

Mapping key soil micronutrients across the Tibetan Plateau

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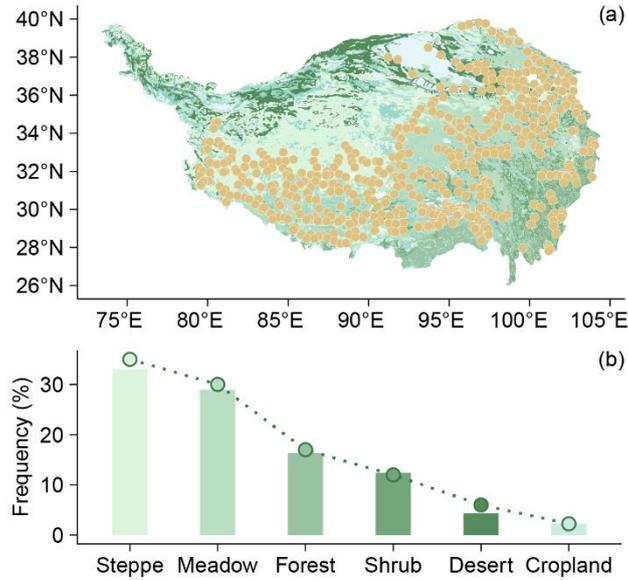
Abstract. Soil micronutrients supply sustains critical ecological functions but exhibit poorly quantified distribution patterns in high-altitude ecosystems. This study bridges this knowledge gap through a large-scale investigation across the Tibetan Plateau, a cold-arid region where cryogenic weathering, aridity, and suppressed pedogenesis interact to govern micronutrient cycling. We assembled a plateau-wide dataset from 526 sites with triplicate surface soils (0-10 cm) per site (n = 1,660). Seven micronutrients (Fe, Mn, Cu, Zn, Ni, V, Mo) were measured and paired with multi-source predictors (climate, vegetation, soil properties, topography, graze disturbance, and weathering proxy). Elemental contents span broad ranges, with site-level summaries (mean \pm SD, mg \cdot kg⁻¹) of Fe 22,864.30 \pm 7,589.01, Mn 576.74 \pm 206.44, Cu 25.32 \pm 9.28, Zn 27.24 \pm 8.55, Ni 49.35 \pm 10.98, V 56.99 \pm 19.33, and Mo 4.63 \pm 1.14. Random Forest model was employed to quantify controls and generate high-resolution spatial maps. Key results reveal that pronounced regional heterogeneity driven primarily by climate-related weathering intensity and topography variable, with secondary modulation from soil and vegetation factors. Element-specific spatial patterns were observed, with Fe enrichment in southeastern/southern plateaus, Mn gradients increasing southeastward and Cu over the western-northern plateau with minima in the southeast. Zn hotspots in central-eastern and western marginal zones, Ni enriched across the northern-central interior and western highlands, and V exhibits a moderate spatial gradient, with higher contents in the southeastern Tibetan Plateau and relatively lower values in the northwest. Mo weakly patterned, slightly higher on northern basins and lower toward the southeast. We provide 1-km maps of all seven micronutrients together with pixel-wise uncertainty layers to support benchmarking of process-based micronutrient cycling models and to inform sustainable ecosystem management under climate change. The dataset is openly available at TPDC (<https://doi.org/10.11888/Terre.tpdc.303049>; Huo et al., 2025).

26 **1 Introduction**

27 As essential yet trace-level components of living systems, micronutrients (e.g., Fe, Mn, Cu, Zn, Ni, V, Mo) sustain
28 fundamental ecological processes, including photosynthesis (Fe, Mn; Fischer et al., 2015; Schmidt et al., 2020), respiration
29 (Fe; Dallman, 1986), enzymatic/redox functions (Cu, Zn, Mn, V; Hänsch et al., 2009), and biological nitrogen fixation (Ni, V,
30 Mo; O'Hara, 2001). Crucially, micronutrient gradients in soils propagate through trophic chains, directly influencing human
31 nutrition and health; deficiencies exacerbate global malnutrition burdens (Fageria et al., 2002; White et al., 2005). Despite
32 their pivotal role in ecosystem stability and food security (Presteele et al., 2016; Stehfest et al., 2019), critical knowledge
33 gaps persist regarding the distribution patterns and drivers of soil micronutrients from regional to global scales.

34 Soil micronutrient supply originates from coupled physicochemical weathering and biological mediation, critically regulated
35 by local climate and topography (Ochoa-Hueso et al., 2020; Hartmann et al., 2023). In cold-arid high-altitude regions,
36 particularly the Tibetan Plateau, extreme environmental interactions uniquely govern micronutrient cycling. Cryogenic
37 processes such as glacial erosion and freeze-thaw cycles, accelerate physical bedrock weathering to mobilize lithogenic
38 micronutrient reservoirs, while aridity concurrently constrains chemical weathering and elemental release (Mu et al., 2020;
39 Mu et al., 2016). Low temperatures suppress biological turnover and synergize with aridity to compromise pedogenesis
40 through clay deficits and diminished mineral reactive sites, thereby reducing elemental retention capacity (Dijkstra et al.,
41 2004). These counteracting processes fundamentally shape micronutrients distribution patterns, yet remain severely
42 understudied. Current research is largely restricted to localized transects (e.g., Heihe River Basin, Tibetan Plateau Highway)
43 with limited spatial representation (Zhang et al., 2012; Guan et al., 2017; Bu et al., 2016).

44 To address these knowledge gaps, we conducted a large-scale field investigation across the Tibetan Plateau, establishing 526
45 sampling sites distributed across representative temperature and moisture gradients (Fig. 1). The sampling design
46 encompassed the plateau's dominant vegetation types and lithological classes. Using this dataset, we analyzed distribution
47 patterns and key controlling factors for seven essential micronutrients (Fe, Mn, Cu, Zn, Ni, V, Mo). We then applied a
48 Random Forest algorithm to generate high-resolution spatial distribution maps of these micronutrients, representing the first
49 comprehensive quantification at this scale and resolution.



50

51 **Figure 1.** Sampling strategy and ecosystem representativeness. (a) Spatial distribution of sampling sites superimposed on China's
 52 1:1,000,000 vegetation map. (b) Areal proportions of ecosystem types (bars) versus sampling point frequency distribution (dots) across
 53 corresponding ecosystems. Similar bar and dot heights indicate that the sampling is proportionally representative.

54 **2 Methods**

55 **2.1 Field survey and soil micronutrients analysis**

56 We analyzed 1,660 topsoil samples collected from 526 locations during a 2019-2021 growing season (July-August) field
 57 survey across the Tibetan Plateau (79-105°E, 27-40°N; **Fig. 1a**). Sampling sites represent major plateau ecosystems: forests,
 58 shrubs, steppes, meadows, deserts, and croplands (**Table 1**). The sites span broad environmental gradients, ranging from 759
 59 to 5565 m in elevation, -7.83 to 18.46 °C in mean annual temperature (MAT), and 23 to 898 mm in mean annual
 60 precipitation (MAP), effectively capturing the plateau's topographic and climatic variability. Site was selected using
 61 standardized criteria: maintaining relative homogeneity in species composition, community structure, and habitat conditions,
 62 and avoiding proximity to roads or areas with frequent human activity. At each location, we established a 15 m transect
 63 collecting triplicate soil samples (0-10 cm depth) at 0 m, 7.5 m, and 15 m positions. This depth was chosen because it
 64 represents the biologically active surface horizon most relevant to plant uptake and microbially mediated cycling, and is
 65 widely used in regional soil inventories for comparability across ecosystems. On the Tibetan Plateau, steppes and meadows
 66 constitute the dominant land cover, so most samples correspond to mineral A horizons; in forested sites, the 0-10 cm layer
 67 may include an organic-rich surface. This consistent protocol ensures cross-site comparability while capturing the variability
 68 of surface layer that most strongly interacts with vegetation and climate. Geographic coordinates, elevation, community type,
 69 and species composition (Cheng et al., 2022) were systematically documented.

70 To ensure the reliability of soil micronutrient measurements, we adopted a two-step analytical strategy. First, in situ
 71 measurements were conducted using a second-generation portable Niton X-ray fluorescence (XRF) analyzer to obtain
 72 preliminary contents under natural field conditions. However, because this instrument could not reliably detect Mo, all soil
 73 samples were subsequently air-dried, sieved (2 mm), and re-analyzed in the laboratory using a third-generation XRF.
 74 Compared with the second-generation device, the third-generation XRF offers an extended detection range and improved
 75 accuracy (Lemière, 2018). All data presented in this study are based on the laboratory XRF measurements.

76 To further validate the laboratory XRF results, a subset of 218 samples was randomly selected and re-analyzed using
 77 Inductively coupled plasma-Mass Spectrometry (ICP-MS), a widely accepted reference method (Simon, 2005). The two
 78 methods showed strong correlations for Fe, Mn, Cu and V ($R^2 = 0.67-0.89$, $P < 0.001$), with values closely aligning with the
 79 1:1 line, indicating reliable quantification. For Zn and Ni, XRF and ICP-MS results were also significantly correlated,
 80 though systematic deviations from the 1:1 line were observed (**Fig. S1**). XRF tended to slightly over- or underestimated
 81 absolute contents of Zn and Ni contents compared with ICP-MS. To address this, we calibrated Zn and Ni, using the
 82 regression relationships between the two methods and then repeated spatial mapping. For Mo, the results obtained from the
 83 two methods were not fully consistent, likely due to its low content. Comparative results of both methods are provided in the
 84 Supplementary Information (**Figs. S2-S4**).

85 **Table 1.** Ecosystem classification and sampling coverage on the Tibetan Plateau.

Biome	Characteristics and dominant plant species	elevation ranges	climate conditions (temperature/precipitation)	No. of samples	No. of locations
Steppe	Alpine steppes, dominated by cold-adapted herbaceous species such as <i>Stipa purpurea</i> , features sparse vegetation adapted to cold-arid conditions.	1961-5151 m	-7.1-8.0 °C; 34.32-785.61 mm	569	180
Meadow	Alpine meadows feature dense, low-stature vegetation sustained by year-round low temperatures, high humidity, and water-retentive soils. These ecosystems thrive on gentle slopes and valley floors at higher elevations, hosting relatively diverse flora with characteristic dominance of sedges including <i>Kobresia pygmaea</i> and <i>K. humilis</i> .	2661-5565 m	-7.8-9.2 °C; 158.42-848.70 mm	499	154
Forest	Forests on the Tibetan Plateau concentrate primarily in the southeastern region, dominated by high-altitude cold-temperate coniferous forests. These humid-adapted ecosystems feature <i>fir (Abies)</i> and <i>spruce (Picea)</i> species as characteristic components.	1453-4237 m	-0.6-16.5 °C; 400.34-898.48 mm	265	87
Shrub	Tibetan shrublands primarily occur in arid and alpine zones, characterized by low-growing, drought-tolerant dwarf shrubs such as <i>Lonicera (honeysuckle)</i> and <i>Rhododendron</i> species adapted to nutrient-poor soils and extreme climatic conditions.	2169-5022 m	-5.2-11.8 °C; 301.40-868.63 mm	190	64

Desert	Alpine deserts occur in extremely arid, cold regions and exhibit extremely sparse vegetation dominated by <i>arid-tolerant dwarf shrubs</i> and <i>herbs</i> .	2108-5158 m	-7.1-5.6 °C; 22.64-443.43 mm	101	29
Cropland	Cropland, concentrated in river valleys and basin floors and is dominated by highland barley (<i>Hordeum vulgare</i> var. <i>nudum</i> , “qingke”), with spring wheat (<i>Triticum aestivum</i>), rapeseed (<i>Brassica napus</i>), and potato (<i>Solanum tuberosum</i>) commonly cultivated; vegetation cover is strongly seasonal and often bare after harvest.	759-4360 m	0.5-18.4 °C; 112.53-783.06 mm	36	12

86 Biomes are grouped by diagnostic characteristics; dominant plant species are italicized. Elevation range gives the minimum-maximum
87 elevation (m) of sampling sites within each biome. Climate conditions report site-level ranges of mean annual temperature (°C) and mean
88 annual precipitation (mm). No. of samples is the number of soil samples analyzed (0-10 cm, three replicates per site), and No. of locations
89 is the number of unique sampling sites.

90 2.2 Soil Properties

91 Soil samples were sifted through 2 mm sieve, discarding visible stones and extracted roots. Soil pH was measured using the
92 potentiometric method, and soil texture analysis, quantifying clay, silt, and sand content fractions, was determined using a
93 laser diffraction particle size analyzer (Mastersizer 2000, Malvern, UK). The sieved samples were air-dried for elemental
94 analysis. Soil organic carbon (SOC) content was quantified via the potassium dichromate oxidation method (Walkley-Black)
95 with external heating. The chemical index of alteration (CIA) was calculated using the molar proportions of Al₂O₃, CaO*,
96 Na₂O, and K₂O according to the formula: $CIA = [Al_2O_3 / (Al_2O_3 + CaO^* + Na_2O + K_2O)] \times 100$ (Fedo et al., 1995). The
97 chemical index of alteration reflects the relative loss of mobile base cations (Ca, Na, K) compared with the enrichment of
98 immobile Al during weathering (McLennan, 1993). Higher chemical index of alteration values indicate stronger chemical
99 weathering and more advanced soil development, whereas lower values suggest weaker weathering and limited leaching of
100 base cations (Nesbitt et al., 1982).

101 2.3 Environmental variables

102 We considered geographic, climatic, biological, and edaphic drivers. Field measurements provided location (longitude,
103 latitude), slope, aspect and vegetation type. For the spatial prediction of micronutrient distributions, however, the gridded
104 predictor variables (e.g., elevation, slope, aspect, vegetation cover) were derived from online datasets in order to upscale
105 site-level observations to the Plateau scale. Slope and aspect data came from the National Tibetan Plateau Data Center
106 (<https://data.tpc.ac.cn>). The digital elevation model (DEM) data were collected from the Resource and Environment
107 Science and Data Center (<https://www.resdc.cn/>). Climate variables, mean annual temperature (MAT) and mean annual
108 precipitation (MAP) were downloaded from the Climate Data Store (<https://cds.climate.copernicus.eu/#!/home>). The Aridity
109 Index (AI), calculated as mean annual precipitation/mean annual reference evapotranspiration, was obtained from the Global

110 Aridity Index dataset (Trabucco et al., 2018), where higher values indicate greater humidity. Vegetation types (Forest, Shrub,
111 Meadow, Steppe, Desert, Cropland) followed the 1:1,000,000 China Vegetation Map classification (Hou, 2019). The
112 normalized difference vegetation index (NDVI) data were obtained from an Earthdata Search (<https://search.earthdata.nasa.gov/search>). The net primary productivity (NPP) data were obtained from the study by Chen et al. (2023) and
113 were calculated using the CASA model (Potter et al., 1993). The grazing activity data were obtained from statistical
114 yearbooks. Based on the lithological data published by Dijkshoorn et al. in 2018, the rock types on the Tibetan Plateau were
115 classified into acidic igneous rock (IA), acidic metamorphic rock (MA), clastic sedimentary rock (SC), carbonate rock (SO),
116 aeolian facies rock (UE), and fluvial facies rock (UF).

118 **2.4 Relative importance analysis and soil micronutrient mapping**

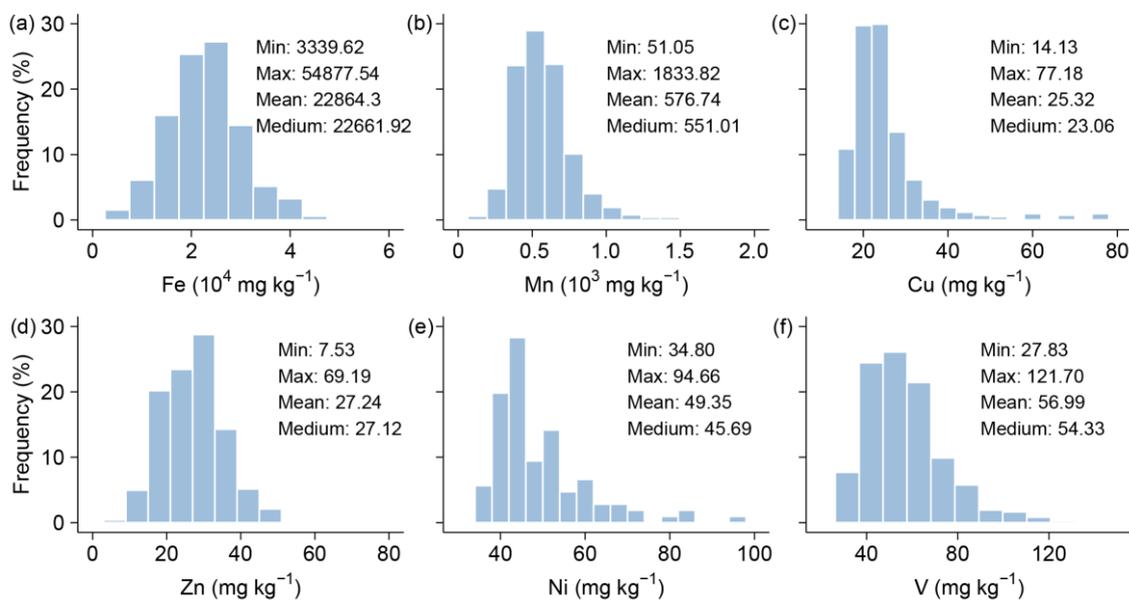
119 Soil micronutrient measurements were preprocessed to detect and remove outliers exceeding the mean \pm 3 standard
120 deviations. After data screening, the retained sample sizes are: Fe 1654 (from 1660), Mn 1630 (from 1646), Cu 920 (from
121 946), Zn 1655 (from 1661), Ni 180 (from 181), V 857 (from 867), and Mo 120 (from 123). To evaluate the relative
122 importance of predictors in explaining soil micronutrient variability across the Tibetan Plateau, we applied the random forest
123 permutation importance (%IncMSE) to independently cross-validate the driver rankings. For spatial prediction, we
124 developed seven area-wide random forest models (each comprising 500 trees) targeting Fe, Mn, Cu, Zn, Ni, V, and Mo
125 contents. The models were trained using a suite of environmental predictors, including topographic features (DEM, slope,
126 aspect), climate variables (MAT, MAP, AI), vegetation indices (NDVI, NPP), soil properties (texture, SOC, pH, CIA), and
127 grazing intensity. Random forest was selected for its ability to model complex, nonlinear relationships and interactions
128 among diverse types of predictors. Model hyperparameters were optimized using grid search combined with tenfold
129 cross-validation. To assess model generalizability, we examined the extent to which the predictor parameter space in the
130 validation set overlapped with that of the original training data. Model performance was evaluated by comparing predicted
131 versus observed values using scatterplots (predicted on the x-axis, observed on the y-axis) following the method of Piñeiro et
132 al. (2008), with models achieving strong predictive performance ($R^2 = 0.6-0.7$). The uncertainty layer is calculated as the
133 inter-pixel variance among tree predictions. All statistical analyses were conducted using R version 3.4.4.

134 **3 Results**

135 **3.1 Soil micronutrients contents**

136 Across all sites, soil micronutrient contents varied widely, with mean value of $22,864.30 \pm 7,589.01$ for Fe (mean \pm SD,
137 $\text{mg}\cdot\text{kg}^{-1}$), 576.74 ± 206.44 for Mn, 25.32 ± 9.28 for Cu, 27.24 ± 8.55 for Zn, 49.35 ± 10.98 for Ni, 56.99 ± 19.33 for V, and

138 4.63 ± 1.14 for Mo. Coefficients of variation (CV = SD/mean) were 33% for Fe, 36% for Mn, 37% for Cu, 31% for Zn, 22%
 139 for Ni, 34% for V and 25% for Mo. Collectively, Fe and Mn dominate in absolute abundance, whereas Cu-Zn-Ni-V occur at
 140 tens of mg·kg⁻¹ and Mo remains low (~5 mg·kg⁻¹), indicating heterogeneous but orderly micronutrient levels across the
 141 Plateau (**Figs. 2 and S5**).



142
 143 **Figure 2.** Frequency distributions of soil micronutrients (Fe, Mn, Cu, Zn, Ni, V) across the Tibetan Plateau.

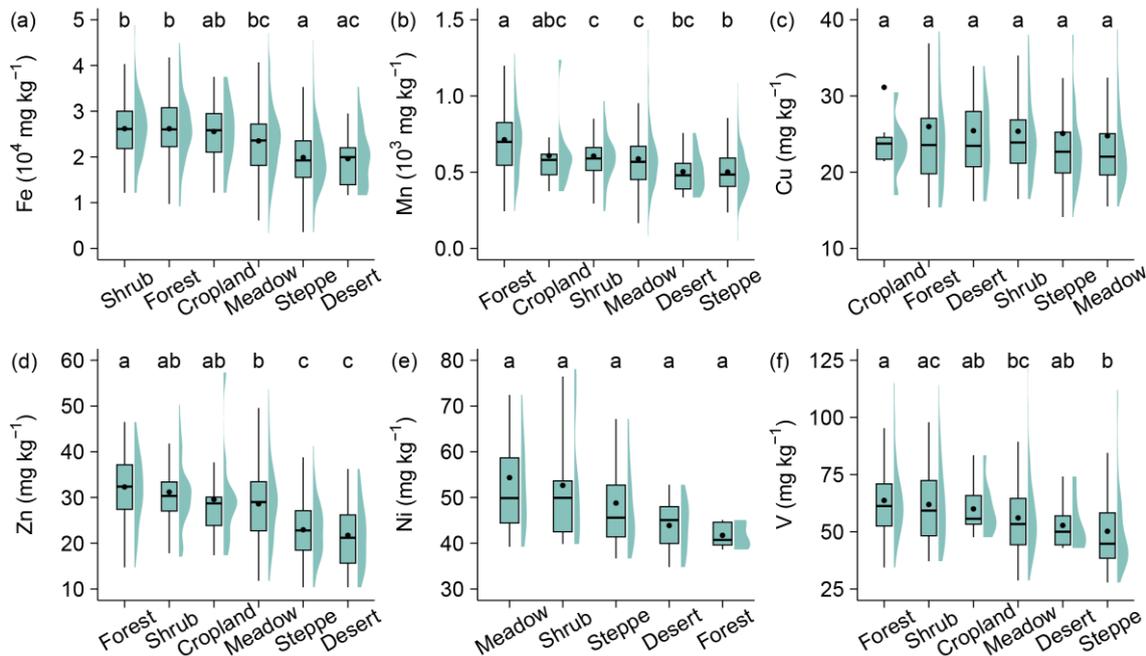
144 3.2 Soil micronutrients across vegetation types

145 Soil micronutrients contents (Fe, Mn, Cu, Zn, Ni, V, and Mo) varied significantly among different vegetation types (**Figs. 3**
 146 **and S6**). Fe contents were highest in shrub and forest ecosystems, with mean values of 26,264.11 mg·kg⁻¹ and 26,090.66
 147 mg·kg⁻¹, respectively, significantly exceeding values observed in desert (19,762.66 mg·kg⁻¹) and steppe ecosystems
 148 (19,852.37 mg·kg⁻¹) by 31-33%. Similarly, Mn contents were greatest in forest (703.22 mg·kg⁻¹) and cropland (608.29
 149 mg·kg⁻¹), followed by shrub (606.33 mg·kg⁻¹) and meadow (591.59 mg·kg⁻¹), while the lowest Mn levels were recorded in
 150 desert and steppe ecosystems (both below 510 mg·kg⁻¹).

151 Soil Cu showed minimal variation across vegetation types, with mean values ranging narrowly between 25.23 and 31.14
 152 mg·kg⁻¹. Cropland had slightly higher Cu levels, but the differences were not statistically significant, suggesting a relatively
 153 uniform spatial distribution of Cu across ecosystems. Zn, however, demonstrated a strong vegetation-dependent variability,
 154 with the highest mean contents found in forest (32.00 mg·kg⁻¹) and shrub (31.15 mg·kg⁻¹) ecosystems. In contrast, Zn levels
 155 were markedly lower in steppe (22.94 mg·kg⁻¹) and desert (21.95 mg·kg⁻¹) soils, with differences exceeding 10 mg·kg⁻¹,
 156 indicating pronounced biogeochemical variation.

157 In contrast to the previous elements, Ni and Mo exhibited distinct patterns across vegetation types, with notably lower
 158 contents in forest and shrub ecosystems. Ni contents were highest in meadow soils ($54.34 \text{ mg}\cdot\text{kg}^{-1}$), followed by shrub
 159 ($52.66 \text{ mg}\cdot\text{kg}^{-1}$), steppe ($48.78 \text{ mg}\cdot\text{kg}^{-1}$), cropland ($48.22 \text{ mg}\cdot\text{kg}^{-1}$), desert ($43.87 \text{ mg}\cdot\text{kg}^{-1}$), and forest ecosystems (41.73
 160 $\text{mg}\cdot\text{kg}^{-1}$). Soil Mo contents showed negligible differences among ecosystems, though cropland exhibited marginally higher
 161 levels. No statistically significant differences were observed between vegetation types, indicating minimal vegetation control
 162 over Mo distribution.

163 The highest soil V mean value occurred in forest soils ($63.70 \text{ mg}\cdot\text{kg}^{-1}$), followed by shrub ($61.94 \text{ mg}\cdot\text{kg}^{-1}$) and cropland
 164 ecosystems ($60.02 \text{ mg}\cdot\text{kg}^{-1}$). These values were markedly greater than those observed in steppe ($50.22 \text{ mg}\cdot\text{kg}^{-1}$) and desert
 165 ($52.80 \text{ mg}\cdot\text{kg}^{-1}$) soils, with differences reaching up to 25%. Meadow soils showed intermediate V levels ($56.07 \text{ mg}\cdot\text{kg}^{-1}$).
 166 Statistical comparisons revealed that V contents in forest and shrub ecosystems were significantly higher than those in steppe
 167 and meadow ecosystems ($p < 0.05$).

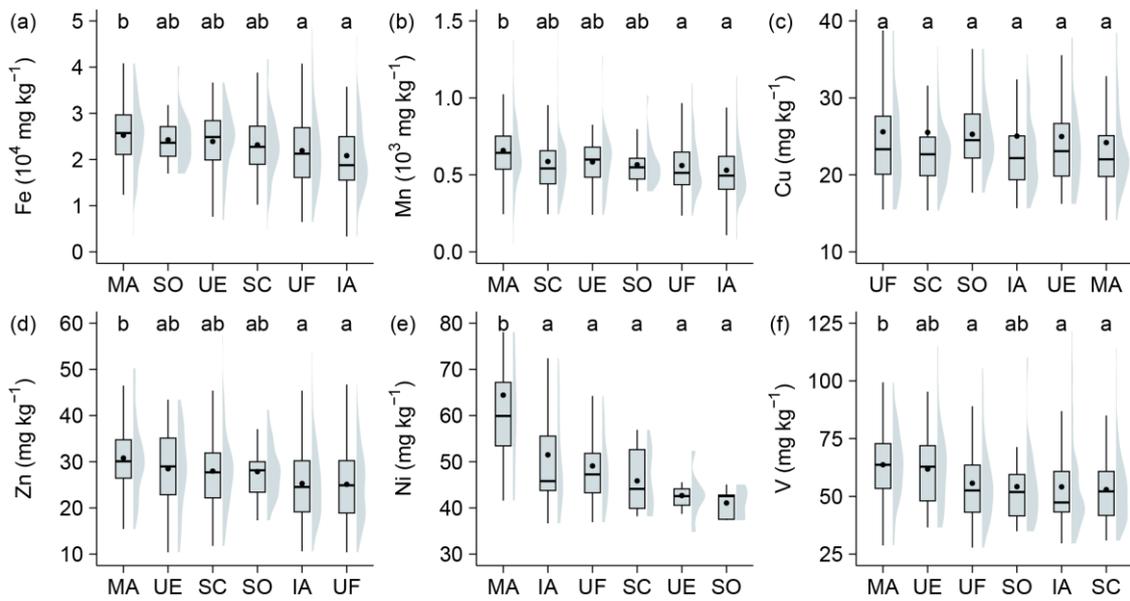


168
 169 **Figure 3.** Variability in soil micronutrient contents (Fe, Mn, Cu, Zn, Ni, V) across Tibetan vegetation types. Boxplots show the data
 170 distributions for each vegetation type. Within each plot, the boxes represent the interquartile range (IQR), the horizontal lines within boxes
 171 indicate the median values, and black dots denote the mean values. The whiskers extend to 1.5 times the IQR. The surrounding shaded
 172 violin shapes indicate the kernel density distribution of the data. Lowercase letters denote significance groups from Tukey's HSD multiple
 173 comparisons ($\alpha = 0.05$): groups sharing at least one letter do not differ significantly, whereas groups with no letters in common differ
 174 significantly (e.g., "ab" is not different from "a" or "b").

175 3.3 Soil micronutrients across lithological classes

176 Lithological class exerted a significant influence on the spatial distribution of certain soil micronutrients (**Figs. 4 and S7**). Fe
 177 contents differed notably across lithological classes, with the highest values observed in soils derived from acidic

178 metamorphic rocks (mean value: 25,252.72 mg·kg⁻¹), followed by those from carbonate rocks (24,260.96 mg·kg⁻¹) and
 179 eolian facies rocks (23,902.77 mg·kg⁻¹). The lowest Fe contents were recorded in soils developed from acidic igneous rocks
 180 (20,830.15 mg·kg⁻¹). A similar geochemical pattern was observed for Mn, with the maximum contents in
 181 acidic-metamorphic-driven soils (658.37 mg·kg⁻¹), and the lowest in acidic-igneous-driven soils (530.45 mg·kg⁻¹).
 182 Zn contents in soils also varied across lithologies, with the highest values associated with acidic metamorphic rocks (mean
 183 value of 30.79 mg·kg⁻¹), followed by eolian facies rocks and clastic sedimentary rocks, while relatively lower levels were
 184 found under acidic igneous rocks (25.27 mg·kg⁻¹) and fluvial facies rocks (25.10 mg·kg⁻¹). In contrast, Cu showed no
 185 significant differences among lithological classes. Cu contents were slightly higher under fluvial facies rocks (25.60
 186 mg·kg⁻¹), but overall differences were minimal. Similarly, Mo contents were relatively uniform, with slightly elevated levels
 187 under acidic metamorphic rocks and eolian facies rocks (both ~4.86 mg·kg⁻¹), and lower values under clastic sedimentary
 188 rocks and carbonate rocks (both <4.48 mg·kg⁻¹), though differences were not statistically significant.



189
 190 **Figure 4.** Variability in soil micronutrient contents (Fe, Mn, Cu, Zn, Ni, V) across Tibetan lithological classes. Boxplots show the data
 191 distributions for each lithological classes. Within each plot, the boxes represent the interquartile range (IQR), the horizontal lines within
 192 boxes indicate the median values, and black dots denote the mean values. The whiskers extend to 1.5 times the IQR. The surrounding
 193 shaded violin shapes indicate the kernel density distribution of the data. Lowercase letters denote significance groups from Tukey's HSD
 194 multiple comparisons ($\alpha = 0.05$): groups sharing at least one letter do not differ significantly, whereas groups with no letters in common
 195 differ significantly (e.g., "ab" is not different from "a" or "b"). Abbreviations of lithological classes: IA = acidic igneous rock, MA = acidic
 196 metamorphic rock, SC = clastic sedimentary rock, SO = carbonate rock, UE = eolian facies rock, UF = fluvial facies rock.

197 Soil Ni contents showed the most pronounced variation among lithologies, with the highest level recorded under acidic
 198 metamorphic rocks (64.43 mg·kg⁻¹), followed by acidic igneous rocks (51.49 mg·kg⁻¹). Soils developed from clastic
 199 sedimentary rocks and carbonate rocks had significantly lower Ni contents (45.85 and 41.06 mg·kg⁻¹, respectively). The

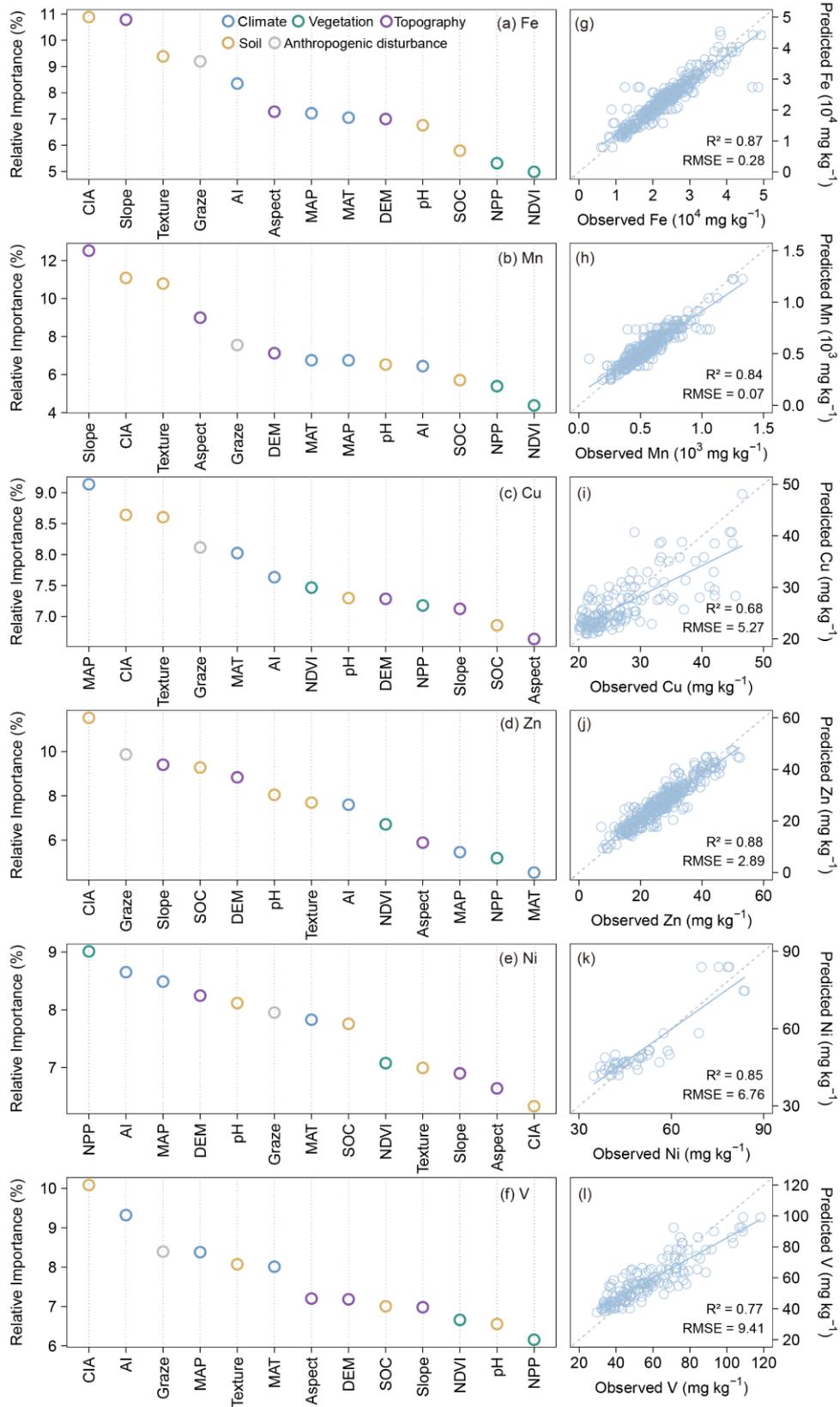
200 highest V contents were found in soils derived from acidic metamorphic rocks (63.67 mg·kg⁻¹), followed by soils developed
201 from eolian facies rocks (61.81 mg·kg⁻¹) and fluvial facies rocks (55.67 mg·kg⁻¹). In contrast, soils originating from
202 carbonate (54.20 mg·kg⁻¹), acidic igneous (54.11 mg·kg⁻¹), and clastic sedimentary rocks (52.92 mg·kg⁻¹) exhibited
203 relatively lower V levels. In summary, Fe, Mn, Zn and V were substantially enriched in soils derived from acidic
204 metamorphic rocks. Cu and Mo showed relatively uniform distributions across lithologies, Ni was notably elevated under
205 acidic metamorphic rocks.

206 **3.4 Drivers of soil micronutrient pattern**

207 Relative importance analysis was conducted for five predictor groups, including climate, vegetation, soil properties,
208 topography, and grazing disturbances. The rankings show clear element-specific controls (**Figs. 5a-f and S8**). Soil properties
209 were the primary contributors to the spatial variation of micronutrient contents, followed by topographic and climatic factors,
210 whereas vegetation and anthropogenic disturbance had relatively minor effects. Among all predictors, the chemical index of
211 alteration, slope, and soil texture exhibited the highest importance, indicating that weathering intensity, terrain, and soil
212 physical structure are the dominant controls on the spatial distribution of soil micronutrients. Specifically, the chemical index
213 of alteration emerged as the dominant factor for Fe, Zn, and V. Topographic factors (slope) primarily influenced Mn, while
214 climatic variables (MAP, AI) and biological productivity (NPP) contributed more to Cu and Ni variations. Mo is mainly
215 influenced by the soil texture.

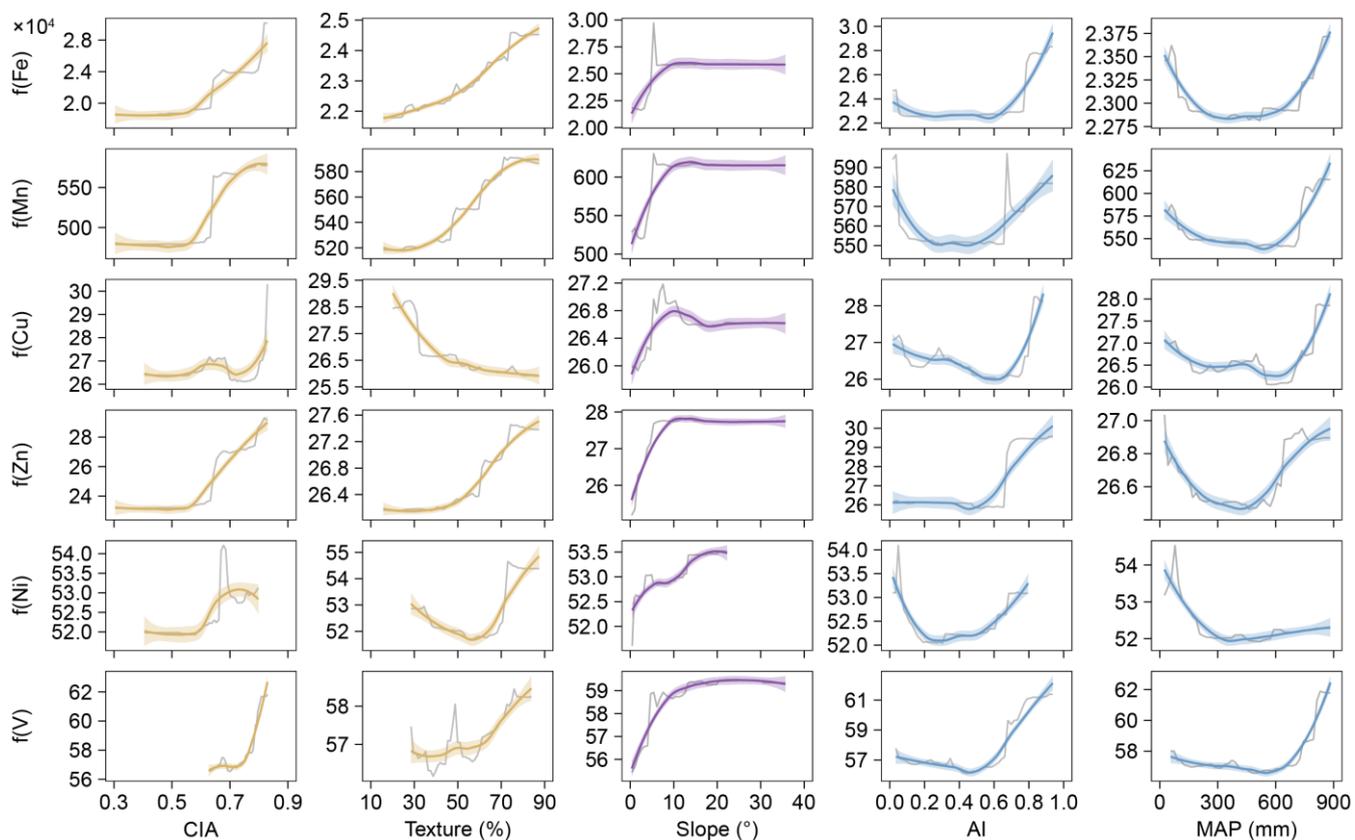
216 The partial dependence plots illustrated the nonlinear responses of seven soil micronutrients (Fe, Mn, Cu, Zn, Ni, V, and Mo)
217 to five key environmental variables (**Figs. 6 and S9**). Overall, the effects of the chemical index of alteration, soil texture,
218 slope, and aridity index varied substantially among elements. The chemical index of alteration showed a generally positive
219 relationship with most micronutrients, particularly Fe, Mn, Zn, V, and Mo, which increased markedly when chemical index
220 of alteration exceeded approximately 0.5. Soil texture exerted a moderate positive influence on micronutrients. In contrast,
221 slope exhibited a saturating effect—nutrient concentrations rose sharply at low slope angles (<10°) and plateaued
222 thereafter—highlighting the limited accumulation potential on steeper terrains. The aridity index demonstrated a clear
223 U-shaped pattern for most micronutrients (Fe, Mn, Cu, Zn, Ni, V, and Mo), with relatively low concentrations under
224 intermediate aridity. Seven micronutrients (Fe, Mn, Cu, Zn, Ni, V, and Mo) exhibited a typical U-shaped relationship with
225 mean annual precipitation, showing higher contents in both low and high precipitation zones, and a clear trough in the
226 intermediate range (approximately 300-500 millimeters). This pattern aligned closely with responses to drought indices,
227 confirming shared moisture sensitivity. Collectively, these patterns emphasize that soil micronutrient distributions across the

228 Tibetan Plateau are strongly shaped by interactions among chemical weathering intensity, soil physical properties, and
 229 climatic aridity.



230

231 **Figure 5.** Relative importance of biotic and abiotic factors for soil micronutrients (Fe, Mn, Cu, Zn, Ni, V) on the Tibetan Plateau (a-f).
 232 Relationship between observed and predicted values of soil micronutrients (Fe, Mn, Cu, Zn, Ni, V) on the Tibetan Plateau based on the
 233 Random Forest model (g-l). The blue solid line represents the fitted relationship using ordinary least squares regression, while the gray
 234 dashed line indicates the 1:1 line between observed and predicted values.



235
 236 **Figure 6.** Partial dependence of six soil micronutrients (Fe, Mn, Cu, Zn, Ni, V) on four predictive variables: chemical index of alteration
 237 (CIA), Texture, Slope, aridity index (AI) and mean annual precipitation (MAP). Gray lines represent the original partial dependence,
 238 while smoothed fits for each micronutrient are shown as colored lines. Shaded areas represent the 95% confidence intervals.

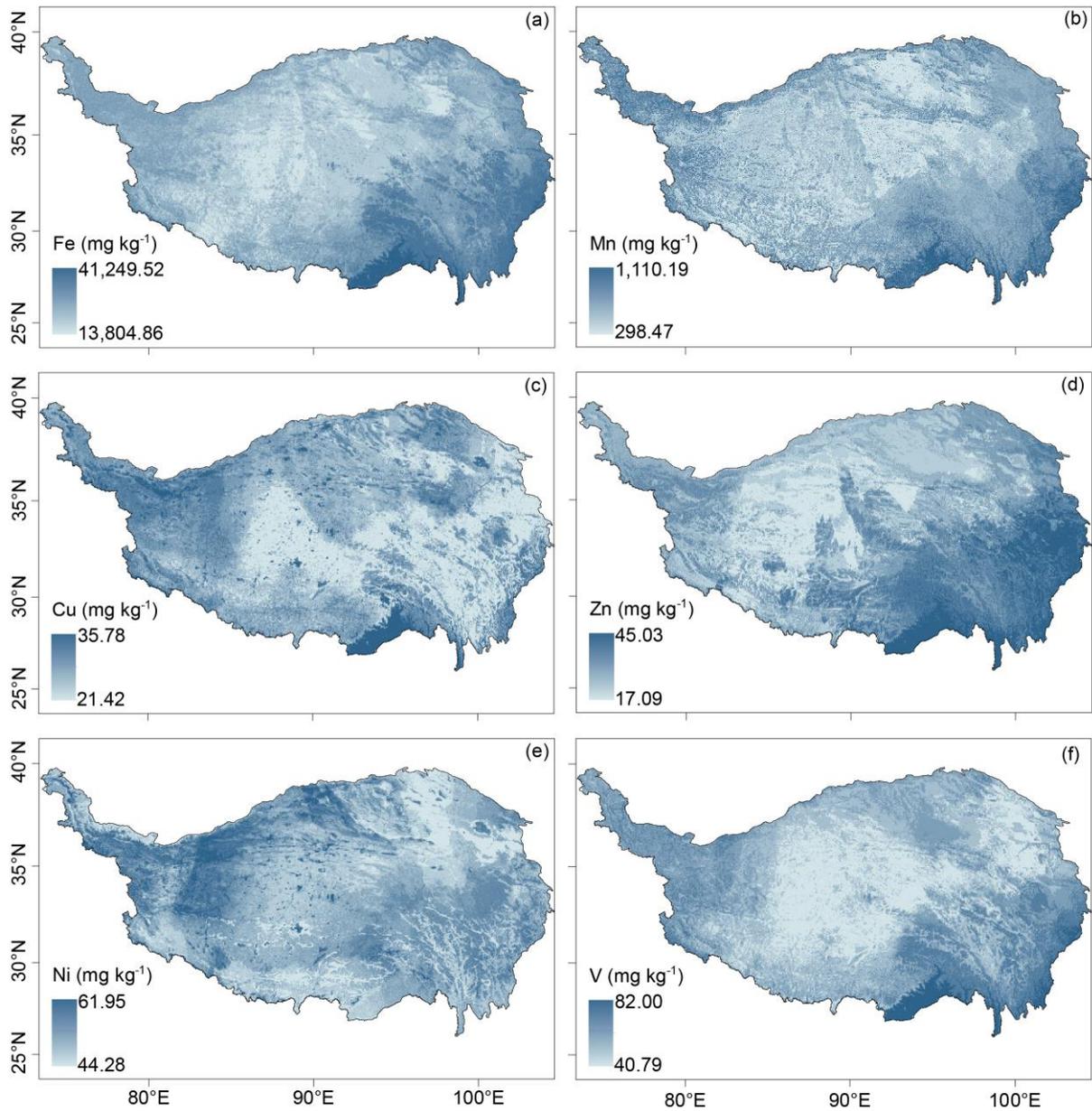
239 3.5 Soil micronutrients maps

240 The random forest (RF) models exhibited strong overall predictive performance for soil micronutrient concentrations across
 241 the Tibetan Plateau, with varying accuracies among elements (**Figs. 5g-l and S8**). The models for Fe, Mn, Zn, Ni, and Mo
 242 performed best, with R^2 values ranging from 0.83 to 0.88 and relatively low RMSE, indicating high consistency between
 243 observed and predicted values. In particular, Zn ($R^2 = 0.88$) and Fe ($R^2 = 0.87$) achieved the highest predictive accuracy,
 244 suggesting that their spatial variability is well captured by the selected environmental predictors. Mo ($R^2 = 0.83$), Mn ($R^2 =$
 245 0.84) and Ni ($R^2 = 0.85$) also showed strong model fits, reflecting robust representation of their spatial patterns. The Cu and
 246 V models displayed slightly lower but still acceptable performance ($R^2 = 0.68$ and 0.77 , respectively), implying that
 247 additional predictors may be required to fully capture their local variability. Overall, these results demonstrate that the

248 random forest models effectively predict the spatial distributions of major soil micronutrients, with particularly high
249 reliability for Fe, Mn, Zn, and Ni.

250 **Figure 7 and Figure S10** illustrates the spatial patterns of soil micronutrients (Fe, Mn, Cu, Zn, Ni, V, and Mo) across the
251 Tibetan Plateau, as predicted by random forest models. The resulting maps reveal significant spatial heterogeneity of these
252 elements. The highest contents of Fe are primarily located in the southeastern region, the southern margins, and parts of the
253 western plateau. Mn shows a distinct gradient, with contents increasing from northeast to southwest. Predicted Mn values
254 range from 298.47 to 1,110.19 mg·kg⁻¹, and the areas with the highest Mn contents are mainly distributed in the humid
255 southeastern and southern parts of the plateau. Cu ranges from 21.42 to 35.78 mg·kg⁻¹, with relatively higher contents over
256 the western-northern interior and highlands and lower values in the humid southeastern valleys and along the southern
257 margin. Zn displays relatively high contents in the central-eastern region and along certain western edges of the plateau, with
258 predicted values ranging from 17.09 to 45.03 mg·kg⁻¹. Ni has a narrower predicted content range (44.28-61.95 mg·kg⁻¹) and
259 shows a more spatially homogeneous distribution. However, localized hotspots of elevated Ni contents are observed in parts
260 of the northeastern and southern plateau. Predicted V contents ranged from 40.79 to 82.00 mg·kg⁻¹, displaying a clear spatial
261 gradient. Higher contents were predominantly found in the southeastern and southern regions of the plateau, where humid
262 climatic conditions and intense chemical weathering promote V enrichment. Moderate levels occurred across the central and
263 eastern plateau, while the lowest concentrations were observed in the arid western and northern areas. In contrast, Mo varies
264 within a narrower interval (3.88-5.87 mg·kg⁻¹) and exhibits a weak plateau-wide gradient, with slight enrichment on the
265 northern plateau/basin areas and comparatively lower levels toward the southeast.

266



267
268 **Figure 7.** Spatial distribution of soil micronutrients (Fe, Mn, Cu, Zn, Ni, V) on the Tibetan Plateau.

269 **4 Discussion**

270 We added comparisons of our results with the few available observations on the Tibetan Plateau, as well as broader datasets
 271 from other regions, to better place our findings in a wider context (**Table S1**). In terms of data range and mean values, the
 272 contents and magnitudes of the soil micronutrients we observed are generally consistent with previous reports from the
 273 Plateau and global grassland soils, indicating the reliability of our datasets. For the Plateau, the mean Zn, and V contents in
 274 this study were slightly lower, while Ni were somewhat higher (**Fig. S11**). These discrepancies are more likely due to
 275 differences in study regions. Our sampling covered a much broader area, whereas previous studies focused mainly on local
 276 regions such as the Heihe Basin (Bu et al., 2016). Given the substantial spatial heterogeneity of soil micronutrients across the

277 Tibetan Plateau, such differences are expected and further highlight the necessity of exploring soil micronutrient patterns at
278 the plateau scale.

279 The distribution of soil micronutrients across the Tibetan Plateau demonstrates significant spatial heterogeneity (**Figs. 2, 7,**
280 **S4, and S10**), aligning with patterns observed in Europe for elements such as Zn (Van et al., 2023). This variability is
281 predominantly governed by the interactions among soil, topography, and climate (**Figs. 5a-f, 6 S8 and S9**). Soils act as the
282 principal sink for micronutrients entering the terrestrial environment through both natural processes and anthropogenic
283 inputs. Consistent with continental-scale assessments, clay content emerged as the dominant control on the spatial
284 distribution of soil Zn in Europe, with coarse-textured soils showing markedly lower Zn concentrations (Van et al., 2023).
285 Similar texture effects have been reported for other elements: soil Cu is positively related to clay and negatively related to
286 sand (Ballabio et al., 2018), sensitivity analyses show that soil Se increases with increasing clay (Jones et al., 2017), whereas
287 surface-soil Hg decreases once sand exceeds ~30 % (Ballabio et al., 2021). These patterns agree with our observation that
288 ecosystems developed on coarse substrates contain the lowest levels of micronutrients. This is because clay serves as an
289 adsorption site for metal elements, whereas sandy soils are relatively weak adsorbents for micronutrients (Mesquita, 1998).

290 In poorly developed soils with low clay content, such as those in arid climates, the lack of capacity to retain soil organic
291 matter leads to a significant reduction in the amount of micronutrients, making them less likely to accumulate
292 (Moreno-Jimenez et al., 2019). In addition to soil properties, factors such as topography also play a significant role in the
293 micronutrient content of surface soils on the Tibetan Plateau. Elevation-related variations in soil micronutrient are limited,
294 and some elements (e.g., Cu, Mo) remain relatively stable; Fe, Mn, Zn, and V show similar patterns, with higher contents at
295 lower elevations than at higher ones (**Fig. S12 and S13**). Contents of Fe, Mn, Zn, and V increased with slope steepness,
296 particularly below 10°, whereas Cu and Ni showed weaker or nonlinear trends. This pattern suggests that topography
297 governs both physical and chemical soil-forming processes by controlling erosion, leaching, and moisture dynamics (Jenny,
298 1941; Montgomery, 2007). Steeper slopes enhance runoff and erosion, removing fine particles enriched in Fe and Mn oxides
299 and promoting elemental redistribution along hillslopes, while also accelerating bedrock weathering and releasing lithogenic
300 elements such as Fe and V (Yoo et al., 2006). In contrast, gentler slopes favor organic matter accumulation and reducing
301 conditions that can immobilize or leach redox-sensitive elements. Overall, slope-driven variations in hydrology, weathering,
302 and redox status collectively shape the spatial heterogeneity of soil micronutrients across the Plateau (Quesada et al., 2010).

303 Notably, precipitation emerges as the primary predictor for micronutrients such as Fe, Mn, Cu, Zn, V, and Mo, with all
304 elements exhibiting characteristic U-shaped responses, with minima occurring at 300-500 mm. This pattern likely reflects
305 distinct weathering regimes across precipitation gradients. In arid regions (<300 mm), physical weathering processes,
306 including freeze-thaw fracturing and aeolian erosion, predominate, allowing micronutrients (e.g., Ni) to accumulate near the

307 surface due to evaporation effects (Pachauri et al., 2014; Moreno-Jimenez et al., 2023). In transitional precipitation regimes
308 (300-500 mm), intensified chemical weathering occurs; however, leaching fluxes surpass the rates of parent material
309 weathering, leading to soil elemental depletion (Anderson, 2019; Bluth et al., 1994; Hartmann et al., 2014). In humid regions
310 (>500 mm), enhanced chemical weathering results in the formation of secondary clay minerals (e.g., montmorillonite, illite),
311 whose negatively charged surfaces facilitate elemental retention through ionic adsorption and co-precipitation mechanisms
312 (Alloway, 2009). To further verify these nonlinear effects, accumulated local effects (ALE) analyses were performed and
313 revealed consistent U-shaped responses for Fe, Mn, Cu, Zn, V, and Mo, confirming the robustness of the partial dependence
314 results (**Figs. S14 and S15**). This pattern likely reflects a trade-off between weathering intensity, leaching losses, and
315 elemental retention under contrasting hydrological conditions. Additionally, MAP showed a significant positive correlation
316 with the chemical index of alteration (**Fig. S16**), indicating that higher precipitation enhances chemical weathering. However,
317 this relationship alone does not fully explain the observed U-shaped nutrient responses, as direct indicators of leaching fluxes
318 and mineral retention are lacking. We therefore consider this U-shaped pattern, supported by multiple lines of evidence, as an
319 open question for future investigation, emphasizing the need for complementary datasets (e.g., redox status, mineral
320 composition) to refine its mechanistic interpretation. Although lithological differences are detectable for some elements, they
321 do not alter the overall spatial patterns (**Figs. 4 and S7**). In summary, lithology modulates the baseline levels of certain
322 micronutrients but is not the dominant factor shaping their regional distributions, which are more strongly governed by
323 external environmental drivers such as climate, weathering intensity, sediment transport, and biotic cycling.

324 Our findings suggest that the aridity index is a significant determinant of soil micronutrient distribution. Specifically,
325 elemental contents tend to decrease when the aridity index falls below a certain threshold (**Figs. 6 and S9**). This trend likely
326 reflects reduced input or retention of elements under arid conditions. Grouped comparisons across aridity classes (humid, dry
327 sub-humid, semi-arid, arid, and hyper-arid) were performed. The results show that the medians and means of Fe, Mn, Cu, Zn,
328 Ni, V, and Mo decrease with increasing aridity, with the largest declines observed for Fe, Mn, Zn, and V (**Fig. S17 and S18**).

329 Drought conditions may modify soil redox states, thereby influencing element speciation, adsorption capacity, mobility, and
330 ultimately, leaching behavior (Brady et al., 2016; Loveland et al., 2003; Carter et al., 1995). Also, arid environments may
331 indirectly impact micronutrients through alterations in soil pH and soil organic matter content (Moreno-Jimenez et al., 2019).
332 Previous research has shown that droughts induced by climate change can restrict the availability of essential micronutrients,
333 such as iron and zinc. This limitation, along with other adverse effects like diminished water availability, poses substantial
334 threats to vital ecological processes and services in drylands, including food production (Gupta et al., 2008; Graham, 1991).
335 Furthermore, ongoing regional warming and associated hydroclimatic shifts may interact with hydrology and redox

336 processes, organic-matter cycling, and vegetation patterns, thereby altering the availability and spatial variability of certain
337 micronutrients (Myers et al., 2014; Pachauri et al., 2014).

338 Multiple biogeochemical processes are sensitive to micronutrient availability. Biological nitrogen fixation relies on Mo and
339 Fe as essential cofactors of nitrogenase, while the methane cycle depends on Ni and Cu through their roles in
340 methanogenesis and methane oxidation, respectively (Thauer et al., 2019; Stefan et al., 2020). Permafrost thaw and
341 thermokarst development can further activate Fe-Mn redox cycling, altering metal mobility (Chauhan et al., 2024).
342 Collectively, soil micronutrients depletion may cascade through nutrient cycling and ecosystem feedbacks, amplifying the
343 impacts of ongoing environmental change across the Plateau.

344 **5 Uncertainty**

345 Although the laboratory XRF measurements used in this study were validated against ICP-MS and generally showed strong
346 correlations, several limitations should be noted. For Fe, Mn, V and Cu, the two methods produced highly consistent results,
347 supporting the robustness of the dataset. In contrast, Zn and Ni, exhibited systematic deviations from the 1:1 line, with XRF
348 tending to slightly over- or underestimate contents relative to ICP-MS (**Fig. S1**). These deviations likely arise from
349 element-specific detection sensitivities of XRF. To address this, we calibrated Zn and Ni, using the regression relationships
350 between the two methods and then repeated spatial mapping. The results before and after calibration were not significantly
351 different, confirming that these deviations do not affect our main findings (**Figs. S2 and S3**). For Mo, the results obtained
352 from the two methods were not fully consistent, likely due to its low content (**Fig. S4**). The XRF-based measurements may
353 involve considerable uncertainties. Nonetheless, we acknowledge that incorporating ICP-based measurements on a larger
354 sample set would further strengthen the accuracy of the dataset, and future efforts should consider combining multiple
355 analytical approaches to minimize methodological uncertainties.

356 Our random forest regression models demonstrated robust predictive capability (e.g. cross-validated R^2 range from 0.64 to
357 0.77 for Fe Mn Zn Ni; **Table S2**). Cu and Mo exhibit weaker predictive performance, which may be attributed to the
358 following reasons: First, analytical limitations may introduce measurement error, as both elements occur at very low contents,
359 especially Mo. Second, key process controls are not fully captured by the current predictors. E.g. Cu is strongly influenced
360 by nonlinear interactions with Fe/Al oxides, whereas Mo is affected by carbonate content, redox conditions (Tack et al.,
361 1995). These mineralogical and redox variables are underrepresented, leading to potential bias. In addition, both elements
362 show hotspot-prone, right-skewed distributions (**Figs. 2 and S5**), likely due to localized anthropogenic sources. The spatial
363 predictions inevitably contain uncertainties arising from both the Random Forest model and the input datasets. Model-based

364 uncertainty was quantified as the inter-tree variance among predictions (RF-SD; **Figs. S11 and S12**), while additional
365 uncertainty may stem from sparse meteorological observations and limited soil sampling across the Tibetan Plateau.
366 Nevertheless, model accuracy could be further improved to more extensive field sampling and refinement of input data.
367 Specifically, targeted collection of soil micronutrient data and associated covariates in underrepresented high-altitude regions
368 of the Tibetan Plateau is necessary to address existing spatial gaps. Additionally, systematic reduction of uncertainties
369 inherent in gridded environmental datasets is essential, as these uncertainties propagate errors into micronutrient predictions.
370 Continued advancement in both field observations and foundational geospatial dataset is crucial for improving the reliability
371 of regional-scale element mapping.

372 We note that ecosystem-specific pedogenesis (e.g., organic-rich surface layers in some forests vs. thin mineral A horizons in
373 grasslands) can contribute to differences in surface micronutrient contents. Consequently, part of the spatial variation we
374 report likely reflects vertical horizon contrasts sampled at a uniform 0-10 cm depth. Users should consider this context when
375 applying the dataset, especially for process inference or when comparing across ecosystems with contrasting surface
376 horizons. Where applications require deeper profiles or horizon-specific interpretation, we recommend integrating our maps
377 with local profile data.

378 **6 Data availability**

379 The gridded soil micronutrient (Fe, Mn, Cu, Zn, Ni, V, and Mo) maps for Tibetan Plateau can be downloaded from
380 <https://doi.org/10.11888/Terre.tpd.303049> (Huo et al., 2025).

381 **7 Conclusions**

382 This study delivers a comprehensive assessment of spatial distribution patterns for seven soil micronutrients (Fe, Mn, Cu, Zn,
383 Ni, V, and Mo) across the Tibetan Plateau, revealing pronounced regional-scale heterogeneity. Soil properties were the
384 primary contributors to the spatial variation of micronutrient contents, followed by topographic and climatic factors, whereas
385 vegetation and grazing disturbance had relatively minor effects. These findings highlight the coupled effects of climate,
386 vegetation, and parent material on micronutrient biogeochemical cycling within the complex environmental context of the
387 Tibetan Plateau. Using five predictor groups (climate, vegetation, soil properties, topography, and grazing disturbances), we
388 generated high-resolution spatial maps for key soil micronutrients (Fe, Mn, Cu, Zn, Ni, V, and Mo) via machine learning.
389 These maps provide validated initial conditions for process-based models simulating micronutrient cycling, advances
390 understanding of elemental distribution in alpine ecosystems.

391 **Author contributions.**

392 JZ conceived the study. HY conducted the field survey and was responsible for data collection and processing. HY prepared
393 the manuscript with contributions from all co-authors.

394 **Competing interests.**

395 The contact author has declared that none of the authors has any competing interests.

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400 **Appendix A: Supplementary material**

401 **Appendix B: Metadata**

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