}^{-1}$ ) by  
 173 31-33%. Similarly, Mn contents were significantly higher in forest ( $703.22 \text{ mg}\cdot\text{kg}^{-1}$ ) than in shrub ( $606.33 \text{ mg}\cdot\text{kg}^{-1}$ ),  
 174 meadow ( $591.59 \text{ mg}\cdot\text{kg}^{-1}$ ), desert and steppe ecosystems (both below  $510 \text{ mg}\cdot\text{kg}^{-1}$ ). Zn demonstrated a strong  
 175 vegetation-dependent variability, with forest ( $32.00 \text{ mg}\cdot\text{kg}^{-1}$ ) exhibiting the highest values and steppe ( $22.94 \text{ mg}\cdot\text{kg}^{-1}$ ) and  
 176 desert ( $21.95 \text{ mg}\cdot\text{kg}^{-1}$ ) the lowest, shrub and cropland were intermediate and not significantly different from forest or  
 177 meadow. V also varied among vegetation types, with forest being significantly higher than meadow and steppe, and shrub  
 178 significantly higher than steppe, whereas cropland and desert displayed intermediate levels.

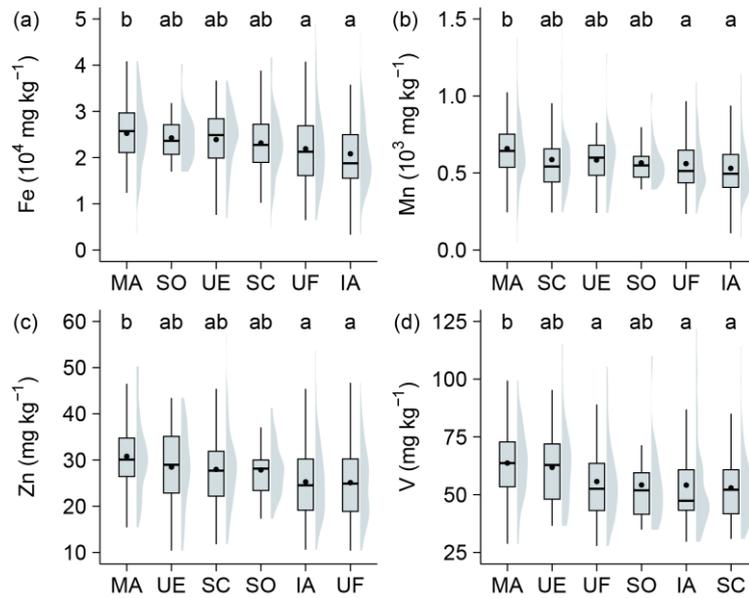


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180 **Figure 3.** Variability in soil micronutrient contents (Fe, Mn, Zn, V) across Tibetan vegetation types. Boxplots show the data distributions  
 181 for each vegetation type. Within each plot, the boxes represent the interquartile range (IQR), the horizontal lines within boxes indicate the  
 182 median values, and black dots denote the mean values. The whiskers extend to 1.5 times the IQR. The surrounding shaded violin shapes  
 183 indicate the kernel density distribution of the data. Lowercase letters denote significance groups from Tukey's HSD multiple comparisons  
 184 ( $\alpha = 0.05$ ): groups sharing at least one letter do not differ significantly, whereas groups with no letters in common differ significantly (e.g.,  
 185 "ab" is not different from "a" or "b").

186 **3.3 Soil micronutrients across lithological classes**

187 Lithological class exerted a significant influence on the distribution of several soil micronutrients (**Fig. 4**). Fe contents  
 188 differed among lithological classes, with soils derived from acidic metamorphic rocks showing the highest mean value  
 189  $25,252.72 \text{ mg}\cdot\text{kg}^{-1}$  and being significantly higher than soils developed from acidic igneous rocks ( $20,830.15 \text{ mg}\cdot\text{kg}^{-1}$ ) and  
 190 fluvial facies rocks ( $21,910.22 \text{ mg}\cdot\text{kg}^{-1}$ ), whereas carbonate ( $24,260.96 \text{ mg}\cdot\text{kg}^{-1}$ ), eolian facies ( $23,902.77 \text{ mg}\cdot\text{kg}^{-1}$ ), and  
 191 clastic sedimentary rocks ( $23,157.66 \text{ mg}\cdot\text{kg}^{-1}$ ) exhibited intermediate levels and were not significantly different from either  
 192 group. A similar pattern was observed for Mn, where acidic-metamorphic-driven soils ( $658.37 \text{ mg}\cdot\text{kg}^{-1}$ ) were significantly  
 193 higher than acidic-igneous-driven soils ( $530.45 \text{ mg}\cdot\text{kg}^{-1}$ ) and fluvial-facies-driven soils ( $560.91 \text{ mg}\cdot\text{kg}^{-1}$ ), with the remaining  
 194 lithologies showing intermediate values. Zn also varied among lithologies, with the highest mean content in acidic  
 195 metamorphic rocks ( $30.79 \text{ mg}\cdot\text{kg}^{-1}$ ), which was significantly higher than acidic igneous rocks ( $25.27 \text{ mg}\cdot\text{kg}^{-1}$ ) and fluvial  
 196 facies rocks ( $25.10 \text{ mg}\cdot\text{kg}^{-1}$ ), while eolian facies, clastic sedimentary, and carbonate rocks did not differ significantly from  
 197 acidic metamorphic rocks or the lower group. For V, acidic-metamorphic-derived soils ( $63.67 \text{ mg}\cdot\text{kg}^{-1}$ ) were significantly  
 198 higher than fluvial facies, acidic igneous, and clastic sedimentary derived soils, whereas eolian facies rocks and carbonate  
 199 rocks showed intermediate levels and were not significantly different from acidic metamorphic rocks or the lower group.



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**Figure 4.** Variability in soil micronutrient contents (Fe, Mn, Zn, V) across Tibetan lithological classes. Boxplots show the data distributions for each lithological classes. Within each plot, the boxes represent the interquartile range (IQR), the horizontal lines within boxes indicate the median values, and black dots denote the mean values. The whiskers extend to 1.5 times the IQR. The surrounding shaded violin shapes indicate the kernel density distribution of the data. Lowercase letters denote significance groups from Tukey's HSD multiple comparisons ( $\alpha = 0.05$ ): groups sharing at least one letter do not differ significantly, whereas groups with no letters in common differ significantly (e.g., "ab" is not different from "a" or "b"). Abbreviations of lithological classes: IA = acidic igneous rock, MA = acidic metamorphic rock, SC = clastic sedimentary rock, SO = carbonate rock, UE = eolian facies rock, UF = fluvial facies rock.

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### 3.4 Drivers of soil micronutrient pattern

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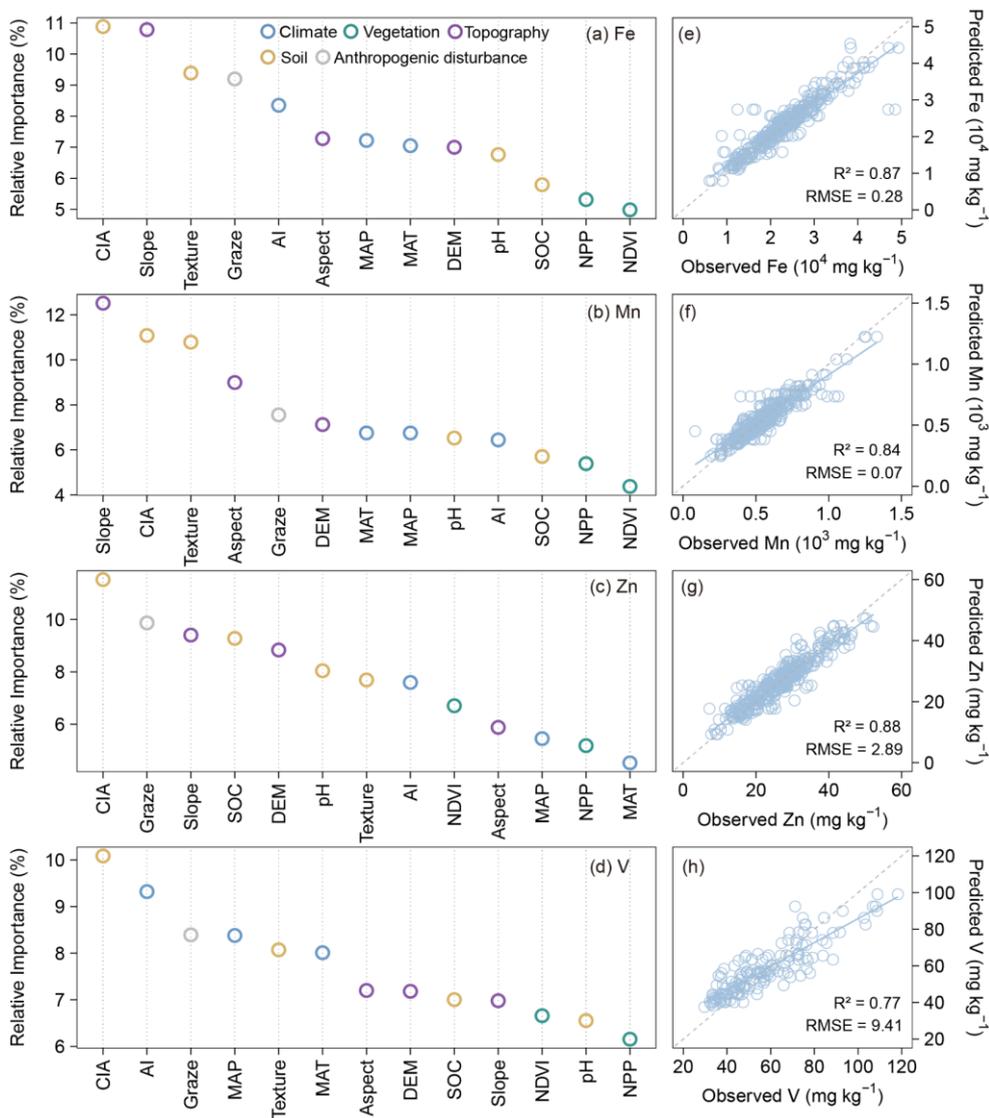
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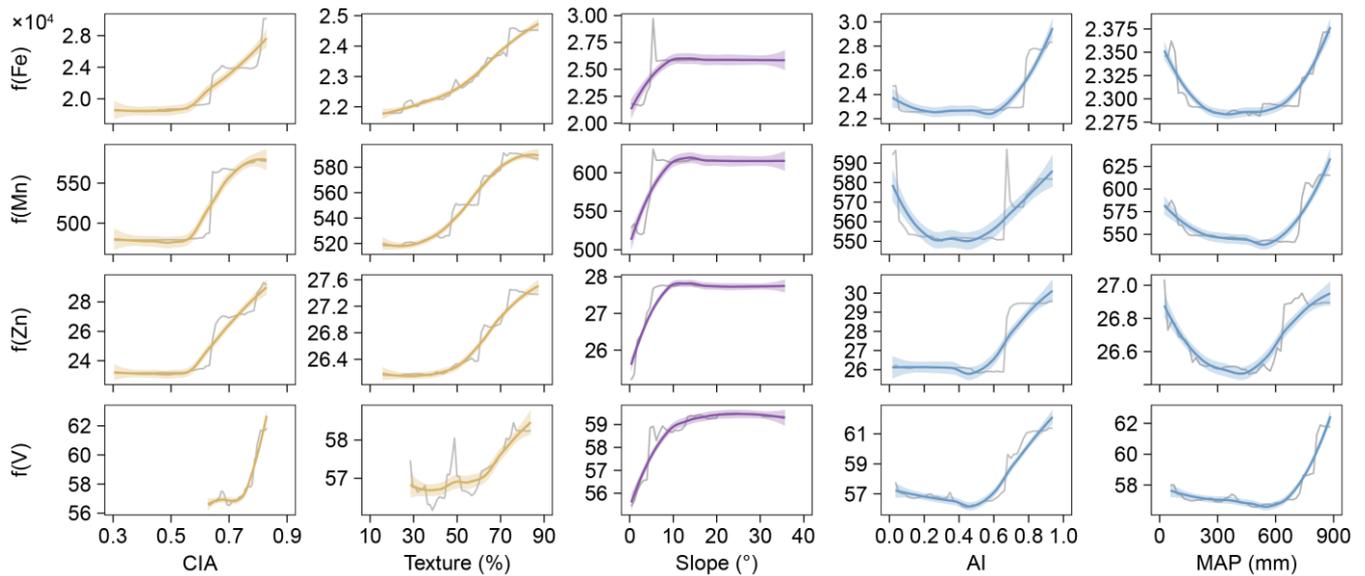
Relative importance analysis was conducted for five predictor groups, including climate, vegetation, soil covariates, topography, and grazing disturbances. The rankings show clear element-specific controls (**Fig. 5a-d**). Soil covariates were the primary contributors to the spatial variation of micronutrient contents, followed by topographic and climatic factors, whereas vegetation and anthropogenic disturbance had relatively minor effects. Among all predictors, CIA, slope, and soil texture exhibited the highest importance, indicating that weathering intensity, terrain, and soil physical structure are the dominant controls on the spatial distribution of soil micronutrients. Specifically, CIA emerged as the dominant factor for Fe, Zn, and V. Topographic factors (slope) primarily influenced Mn.

The partial dependence plots illustrated the nonlinear responses of four soil micronutrients (Fe, Mn, Zn, and V) to five key environmental variables (**Fig. 6**). Overall, the effects of CIA, soil texture, slope, and moisture varied substantially among elements. CIA showed a generally positive relationship with most micronutrients, which increased markedly when chemical index of alteration exceeded approximately 0.5. Soil texture exerted a moderate positive influence on micronutrients. In contrast, slope exhibited a saturating effect—nutrient concentrations rose sharply at low slope angles ( $<10^\circ$ ) and plateaued thereafter—highlighting the limited accumulation potential on steeper terrains. The aridity index demonstrated a clear

222 U-shaped pattern for most micronutrients, with relatively low concentrations under intermediate aridity. Four micronutrients  
 223 exhibited a typical U-shaped relationship with mean annual precipitation, showing higher contents in both low and high  
 224 precipitation zones, and a clear trough in the intermediate range (approximately 300-500 millimeters). This pattern aligned  
 225 closely with responses to drought indices, confirming shared moisture sensitivity. Collectively, these patterns emphasize that  
 226 soil micronutrient distributions across the Tibetan Plateau are strongly shaped by interactions among chemical weathering  
 227 intensity, soil physical properties, and climatic aridity.



228  
 229 **Figure 5.** Relative importance of biotic and abiotic factors for soil micronutrients (Fe, Mn, Zn, V) on the Tibetan Plateau (a-d).  
 230 Relationship between observed and predicted values of soil micronutrients (Fe, Mn, Zn, V) on the Tibetan Plateau based on the Random  
 231 Forest model (e-h). The blue solid line represents the fitted relationship using ordinary least squares regression, while the gray dashed line  
 232 indicates the 1:1 line between observed and predicted values.



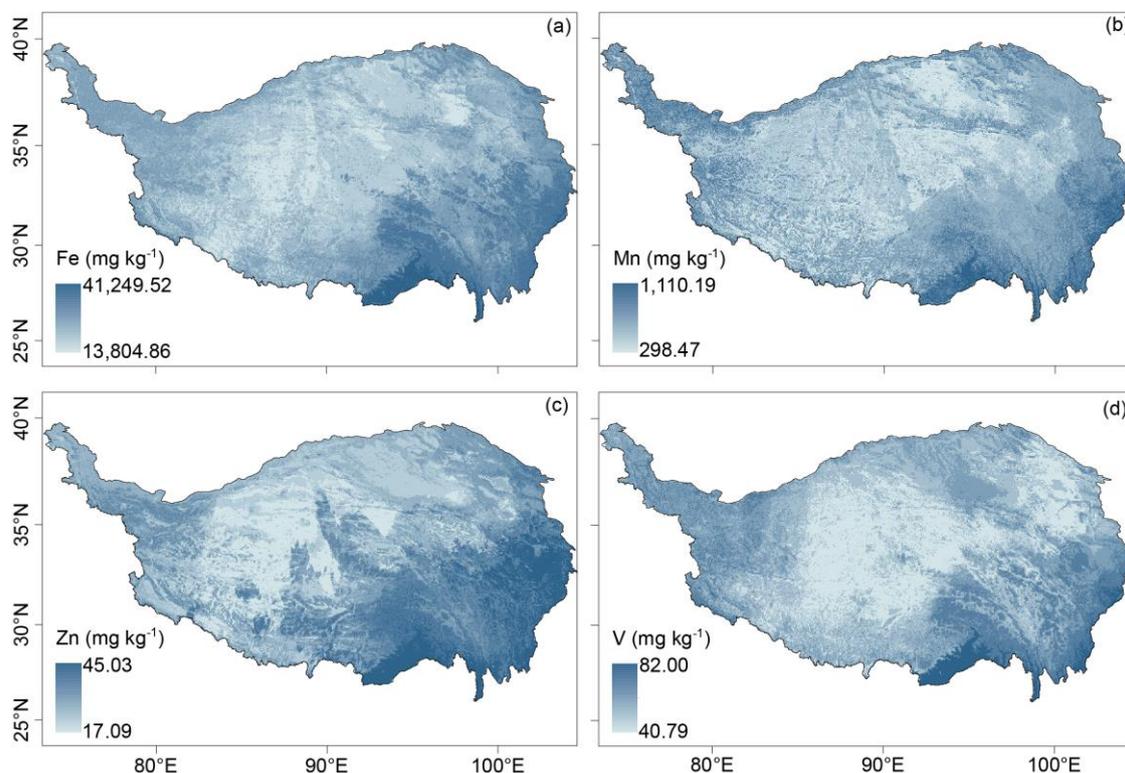
233  
 234 **Figure 6.** Partial dependence of four soil micronutrients (Fe, Mn, Zn, V) on five predictive variables: chemical index of alteration (CIA),  
 235 Texture, Slope, aridity index (AI) and mean annual precipitation (MAP). Gray lines represent the original partial dependence, while  
 236 smoothed fits for each micronutrient are shown as colored lines. Shaded areas represent the 95% confidence intervals.

### 237 3.5 Soil micronutrients maps

238 The random forest (RF) models exhibited strong overall predictive performance for soil micronutrient contents across the  
 239 Tibetan Plateau, with varying accuracies among elements (**Fig. 5e-h**). The models for Fe, Mn, and Zn performed best, with  
 240  $R^2$  values ranging from 0.84 to 0.88 and relatively low RMSE, indicating high consistency between observed and predicted  
 241 values. In particular, Zn ( $R^2 = 0.88$ ) and Fe ( $R^2 = 0.87$ ) achieved the highest predictive accuracy, suggesting that their spatial  
 242 variability is well captured by the selected environmental predictors. Mn ( $R^2 = 0.84$ ) also showed strong model fits,  
 243 reflecting robust representation of their spatial patterns. The V models displayed slightly lower but still acceptable  
 244 performance ( $R^2 = 0.77$ ), implying that additional predictors may be required to fully capture their local variability. Overall,  
 245 these results demonstrate that the random forest models effectively predict the spatial distributions of major soil  
 246 micronutrients, with particularly high reliability for Fe, Mn, and Zn.

247 **Figure 7** illustrates the spatial patterns of soil micronutrients (Fe, Mn, Zn, and V) across the Tibetan Plateau, as predicted by  
 248 random forest models. The resulting maps reveal significant spatial heterogeneity of these elements. The highest contents of  
 249 Fe are primarily located in the southeastern region, the southern margins, and parts of the western plateau. Mn shows a  
 250 distinct gradient, with contents increasing from northeast to southwest. Predicted Mn values range from 298.47 to 1,110.19  
 251  $\text{mg}\cdot\text{kg}^{-1}$ , and the areas with the highest Mn contents are mainly distributed in the humid southeastern and southern parts of  
 252 the plateau. Zn displays relatively high contents in the central-eastern region and along certain western edges of the plateau,  
 253 with predicted values ranging from 17.09 to 45.03  $\text{mg}\cdot\text{kg}^{-1}$ . Predicted V contents ranged from 40.79 to 82.00  $\text{mg}\cdot\text{kg}^{-1}$ ,  
 254 displaying a clear spatial gradient. Higher contents were predominantly found in the southeastern and southern regions of the

255 plateau, where humid climatic conditions and intense chemical weathering promote V enrichment. Moderate levels occurred  
256 across the central and eastern plateau, while the lowest concentrations were observed in the arid western and northern areas.



257  
258 **Figure 7.** Spatial distribution of soil micronutrients (Fe, Mn, Zn, V) on the Tibetan Plateau.

#### 259 **4 Discussion**

260 We added comparisons of our results with the few available observations on the Tibetan Plateau, as well as broader datasets  
261 from other regions, to better place our findings in a wider context (**Table S2**). In terms of data range and mean values, the  
262 contents and magnitudes of the soil micronutrients we observed are generally consistent with previous reports from the  
263 Plateau and global grassland soils, indicating the reliability of our datasets. For the Plateau, the mean Zn, and V contents in  
264 this study were slightly lower (**Fig. S2**). These discrepancies are more likely due to differences in study regions. Our  
265 sampling covered a much broader area, whereas previous studies focused mainly on local regions such as the Heihe Basin  
266 (Bu et al., 2016). Given the substantial spatial heterogeneity of soil micronutrients across the Tibetan Plateau, such  
267 differences are expected and further highlight the necessity of exploring soil micronutrient patterns at the plateau scale.

268 The vegetation-type contrasts likely reflect ecosystem controls on topsoil (0-10 cm) geochemistry (**Fig. 3**). These contrasts  
269 are consistent with ecosystem-mediated controls on topsoil geochemistry, where denser vegetation generally enhances  
270 near-surface nutrient cycling through greater biomass inputs, rhizosphere processes, and organic matter accumulation  
271 (Hinsinger, 2000). Increased organic inputs and reactive organo-mineral associations can promote retention of Fe, Mn, and

272 Zn and, indirectly, V that is often associated with Fe-bearing phases (Lehmann et al., 2015). In contrast, arid steppe and  
273 desert systems typically exhibit weaker soil development and lower organic inputs, reducing reactive surfaces and binding  
274 sites for micronutrients, resulting in lower surface pools (Kabata-Pendias, 2010). Cropland tends to show intermediate levels,  
275 likely reflecting mixing within the plough layer and partial homogenization of surface concentrations (Alloway, 2008).  
276 Lithological class imposes a first order control on surface soil micronutrient variability (**Fig. 4**), consistent with the  
277 well-established role of parent material in determining baseline micronutrient pools through differences in primary mineral  
278 assemblages and weathering products. In our dataset, acidic metamorphic-derived soils are consistently enriched in Fe, Mn,  
279 Zn, and V relative to several other lithologies, whereas acidic igneous and fluvial facies substrates tend to show lower  
280 contents, with carbonate, eolian facies, and clastic sedimentary rocks generally occupying intermediate levels. Such contrasts  
281 are expected where lithology governs not only the supply of Fe and Mn-bearing minerals but also the abundance of  
282 secondary oxides and clay surfaces formed during pedogenesis, which strongly influence trace-element retention and  
283 depletion intensity at regional scales (Lara et al., 2018; Spinola et al., 2022). The coherent behaviour of V alongside Fe is  
284 also mechanistically plausible because vanadate is efficiently retained by Fe and Mn oxides, making V distributions closely  
285 coupled to the abundance and reactivity of these mineral phases (Abernathy et al., 2022; Larsson et al., 2017).

286 The distribution of soil micronutrients across the Tibetan Plateau demonstrates significant spatial heterogeneity (**Figs. 2 and**  
287 **7**), aligning with patterns observed in Europe for elements such as Zn (Van et al., 2023). This variability is predominantly  
288 governed by the interactions among soil, topography, and climate (**Figs. 5a-d and 6**). Soils act as the principal sink for  
289 micronutrients entering the terrestrial environment through both natural processes and anthropogenic inputs. Consistent with  
290 continental-scale assessments, clay content emerged as the dominant control on the spatial distribution of soil Zn in Europe,  
291 with coarse-textured soils showing markedly lower Zn concentrations (Van et al., 2023). Similar texture effects have been  
292 reported for other elements: soil Cu is positively related to clay and negatively related to sand (Ballabio et al., 2018),  
293 sensitivity analyses show that soil Se increases with increasing clay (Jones et al., 2017), whereas surface-soil Hg decreases  
294 once sand exceeds ~30 % (Ballabio et al., 2021). These patterns agree with our observation that ecosystems developed on  
295 coarse substrates contain the lowest levels of micronutrients. This is because clay serves as an adsorption site for metal  
296 elements, whereas sandy soils are relatively weak adsorbents for micronutrients (Mesquita, 1998). In poorly developed soils  
297 with low clay content, such as those in arid climates, the lack of capacity to retain soil organic matter leads to a significant  
298 reduction in the amount of micronutrients, making them less likely to accumulate (Moreno-Jimenez et al., 2019). In addition  
299 to soil covariates, factors such as topography also play a significant role in the micronutrient content of surface soils on the  
300 Tibetan Plateau. Accumulated local effect (ALE) curves indicate pronounced, nonlinear elevation responses in soil  
301 micronutrients, with Fe and Mn generally decreasing with elevation, whereas Zn and V show threshold-like behaviour,

302 remaining relatively stable at mid-elevations but changing sharply at higher elevations (**Fig. S3**). Contents of Fe, Mn, Zn,  
303 and V increased with slope steepness, particularly below 10°. This pattern suggests that topography governs both physical  
304 and chemical soil-forming processes by controlling erosion, leaching, and moisture dynamics (Jenny, 1941; Montgomery,  
305 2007). Steeper slopes enhance runoff and erosion, removing fine particles enriched in Fe and Mn oxides and promoting  
306 elemental redistribution along hillslopes, while also accelerating bedrock weathering and releasing lithogenic elements such  
307 as Fe and V (Yoo et al., 2006). In contrast, gentler slopes favor organic matter accumulation and reducing conditions that can  
308 immobilize or leach redox-sensitive elements. Overall, slope-driven variations in hydrology, weathering, and redox status  
309 collectively shape the spatial heterogeneity of soil micronutrients across the Plateau (Quesada et al., 2010).

310 Notably, precipitation emerges as the primary predictor for micronutrients such as Fe, Mn, Zn, and V, with all elements  
311 exhibiting characteristic U-shaped responses, with minima occurring at 300-500 mm. This pattern likely reflects distinct  
312 weathering regimes across precipitation gradients. In arid regions (<300 mm), physical weathering processes, including  
313 freeze-thaw fracturing and aeolian erosion, predominate, allowing micronutrients to accumulate near the surface due to  
314 evaporation effects (Pachauri et al., 2014; Moreno-Jimenez et al., 2023). In transitional precipitation regimes (300-500 mm),  
315 intensified chemical weathering occurs; however, leaching fluxes surpass the rates of parent material weathering, leading to  
316 soil elemental depletion (Anderson, 2019; Bluth et al., 1994; Hartmann et al., 2014). In humid regions (>500 mm), enhanced  
317 chemical weathering results in the formation of secondary clay minerals (e.g., montmorillonite, illite), whose negatively  
318 charged surfaces facilitate elemental retention through ionic adsorption and co-precipitation mechanisms (Alloway, 2009).

319 To further verify these nonlinear effects, accumulated local effects analyses were performed and revealed consistent  
320 U-shaped responses for Fe, Mn, Zn, and V, confirming the robustness of the partial dependence results (**Fig. S4**). This pattern  
321 likely reflects a trade-off between weathering intensity, leaching losses, and elemental retention under contrasting  
322 hydrological conditions. Additionally, MAP showed a significant positive correlation with CIA (**Fig. S5**), indicating that  
323 higher precipitation enhances chemical weathering. However, this relationship alone does not fully explain the observed  
324 U-shaped nutrient responses, as direct indicators of leaching fluxes and mineral retention are lacking. We therefore consider  
325 this U-shaped pattern, supported by multiple lines of evidence, as an open question for future investigation, emphasizing the  
326 need for complementary datasets (e.g., redox status, mineral composition) to refine its mechanistic interpretation. Although  
327 lithological differences are detectable for some elements, they do not alter the overall spatial patterns (**Fig. 4**). In summary,  
328 lithology modulates the baseline levels of certain micronutrients but is not the dominant factor shaping their regional  
329 distributions, which are more strongly governed by external environmental drivers such as climate, weathering intensity,  
330 sediment transport, and biotic cycling.

331 Our findings suggest that the aridity index is a significant determinant of soil micronutrient distribution. Specifically,  
332 elemental contents tend to decrease when the aridity index falls below a certain threshold (**Fig. 6**). This trend likely reflects  
333 reduced input or retention of elements under arid conditions. Grouped comparisons across aridity classes (humid, dry  
334 sub-humid, semi-arid, arid, and hyper-arid) were performed. The results show that the medians and means of Fe, Mn, Zn,  
335 and V decrease with increasing aridity (**Fig. S6**). Drought conditions may modify soil redox states, thereby influencing  
336 element speciation, adsorption capacity, mobility, and ultimately, leaching behavior (Brady et al., 2016; Loveland et al., 2003;  
337 Carter et al., 1995). Also, arid environments may indirectly impact micronutrients through alterations in soil pH and soil  
338 organic matter content (Moreno-Jimenez et al., 2019). Previous research has shown that droughts induced by climate change  
339 can restrict the availability of essential micronutrients, such as iron and zinc (Moreno-Jimenez et al., 2019; Bista et al., 2018).  
340 This limitation, along with other adverse effects like diminished water availability, poses substantial threats to vital  
341 ecological processes and services in drylands, including food production (Gupta et al., 2008; Graham, 1991). Furthermore,  
342 ongoing regional warming and associated hydroclimatic shifts may interact with hydrology and redox processes,  
343 organic-matter cycling, and vegetation patterns, thereby altering the availability and spatial variability of certain  
344 micronutrients (Myers et al., 2014; Pachauri et al., 2014).

345 Multiple biogeochemical processes are sensitive to micronutrient availability. Biological nitrogen fixation relies on Mo and  
346 Fe as essential cofactors of nitrogenase, while the methane cycle depends on Ni and Cu through their roles in  
347 methanogenesis and methane oxidation, respectively (Thauer et al., 2019; Stefan et al., 2020). Permafrost thaw and  
348 thermokarst development can further activate Fe-Mn redox cycling, altering metal mobility (Chauhan et al., 2024).  
349 Collectively, soil micronutrients depletion may cascade through nutrient cycling and ecosystem feedbacks, amplifying the  
350 impacts of ongoing environmental change across the Plateau.

## 351 **5 Uncertainty**

352 Although the laboratory XRF measurements used in this study were validated against ICP-MS and generally showed strong  
353 correlations, several limitations should be noted. For Fe, Mn, and V, the two methods produced highly consistent results,  
354 supporting the robustness of the dataset. In contrast, Zn exhibited systematic deviations from the 1:1 line, with XRF tending  
355 to slightly underestimate contents relative to ICP-MS (**Fig. S1**). These deviations likely arise from element-specific detection  
356 sensitivities of XRF. To address this, we calibrated Zn, using the regression relationships between the two methods and then  
357 repeated spatial mapping. The results before and after calibration were not significantly different, confirming that these  
358 deviations do not affect our main findings (**Figs. S7 and S8**). Furthermore, the primary scientific purpose of this dataset is to

359 characterize spatial patterns, ecological gradients, and multivariate relationships across the Plateau, rather than to provide  
360 ultra-precise quantification of element contents at the individual-sample level. For these objectives, XRF-derived contents,  
361 supported by calibration, cross-validation, and transparent uncertainty reporting, provide reliable and ecologically  
362 interpretable information. Nonetheless, we acknowledge that incorporating ICP-based measurements on a larger sample set  
363 would further strengthen the accuracy of the dataset, and future efforts should consider combining multiple analytical  
364 approaches to minimize methodological uncertainties.

365 Our random forest regression models demonstrated robust predictive capability (e.g. cross-validated  $R^2$  range from 0.65 to  
366 0.77 for Fe, Mn, and Zn; **Table S3**), V exhibit weaker predictive performance ( $R^2 = 0.39$ ). The spatial predictions inevitably  
367 contain uncertainties arising from both the Random Forest model and the input datasets. Model-based uncertainty was  
368 quantified as the inter-tree variance among predictions (RF-SD; **Fig. S9**), while additional uncertainty may stem from sparse  
369 meteorological observations and limited soil sampling across the Tibetan Plateau. Large regions in the west (including parts  
370 of the Changtang Plateau and Ngari) are sparsely inhabited and difficult to access, with extensive high-elevation terrain and  
371 access restrictions that constrain fieldwork. Consequently, model estimates in these areas are supported by fewer  
372 observations and should be interpreted with greater caution. We therefore recommend that users consider the accompanying  
373 uncertainty layer when applying the gridded products, especially for analyses focusing on the western Plateau or on  
374 fine-scale local interpretation. Nevertheless, model accuracy could be further improved to more extensive field sampling and  
375 refinement of input data. Specifically, targeted collection of soil micronutrient data and associated covariates in  
376 underrepresented high-altitude regions of the Tibetan Plateau is necessary to address existing spatial gaps. Additionally,  
377 systematic reduction of uncertainties inherent in gridded environmental datasets is essential, as these uncertainties propagate  
378 errors into micronutrient predictions. Continued advancement in both field observations and foundational geospatial dataset  
379 is crucial for improving the reliability of regional-scale element mapping.

380 We note that the reported contents represent a fixed-depth (0-10 cm) surface layer, which may correspond to different  
381 dominant horizons among land uses (organic-rich surface in some forests, mineral topsoil in grasslands/meadows, and the  
382 plough layer in croplands). Users should consider this context when applying the dataset, especially for process inference or  
383 when comparing across ecosystems with contrasting surface horizons. Where applications require deeper profiles or  
384 horizon-specific interpretation, we recommend integrating our maps with local profile data. Because croplands occupy a very  
385 small fraction of the Plateau and are represented by a limited number of samples in our area-weighted design,  
386 cropland-related results should be interpreted primarily as a cross-ecosystem contrast at the plateau scale. Users addressing  
387 cropland-specific questions are advised to use these values with caution and, where possible, complement them with targeted  
388 cropland sampling or independent field observations.

## 389 **6 Data availability**

390 The gridded soil micronutrient (Fe, Mn, Zn, and V) maps for Tibetan Plateau can be downloaded from [https://doi.org/](https://doi.org/10.11888/terre.tpd.303242)  
391 [10.11888/terre.tpd.303242](https://doi.org/10.11888/terre.tpd.303242) (Huo et al., 2025). Users should note that laboratory XRF provides robust estimates for major  
392 elements (e.g., Fe), whereas trace-level elements (particularly Zn) carry higher uncertainty even after ICP-MS calibration.  
393 The dataset is intended primarily for plateau-scale analyses and cross-ecosystem spatial comparisons, caution is advised  
394 when using absolute Zn contents for applications requiring high analytical precision (e.g., bioavailability thresholds, detailed  
395 mass-balance calculations, or benchmarking requiring near-reference values), and users should consult the accompanying  
396 calibration statistics and uncertainty diagnostics.

## 397 **7 Conclusions**

398 This study delivers a comprehensive assessment of spatial distribution patterns for four soil micronutrients (Fe, Mn, Zn, and  
399 V) across the Tibetan Plateau, revealing pronounced regional-scale heterogeneity. Soil covariates were the primary  
400 contributors to the spatial variation of micronutrient contents, followed by topographic and climatic factors, whereas  
401 vegetation and grazing disturbance had relatively minor effects. These findings highlight the coupled effects of climate,  
402 vegetation, and parent material on micronutrient biogeochemical cycling within the complex environmental context of the  
403 Tibetan Plateau. Using five predictor groups (climate, vegetation, soil covariates, topography, and grazing disturbances), we  
404 generated high-resolution spatial maps for key soil micronutrients (Fe, Mn, Zn, and V) via machine learning. These maps  
405 provide validated initial conditions for process-based models simulating micronutrient cycling, advances understanding of  
406 elemental distribution in alpine ecosystems.

### 407 **Author contributions.**

408 JZ conceived the study. HY conducted the field survey and was responsible for data collection and processing. HY prepared  
409 the manuscript with contributions from all co-authors.

### 410 **Competing interests.**

411 The contact author has declared that none of the authors has any competing interests.

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#### 416 **Appendix A: Supplementary material**

#### 417 **Appendix B: Metadata**

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