



1 **PMF-LP: the first 10 m plastic-mulched farmland distribution map (2019-2021) in**
2 **the Loess Plateau of China generated using training sample generation and**
3 **classifier transfer method**

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17 **Abstract.** Plastic film mulching has been extensively used to increase crop yields in arid and semi-arid
18 regions, but it also altered agricultural landscapes and caused severe environmental pollution. Therefore,
19 accurate and timely mapping of plastic-mulched farmland (PMF) distributions is crucial for planning
20 agricultural production and preventing micro-plastic pollution. However, the scarcity of sufficient and
21 representative training samples hinders large-scale supervised classification and extraction of PMF.
22 Additionally, it remains unclear whether a pre-trained classifier can be directly applied to different regions
23 and years for rapid PMF mapping. To address these challenges, we proposed a new framework that
24 simultaneously takes advantages of the automation of index-based method and the generalization ability of
25 supervised classifier-based approach for PMF mapping. Based on the distinctive spectral responses induced
26 by plastic film deployment events, two novel and robust PMF indices—the Max Blue Band-based Plastic-
27 mulched Farmland Index (MBPMFI) and the Blue Band-based Plastic-mulched Farmland Index
28 (BPMFI)—were initially designed to automatically and rapidly extract PMF pixels in cloud-free areas as
29 candidate training samples. Additionally, the transferability of classifiers pre-trained with these
30 automatically generated samples and optimal features was further evaluated in spatial and spatial-temporal
31 transferability scenarios using F1 values. Finally, by coupling the index-based training sample generation
32 method with the temporal classifier transfer approach, PMF distributions were rapidly produced for the
33 Loess Plateau of China (PMF-LP) for 2019–2021. The results showed that the two newly established
34 indices, MBPMFI and BPMFI, were more robust than the existing PMF indices in enhancing PMF
35 information and suppressing complicated backgrounds. The temporal classifier transfer proved suitable for
36 directly and rapidly mapping PMF across multiple years without additional training samples. Using the
37 locally adaptive classifiers as a reference, the average accuracy decrease of the transferred classifiers was
38 less than 7.0% under the temporal transferability scenario. Our mapping framework achieved F1 values of
39 0.80–0.86 in recognizing PMF distributions for the Loess Plateau, highlighting its ability to delineate large-
40 scale spatial patterns of PMF. Additionally, the estimated PMF areas based on the PMF-LP aligned well
41 with the agricultural census data at municipal level ($R^2 > 0.87$). The framework developed in this study lays
42 a foundation for future monitoring of PMF distributions and agricultural micro-plastic pollution on a large
43 scale. The full archive of PMF-LP is freely available at <https://doi.org/10.5281/zenodo.13369426> (Zhao et
44 al., 2024).



45 1 Introduction

46 Plastic film mulching has been widely promoted and applied in China since 1978, due to its ability to
47 improve grain crop yields and water use efficiency through conserving water, maintaining soil moisture,
48 and increasing soil temperature (Liu et al., 2014; Sun et al., 2020; Zhao et al., 2023). Over the past 20 years,
49 China has consistently ranked first in the world in terms of the usage and area of plastic film (Yang et al.,
50 2015; Yan et al., 2014). The areas of plastic film mulching in China have increased from 11.0 million ha in
51 2000 to 17.6 million ha in 2020 (Zhang et al., 2022d). However, the extensive application of plastic films
52 has also caused severe environmental issues. Plastic films, mainly composed of polyvinyl chloride, always
53 have residues that are highly difficult to degrade in the soil, leading to severe “white pollution” (Liu et al.,
54 2014; Kumar et al., 2020; Gao et al., 2019). Furthermore, plastic film mulching might have impacts on the
55 regional climates, since the high-reflectivity and gas-tightness of the plastic film can alter the material and
56 energy exchange between the land surface and the atmosphere (Lu et al., 2014; Zhao et al., 2023). Therefore,
57 precise information about the spatial distributions of plastic-mulched farmland (PMF) over large areas is
58 essential for planning agriculture production, mitigating plastic residue pollution, and understanding water
59 and energy cycles in the agroecosystems (Veetil et al., 2023; Kumar et al., 2020).

60 Plastic film mulching data are usually derived from labor-intensive and time-consuming field surveys,
61 and then documented by the statistical yearbooks. However, statistical data often lack accurate information
62 about the locations and distributions of PMF. Benefiting from advances in Earth observation techniques,
63 the quantity and accessibility of remote sensing images have increased, enabling effective large-scale
64 monitoring of agriculture in a cost-effective and timely manner. (Maselli et al., 2020; Weiss et al., 2020;
65 Phiri et al., 2020). In particular, Sentinel-2 satellites launched by the European Space Agency (ESA),
66 provide data with a 10-m spatial resolution and a five-day revisit cycle, facilitating accurate monitoring of
67 the Earth’s surface (Drusch et al., 2012; Chaves et al., 2020). Furthermore, the emergence of cloud-based
68 geospatial processing platforms, particularly the Google Earth Engine (GEE) (Gorelick et al., 2017), has
69 significantly enhanced the processing capabilities of remote sensing data (Tamiminia et al., 2020; Pham-
70 Duc et al., 2023). These advancements have supported the establishment of region- and nation-wide land
71 cover maps, such as the China Land Cover Dataset (CLCD) (Yang and Huang, 2021) for China, and the
72 Annual Crop Inventory (ACI) for Canada (Fisette et al., 2013). However, up to date, PMF has received
73 comparatively less attention than crops, and regional or national PMF maps are rarely publicly available,
74 particularly in China, where has the largest usage and area of plastic film.

75 Most optical satellite-based PMF mapping methods rely on spectral characteristics, as PMF can be
76 easily differentiated from other land cover types during the crop sowing period (Hasituya et al., 2016; Xiong
77 et al., 2019; Zheng et al., 2022). In this period, plastic films are applied to the surface of cultivated lands,
78 resulting in a distinctive bright-white color characteristic for PMF (Lu et al., 2014; Hasituya et al., 2020).
79 Thus, several studies have attempted to mapping PMF based on the unique color and spectral changes



80 caused by temporal variations in the “plastic film–vegetation–soil” composition (Lu et al., 2015; Xiong et
81 al., 2019; Hao et al., 2019; Fu et al., 2022; Zhou et al., 2023). Additionally, spectral indices derived from
82 images taken during the crop sowing period, such as the Plastic-mulched Landcover Index (PMLI) (Lu et
83 al., 2014) and the Modified Plastic-mulched Cropland Index (mPMCI) (Fu et al., 2022), have been used to
84 recognize PMF. Unfortunately, the time window for PMF recognition is very short (approximately one
85 month) and mainly spans from April to May in northern China (Xiong et al., 2019; Cheng et al., 2023). As
86 a result, obtaining seamless optical images over large areas during this period for index-based PMF
87 mapping is challenging due to the inevitable cloud contamination.

88 Another approach for PMF mapping is machine learning, which exhibits less dependence on images
89 obtained during the crop sowing periods. Owing to the predictive abilities of machine learning, even if a
90 PMF pixel is cloud-contaminated during the crop sowing period, machine learning classifiers can still
91 effectively identify it based on clear-sky images from other times (Zhang et al., 2022c; Wang et al., 2019;
92 Gao et al., 2023). Thus, machine learning classifiers such as the Random Forest (RF) and Support Vector
93 Machine (SVM) have been widely used for PMF mapping and achieved satisfactory accuracy (Hasituya
94 and Chen, 2017; Lu et al., 2018; Zheng et al., 2022). Additionally, the predictive capabilities of machine
95 learning classifiers combined with multi-temporal satellite observations also facilitated the seamless land
96 cover mapping at regional and national scales (Zhang et al., 2022c; Zhang et al., 2021a; Liu and Zhang,
97 2023). However, implementing machine learning classifiers for large-scale land cover classification,
98 particularly for PMF, remains a significant challenge due to the lack of abundant and representative training
99 samples (Skakun et al., 2017; Foody and Arora, 1997; Wen et al., 2022).

100 To train machine learning classifiers, it is an efficient method to leverage existing land use/cover maps
101 to obtain abundant training samples (Zhang et al., 2022b; Xuan et al., 2023; Zhang et al., 2022a; Zhang and
102 Roy, 2017). Unfortunately, there are no publicly available PMF maps in China, which limits the application
103 of this method. Since the index-based methods can rapidly and automatically identify specific land cover
104 types in areas with high-quality image observations (Deng and Wu, 2012; Zhang et al., 2022c), some
105 researchers have attempted to generate training samples for certain land cover types using index-based
106 methods to address the lack of base maps for training sample generation. Thus, index-based approaches,
107 combined with machine learning classifiers, have been gradually employed for automatic and seamless land
108 cover classification (Zang et al., 2023; Zhang et al., 2022c; Yang et al., 2023). For example, Zhang et al.
109 (2022c) demonstrated that integrating Winter Rapeseed Index-derived training samples with the RF
110 classifier can seamlessly and automatically recognize rapeseed in large cloudy regions. For PMF
111 recognition, most existing PMF indices were established in specific regions in Xinjiang Province, China
112 (Lu et al., 2014; Xiong et al., 2019; Hao et al., 2019; Fu et al., 2022). However, different from the large and
113 uniformly structured land parcels in Xinjiang, most regions in China typically have smaller and more
114 scattered land parcels. Thus, the existing PMF indices struggle to distinguish PMF from the complex and
115 fragmented backgrounds in these areas. Therefore, to fully leverage the strengths of index-based methods



116 and machine learning classifiers for automatic and seamless PMF mapping, a more robust PMF index is
117 urgently needed to separate PMF from other land cover types under diverse background conditions.

118 Moreover, classifier transfer, which involves applying classifiers trained in a source domain to
119 accomplish related tasks in a target domain, is also employed to address the scarcity of training samples
120 (Pan and Yang, 2009; Ma et al., 2024). This approach supports more rapid land cover mapping across large
121 scales and multiple years compare to locally adaptive classifiers. Consequently, it has been employed to
122 swiftly identify crop distributions in regions lacking ground reference samples and to retrace the spatial
123 distributions of historical crop types in the same region (Wang et al., 2019; Hu et al., 2022; Orynbaikyzy et
124 al., 2022; Qadir et al., 2024b; Qadir et al., 2024a). However, existing studies predominantly focused on
125 transferring supervised classifiers for crop classification. It remains unclear whether this approach is
126 applicable for the rapid mapping of PMF at large scales. And the factors influencing the accuracy of PMF
127 identification using this method are also unknown.

128 This study aimed to propose a novel framework for the rapid and automatic mapping of PMF
129 distributions across different regions and years at regional scales. Subsequently, we applied this framework
130 to mapping PMF distributions in the Loess Plateau of China from 2019 to 2021. The main objectives were:
131 (1) to automatically generate training samples based on newly proposed PMF indices that can robustly
132 enhance PMF information and suppress background noise under different background conditions; (2) to
133 assess the spatial–temporal transferability of the pre-trained classifiers for PMF mapping and identify the
134 factors that influence the transferability; (3) to rapidly and automatically produce seamless PMF
135 distribution maps (PMF-LP) for the Loess Plateau of China from 2019 to 2021 using the automatically
136 generated training samples and classifier transfer methods; and (4) to validate the accuracy of the PMF-LP
137 with independent ground samples and survey statistics.

138 **2 Study area and data**

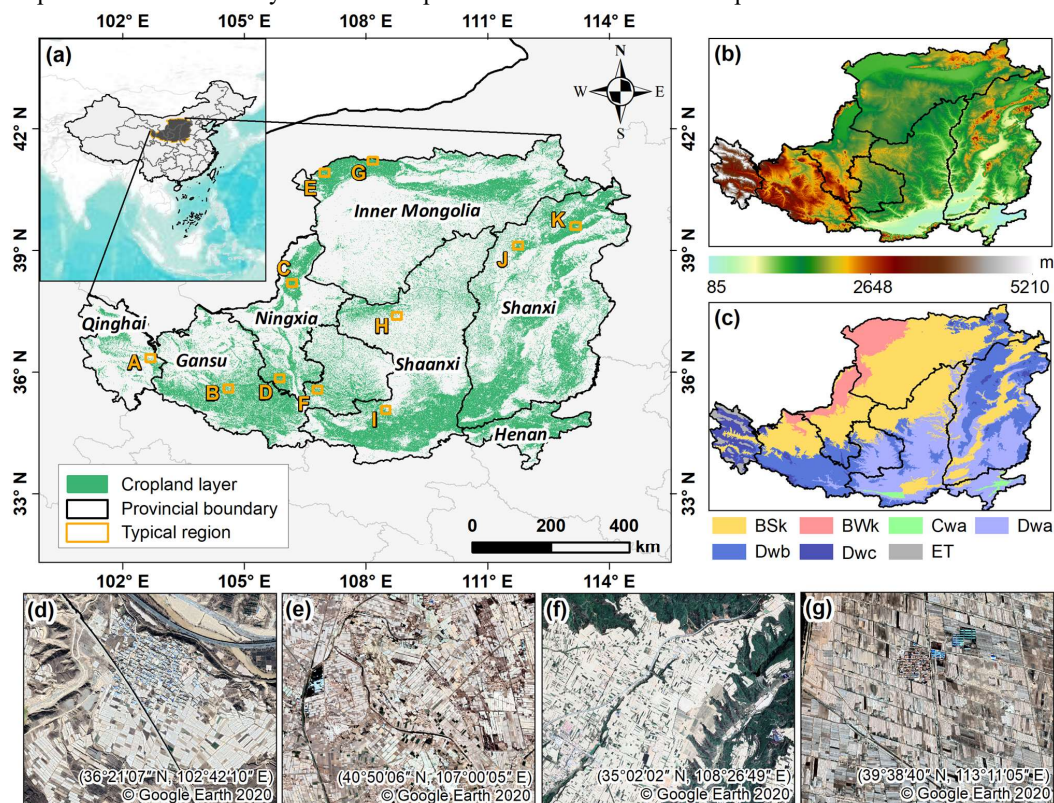
139 **2.1 Study area**

140 This study identified PMF distributions from 2019 to 2021 in the Loess Plateau (33°43' N–41°16' N,
141 100°54' E–114°33' E) of China. The Loess Plateau covers an area of approximately 6.4×10^5 km², spanning
142 seven provinces: Qinghai, Gansu, Ningxia, Inner Mongolia, Shaanxi, Shanxi, and Henan (Fig. 1 (a)). This
143 region features diverse topography, with elevations ranging from about 300 meters above sea level in the
144 Guanzhong Plain to 5,000 meters in Qinghai Province (Fig. 1 (b)). According to the Köppen-Geiger climate
145 classification (Beck et al., 2023), the Loess Plateau experiences a warm temperate continental monsoon
146 climate, characterized by hot-rainy summers, and cold-dry winters (Fig. 1 (c)). The mean annual
147 precipitation varies across the region, from 200 mm in the northwest to 800 mm in the southeast, with 60–
148 70% of the total precipitation falling between July and October. The annual evapotranspiration ranges from
149 1,400 to 2,000 mm, and the average annual temperature varies between 3.6°C and 14.3°C (Li et al., 2024;
150 Tang et al., 2018; Li et al., 2012). Plastic-film mulching, an effective field management practice for storing



151 and maintaining soil moisture, is widely used across the Loess Plateau (Fig. 1 (d)~(g)) (Zhang et al., 2022f;
152 Wang et al., 2023). Notably, transparent plastic film predominates in this region, covering about 90% of the
153 total plastic-mulched fields (Lu et al., 2014; Wang et al., 2016; Hasituya et al., 2020). Thus, this study
154 focused solely on farmland mulched with transparent plastic film. To balance the computational cost and
155 the need for adequate training samples, the study area was divided into 32 mapping units based on the
156 municipal administrative boundaries (Fig. S1).

157 Three major crops—maize, potato, and winter wheat—are extensively cultivated in the Loess Plateau.
158 Maize and potato, the primary plastic-mulched crops, are typically sown in mid-April and harvested in late
159 September, with plastic film applied in a short time before and after sowing. Winter wheat is sown in
160 October of the prior year and harvested in late June. Only a very small portion of winter wheat is mulched
161 with plastic film in the study area. The crop calendars for these main crops were illustrated in Table 1.



162
163 **Figure 1.** Overview of the study area. (a) Location of the Loess Plateau in China, spatial distributions of croplands.
164 Rectangular regions (A~K) are representative mulched regions where high-resolution Google Earth images are available
165 for visual interpretation. (b) Topography across the Loess Plateau. Digital Elevation Model (DEM) with a spatial resolution
166 of 30 m was obtained from the Shuttle Radar Topography Mission (SRTM) (Farr et al., 2007). (c) Köppen-Geiger climate
167 classification in the Loess Plateau. BSk: arid, steppe, cold; BWk: arid, desert, cold; Cwa: temperate, dry winter, hot summer;
168 Dwa: cold, dry winter, hot summer; Dwb: cold, dry winter, warm summer; Dwc: cold, dry winter, cold summer; and ET:
169 polar, tundra. (d)~(g) Zoom-in views of Google Earth images of plastic-mulched farmland during the mulching stage in
170 2020 for the rectangular regions of A, E, I, and K, respectively.



171 **Table 1.**
 172 Crop calendars illustrating growing phases of three major crops (i.e., maize, potato, and wheat) in the Loess Plateau of
 173 China. The labels “E”, “M”, and “L” denote the early, middle, and last 10-day phases of a month.

| Crops | Month | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
|--------|---------|---|---|----------|---|---|--------|----------|---|--------|---------|---|-----|-----------|---------|-----------|----------------|----------|---------|---|---|-----|---|---|-----|---|---|-----|---|---|-----|---|---|-----|---|---|---|---|---|
| | Jan | | | Feb | | | Mar | | | Apr | | | May | | | Jun | | | Jul | | | Aug | | | Sep | | | Oct | | | Nov | | | Dec | | | | | |
| | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L | E | M | L |
| Maize | | | | | | | | | | Sowing | Seeding | | | Jointing | Heading | Milk-ripe | Maturity | | | | | | | | | | | | | | | | | | | | | | |
| Potato | | | | | | | | | | Sowing | Seeding | | | Flowering | | | Tuber-swelling | Maturity | | | | | | | | | | | | | | | | | | | | | |
| Wheat | Seeding | | | Jointing | | | Earing | Maturity | | | | | | | | | | Sowing | Seeding | | | | | | | | | | | | | | | | | | | | |

174
 175 **2.2 Data**

176 **2.2.1 Sentinel-2 data**

177 The Sentinel-2 satellite, equipped with multispectral instruments, has 13 spectral bands covering
 178 visible (10 m/pixel), near-infrared (20 m/pixel), and shortwave-infrared (20 m/pixel) wavelengths (Han et
 179 al., 2021; Drusch et al., 2012). Its 5-day revisit cycle at the equator provides high-frequency Earth
 180 observation data for monitoring rapid changes on the Earth’s surface (Zhang et al., 2020b). In this study,
 181 all available Sentinel-2 surface reflectance (SR) data covering the Loess Plateau from 2019 to 2021,
 182 archived in GEE, were employed as input data since SR data are unavailable for this region before 2019.
 183 The Sentinel-2 SR products in GEE platform have been pre-processed with radiometric and atmospheric
 184 corrections. Considering the growth patterns of plastic-mulched crops (Table 1), Sentinel-2 data were
 185 further limited to March to October, and quality assessment (QA) bands were used to mask clouds and
 186 cirrus pixels.

187 Given that at least five points were required to fit a second-order harmonic regression curve whose
 188 coefficients served as the input variables for the classifiers (Section 3.2), at least five images per pixel were
 189 needed. We counted the number of yearly cloud-free images per pixel from 2019 to 2021 (Fig. S2), and
 190 found that nearly 100% of the pixels in the study area had more than five observations each year. The mean
 191 numbers of cloud-free images per pixel were 46 in 2019, 45 in 2020, and 43 in 2021, respectively. Therefore,
 192 the Sentinel-2 time series images provided ample data for the image processing requirements of this study.

193 **2.2.2 Cropland mask**

194 To simplify the process of PMF mapping and reduce commission errors caused by other land cover
 195 types (Zhang et al., 2023; You et al., 2023; You et al., 2021; Defourny et al., 2019), two cropland layers,
 196 including the ESA WorldCover 2020 (Zanaga et al., 2022) and the China Land Cover Dataset (CLCD)
 197 (Yang and Huang, 2021) of 2020, were overlapped and utilized to exclude non-cropland pixels in this study.
 198 The idea to incorporate two different cropland layers was driven by the assumption that relying on a single
 199 cropland layer could introduce uncertainties, while combining two layers could generate cropland pixels
 200 with a higher confidence level (You et al., 2023). Since the CLCD has a spatial resolution of 30 m, we re-
 201 sampled it to a 10-m resolution using the nearest-neighbor method. Given the relatively stable nature of



202 cropland across adjacent years (Gong et al., 2019), we applied the overlapped cropland layer in 2020 to
203 mask Sentinel-2 images for each year from 2019 to 2021.

204 **2.2.3 Google Earth high-resolution images (GE-HRIs)**

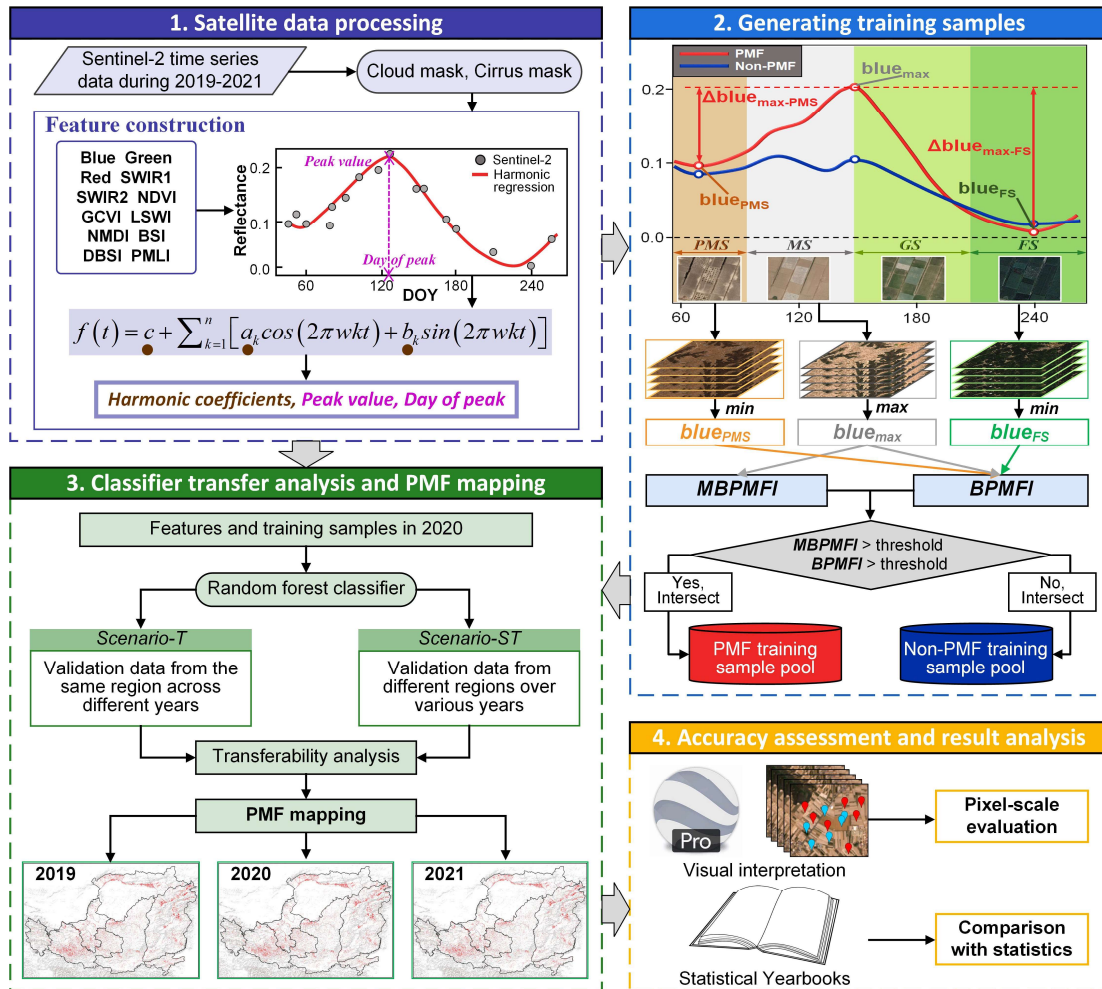
205 Google Earth high-resolution images, sourced from the Google Earth platform, provide satellite and
206 aerial imagery of the Earth's surface with a sub-meter spatial resolution of 0.6 m, which facilitates sample
207 collection through visual interpretation. Therefore, the GE-HRIs were utilized to assist in obtaining
208 validation samples in this study. Additionally, two representative rectangular regions (20 km × 20 km) with
209 at least one GE-HRI image available during the mulching stage were selected in each province (yellow
210 areas A~K in Fig. 1(a)). The samples interpreted in these rectangular regions (300 PMF/Non-PMF points
211 for each region) were employed to determine optimal thresholds for the newly proposed PMF indices in
212 Section 3.1.2.

213 **2.2.4 Agricultural statistics**

214 The PMF census data from 2019 to 2021, sourced from municipal-level agricultural statistical offices,
215 were collected to assess the mapped PMF areas. It is noteworthy that plastic film usage in the statistical
216 yearbooks of most cities in the Loess Plateau is quantified in tons. Consequently, in this study, we only
217 obtained statistical data for 12 cities (Fig. S3), where plastic film usage was measured in terms of area.
218 Since these 12 cities account for nearly one-third of the total area of the Loess Plateau, the accuracy
219 assessment of the mapped PMF areas based on the municipal statistical data was considered representative
220 and reliable.

221 **3 Methodology**

222 The workflow of this study was summarized in Fig. 2. First, cloud-free Sentinel-2 time series data
223 were used to fit harmonic curves, generating feature variables that served as input variables for the
224 classifiers. Next, a temporal signature analysis of various cropland-related land covers was conducted.
225 Distinctive signatures during the mulching period guided the design of PMF indices, which effectively
226 separated PMF from other land cover types. These indices were then employed to identify PMF in cloud-
227 free areas and automatically generate training samples. Third, random forest classifiers were developed to
228 recognize PMF based on the feature variables and the automatically generated training samples. The
229 transferability of these classifiers was then evaluated in terms of temporal and spatial-temporal
230 transferability. Finally, the PMF mapping results were assessed against ground reference samples and
231 statistical PMF area data.



232
 233 **Figure 2.** Flowchart depicting the integration of index-based training sample generation and classifier transfer for plastic-
 234 mulched farmland mapping. PMF: Plastic-mulched Farmland, Non-PMF: Non-plastic-mulched Farmland. The entire
 235 growing season of plastic-mulched crops was divided into four stages: pre-mulching stage (PMS), mulching stage (MS),
 236 growing stage (GS), and flourishing stage (FS) (Section 3.1.2). The MBPMFI (Max Blue Band-based Plastic-mulched
 237 Farmland Index) and BPMFI (Blue Band-based Plastic-mulched Farmland Index) are the two novel PMF indices proposed
 238 in this study (Section 3.1.2).

239 **3.1 Automatically generating training samples**

240 **3.1.1 Optical characteristics of the PMF**

241 The multi-temporal Sentinel-2 images with true-color composites from a representative sub-region in
 242 Pingliang City, Gansu Province, were shown in Fig. 3 (c)–(j) to analyze the optical characteristics of PMF.
 243 The PMF, plastic greenhouses (PGs), and vegetation (mainly winter crop) within this sub-region
 244 represented the primary agricultural-related land cover types on the cultivated farmland in the Loess Plateau.
 245 During the crop sowing period, PMF exhibited a bright-white color due to the plastic films, making it easily

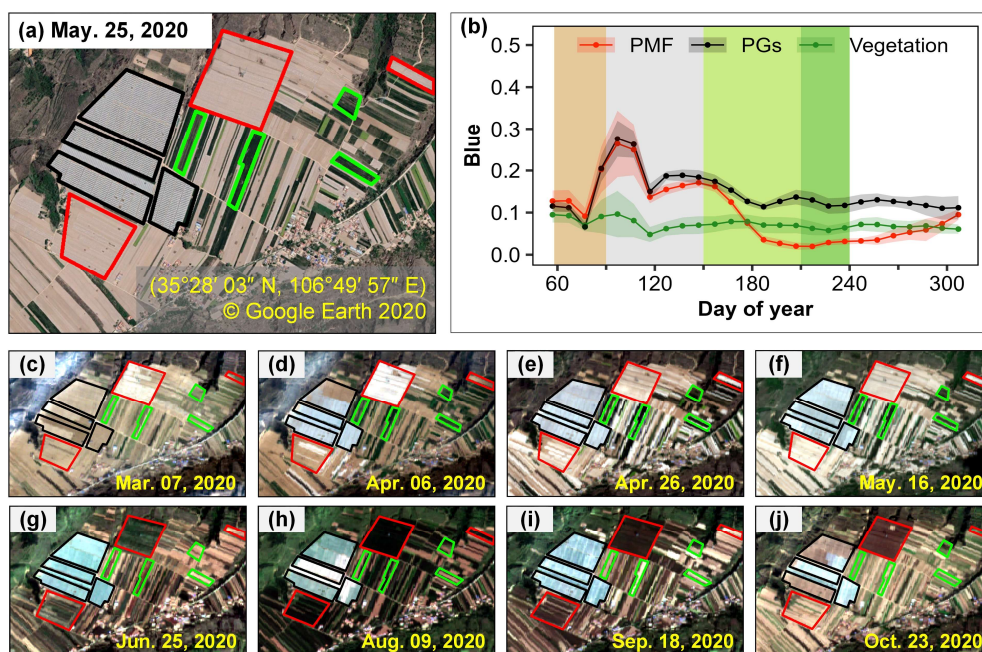


246 distinguishable from vegetation (dark-green color) but potentially confusing it with PGs (blue-gray or white
247 color) (Fig. 3 (c)~(f)). Subsequently, from June to September (Fig. 3 (g)~(i)), the crop canopy covered the
248 PMF, giving it a green appearance similar to vegetation. Meanwhile, since crops grew inside the PGs, the
249 PGs maintained blue-gray or white appearance consistently. Thus, this period was ideal for differentiating
250 between PGs and PMF.

251 Based on this analysis, we observed that PMF exhibited a bright-white appearance during the crop
252 sowing period. Theoretically, the conspicuous bright white characteristic indicated the substantial
253 reflectance of PMF in the visible spectrum. Temporal profiles of the Sentinel-2 blue band revealed
254 distinguishable patterns among PMF, PGs, and vegetation (Fig. 3 (b)). Specifically, during the crop sowing
255 period, areas covered by plastic films had higher reflectance in the blue band compared to those covered
256 by the crop canopy. Subsequently, as crops grew in the PMF, the increased absorption in the visible
257 spectrum led to a gradual decrease in PMF reflectance in the blue band. Particularly, during the peak
258 growing stage of crops, PMF reached its lowest reflectance in the blue band.

259 In addition to the blue band, PMF demonstrated clear differentiation from other land cover types in
260 the green and red bands, as well as in the red-edge band (red edge1) (Fig. S4). Notably, this study focused
261 exclusively on the blue band due to the consistent temporal profiles observed across these four bands.
262 Furthermore, the distinct PMF characteristics in the blue band were also tested across different regions in
263 the Loess Plateau (Fig. S5). The results showed that PMF reflectance in the blue band exhibited the same
264 temporal profiles as those illustrated in Fig. 3 (b), characterized by a rapid increase during the crop sowing
265 period, followed by a swift decrease during the flourishing period.

266
267



268

269 **Figure 3.** Time-series profiles of plastic-mulched farmland (PMF), plastic greenhouses (PGs), and vegetation in a
270 representative sub-region of Pingliang City, Gansu Province, China. (a) High-resolution Google Earth imagery during the
271 crop sowing period. The red, black, and green outlined plots denote the PMF, PGs, and vegetation, respectively. (b) Time
272 series of PMF, PGs, and vegetation in the Sentinel-2 blue band. The red, gray, and blue buffers indicate one standard
273 deviation. The brown, gray, yellowish-green, and green rectangular areas denote the pre-mulching stage (PMS), mulching
274 stage (MS), growing stage (GS), and flourishing stage (FS) as defined in Section 3.1.2, respectively. (c)~(j) Sentinel-2
275 images with a true-color composite of blue, green, and red bands on different dates across the entire crop growing season.

276

3.1.2 Developing PMF indices to automatically generate training samples

277

Following the temporal profile analyses in Section 3.1.1, the entire growing season of plastic-mulched
278 crops was divided into four distinct stages (Fig. 4). The pre-mulching stage (PMS) (March to early April)
279 involves primarily field plowing, with seldom plastic film coverage on the farmland. The mulching stage
280 (MS) (mid-April to May) represents the gradual coverage of farmland with plastic films. During the
281 growing stage (GS) (June to July), crops begin to grow but do not yet fully cover the cropland. The
282 flourishing stage (FS) (August to September) is characterized by vigorous crop growth, with the crop
283 canopy fully sheltering the cropland.

284

Across these stages, the distinctive spectral responses of PMF in the blue band were summarized as
285 follows: a relatively large range of reflectance dynamics from PMS to MS ($\Delta\text{blue}_{\text{max-PMS}}$) due to the gradual
286 deployment of plastic film, and the maximal range of reflectance dynamics from GS to FS ($\Delta\text{blue}_{\text{max-FS}}$)
287 due to the abundant visible spectrum absorption for photosynthesis. Considering that both $\Delta\text{blue}_{\text{max-PMS}}$
288 and $\Delta\text{blue}_{\text{max-FS}}$ for PMF were relatively higher than those for other land cover types, multiplying these
289 two metrics could effectively highlight PMF while suppressing other land cover types. Specifically, two

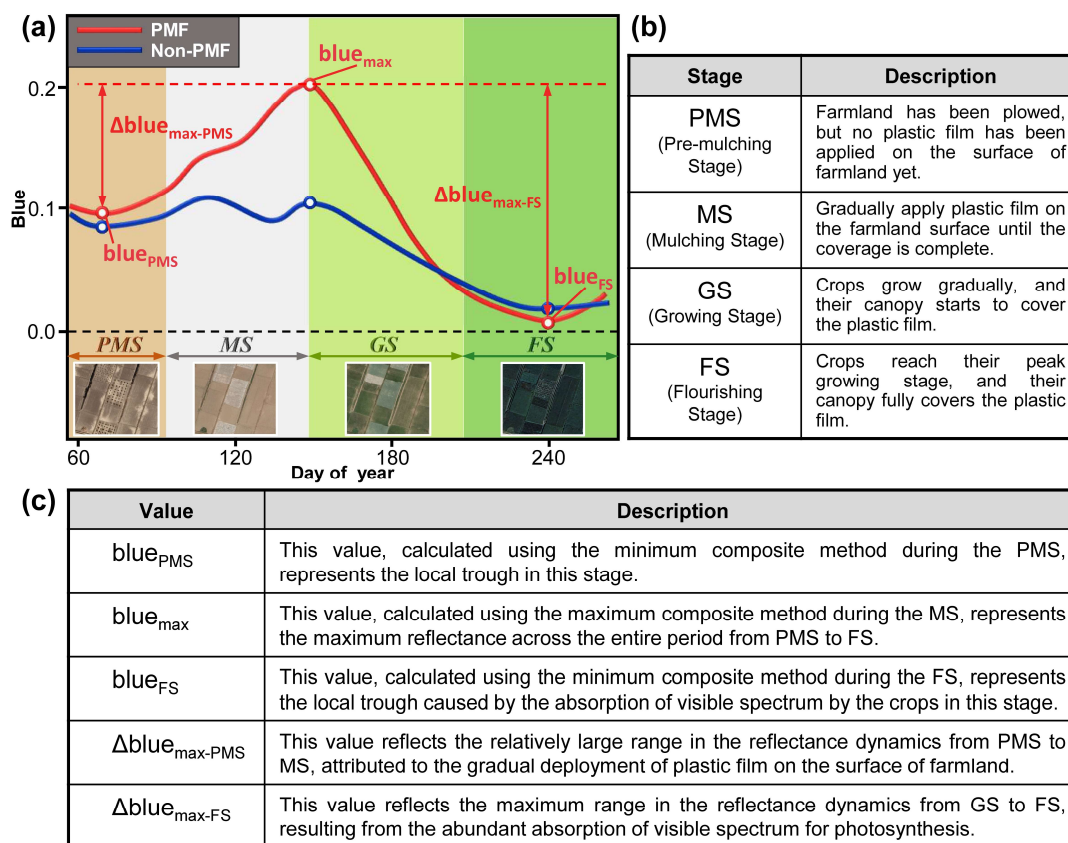


290 local troughs ($blue_{PMS}$ and $blue_{FS}$) are the minimum reflectance values of the blue band during the PMS and
 291 FS, respectively. The local peak ($blue_{max}$) refers to the maximum reflectance value of the blue band during
 292 the MS. Additionally, the metric $blue_{max}$ was also employed to enhance PMF signals, as it exhibited the
 293 highest reflectance value throughout the period from PMS to FS. Finally, two PMF indices, the Max Blue
 294 Band-based Plastic-mulched Farmland Index (MBPMFI) (Eq. (1)) and the Blue Band-based Plastic-
 295 mulched Farmland Index (BPMFI) (Eq. (2)) were defined.

$$296 \quad MBPMFI = blue_{max} \quad (1)$$

$$297 \quad BPMFI = 100 \times \Delta blue_{max-PMS} \times \Delta blue_{max-FS} = 100 \times (blue_{max} - blue_{PMS}) \times (blue_{max} - blue_{FS}) \quad (2)$$

298 where the multiplier 100 was used to linearly stretch the BPMFI to a wider range.

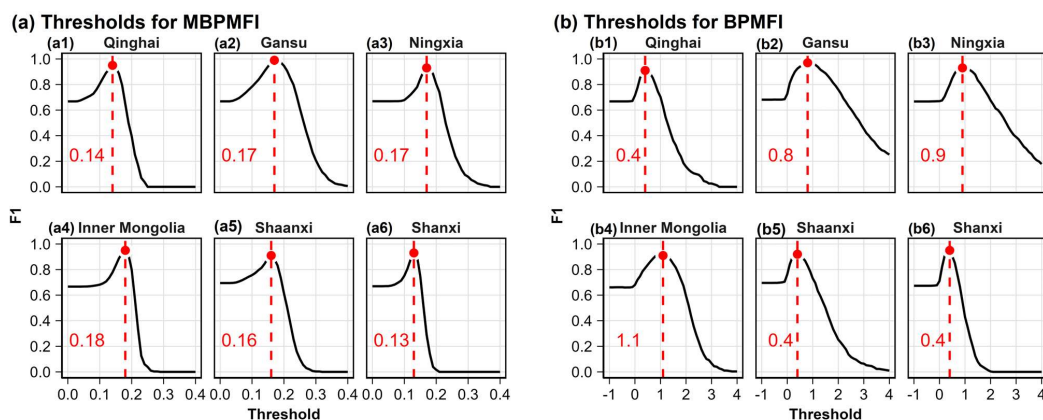


299
 300 **Figure 4.** Schematic diagrams illustrating the construction of the plastic-mulched farmland indices. (a) Time-series
 301 profiles of the plastic-mulched farmland (PMF) and non-plastic-mulched farmland (Non-PMF) in the Sentinel-2 blue band.
 302 (b) Defined stages used to divide the entire growing season of plastic-mulched crops into four parts. (c) Explanation of the
 303 variables used to represent the unique characteristics of PMF.

304

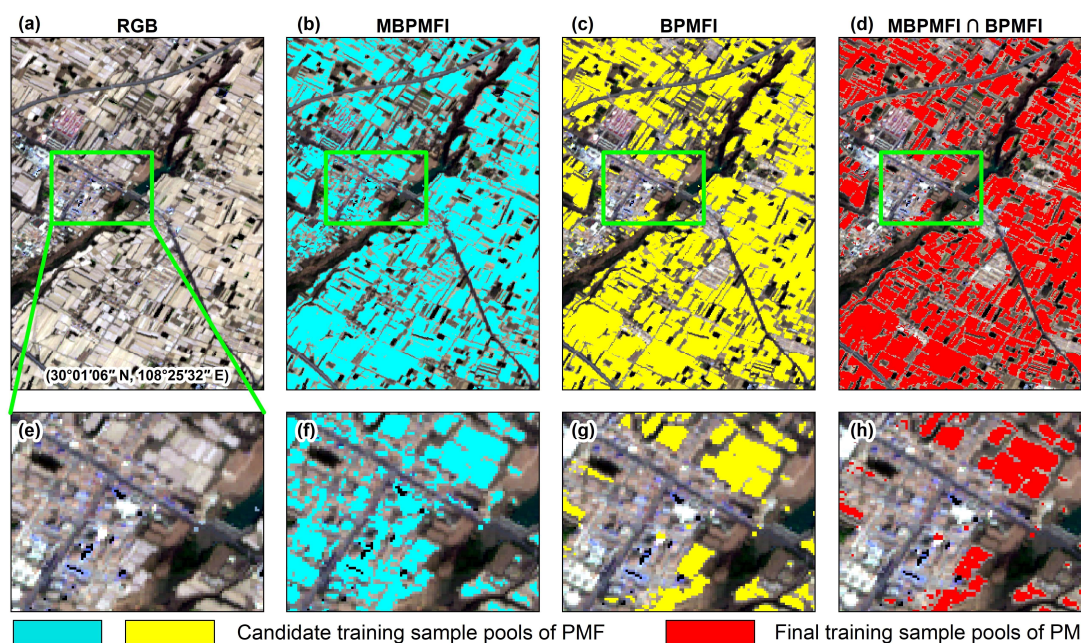


305 Notably, since the primary objective of this study was to recognize PMF and given the relative scarcity
306 of PGs compared with PMF (Feng et al., 2021; Tong et al., 2024), land cover types were categorized into
307 two classes: PMF and Non-PMF. Table S1 presented the statistical analysis results of the MBPMFI and
308 BPMFI indices for both PMF and Non-PMF samples collected from the rectangular regions (Section 2.2.3).
309 Generally, the *p-values* for the two PMF indices across different provinces were equal to 0, indicating the
310 great potential of MBPMFI and BPMFI to enhance PMF signals while suppressing Non-PMF signals.
311 Referring to previous studies (Zhang et al., 2022g; Zhou et al., 2024), sample points from the rectangular
312 regions were exploited to establish province-specific thresholds for each PMF index. The optimal
313 thresholds for the two novel indices were separately determined based on accuracy information with
314 different thresholds (Fig. 5). Particularly, for Henan Province, we adopted the same thresholds as those used
315 in Shanxi Province. Finally, PMF and Non-PMF pixels in cloud-free areas were automatically extracted.
316 To improve the reliability of the training samples, only cropland classified as PMF or Non-PMF by both
317 PMF indices was included in the final training sample pools (Fig. 6).



318

319 **Figure 5.** Classification accuracy curves of various plastic-mulched farmland indices at different thresholds.



320
321 **Figure 6.** Zoomed-in views for illustrating the detailed process of training sample generation. (a) Sentinel-2 images with
322 a true-color composite of blue, green, and red bands. (b) Candidate training sample pool derived based on MBPMFI (Max
323 Blue Band-based Plastic-mulched Farmland Index). (c) Candidate training sample pool derived based on BPMFI (Blue
324 Band-based Plastic-mulched Farmland Index). (d) Final training sample pool. “ \cap ” means the intersection is implemented
325 for sample pools of MBPMFI and BPMFI. (e)~(h) Zoomed-in views from the green frame in (a) for presenting the
326 misclassification between plastic-mulched farmland (PMF) and impervious surfaces.

327 3.1.3 Training sample generation and refinement

328 Based on the candidate sample pools generated in Section 3.1.2, a stratified random sampling approach
329 was adopted to select 2000 pixels for PMF and Non-PMF within every mapping unit for the year 2020. To
330 further ensure the reliability of these samples and eliminate potential errors, a strict spatial filter (8-neighbor
331 filter) was conducted for these samples (Zhang et al., 2022a; Zhang et al., 2023; Zhang and Roy, 2017).
332 Only pixels sharing the same class type as the surrounding eight pixels were considered as high-quality
333 samples. These samples were utilized as training data for subsequent classifiers to identify PMF.

334 3.2 Feature calculation and selection

335 The visible bands (Blue, Green, and Red) and shortwave-infrared bands (SWIR1 and SWIR2) from
336 Sentinel-2 were selected as input features for classifiers, due to their high potential to separate PMF from
337 other land cover types (Hasituya et al., 2016; Xiong et al., 2019; Hao et al., 2019). Since plastic films can
338 influence energy balance and water cycles on the land surface (Lu et al., 2014; Hasituya and Chen, 2017),
339 several remote-sensing vegetation indices associated with vegetation and soil conditions were also included
340 as input features (Table S2): NDVI (Tucker, 1979), GCVI (Gitelson et al., 2005), LSWI (Xiao et al., 2002),
341 NMDI (Wang and Qu, 2007), BSI (Rikimaru et al., 2002), DBSI (Rasul et al., 2018), PMLI (Lu et al., 2014).



342 Moreover, the newly proposed PMF indices, MBPMFI and PMFI, were also incorporated.

343 Compared with Non-PMF, PMF exhibited unique temporal profiles resembling sine waves in the blue
344 band (Fig. 4 (a), Fig. S5). To characterize these profiles and fill missing values caused by cloud
345 contaminations, harmonic regression (Jakubauskas et al., 2001; Zhou et al., 2022) (Eq. (3)) was conducted
346 to fit time series curves for PMF across the five surface reflectance bands and seven vegetation indices. The
347 time window for fitting the time-series curves extended from March 1st to October 31st, covering the entire
348 growing season of plastic-mulched crops. Each band was treated as a time-dependent function, denoted as
349 $f(t)$.

$$350 \quad f(t) = c + \sum_{k=1}^n [a_k \cos(2\pi k \omega t) + b_k \sin(2\pi k \omega t)] \quad (3)$$

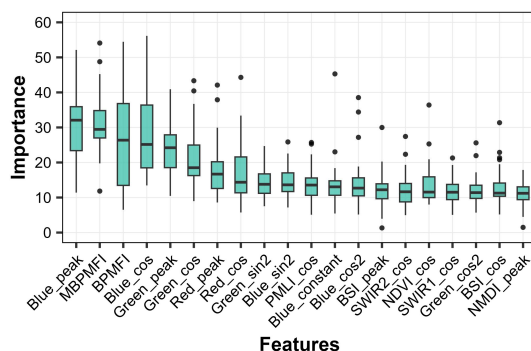
351 where the independent variable t is the day of year expressed as a fraction between 0 (January 1st) and 1
352 (December 31st) for a satellite image, c is the intercept term, n is the order of harmonic series, a_k are the
353 cosine coefficients, b_k are the sine coefficients, and ω is the angular frequency.

354 In the above harmonic regression formula, parameters n and ω need to be adjusted to balance the fitting
355 closeness to the observation points and prevent overfitting. Based on the ground reference PMF samples
356 within the rectangular regions (yellow areas A~K in Fig. 1 (a)), we picked $n = 2$ and $\omega = 1.5$ as the optimal
357 parameters by evaluating the root mean square error (RMSE) of the fitting progress (Fig. S6). Therefore,
358 we obtained the final harmonic regression formula (Eq. (4)).

$$359 \quad f(t) = c + a_1 \cos(3\pi t) + b_1 \sin(3\pi t) + a_2 \cos(6\pi t) + b_2 \sin(6\pi t) \quad (4)$$

360 After the harmonic regression, the time series of each band was represented by five coefficients: c , a_1 , b_1 ,
361 a_2 , and b_2 . Furthermore, the peak value (peak) of each band and their corresponding dates (timing) were
362 also extracted from the fitted harmonic regression curves. These coefficients combined with MBPMFI and
363 BPMFI resulted in a total of 86 ($12 \times 7 + 2 = 86$) features.

364 To select the important features and discard unimportant ones which may adversely affected the
365 accuracy and computational cost of supervised classifiers (Zou et al., 2015; Wang et al., 2022b), a
366 combination of time-series correlation analysis and random forest feature importance analysis was
367 employed to reduce the number of features (Fig. S7). The reduction process included: (1) grouping bands
368 with correlation coefficients exceeding 0.90, retaining only the band with the highest feature importance
369 (retained bands: Blue, SWIR2, BSI, DBSI, GCVI, NMDI, and PMLI); (2) further limiting the seven
370 harmonic coefficients to the peak, a_1 , and b_2 terms based on their importance evaluation results; (3) retaining
371 MBPMFI and BPMFI due to their higher importance compared with other features (Fig. 7). Ultimately, a
372 total of 23 features were adopted as input variables for the classifiers in this study. A comparison of PMF
373 identification accuracies using 86 features versus 23 features revealed no significant difference ($\alpha = 0.05$)
374 in classification accuracy between the two sets of features (Fig. S8).



375
376 **Figure 7.** Feature importance of top 20 features calculated by the random forest classifiers trained in each city in the
377 Loess Plateau. Since the feature importance in each city is not comparable, we normalized the individual value by dividing
378 the total importance of all features in each city and scaled them by multiplying 1000.

379 3.3 Classifier transferability analysis

380 We employed the pixel-based random forest (RF) classifier (Breiman, 2001) available on the GEE
381 platform to distinguish PMF from Non-PMF for each mapping unit, based on the selected optimal features
382 and automatically generated samples mentioned above. As an ensemble algorithm comprising numerous
383 decision trees, RF has been widely used for large-scale land cover classification (Zhang et al., 2021c; Liu
384 et al., 2020), crop type mapping (Yang et al., 2023; Wang et al., 2020), and classifier transferability analysis
385 (Wijesingha et al., 2024; Qadir et al., 2024b), due to its capabilities in handling high-dimensional data,
386 tolerating noise, and preventing overfitting (Belgiu and Drăguț, 2016; Sheykhmousa et al., 2020). Two key
387 hyperparameters need to be set when using RF, which are the number of trees (*Ntree*) to be generated and
388 the number of features (*Mtry*) used for testing the best split when growing the trees (Belgiu and Drăguț,
389 2016). The former parameter was set as 100 since the out-of-bag errors in different mapping units ceased
390 to decrease beyond 100 trees (Fig. S9); the latter one was set as the default value (i.e., the square root of
391 the number of features).

392 Revealing the temporal and spatial-temporal transferability of classifiers can facilitate multi-year PMF
393 mapping without the need to collect domain- and year-specific training samples. The classifiers trained with
394 the optimal features and automatically generated samples in each mapping unit in 2020 (reference scenario,
395 Scenario-Ref) were tested in two transferability scenarios to evaluate their temporal and spatial-temporal
396 transferability: (1) classifiers trained in each mapping unit in 2020 were applied to recognize PMF within
397 the same units for 2019 and 2021 to assess their temporal transferability (Scenario-T); and (2) classifiers
398 trained in each mapping unit in 2020 were employed to recognize PMF in all other remaining units for 2019
399 and 2021 to test their spatial-temporal transferability (Scenario-ST). Note that classifiers trained with data
400 from 2020 were not assessed in other spatial regions in the same year of 2020 because the primary goal of
401 classifier transfer in this study is to rapidly retrace historical PMF distributions. Additionally, we trained
402 classifiers in each mapping unit (municipal level) because sample point collection in practice is usually



403 conducted within administrative boundaries rather than in agro-ecological boundaries (Hao et al., 2020;
404 Wang et al., 2019). In each scenario, the F1-score (F1) was calculated based on validation samples as
405 accuracy evaluation metric. Furthermore, to assess changes of F1 in each transferability scenario against
406 reference scenario, a percentage change of F1 values ($F1_{change}$) was also computed (Eq. (5)).

$$407 \quad F1_{change} = \frac{F1_{transferability} - F1_{reference}}{F1_{reference}} \times 100 \quad (5)$$

408 3.4 Accuracy assessment

409 We assessed the quality of the generate PMF maps with two measures. First, we performed a pixel-
410 wise quantitative assessment for the PMF maps in 2019–2021, using more than 5000 ground reference
411 samples each year (Table S3). Four accuracy metrics were adopted: producer accuracy (PA), user accuracy
412 (UA), overall accuracy (OA), and F1-score (F1). Next, the coefficient of determination (R^2) was used to
413 quantitatively measure the consistency of the PMF areas derived from agricultural statistics and those from
414 the resultant maps. Due to the lack of publicly available PMF products in China, inter-comparison with
415 existing products was not possible in this study.

416 The validation samples were collected by visually interpreting GE-HRIs and Sentinel-2 images, based
417 on the unique characteristics of PMF (Section 3.1.1). A two-step strategy was designed to generate the
418 validation samples: (1) croplands exhibiting a white hue during the mulching stage (MS) and a dark green
419 hue during the flourishing stage (FS) in the true-color composite images were more likely to be PMF than
420 those showing other hue changes; (2) the time-series curves of PMF in the blue band displayed more
421 pronounced peak values during the mulching stage (MS) and trough values during the flourishing stage (FS)
422 compared to other land cover types. Pixels meeting above two rules were finally labeled as PMF; otherwise,
423 they were grouped into Non-PMF. Ultimately, we obtained 7,091, 12,140, and 6,714 validation samples for
424 2019, 2020, and 2021, respectively (Table S3, Fig. S10).

425 Moreover, five additional PMF indices (Table 2), including PMLI (Lu et al., 2014) (Eq. (6)), PMFI_{first}
426 (Xiong et al., 2019) (Eq. (7)), PMFI_{second} (Xiong et al., 2019) (Eq. (8)), PMLI_{SWIR} (Hao et al., 2019) (Eq.
427 (9)), and mPMCI (Fu et al., 2022) (Eq. (10)), were utilized to compare with the newly established MBPMFI
428 and BPMFI indices in enhancing PMF information while suppressing Non-PMF information. Except for
429 the PMLI, higher values of these indices are more related to PMF. To test the robustness of these indices,
430 three representative regions with diverse environmental conditions were selected: (1) Region I, located in
431 Bayan Nur City, Inner Mongolia Province, exhibits a large and densely distributed PMF area in this region
432 (Fig. 8 (a)); (2) Region II, located in Linxia City, Gansu Province, features an irregularly shaped PMF with
433 sparse distribution (Fig. 9 (a)); and (3) Region III is located in Xianyang City, Shaanxi Province, where
434 some plastic greenhouses and buildings exhibit similar spectral profiles to PMF (Fig. 10 (a)). A visual
435 evaluation was conducted to intuitively compare the performance of these indices. Since the presence of
436 noises in the index images could affect their display and subsequently influence the visual comparison



437 results, all index images were displayed with a 1% linear stretch (Zhang et al., 2022g) to ensure the fairness
 438 of visual comparison among them.

439 **Table 2.**

440 Summary of the existing plastic-mulched farmland indices used for comparison with the newly established MBPMFI and
 441 BPMFI indices in this study.

| Name | Equation* | Equation index |
|---|--|----------------|
| PMLI (Plastic-mulched Landcover Index) | $\frac{\rho_{SWIR1} - \rho_{Red}}{\rho_{SWIR1} + \rho_{Red}}$ | (6) |
| PMFI _{first} (Plastic-mulched Farmland Index First) | $\frac{\rho_{SWIR2}}{\rho_{NIR}}$ | (7) |
| PMFI _{second} (Plastic-mulched Farmland Index Second) | $\frac{\rho_{SWIR2}}{\rho_{Blue}}$ | (8) |
| PMLI _{SWIR} (Plastic-mulched Land Index with Shortwave- infrared Band) | $\frac{(\rho_{RE4} + \rho_{NIR} + \rho_{RE3}) - (\rho_{SWIR1} + \rho_{SWIR2})}{\rho_{SWIR1} + \rho_{SWIR2}}$ | (9) |
| mPMCI (Modified Plastic-mulched Cropland Index) | $\frac{\rho_{SWIR1} + \rho_{NIR}}{\rho_{SWIR1} - \rho_{NIR}}$ | (10) |

442 * ρ_{blue} , ρ_{red} , ρ_{nir} , ρ_{RE3} , ρ_{RE4} , ρ_{SWIR1} , and ρ_{SWIR2} represent the Sentinel-2 reflectance in the blue band (496 nm), red band
 443 (665 nm), near-infrared band (833 nm), red-edge 3 band (779 nm), red-edge 4 band (864 nm), shortwave-infrared band
 444 1 (1610 nm), and shortwave-infrared band 2 (2185 nm), respectively.

445 4 Results

446 4.1 Comparisons with existing PMF indices

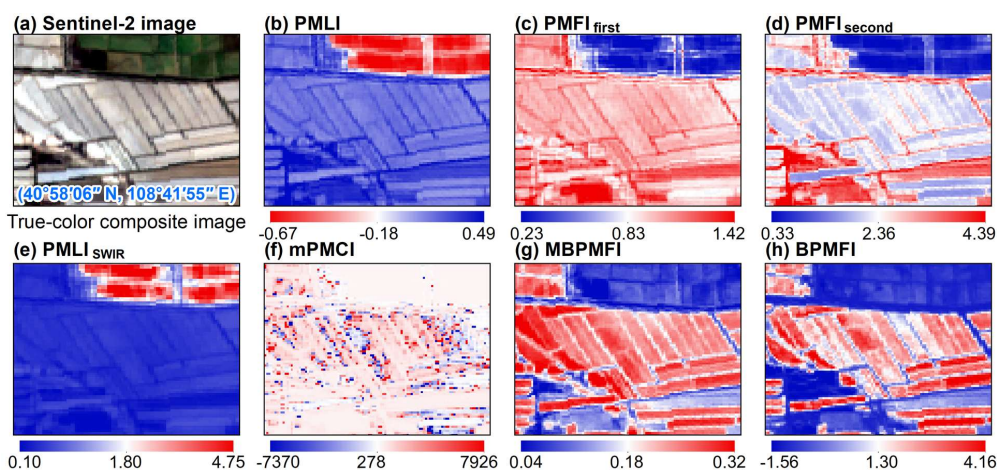
447 In this section, we presented detailed visual comparison results for all the seven PMF indices under
 448 three different environmental conditions. High values of the PMFI_{first}, PMFI_{second}, PMLI_{swir}, mPMCI,
 449 MBPMFI, and BPMFI correspond to PMF objects, while a high value of the PMLI is related to Non-PMF
 450 objects. For the convenience of visual comparisons between the PMLI and the other six indices, the color
 451 bar of PMLI was reversed. Consequently, in all seven index images, the red color is associated with PMF,
 452 and the blue color is related to Non-PMF.

453 4.1.1 Visual comparison in Region I

454 In Region I, the MBPMFI and BPMFI indices outperformed the other five indices in highlighting PMF
 455 and suppressing background noise (Fig. 8 (g) and (h)). The mPMCI performed the poorest, barely
 456 recognizing PMF (Fig. 8 (f)). PMLI and PMLI_{swir} showed similar performance, both misclassifying



457 vegetation as PMF. Region I is located in the Hetao Irrigation District of Inner Mongolia, where croplands
458 are extensively irrigated through the integral irrigation systems. The shortwave-infrared spectrum is highly
459 absorbed by water. Although PMF had a higher reflectance value in the red spectrum than vegetation, the
460 wet soil caused by irrigation can lead to lower reflectance values in shortwave-infrared spectrum for
461 vegetation than for PMF. In this context, both PMF and vegetation might exhibit a low reflectance value in
462 PMLI (Eq. (6)), leading to misclassification between them (Zhang et al., 2022g). While $PMLI_{SWIR}$ (Eq. (9))
463 was primarily designed to separate PMF from bare land, it failed to suppress vegetation, which also had a
464 high reflectance value in the near-infrared spectrum and low reflectance value in the shortwave-infrared
465 spectrum similar to PMF (Fig. 8 (e)). $PMFI_{first}$ and $PMFI_{second}$ could highlight PMF to a certain extent, but
466 wrongly highlighted some background objects, such as roads and bare lands (Fig. 8 (c) and (d)). Overall,
467 MBPMPFI and BPMFI showed comparable performance in highlighting PMF signals, with BPMFI having
468 an advantage in suppressing background noise, especially roads.



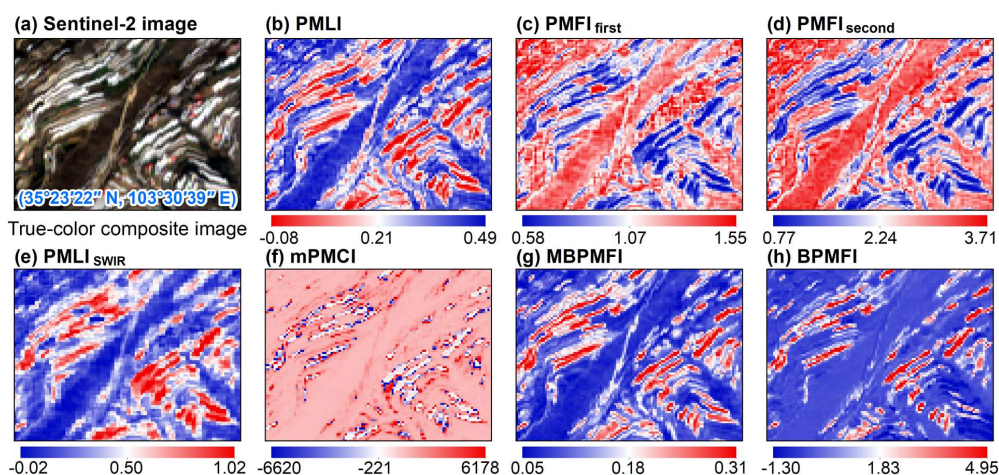
469
470 **Figure 8.** Visual evaluation results based on the plastic-mulched farmland indices of PMLI (b), $PMFI_{first}$ (c), $PMFI_{second}$
471 (d), $PMLI_{SWIR}$ (e), mPMCI (f), MBPMPFI (g), and BPMFI (h) in a representative area of Bayan Nur City, Inner Mongolia
472 Province (a). The red color is related to PMF, and the blue color is associated with backgrounds under all of the seven
473 indices.

474 4.1.2 Visual comparison in Region II

475 Unlike Region I, Region II is dominated by bare land and PMF, with PMF distributed sparsely in long
476 strips. MBPMPFI and BPMFI continued to demonstrate the best performance in highlighting PMF and
477 suppressing background noise (Fig. 9 (g) and (h)). Since PMLI and $PMLI_{SWIR}$ were primarily proposed to
478 differentiate PMF from bare land (Lu et al., 2014; Hao et al., 2019), they could roughly identify PMF but
479 failed to suppress some backgrounds, such as roads and buildings (Fig. 9 (b) and (e)). In contrast to Region
480 I (Fig. 8), the reflectance values of PMF in $PMFI_{first}$ and $PMFI_{second}$ were both lower than those of Non-



481 PMF, indicating that Non-PMF was highlighted, while PMF was suppressed (Fig. 9 (c) and (d)). This was
482 because, compared with bare land, PMF had higher reflectance values in both the blue and near infrared
483 spectra (Xiong et al., 2019; Hasituya et al., 2016; Hao et al., 2019), but no obvious difference was observed
484 in the shortwave-infrared spectrum (Hasituya et al., 2016). Consequently, PMF might have lower values in
485 $PMFI_{first}$ and $PMFI_{second}$ than bare land. The $mPMCI$ continued to exhibit the poorest performance among
486 the seven indices (Fig. 9 (f)). Compared with MBPMFI, BPMFI was more effective in suppressing Non-
487 PMF land covers, such as bare land and buildings (Fig. 9 (g) and (h)). In general, BPMFI still performed
488 the best in Region II.



489
490 **Figure 9.** Visual evaluation results based on the plastic-mulched farmland indices of PMLI (b), $PMFI_{first}$ (c), $PMFI_{second}$
491 (d), $PMLI_{SWIR}$ (e), $mPMCI$ (f), MBPMFI (g), and BPMFI (h) in a representative area of Linxia City, Gansu Province (a).
492 The red color is related to PMF, and the blue color is associated with backgrounds under all of the seven indices.

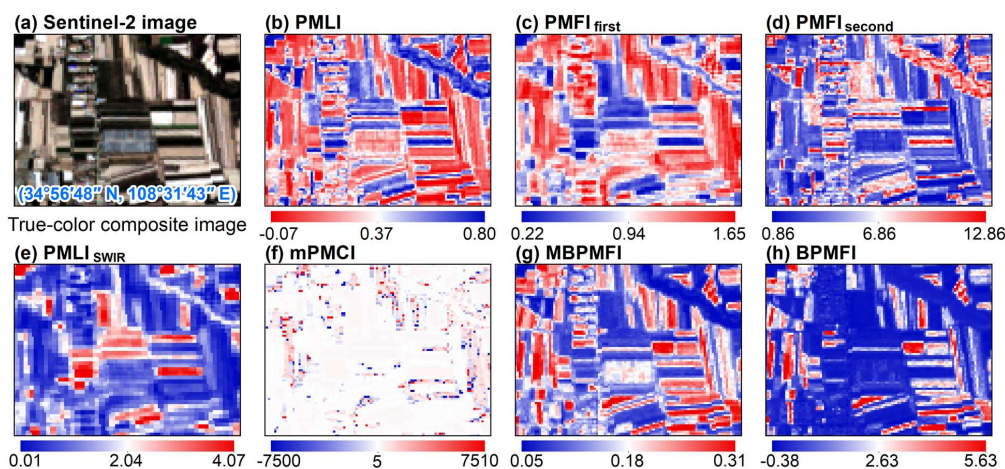
493 4.1.3 Visual comparison in Region III

494 Differing from Region I and Region II, Region III featured a more complex background environment,
495 including PMF, vegetation, bare land, buildings, and plastic greenhouses. Among the seven indices, BPMFI
496 demonstrated the best performance in identifying PMF and suppressing background noise (Fig. 10 (h)).
497 While PMLI, $PMFI_{first}$, and MBPMFI were also able to highlight PMF, they failed to effectively suppress
498 certain background elements, particularly buildings and plastic greenhouses (Fig. 10 (b), (c), and (g)).
499 $PMFI_{second}$ and $PMLI_{SWIR}$ seemed to wrongly highlight backgrounds while suppressing PMF (Fig. 10 (d)
500 and (e)). The $mPMCI$ continued to perform the worst in recognizing PMF (Fig. 10 (f)). Overall, BPMFI
501 effectively separated PMF from background elements, whereas PMF detected by other indices was easily
502 confused with the background.

503 The evaluation results indicated that BPMFI had the highest robustness in identifying PMF under



504 various environmental conditions. It could more effectively highlight PMF and suppress Non-PMF than the
505 other six indices. The performance of MBPMFI ranked only next to BPMFI. Although MBPMFI was useful
506 for identifying PMF in Region I and Region II, it failed to suppress the complex backgrounds in Region III.
507 The other five indices were only effective in identifying PMF in specific regions, lacking the capability to
508 consistently recognize PMF under complicated background conditions.



509
510 **Figure 10.** Visual evaluation results based on the plastic-mulched farmland indices of PMLI (b), PMFI_{first} (c), PMFI_{second}
511 (d), PMLI_{SWIR} (e), mPMCI (f), MBPMFI (g), and BPMFI (h) in a representative area of Xianyang City, Shaanxi Province
512 (a). The red color is related to PMF, and the blue color is associated with backgrounds under all of the seven indices.

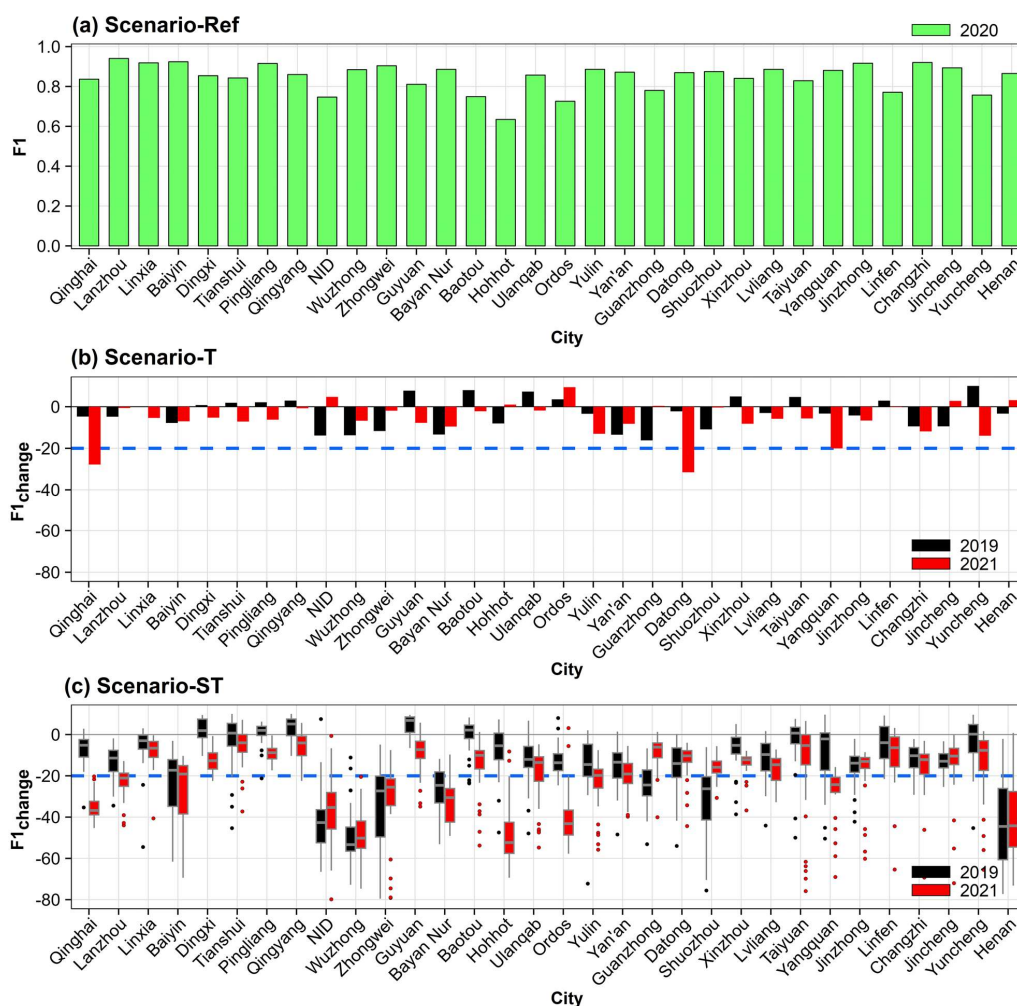
513 4.2 Classifier performance for the transferability scenarios

514 All locally adaptive classifiers in the reference scenario (Scenario-Ref) showed satisfactory accuracy
515 in PMF mapping (Fig. 11 (a)). In this scenario, the F1 values of PMF ranged from 0.64 to 0.94, with an
516 average of 0.85. For all cities, the F1 values were close to or above 0.80, except in the NID (0.75), Baotou
517 (0.75), Hohhot (0.64), and Ordos (0.73). The high F1 values indicated that the locally adaptive classifiers
518 trained in this study can meet the practical demands for actual PMF mapping.

519 The F1 values from each transferability scenario (Scenario-T, Scenario-ST) were compared with
520 those from the reference scenario (Scenario-Ref) in each city, and the percentage change in the F1 values
521 (F1_{change}) was also computed (Fig. 11 (b) and (c)). Generally, the comparison results showed that for all
522 cities, the F1 values decreased from the Scenario-Ref to Scenario-ST in an orderly manner (Scenario-ST >
523 Scenario-T > Scenario-Ref). In Scenario-T (Fig. 11 (b)), the average F1_{change} was -2.92% for 2019 and -
524 6.07% for 2021. Except in Qinghai, Datong, and Yangquan, where the F1_{change} values were below -20% in
525 2021, the F1_{change} values for all other cities were above -20%. In Scenario-ST (Fig. 11 (c)), the average
526 F1_{change} was -13.86% for 2019 and -21.94% for 2021, which was absolutely 10.94% and 15.87% lower than
527 in Scenario-T for 2019 and 2021, respectively. Furthermore, one-third of the cities in the Loess Plateau
528 showed a percentage decline in the F1 values over -20% in Scenario-ST. The greater decrease in F1 values



529 in Scenario-ST compared to Scenario-T indicated that Scenario-ST was hardly suitable for the mapping
 530 of PMF distributions across different years.



531
 532 **Figure 11.** Transferability evaluation of the pre-trained classifiers for plastic-mulched farmland (PMF) mapping. (a) F1
 533 values of PMF mapping based on locally adaptive classifiers trained with data in 2020 in each city of the Loess Plateau
 534 (Scenario-Ref). (b) Percentage change of F1 values ($F1_{\text{change}}$) for temporal transferability scenario (Scenario-T). (c)
 535 $F1_{\text{change}}$ for spatial-temporal transferability scenario (Scenario-ST).

536 4.3 Accuracy assessment for resultant maps

537 Given the strong performance of the temporal classifier transfer (Scenario-T) in multi-year PMF
 538 mapping, classifiers trained on automatically generated samples and optimal features in 2020 were
 539 exploited to mapping PMF distributions (PMF-LP) for 2019–2021. The accuracy assessment for the three
 540 years was implemented based on the validation samples of the entire study area (Table 3). The OA of the
 541 resultant maps for all three years varied from 0.82 (2021) to 0.87 (2020). PMF was accurately identified,



542 with F1 values ranging from 0.80 in 2021 to 0.86 in 2020, averaging 0.83 over the three years. For all three-
543 year classification results, PMF exhibited higher PA than UA, indicating that commission errors in PMF
544 recognition were higher than the omission errors. The higher commission errors indicated that some Non-
545 PMF pixels were incorrectly classified as PMF, potentially leading to an overestimation of PMF areas.
546 Detailed pixel-wise accuracy assessments for each city in the Loess Plateau were presented in Table S4–S6.

547

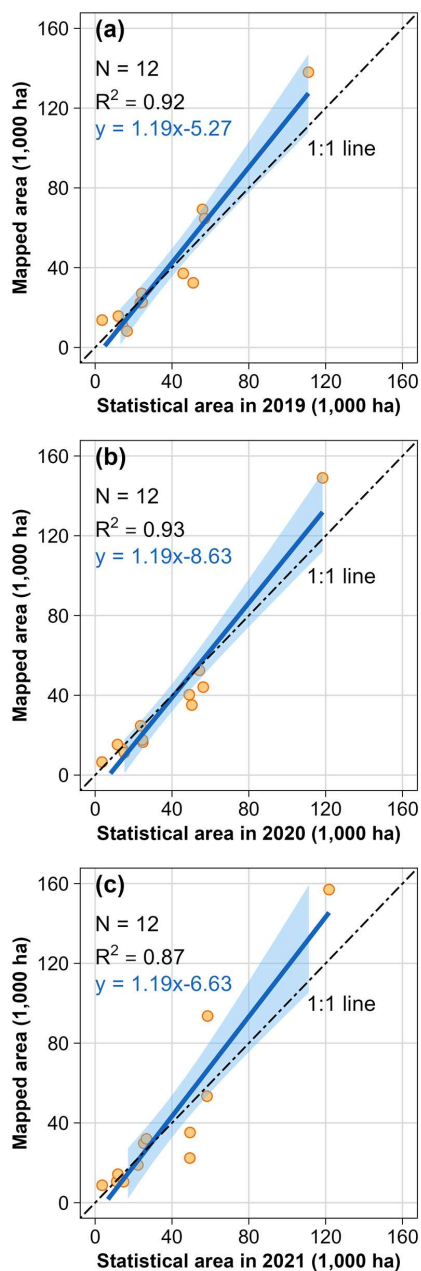
548 **Table 3.**

549 Confusion matrix of the plastic-mulched farmland maps of the Loess Plateau (PMF-LP) for the year 2019–2021. Reference
550 categories are rows while map categories are columns. The numbers in the confusion matrix are the number of ground
551 reference samples. The producer accuracy (PA), user accuracy (UA), F1-score (F1), and overall accuracy (OA) were also
552 displayed in the table. PMF: Plastic-mulched Farmland, Non-PMF: Non-plastic-mulched Farmland.

| Year | | PMF | Non-PMF | PA | UA | F1 | OA |
|------|---------|------|---------|------|------|------|------|
| 2019 | PMF | 2685 | 414 | 0.87 | 0.79 | 0.83 | 0.84 |
| | Non-PMF | 714 | 3278 | 0.82 | 0.89 | 0.85 | |
| 2020 | PMF | 4678 | 664 | 0.88 | 0.84 | 0.86 | 0.87 |
| | Non-PMF | 918 | 5880 | 0.87 | 0.90 | 0.88 | |
| 2021 | PMF | 2468 | 455 | 0.84 | 0.76 | 0.80 | 0.82 |
| | Non-PMF | 768 | 3023 | 0.80 | 0.87 | 0.83 | |

553

554 The estimated areas derived from our maps were compared with the statistical data from yearbooks of
555 12 cities for 2019–2021 (Fig. 12). The PMF areas derived from our maps exhibited high consistency with
556 the statistics, with coefficients of determination (R^2) of 0.92, 0.93, and 0.87 for the year 2019, 2020, and
557 2021, respectively. Notably, the resultant maps for 2021 (Fig. 12 (c)) tended to overestimate PMF areas
558 compared to the statistics, particularly in Qinghai and northern Shanxi, likely due to snow and cloud
559 residuals being misclassified as PMF (discussed in Section 5.3). The strong consistency between the
560 estimated areas and the statistical data underscored the reliability of the maps produced in this study.



561

562 **Figure 12.** Comparison of the mapped PMF (Plastic-mulched Farmland) areas with the statistics for 12 cities with
563 PMF area statistics for the year 2019 (a), 2020 (b), and 2021 (c).

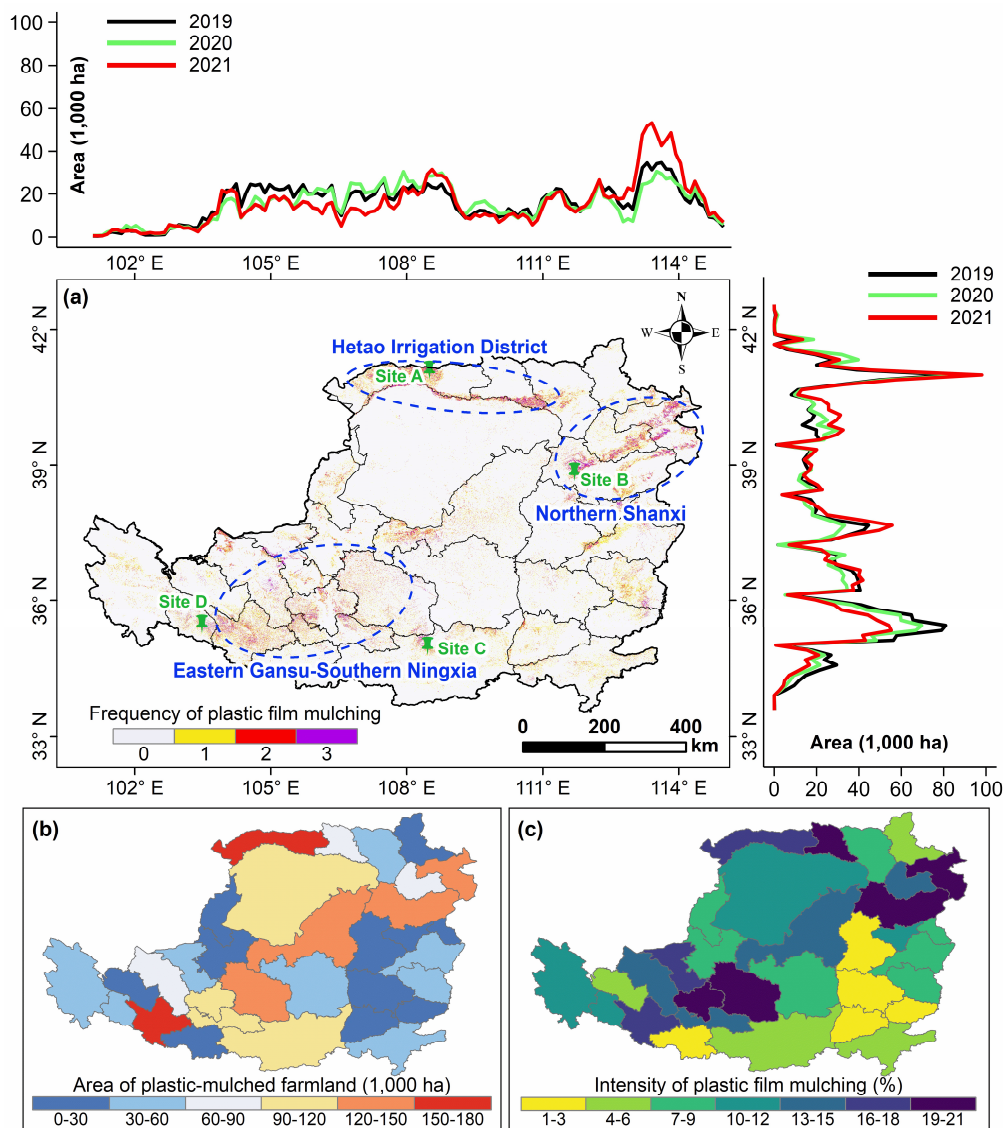
564 4.4 Spatial distributions of PMF

565 The final PMF mapping results were shown in Fig. 13. To quantify the intensity of plastic film usage
566 across different cities in the Loess Plateau, we calculated the ratio of plastic-mulched area (derived from



567 the resultant maps) to the total cultivated land area (derived from cropland layers in Section 2.2.2) for each
568 city (Fig. 13 (c)). PMF was widely distributed across the Loess Plateau, exhibiting a pattern of extensive
569 dispersion and localized clustering (Fig. 13 (a)). The Hetao Irrigation District, including the Bayan Nur and
570 Baotou cities in Inner Mongolia Province, exhibited the highest intensity of plastic film usage at 18%,
571 followed by Northern Shanxi (including Datong, Shuozhou, and Xinzhou) at 17%, and the Eastern Gansu-
572 Southern Ningxia region (including Dingxi, Baiyin, Pingliang, Qingyang, Zhongwei, and Guyuan) at 16%
573 (Fig. 13 (c)). In these arid and cold regions, plastic films have been extensively used for decades to ensure
574 crop yields by regulating temperature and conserving moisture (Wang et al., 2022a; Yan et al., 2014; Zhao
575 et al., 2023). The total PMF areas for each 0.10° longitude and latitude bin were also presented on the top
576 and right sides of the 2020 resultant map (Fig. 13 (a)). Plastic film was extensively used in latitude ranges
577 of 35° – 37° N and 40° – 41° N, and in the longitude ranges of 106° – 109° E and 112° – 114° E. However,
578 plastic film usage in the latitude range of 38° – 40° N and longitude range of 109° – 111° E was comparatively
579 lower than other regions. This disparity could be attributed to the prevalent local land cover types in Ordos
580 of Inner Mongolia Province, where grassland and bare land account for about 90% of the total area (Zhang
581 et al., 2021c; Yang and Huang, 2021).

582 To further validate the PMF mapping results, a more detailed visual evaluation was conducted by
583 selecting representative areas in different years (Fig. 14). Obviously, PMF was effectively distinguished
584 from other land cover types, regardless of various cropland sizes and shapes. In site A, Bayan Nur, Inner
585 Mongolia Province, PMF was densely distributed in regular rectangles (Fig. 14 (a)). There were clear
586 separations between PMF and roads in the mapping results. When the classifiers trained in 2020 were
587 applied to 2019 and 2021, some PMF pixels in site A were not identified in 2019 (Fig. 14 (b)), while in
588 2021, there were instances of erroneously classifying Non-PMF pixels as PMF (Fig. 14 (f)). Site B in
589 Xinzhou of Shanxi Province, had large fields with irregular shapes compared to site A (Fig. 14 (g))
590 compared with site A. The PMF mapping results from 2019 to 2021 at this site showed strong spatial
591 agreement with the Sentinel-2 true-color composite images (Fig. 14 (h), (j), and (l)). In site C, Xianyang,
592 Shaanxi Province, PMF was scattered among winter crops (Fig. 14 (m)). The mapping results showed good
593 separability between winter crops and PMF across different years (Fig. 14 (n), (p), and (r)). For site D in
594 Linxia of Gansu Province, the cultivated farmland was primarily characterized by terraced fields, presenting
595 elongated and irregular shapes (Fig. 14 (s)). These terraced fields were interspersed with vegetations,
596 buildings, and bare lands. The mapping results were also satisfactory in this site (Fig. 14 (t), (v), and (x)).



597

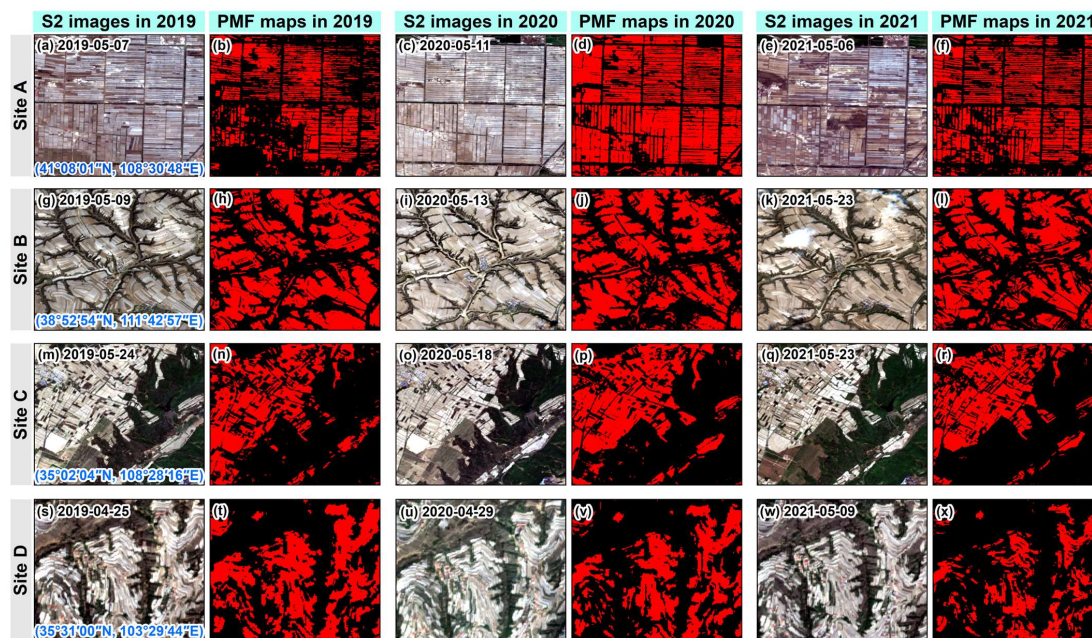
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Figure 13. Spatial patterns of plastic-mulched farmland (PMF) in the Loess Plateau from 2019 to 2021. (a) Frequency of plastic film mulching. The blue dashed ellipses represent the regions with dense distributions of PMF. Site A, B, C, and D are the zoom-in view cases in Fig. 14. (b) Area of PMF in each city in 2020. (c) Intensity of plastic film mulching in 2020 across various cities. Intensity of plastic film mulching is defined as the ratio of PMF area to total cultivated area.



602
603 **Figure 14.** Spatial details of the plastic-mulched farmland (PMF) map subsets from different regions in the years of 2019,
604 2020, and 2021. Site A, B, C, and D correspond to the four typical regions marked in Fig 13. The 1st, 3rd, and 5th columns
605 are the Sentinel-2 (S2) images with a true-color composite of red, green, blue bands in the years of 2019, 2020, and 2021,
606 respectively. The 2nd, 4th, and 6th columns are the PMF maps established in this study.

607 5 Discussion

608 5.1 Feasibility of MBPMFI and BPMFI for automatic training sample generation

609 Image sources, algorithms, and training samples are three crucial factors influencing any land cover
610 classification task (Huang et al., 2020; Wen et al., 2022; Zhang et al., 2021b). The advancement of remote
611 sensing technology has provided access to massive open datasets and powerful cloud-based computing
612 platforms (Song et al., 2017; Yang et al., 2019; Tamiminia et al., 2020). However, large-scale and multi-
613 year land cover classification is still constrained by the lack of training samples (Huang et al., 2020; Wen
614 et al., 2022). This issue is particularly pronounced for land cover types like PMF, which have received less
615 attention in the past, making training sample collection more challenging. To address this gap, we developed
616 two novel indices, MBPMFI and BPMFI, to automatically generate training samples. Compared with
617 traditional field surveys and visual interpretation, the proposed indices offer a more convenient and labor-
618 saving alternative for obtaining training samples. The prior information, such as mulching and flourishing
619 dates, is not strictly time-specific and can be easily obtained from local management experiences and
620 relevant studies. Furthermore, because the two indices rely solely on the blue band, their calculation is
621 simple and fast.



622 Additionally, the visual comparison results under various environmental conditions in this study
623 demonstrated the superior robustness of MBPMFI and BPMFI over the other existing indices in
624 highlighting PMF signals and suppressing background noises. Previous studies have shown that PMF can
625 be effectively distinguished from other land cover types across the visible to shortwave-infrared spectrum
626 during the mulching stage (Lu et al., 2014; Hasituya et al., 2016; Hao et al., 2019). However, PMF indices
627 based solely on optical images from the mulching stage often struggle to differentiate PMF from bare land
628 and impervious surfaces (Lu et al., 2014; Hao et al., 2019) (Fig. 6 (e)~(h)), which limits their applicability
629 in complex environments. In this study, we considered PMF characteristics across the pre-mulching,
630 mulching, and flourishing stages to develop two novel PMF indices. Both indices exhibited greater
631 robustness than the existing ones in distinguishing PMF from various complex backgrounds. Of the two
632 indices, BPMFI was more effective in highlighting PMF signals and suppressing background noise, making
633 it particularly promising for rapid and precise PMF recognition in areas with seamless satellite imagery.
634 However, because the calculation of BPMFI requires images from three different periods, which might be
635 unavailable in cloudy regions, MBPMFI is more practical for PMF recognition when only mulching-stage
636 images are available. To further improve the quality of training samples in this study, we collected samples
637 classified as PMF by both indices, and then conducted strict spatial filtering. The high PMF mapping
638 accuracy ($F1 = 0.86$) and the strong correlation ($R^2 = 0.93$) between estimated and statistical PMF areas in
639 2020 confirm the quality of these training samples.

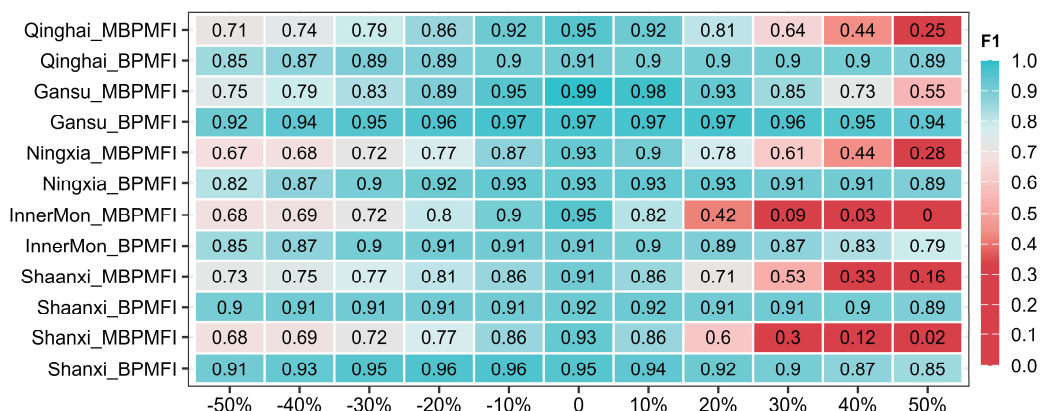
640 **5.2 Threshold stability of the MBPMFI and BPMFI**

641 Threshold used for index-based classification is typically determined using statistical data (Zhang et
642 al., 2022e; Zhang et al., 2022b), ground truth data (Zhang et al., 2022g; Zhou et al., 2024), or automatic
643 threshold determination methods (Yang et al., 2023). In this study, we employed ground truth samples from
644 rectangular regions (Fig. 1 (a)) to determine the threshold for each province of the Loess Plateau, because
645 statistical data were not available in most cities and automatic methods could require significant
646 computational resources at large scales. To evaluate the stability of MBPMFI and BPMFI thresholds, we
647 tested the mapping accuracy in the rectangular regions based on thresholds -50%–50% off the best threshold
648 (Fig. 15). Compared with MBPMFI, BPMFI can achieve more stable performance even with a threshold
649 largely off from the best threshold. The F1 values ranged from 0 to 0.98 for MBPMFI and from 0.80 to
650 0.97 for BPMFI with the thresholds $\pm 50\%$ off from the best threshold across all regions. The stable F1
651 values of BPMFI indicated that it is not sensitive to threshold variations, which makes it well-suited for
652 large-scale PMF mapping.

653 According to the results of our study, the best BPMFI threshold was 0.40 for Qinghai, Shaanxi, and



654 Shanxi, and 0.80–1.10 for Gansu, Ningxia, and Inner Mongolia. The relatively large difference in the best
 655 thresholds may be attributed to variations in plastic materials (Xiong et al., 2019; Zhang et al., 2022g) and
 656 crop species grown in the fields. The former affects the peak reflectance of PMF during the mulching stage,
 657 and the latter influences the absorption of the visible spectrum during the flourishing stage. The more
 658 detailed relationship between these factors and the spectral signatures of PMF deserves to be explored in
 659 the future. Additionally, based on the threshold stability assessment in Fig. 15, we recommend using a
 660 BPMFI threshold of 0.40–0.60 (i.e., approximately 50%–150% of the optimal threshold) and an MBPMFI
 661 threshold of 0.14–0.15 (i.e., approximately 90%–110% of the optimal threshold) for PMF recognition
 662 across the entire Loess Plateau. But for high-precision PMF mapping, it is suggested to segment the whole
 663 area into sub-regions and determine the optimal thresholds separately for each region.



664
 665 **Figure 15.** Plastic-mulched farmland (PMF) mapping accuracies (F1-score) based on thresholds -50%–50% off the best
 666 threshold in the two novel PMF indices of MBPMFI and BPMFI. “±n%” indicates threshold = best threshold ±n% × best
 667 threshold.

668 5.3 Feasibility and error of classifier transfer for PMF mapping

669 Classifiers, pre-trained in specific areas/years with sufficient samples, have great potential for rapid
 670 land cover and crop type mapping in new areas/years where training data are lacking (Ma et al., 2024).
 671 Previous studies on classifier transfer primarily focused on crop type mapping (Wang et al., 2019; Hao et
 672 al., 2020; Yang et al., 2023). In contrast, ground reference samples for PMF are more limited, and the
 673 feasibility of employing classifier transfer for PMF mapping has not been thoroughly explored.
 674 Understanding the extent of performance loss in classifier transfer is crucial for large-scale and multi-year
 675 PMF mapping. Therefore, this study quantitatively evaluated the performance loss when the trained
 676 classifier is directly applied in two transferability scenarios (Scenario-T, Scenario-ST) for PMF mapping.
 677 The findings demonstrated decreased performance in both transferability scenarios compared to the no-
 678 transfer case (Scenario-Ref). The temporal classifier transfer (Scenario-T) was more suitable for retracing



679 historical PMF distributions than spatial–temporal classifier transfer (Scenario–ST), with percentage
680 changes in F1 being less than 7.0% on average compared to the no-transfer case. Several potential factors
681 that may cause performance reduction in Scenario–T and Scenario–ST (Fig. 11) could be categorized into
682 three main groups: (1) missing clear Sentinel-2 data during the mulching stage; (2) snow and cloud residual
683 contaminations during the pre-mulching stage; and (3) variations in mulching dates across different regions
684 and years.

685 First, high-quality Sentinel-2 data during the mulching stage is crucial for enlarging the inter-class
686 variations between PMF and Non-PMF as illustrated in Section 3.1.2. Particularly, the local peak values in
687 the blue band of time-series Sentinel-2 (i.e., Blue_peak) is the most important feature for PMF mapping
688 (Fig. 7). Missing data during the mulching stage could result in the disappearance or distortion of the local
689 peak values, causing classifiers trained on regions/years with high-quality data to fail in mapping PMF in
690 regions/years with missing data. For example, during the mulching stage, more than half of the pixels in
691 Qingyang had more than four clear-sky observations (Fig. 16 (a)), while nearly 75% of the pixels in Henan
692 had fewer than three clear-sky observations (Fig. 16 (b)). Consequently, the local peaks of the harmonic
693 regression curves in Henan (Fig. 16 (e)) were lower than those in Qingyang (Fig. 16 (d)) during 2019–2021,
694 which led to low accuracy when using classifiers trained with Qingyang data from 2020 to recognize PMF
695 in Henan for 2019 ($F1 = 0.64$) and 2021 ($F1 = 0.49$). Although the Sentinel-2 satellite provides data with a
696 5-day revisit cycle and cloud contamination in northern China is less severe than in southern China, the
697 observable duration of plastic film is short and influenced by crop canopy, which leads to the unavailability
698 of images containing plastic film signals. Multi-sensor satellite data could be employed to increase the
699 density of observations to account for this limitation.

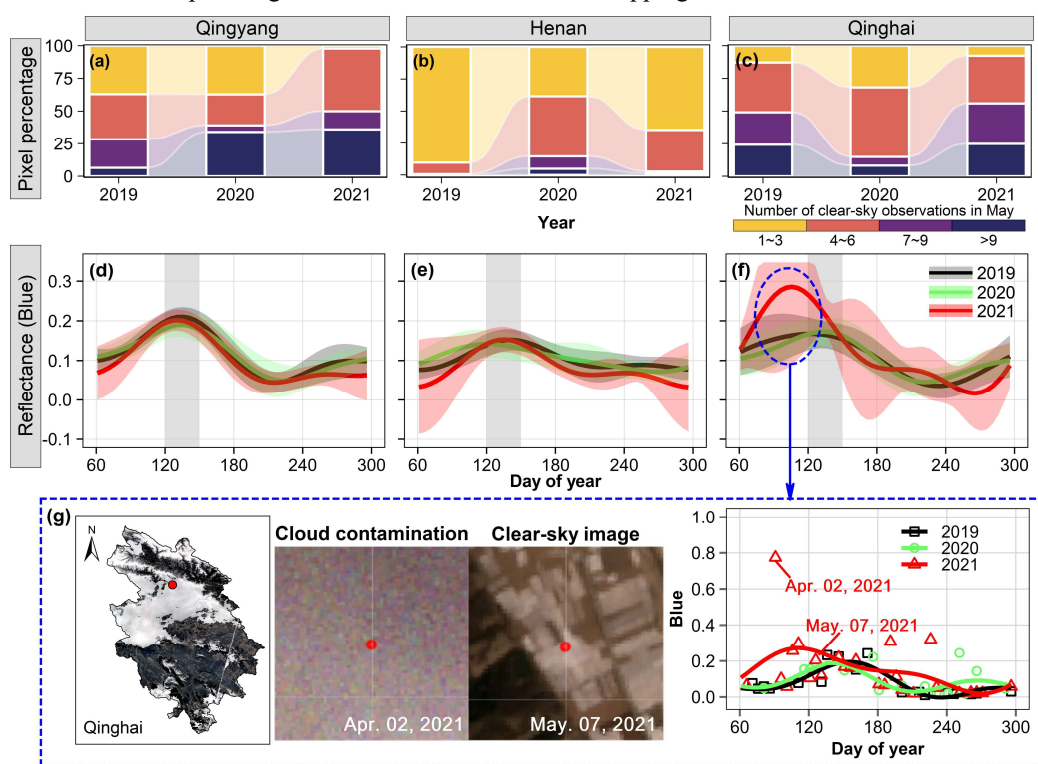
700 Next, clouds and snow, which exhibit similar bright characteristics in all visible bands as PMF, are
701 difficult to be thoroughly masked out using only the quality assessment (QA) bands of Sentinel-2 (You and
702 Dong, 2020). For example, although the percentage of clear-sky observations in Qinghai was the same as
703 in Qingyang (Fig. 16 (a) and (c)), snow and cloud residuals distorted the harmonic regression curves in
704 Qinghai in 2021 (Fig. 16 (f) and (g)), leading to the performance loss when using classifiers trained with
705 Qinghai (or other city) data from 2020 to recognize PMF in the same region for 2021 (Fig. 11 (b) and (c)).
706 To mitigate the negative effects of clouds and snow on classifier transfer for PMF mapping, more effective
707 cloud detection algorithms should be implemented. Inspired by You and Dong (2020), two spectral indices
708 (Normalized Difference Moisture Index (NDMI) and Normalized Difference Snow Index (NDSI)) could
709 be used to mask out clouds to improve the performance of classifier transfer in the future.

710 Finally, variations in sowing dates of plastic-mulched crops among different regions could lead to
711 changes in mulching dates, which in turn affect the occurrence dates of local peaks and the shape of
712 harmonic regression curves. In this study, in the Ningxia Irrigation District (NID, Fig. S1) and Wuzhong of



713 Ningxia Province, mulching dates were earlier than in other cities (Fig. S11) due to the widespread
 714 cultivation of watermelon, which is mainly sown in March. Consequently, the performance exhibited a
 715 significant decrease (median $F1_{\text{change}} \approx -50\%$) when using classifiers trained with data from other city in
 716 2020 to recognized PMF in the two cities (Scenario-ST) for 2019 and 2021.

717 Apart from the main factors mentioned above, climate variables (e.g., rainfall, temperature), farming
 718 practices (e.g., sowing and harvesting practices), technological evolution, and policy implementation could
 719 also impact the mulching practices, which may adversely affect the performance of classifier transfer. The
 720 links between classifier transferability and these factors require further exploration. Additionally, multi-
 721 year/multi-region data have been employed to enhance classifier transferability for crop type mapping
 722 (Wijesingha et al., 2024; Orynbaikyzy et al., 2022; Luo et al., 2022). Considering the increasing focus on
 723 PMF mapping (Veettil et al., 2023), there will likely be more training data from different years and regions
 724 available to develop more generalized classifiers for PMF mapping in the future.



725
 726 **Figure 16.** Explanations for the failures of Plastic-mulched farmland (PMF) mapping caused by data missing and cloud
 727 contamination. (a)~(c) Percentage of clear-sky observations at each cropland pixel during the mulching stage in Qingyang,
 728 Henan, and Qinghai. (d)~(f) Harmonic regression curves of PMF in Qingyang, Henan, and Qinghai for the year 2019,
 729 2020 and 2021. The gray buffer represents mulching period. The black, green, and red buffers indicate one standard
 730 deviation. (g) A representative example in Qinghai for depicting the adverse effects of snow and clouds on PMF mapping.

731



732 **5.4 Advantages and uncertainties of the new framework**

733 In this study, we proposed a new framework that combines index-based methods with supervised
734 classifier transfer for large-scale PMF mapping. This framework offers two main advantages. First, we
735 introduced two novel and robust PMF indices (MBPMFI and BPMFI) for generating training samples,
736 significantly enhancing the efficiency of sample collection compared to field surveys and visual
737 interpretation. Second, temporally transfer pre-trained classifiers to mapping multi-year PMF distributions
738 further facilitates the long-term PMF monitoring without training domain- and year-specific classifiers. The
739 automatic generation of training samples and classifier transfer make this framework highly promising for
740 rapidly mapping large-scale and multi-year PMF distributions in practice. The satisfactory mapping
741 accuracy of PMF ($F1 > 0.80$) in the Loess Plateau during 2019–2021 further demonstrated the reliability of
742 this framework. Additionally, since the reflectance bands used for training sample generation and feature
743 extraction are similar to those of Landsat satellites, this framework is also believed to be applicable for
744 PMF mapping with Landsat data.

745 Although the new framework successfully identified PMF across different regions and years in the
746 Loess Plateau of China, there still exists some room for improvement. First, two different cropland layers
747 from 2020 were integrated to exclude the non-cropland pixels for 2019–2021 before PMF recognition.
748 However, errors in these products may affect PMF mapping results. For instance, unrecognized cropland
749 pixels in the products could lead to the loss of potential PMF pixels, thereby increasing the omission errors
750 in the PMF mapping results. One possible solution is to integrate more state-of-the-art land cover products
751 (Tu et al., 2024; Zhang et al., 2024) to enhance the reliability of the cropland mask.

752 Next, although Sentinel-2 has a 5-day revisit cycle, clear-sky observations are still often limited in
753 some regions, especially in southern China, due to inevitable cloud contaminations. The large-scale loss of
754 images during the mulching stage could constrain the application of the proposed framework. To address
755 this issue, more optical satellite imagery, such as Landsat-4/5/7/8/9, could be integrated to provide denser
756 time-series observations in the future. Furthermore, this study did not include radar data which is
757 impervious to atmospheric interference in PMF mapping. Incorporating radar data for PMF mapping will
758 be the next step of our research.

759 Third, this study focused on evaluating temporal and spatial-temporal transferability of the RF
760 classifier for PMF mapping. Due to the limitations of GEE platform, deep learning algorithms such as Deep
761 Neural Network (DNN) and Convolution Neural Network (CNN), which exhibit better performance than
762 the traditional supervised classifiers (Zhong et al., 2019; Wang et al., 2021), were not explored in this study.
763 However, employing deep learning algorithms for PMF mapping presents challenges related to



764 computational costs and the need for more training data (Xu et al., 2021; Zhang et al., 2020a; Wijesingha
765 et al., 2024).

766 **6 Data availability**

767 The PMF-LP product generated in this paper is openly available at
768 <https://doi.org/10.5281/zenodo.13369426> (Zhao et al., 2024). The dataset includes a set of GeoTIFF images
769 in the EPSG:4326 spatial reference system. The value 0 and 1 represent PMF and Non-PMF, respectively.
770 We encourage users to independently verify the PMF maps.

771 **7 Conclusions**

772 Plastic film mulching technology can effectively enhance crop yield, but also poses significant
773 environmental issues, such as “white pollution” and soil microplastic contamination. Unfortunately, there
774 is no method currently available for the rapid and automated mapping of plastic-mulched farmland (PMF)
775 across large areas and multiple years. In this study, to address the challenge of lacking ground reference
776 samples to train supervised classifiers, we developed an automatic training sample generation method based
777 on two newly proposed PMF indices: MBPMFI and BPMFI. To rapidly retrace historical PMF distributions,
778 we thoroughly explored the spatial–temporal transferability of pre-trained classifiers and the factors
779 influencing their transferability. Finally, PMF distribution maps for the Loess Plateau of China were
780 produced by integrating the automatic training sample generation method with the classifier transfer
781 approach.

782 The results of this study indicated that the novel PMF indices, MBPMFI and BPMFI, outperformed
783 the existing PMF indices in PMF recognition. The novel indices could not only highlight PMF information
784 but also suppress complex background signals under various environmental conditions. Temporal classifier
785 transfer (Scenario–T) showed great potential for rapid mapping of multi-year PMF distributions in regions
786 and years lacking ground reference samples, with an average performance loss less than 7.0% compared to
787 the locally adaptive classifiers. Based on the index-based training samples generation method and classifier
788 transfer, PMF distributions were generated across the Loess Plateau (PMF-LP) and achieved satisfactory
789 accuracy, with F1 values of 0.83, 0.86, and 0.80 for 2019, 2020, and 2021, respectively. The estimated PMF
790 areas also agreed well with the municipal-level agricultural census data ($R^2 \geq 0.87$). In summary, the newly
791 proposed PMF mapping framework has great potential to rapidly mapping PMF distributions across large
792 areas and multiple years. Future work will apply this framework to monitor the spatiotemporal dynamics
793 of PMF distributions in China.

794



795 **Supplement.**

796 The supplement related to this article is available online at: <https://doi.org/>

797 **Author contributions.**

798 CZ and JH conceptualized the study. CZ designed and carried out the experiment. CZ wrote the
799 original manuscript. YL, XC, and LW collected dataset and implemented formal analysis. ZW, HF, and QY
800 provided direction and comments. JH was responsible for funding acquisition. All authors discussed the
801 results and revised the manuscript.

802 **Competing interests.**

803 The contact author has declared that none of the authors has any competing interests.

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806 published maps and institutional affiliations.

807 **Acknowledgements.**

808 The author gratefully acknowledges all data providers whose data have been used in this study.
809 Then we thank the support of the Google Earth Engine platform, which provided essential data and
810 computational resources for this study.

811 **Financial support**

812 This research was supported by National Natural Science Foundation of China (No. 52079115), the
813 National Key Research and Development Program of China (No. 2021YFD1900700), and the "111 Project"
814 (No. B12007) of China.

815

816 **References**

- 817 Beck, H. E., McVicar, T. R., Vergopolan, N., Berg, A., Lutsko, N. J., Dufour, A., Zeng, Z., Jiang, X., van Dijk, A. I., and Miralles,
818 D. G.: High-resolution (1 km) Köppen-Geiger maps for 1901–2099 based on constrained CMIP6 projections, *Sci. Data*,
819 10, 724, <https://doi.org/10.1038/s41597-023-02549-6>, 2023.
- 820 Belgiu, M. and Drăguț, L.: Random forest in remote sensing: A review of applications and future directions, *ISPRS J.*
821 *Photogramm. Remote Sens.*, 114, 24–31, <https://doi.org/10.1016/j.isprsjprs.2016.01.011>, 2016.
- 822 Breiman, L.: Random forests, *Mach. Learn.*, 45, 5–32, 2001.
- 823 Chaves, M. E. D., Picoli, M. C. A., and Sanches, I. D.: Recent applications of Landsat 8/OLI and Sentinel-2/MSI for land use
824 and land cover mapping: A systematic review, *Remote Sens.*, 12, 3062, <https://doi.org/10.3390/rs12183062>, 2020.
- 825 Cheng, L., Fu, C., Xi, H., Chang, L., Li, S., and Qin, S.: Spatiotemporal variations of plastic-mulched cropland in Hexi Corridor
826 using multi-source remote sensing data (in Chinese), *Transactions of the Chinese Society of Agricultural Engineering*,



- 827 39, 124-131, <https://doi.org/10.11975/j.issn.1002-6819.202303209>, 2023.
- 828 Defourny, P., Bontemps, S., Bellemans, N., Cara, C., Dedieu, G., Guzzonato, E., Hagolle, O., Inglada, J., Nicola, L., Rabaute, T.,
829 Savinaud, M., Udroui, C., Valero, S., Bégué, A., Dejoux, J.-F., Harti, A. E., Ezzahar, J., Kussul, N., Labbassi, K.,
830 Lebourgeois, V., Miao, Z., Newby, T., Nyamugama, A., Salh, N., Shelestov, A., Simonneaux, V., Traore, P. S., Traore,
831 S. S., and Koetz, B.: Near real-time agriculture monitoring at national scale at parcel resolution: Performance assessment
832 of the Sen2-Agri automated system in various cropping systems around the world, *Remote Sens. Environ.*, 221, 551-
833 568, <https://doi.org/10.1016/j.rse.2018.11.007>, 2019.
- 834 Deng, C. and Wu, C.: BCI: A biophysical composition index for remote sensing of urban environments, *Remote Sens. Environ.*,
835 127, 247-259, <https://doi.org/10.1016/j.rse.2012.09.009>, 2012.
- 836 Drusch, M., Del Bello, U., Carlier, S., Colin, O., Fernandez, V., Gascon, F., Hoersch, B., Isola, C., Laberinti, P., and Martimort,
837 P.: Sentinel-2: ESA's optical high-resolution mission for GMES operational services, *Remote Sens. Environ.*, 120, 25-
838 36, <https://doi.org/10.1016/j.rse.2011.11.026>, 2012.
- 839 Farr, T. G., Rosen, P. A., Caro, E., Crippen, R., Duren, R., Hensley, S., Kobrick, M., Paller, M., Rodriguez, E., Roth, L., Seal, D.,
840 Shaffer, S., Shimada, J., Umland, J., Werner, M., Oskin, M., Burbank, D., and Alsdorf, D.: The shuttle radar topography
841 mission, *Rev. Geophys.*, 45, <https://doi.org/10.1029/2005RG000183>, 2007.
- 842 Feng, Q., Niu, B., Zhu, D., Yao, X., Liu, Y., Ou, C., Cheng, B., Yang, J., Guo, H., and Liu, J.: A dataset of remote sensing-based
843 classification for agricultural plastic greenhouses in China in 2019 (in Chinese), *China Scientific Data*, 6,
844 <https://doi.org/10.11922/noda.2021.0009.zh>, 2021.
- 845 Fiset, T., Rollin, P., Aly, Z., Campbell, L., Daneshfar, B., Filyer, P., Smith, A., Davidson, A., Shang, J., and Jarvis, I.: AAFC
846 annual crop inventory, 2013 Second International Conference on Agro-Geoinformatics (Agro-Geoinformatics), 270-
847 274,
- 848 Foody, G. M. and Arora, M. K.: An evaluation of some factors affecting the accuracy of classification by an artificial neural
849 network, *Int. J. Remote Sens.*, 18, 799-810, <https://doi.org/10.1080/014311697218764>, 1997.
- 850 Fu, C., Cheng, L., Qin, S., Tariq, A., Liu, P., Zou, K., and Chang, L.: Timely plastic-mulched cropland extraction method from
851 complex mixed surfaces in arid regions, *Remote Sens.*, 14, 4051, <https://doi.org/10.3390/rs14164051>, 2022.
- 852 Gao, H. H., Yan, C. R., Liu, Q., Ding, W. L., Chen, B. Q., and Li, Z.: Effects of plastic mulching and plastic residue on agricultural
853 production: A meta-analysis, *Sci. Total Environ.*, 651, 484-492, <https://doi.org/10.1016/j.scitotenv.2018.09.105>, 2019.
- 854 Gao, Y., Pan, Y., Zhu, X., Li, L., Ren, S., Zhao, C., and Zheng, X.: FARM: A fully automated rice mapping framework combining
855 Sentinel-1 SAR and Sentinel-2 multi-temporal imagery, *Comput. Electron. Agric.*, 213, 108262,
856 <https://doi.org/10.1016/j.compag.2023.108262>, 2023.
- 857 Gitelson, A. A., Viña, A., Ciganda, V., Rundquist, D. C., and Arkebauer, T. J.: Remote estimation of canopy chlorophyll content
858 in crops, *Geophys. Res. Lett.*, 32, <https://doi.org/10.1029/2005GL022688>, 2005.
- 859 Gong, P., Liu, H., Zhang, M., Li, C., Wang, J., Huang, H., Clinton, N., Ji, L., Li, W., Bai, Y., Chen, B., Xu, B., Zhu, Z., Yuan, C.,
860 Ping Suen, H., Guo, J., Xu, N., Li, W., Zhao, Y., Yang, J., Yu, C., Wang, X., Fu, H., Yu, L., Dronova, I., Hui, F., Cheng,
861 X., Shi, X., Xiao, F., Liu, Q., and Song, L.: Stable classification with limited sample: transferring a 30-m resolution
862 sample set collected in 2015 to mapping 10-m resolution global land cover in 2017, *Sci. Bull.*, 64, 370-373,
863 <https://doi.org/10.1016/j.scib.2019.03.002>, 2019.
- 864 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., and Moore, R.: Google Earth Engine: Planetary-scale
865 geospatial analysis for everyone, *Remote Sens. Environ.*, 202, 18-27, <https://doi.org/10.1016/j.rse.2017.06.031>, 2017.
- 866 Han, J., Zhang, Z., Luo, Y., Cao, J., Zhang, L., Zhang, J., and Li, Z.: The RapeseedMap10 database: annual maps of rapeseed at
867 a spatial resolution of 10 m based on multi-source data, *Earth Syst. Sci. Data*, 13, 2857-2874,
868 <https://doi.org/10.5194/essd-13-2857-2021>, 2021.
- 869 Hao, P., Di, L., Zhang, C., and Guo, L.: Transfer Learning for Crop classification with Cropland Data Layer data (CDL) as
870 training samples, *Sci. Total Environ.*, 733, 138869, <https://doi.org/10.1016/j.scitotenv.2020.138869>, 2020.
- 871 Hao, P., Chen, Z., Tang, H., Li, D., and Li, H.: New workflow of plastic-mulched farmland mapping using multi-temporal
872 Sentinel-2 data, *Remote Sens.*, 11, 1353, <https://doi.org/10.3390/rs11111353>, 2019.



- 873 Hasituya and Chen, Z.: Mapping plastic-mulched farmland with multi-temporal Landsat-8 data, *Remote Sens.*, 9, 557,
874 <https://doi.org/10.3390/rs9060557>, 2017.
- 875 Hasituya, Chen, Z. X., Li, F., and Hu, Y. C.: Mapping plastic-mulched farmland by coupling optical and synthetic aperture radar
876 remote sensing, *Int. J. Remote Sens.*, 41, 7757-7778, 2020.
- 877 Hasituya, Chen, Z., Wang, L., Wu, W., Jiang, Z., and Li, H.: Monitoring plastic-mulched farmland by Landsat-8 OLI imagery
878 using spectral and textural features, *Remote Sens.*, 8, 353, <https://doi.org/10.3390/rs8040353>, 2016.
- 879 Hu, Y., Zeng, H., Tian, F., Zhang, M., Wu, B., Gilliams, S., Li, S., Li, Y., Lu, Y., and Yang, H.: An interannual transfer learning
880 approach for crop classification in the Hetao Irrigation District, China, *Remote Sens.*, 14, 1208,
881 <https://doi.org/10.3390/rs14051208>, 2022.
- 882 Huang, H., Wang, J., Liu, C., Liang, L., Li, C., and Gong, P.: The migration of training samples towards dynamic global land
883 cover mapping, *ISPRS J. Photogramm. Remote Sens.*, 161, 27-36, <https://doi.org/10.1016/j.isprsjprs.2020.01.010>, 2020.
- 884 Jakubauskas, M. E., Legates, D. R., and Kastens, J. H.: Harmonic analysis of time-series AVHRR NDVI data, *Photogramm. Eng.*
885 *Remote Sens.*, 67, 461-470, 2001.
- 886 Kumar, M., Xiong, X., He, M., Tsang, D. C. W., Gupta, J., Khan, E., Harrad, S., Hou, D., Ok, Y. S., and Bolan, N. S.: Microplastics
887 as pollutants in agricultural soils, *Environ. Pollut.*, 265, 114980, <https://doi.org/10.1016/j.envpol.2020.114980>, 2020.
- 888 Li, P., Li, H., Si, B., Zhou, T., Zhang, C., and Li, M.: Mapping planted forest age using LandTrendr algorithm and Landsat 5–8
889 on the Loess Plateau, China, *Agric. For. Meteorol.*, 344, 109795, <https://doi.org/10.1016/j.agrformet.2023.109795>, 2024.
- 890 Li, Z., Zheng, F., Liu, W., and Jiang, D.: Spatially downscaling GCMs outputs to project changes in extreme precipitation and
891 temperature events on the Loess Plateau of China during the 21st Century, *Glob. Planet. Change*, 82-83, 65-73,
892 <https://doi.org/10.1016/j.gloplacha.2011.11.008>, 2012.
- 893 Liu, E. K., He, W. Q., and Yan, C. R.: ‘White revolution’ to ‘white pollution’—agricultural plastic film mulch in China, *Environ.*
894 *Res. Lett.*, 9, 091001, <https://doi.org/10.1088/1748-9326/9/9/091001>, 2014.
- 895 Liu, H., Gong, P., Wang, J., Clinton, N., Bai, Y., and Liang, S.: Annual dynamics of global land cover and its long-term changes
896 from 1982 to 2015, *Earth Syst. Sci. Data*, 12, 1217-1243, 2020.
- 897 Liu, W. and Zhang, H.: Mapping annual 10 m rapeseed extent using multisource data in the Yangtze River Economic Belt of
898 China (2017–2021) on Google Earth Engine, *Int. J. Appl. Earth Obs. Geoinf.*, 117, 103198,
899 <https://doi.org/10.1016/j.jag.2023.103198>, 2023.
- 900 Lu, L., Hang, D., and Di, L.: Threshold model for detecting transparent plastic-mulched landcover using moderate-resolution
901 imaging spectroradiometer time series data: a case study in southern Xinjiang, China, *J. Appl. Remote Sens.*, 9, 097094,
902 <https://doi.org/10.1117/1.JRS.9.097094>, 2015.
- 903 Lu, L., Tao, Y., and Di, L.: Object-based plastic-mulched landcover extraction using integrated Sentinel-1 and Sentinel-2 data,
904 *Remote Sens.*, 10, 1820, <https://doi.org/10.3390/rs10111820>, 2018.
- 905 Lu, L. Z., Di, L. P., and Ye, Y. M.: A decision-tree classifier for extracting transparent plastic-mulched landcover from Landsat-
906 5 TM images, *IEEE J. Sel. Top. Appl. Earth Observ. Remote Sens.*, 7, 4548-4558,
907 <https://doi.org/10.1109/JSTARS.2014.2327226>, 2014.
- 908 Luo, Y., Zhang, Z., Zhang, L., Han, J., Cao, J., and Zhang, J.: Developing high-resolution crop maps for major crops in the
909 European Union based on transductive transfer learning and limited ground data, *Remote Sens.*, 14, 1809,
910 <https://doi.org/10.3390/rs14081809>, 2022.
- 911 Ma, Y., Chen, S., Ermon, S., and Lobell, D. B.: Transfer learning in environmental remote sensing, *Remote Sens. Environ.*, 301,
912 113924, <https://doi.org/10.1016/j.rse.2023.113924>, 2024.
- 913 Maselli, F., Battista, P., Chiesi, M., Rapi, B., Angeli, L., Fibbi, L., Magno, R., and Gozzini, B.: Use of Sentinel-2 MSI data to
914 monitor crop irrigation in Mediterranean areas, *Int. J. Appl. Earth Obs. Geoinf.*, 93, 102216,
915 <https://doi.org/10.1016/j.jag.2020.102216>, 2020.
- 916 Orynbaikyzy, A., Gessner, U., and Conrad, C.: Spatial transferability of random forest models for crop type classification using
917 Sentinel-1 and Sentinel-2, *Remote Sens.*, 14, 1493, <https://doi.org/10.3390/rs14061493>, 2022.
- 918 Pan, S. J. and Yang, Q.: A survey on transfer learning, *IEEE Trans. Knowl. Data Eng.*, 22, 1345-1359,



- 919 <https://doi.org/10.1109/TKDE.2009.191>, 2009.
- 920 Pham-Duc, B., Nguyen, H., Phan, H., and Tran-Anh, Q.: Trends and applications of google earth engine in remote sensing and
921 earth science research: A bibliometric analysis using scopus database, *Earth Sci. Inform.*, 16, 2355-2371,
922 <https://doi.org/10.1007/s12145-023-01035-2>, 2023.
- 923 Phiri, D., Simwanda, M., Salekin, S., Nyirenda, V. R., Murayama, Y., and Ranagalage, M.: Sentinel-2 data for land cover/use
924 mapping: A review, *Remote Sens.*, 12, 2291, <https://doi.org/10.3390/rs12142291>, 2020.
- 925 Qadir, A., Skakun, S., Becker-Reshef, I., Kussul, N., and Shelestov, A.: Estimation of sunflower planted areas in Ukraine during
926 full-scale Russian invasion: Insights from Sentinel-1 SAR data, *Science of Remote Sensing*, 10, 100139,
927 <https://doi.org/10.1016/j.srs.2024.100139>, 2024a.
- 928 Qadir, A., Skakun, S., Kussul, N., Shelestov, A., and Becker-Reshef, I.: A generalized model for mapping sunflower areas using
929 Sentinel-1 SAR data, *Remote Sens. Environ.*, 306, 114132, <https://doi.org/10.1016/j.rse.2024.114132>, 2024b.
- 930 Rasul, A., Balzter, H., Ibrahim, G. R. F., Hameed, H. M., Wheeler, J., Adamu, B., Ibrahim, S. a., and Najmaddin, P. M.: Applying
931 built-up and bare-soil indices from Landsat 8 to cities in dry climates, *Land*, 7, 81, <https://doi.org/10.3390/land7030081>,
932 2018.
- 933 Rikimaru, A., Roy, P. S., and Miyatake, S.: Tropical forest cover density mapping, *Trop. Ecol.*, 43, 39-47, 2002.
- 934 Sheykhmousa, M., Mahdianpari, M., Ghanbari, H., Mohammadimanes, F., Ghamisi, P., and Homayouni, S.: Support vector
935 machine versus random forest for remote sensing image classification: A meta-analysis and systematic review, *IEEE J.*
936 *Sel. Top. Appl. Earth Observ. Remote Sens.*, 13, 6308-6325, <https://doi.org/10.1109/JSTARS.2020.3026724>, 2020.
- 937 Skakun, S., Franch, B., Vermote, E., Roger, J. C., Becker-Reshef, I., Justice, C., and Kussul, N.: Early season large-area winter
938 crop mapping using MODIS NDVI data, growing degree days information and a Gaussian mixture model, *Remote Sens.*
939 *Environ.*, 195, 244-258, <https://doi.org/10.1016/j.rse.2017.04.026>, 2017.
- 940 Song, X., Potapov, P. V., Krylov, A., King, L., Bella, C. M. D., Hudson, A., Khan, A., Adusei, B., Stehman, S. V., and Hansen,
941 M. C.: National-scale soybean mapping and area estimation in the United States using medium resolution satellite
942 imagery and field survey, *Remote Sens. Environ.*, 190, 383-395, <https://doi.org/10.1016/j.rse.2017.01.008>, 2017.
- 943 Sun, D., Li, H., Wang, E., He, W., Hao, W., Yan, C., Li, Y., Mei, X., Zhang, Y., and Sun, Z.: An overview of the use of plastic-
944 film mulching in China to increase crop yield and water-use efficiency, *Natl. Sci. Rev.*, 7, 1523-1526,
945 <https://doi.org/10.1093/nsr/nwaa146>, 2020.
- 946 Tamiminia, H., Salehi, B., Mahdianpari, M., Quackenbush, L., Adeli, S., and Brisco, B.: Google Earth Engine for geo-big data
947 applications: A meta-analysis and systematic review, *ISPRS J. Photogramm. Remote Sens.*, 164, 152-170,
948 <https://doi.org/10.1016/j.isprsjprs.2020.04.001>, 2020.
- 949 Tang, X., Miao, C., Xi, Y., Duan, Q., Lei, X., and Li, H.: Analysis of precipitation characteristics on the loess plateau between
950 1965 and 2014, based on high-density gauge observations, *Atmos. Res.*, 213, 264-274,
951 <https://doi.org/10.1016/j.atmosres.2018.06.013>, 2018.
- 952 Tong, X., Zhang, X., Fensholt, R., Jensen, P. R. D., Li, S., Larsen, M. N., Reiner, F., Tian, F., and Brandt, M.: Global area boom
953 for greenhouse cultivation revealed by satellite mapping, *Nat. Food*, 5, 513-523, <https://doi.org/10.1038/s43016-024-00985-0>, 2024.
- 954
- 955 Tu, Y., Wu, S., Chen, B., Weng, Q., Bai, Y., Yang, J., Yu, L., and Xu, B.: A 30 m annual cropland dataset of China from 1986 to
956 2021, *Earth Syst. Sci. Data*, 16, 2297-2316, <https://doi.org/10.5194/essd-16-2297-2024>, 2024.
- 957 Tucker, C. J.: Red and photographic infrared linear combinations for monitoring vegetation, *Remote Sens. Environ.*, 8, 127-150,
958 [https://doi.org/10.1016/0034-4257\(79\)90013-0](https://doi.org/10.1016/0034-4257(79)90013-0), 1979.
- 959 Veetil, B. K., Van, D. D., Quang, N. X., and Hoai, P. N.: Remote sensing of plastic-covered greenhouses and plastic-mulched
960 farmlands: Current trends and future perspectives, *Land Degrad. Dev.*, 34, 591-609, <https://doi.org/10.1002/ldr.4497>,
961 2023.
- 962 Wang, J., Xie, J., Li, L., and Adingo, S.: Review on the fully mulched ridge-furrow system for sustainable maize production on
963 the semi-arid Loess Plateau, *J. Integr. Agric.*, 22, 1277-1290, <https://doi.org/10.1016/j.jia.2022.09.023>, 2023.
- 964 Wang, L. and Qu, J. J.: NMDI: A normalized multi-band drought index for monitoring soil and vegetation moisture with satellite



- 965 remote sensing, *Geophys. Res. Lett.*, 34, <https://doi.org/10.1029/2007GL031021>, 2007.
- 966 Wang, P., Fan, E., and Wang, P.: Comparative analysis of image classification algorithms based on traditional machine learning
967 and deep learning, *Pattern Recognit. Lett.*, 141, 61-67, <https://doi.org/10.1016/j.patrec.2020.07.042>, 2021.
- 968 Wang, S., Azzari, G., and Lobell, D. B.: Crop type mapping without field-level labels: Random forest transfer and unsupervised
969 clustering techniques, *Remote Sens. Environ.*, 222, 303-317, <https://doi.org/10.1016/j.rse.2018.12.026>, 2019.
- 970 Wang, S., Di Tommaso, S., Deines, J. M., and Lobell, D. B.: Mapping twenty years of corn and soybean across the US Midwest
971 using the Landsat archive, *Sci. Data*, 7, 307, 2020.
- 972 Wang, S., Fan, T., Cheng, W., Wang, L., Zhao, G., Li, S., Dang, Y., and Zhang, J.: Occurrence of macroplastic debris in the long-
973 term plastic film-mulched agricultural soil: A case study of Northwest China, *Sci. Total Environ.*, 831, 154881,
974 <https://doi.org/10.1016/j.scitotenv.2022.154881>, 2022a.
- 975 Wang, S., Feng, W., Quan, Y., Li, Q., Dauphin, G., Huang, W., Li, J., and Xing, M.: A heterogeneous double ensemble algorithm
976 for soybean planting area extraction in Google Earth Engine, *Comput. Electron. Agric.*, 197, 106955,
977 <https://doi.org/10.1016/j.compag.2022.106955>, 2022b.
- 978 Wang, Y. P., Li, X. G., Zhu, J., Fan, C. Y., Kong, X. J., Turner, N. C., Siddique, K. H. M., and Li, F.-M.: Multi-site assessment of
979 the effects of plastic-film mulch on dryland maize productivity in semiarid areas in China, *Agric. For. Meteorol.*, 220,
980 160-169, <https://doi.org/10.1016/j.agrformet.2016.01.142>, 2016.
- 981 Weiss, M., Jacob, F., and Duveiller, G.: Remote sensing for agricultural applications: A meta-review, *Remote Sens. Environ.*,
982 236, 111402, <https://doi.org/10.1016/j.rse.2019.111402>, 2020.
- 983 Wen, Y., Li, X., Mu, H., Zhong, L., Chen, H., Zeng, Y., Miao, S., Su, W., Gong, P., Li, B., and Huang, J.: Mapping corn dynamics
984 using limited but representative samples with adaptive strategies, *ISPRS J. Photogramm. Remote Sens.*, 190, 252-266,
985 <https://doi.org/10.1016/j.isprsjprs.2022.06.012>, 2022.
- 986 Wijesingha, J., Dzene, I., and Wachendorf, M.: Evaluating the spatial-temporal transferability of models for agricultural land
987 cover mapping using Landsat archive, *ISPRS J. Photogramm. Remote Sens.*, 213, 72-86,
988 <https://doi.org/10.1016/j.isprsjprs.2024.05.020>, 2024.
- 989 Xiao, X., Boles, S., Liu, J., Zhuang, D., and Liu, M.: Characterization of forest types in Northeastern China, using multi-temporal
990 SPOT-4 VEGETATION sensor data, *Remote Sens. Environ.*, 82, 335-348, [https://doi.org/10.1016/S0034-4257\(02\)00051-2](https://doi.org/10.1016/S0034-4257(02)00051-2), 2002.
- 992 Xiong, Y., Zhang, Q., Chen, X., Bao, A., Zhang, J., and Wang, Y.: Large scale agricultural plastic mulch detecting and monitoring
993 with multi-source remote sensing data: A case study in Xinjiang, China, *Remote Sens.*, 11, 2088,
994 <https://doi.org/10.3390/rs11182088>, 2019.
- 995 Xu, J., Yang, J., Xiong, X., Li, H., Huang, J., Ting, K. C., Ying, Y., and Lin, T.: Towards interpreting multi-temporal deep learning
996 models in crop mapping, *Remote Sens. Environ.*, 264, 112599, <https://doi.org/10.1016/j.rse.2021.112599>, 2021.
- 997 Xuan, F., Dong, Y., Li, J., Li, X., Su, W., Huang, X., Huang, J., Xie, Z., Li, Z., Liu, H., Tao, W., Wen, Y., and Zhang, Y.: Mapping
998 crop type in Northeast China during 2013–2021 using automatic sampling and tile-based image classification, *Int. J. Appl. Earth Obs. Geoinf.*, 117, 103178, <https://doi.org/10.1016/j.jag.2022.103178>, 2023.
- 1000 Yan, C., Liu, E., Shu, F., Liu, Q., Liu, S., and He, W.: Review of agricultural plastic mulching and its residual pollution and
1001 prevention measures in China (in Chinese), *Journal of Agricultural Resources and Environment*, 31, 95-102,
1002 <https://doi.org/10.13254/j.jare.2013.0223>, 2014.
- 1003 Yang, G., Li, X., Liu, P., Yao, X., Zhu, Y., Cao, W., and Cheng, T.: Automated in-season mapping of winter wheat in China with
1004 training data generation and model transfer, *ISPRS J. Photogramm. Remote Sens.*, 202, 422-438,
1005 <https://doi.org/10.1016/j.isprsjprs.2023.07.004>, 2023.
- 1006 Yang, J. and Huang, X.: The 30 m annual land cover dataset and its dynamics in China from 1990 to 2019, *Earth Syst. Sci. Data*,
1007 13, 3907-3925, <https://doi.org/10.5194/essd-13-3907-2021>, 2021.
- 1008 Yang, N., Liu, D., Feng, Q., Xiong, Q., Zhang, L., Ren, T., Zhao, Y., Zhu, D., and Huang, J.: Large-scale crop mapping based on
1009 machine learning and parallel computation with grids, *Remote Sens.*, 11, 1500, <https://doi.org/10.3390/rs11121500>,
1010 2019.



- 1011 Yang, N., Sun, Z., Feng, L., Zheng, M., Chi, D., Meng, W., Hou, Z., Bai, W., and Li, K.: Plastic film mulching for water-efficient
1012 agricultural applications and degradable films materials development research, *Mater. Manuf. Process.*, 30, 143-154,
1013 <https://doi.org/10.1080/10426914.2014.930958>, 2015.
- 1014 You, N. and Dong, J.: Examining earliest identifiable timing of crops using all available Sentinel 1/2 imagery and Google Earth
1015 Engine, *ISPRS J. Photogramm. Remote Sens.*, 161, 109-123, <https://doi.org/10.1016/j.isprsjprs.2020.01.001>, 2020.
- 1016 You, N., Dong, J., Li, J., Huang, J., and Jin, Z.: Rapid early-season maize mapping without crop labels, *Remote Sens. Environ.*,
1017 290, 113496, <https://doi.org/10.1016/j.rse.2023.113496>, 2023.
- 1018 You, N., Dong, J., Huang, J., Du, G., Zhang, G., He, Y., Yang, T., Di, Y., and Xiao, X.: The 10-m crop type maps in Northeast
1019 China during 2017–2019, *Sci. Data*, 8, 41, <https://doi.org/10.1038/s41597-021-00827-9>, 2021.
- 1020 Zanaga, D., Van De Kerchove, R., Daems, D., De Keersmaecker, W., Brockmann, C., Kirches, G., Wevers, J., Cartus, O., Santoro,
1021 M., and Fritz, S.: ESA WorldCover 10 m 2021 v200, <https://doi.org/10.5281/zenodo.5571936>, 2022.
- 1022 Zang, Y., Qiu, Y., Chen, X., Chen, J., Yang, W., Liu, Y., Peng, L., Shen, M., and Cao, X.: Mapping rapeseed in China during
1023 2017-2021 using Sentinel data: an automated approach integrating rule-based sample generation and a one-class
1024 classifier (RSG-OC), *GISci. Remote Sens.*, 60, 2163576, <https://doi.org/10.1080/15481603.2022.2163576>, 2023.
- 1025 Zhang, C., Dong, J., and Ge, Q.: Irrimap_CN: Annual irrigation maps across China in 2000–2019 based on satellite observations,
1026 environmental variables, and machine learning, *Remote Sens. Environ.*, 280, 113184,
1027 <https://doi.org/10.1016/j.rse.2022.113184>, 2022a.
- 1028 Zhang, C., Zhang, H., and Tian, S.: Phenology-assisted supervised paddy rice mapping with the Landsat imagery on Google
1029 Earth Engine: Experiments in Heilongjiang Province of China from 1990 to 2020, *Comput. Electron. Agric.*, 212,
1030 108105, <https://doi.org/10.1016/j.compag.2023.108105>, 2023.
- 1031 Zhang, C., Zhang, H., and Zhang, L.: Spatial domain bridge transfer: An automated paddy rice mapping method with no training
1032 data required and decreased image inputs for the large cloudy area, *Comput. Electron. Agric.*, 181, 105978,
1033 <https://doi.org/10.1016/j.compag.2020.105978>, 2021a.
- 1034 Zhang, C., Dong, J., Xie, Y., Zhang, X., and Ge, Q.: Mapping irrigated croplands in China using a synergetic training sample
1035 generating method, machine learning classifier, and Google Earth Engine, *Int. J. Appl. Earth Obs. Geoinf.*, 112, 102888,
1036 <https://doi.org/10.1016/j.jag.2022.102888>, 2022b.
- 1037 Zhang, D., Pan, Y., Zhang, J., Hu, T., Zhao, J., Li, N., and Chen, Q.: A generalized approach based on convolutional neural
1038 networks for large area cropland mapping at very high resolution, *Remote Sens. Environ.*, 247, 111912,
1039 <https://doi.org/10.1016/j.rse.2020.111912>, 2020a.
- 1040 Zhang, H., Liu, W., and Zhang, L.: Seamless and automated rapeseed mapping for large cloudy regions using time-series optical
1041 satellite imagery, *ISPRS J. Photogramm. Remote Sens.*, 184, 45-62, <https://doi.org/10.1016/j.isprsjprs.2021.12.001>,
1042 2022c.
- 1043 Zhang, H., Wang, Y., Shang, J., Liu, M., and Li, Q.: Investigating the impact of classification features and classifiers on crop
1044 mapping performance in heterogeneous agricultural landscapes, *Int. J. Appl. Earth Obs. Geoinf.*, 102, 102388,
1045 <https://doi.org/10.1016/j.jag.2021.102388>, 2021b.
- 1046 Zhang, H. K. and Roy, D. P.: Using the 500 m MODIS land cover product to derive a consistent continental scale 30 m Landsat
1047 land cover classification, *Remote Sens. Environ.*, 197, 15-34, <https://doi.org/10.1016/j.rse.2017.05.024>, 2017.
- 1048 Zhang, H. Y., Kang, J. Z., Xu, X., and Zhang, L. P.: Accessing the temporal and spectral features in crop type mapping using
1049 multi-temporal Sentinel-2 imagery: A case study of Yi'an Country, Heilongjiang province, China, *Comput. Electron.
1050 Agric.*, 176, 105618, <https://doi.org/10.1016/j.compag.2020.105618>, 2020b.
- 1051 Zhang, J., Ren, S., Dai, J., Ding, F., Xiao, M., Liu, X., Yan, C., Ge, T., Wang, J., Liu, Q., Wang, K., and Zhang, F.: Influence of
1052 plastic film on agricultural production and its pollution control (in Chinese), *Scientia Agricultura Sinica*, 55, 3983-3996,
1053 <https://doi.org/10.3864/j.issn.0578-1752.2022.20.010>, 2022d.
- 1054 Zhang, L., Zhang, K., Zhu, X., Chen, H., and Wang, W.: Integrating remote sensing, irrigation suitability and statistical data for
1055 irrigated cropland mapping over mainland China, *J. Hydrol.*, 613, 128413,
1056 <https://doi.org/10.1016/j.jhydrol.2022.128413>, 2022e.



- 1057 Zhang, L., Wang, W., Ma, Q., Hu, Y., Ma, H., and Zhao, Y.: CCropLand30: High-resolution hybrid cropland maps of China
1058 created through the synergy of state-of-the-art remote sensing products and the latest national land survey, *Comput.*
1059 *Electron. Agric.*, 218, 108672, <https://doi.org/10.1016/j.compag.2024.108672>, 2024.
- 1060 Zhang, L., Qin, R., Wei, H., Zhang, K., Yu, C., Li, F., and Zhang, F.: Optimum plastic mulching application to reduce greenhouse
1061 gas emissions without compromising on crop yield and farmers' income, *Sci. Total Environ.*, 809, 151998,
1062 <https://doi.org/10.1016/j.scitotenv.2021.151998>, 2022f.
- 1063 Zhang, P., Du, P., Guo, S., Zhang, W., Tang, P., Chen, J., and Zheng, H.: A novel index for robust and large-scale mapping of
1064 plastic greenhouse from Sentinel-2 images, *Remote Sens. Environ.*, 276, 113042,
1065 <https://doi.org/10.1016/j.rse.2022.113042>, 2022g.
- 1066 Zhang, X., Liu, L., Chen, X., Gao, Y., Xie, S., and Mi, J.: GLC_FCS30: Global land-cover product with fine classification system
1067 at 30 m using time-series Landsat imagery, *Earth Syst. Sci. Data*, 13, 2753-2776, [https://doi.org/10.5194/essd-13-2753-](https://doi.org/10.5194/essd-13-2753-2021)
1068 [2021](https://doi.org/10.5194/essd-13-2753-2021), 2021c.
- 1069 Zhao, C., Luo, Y., Chen, X., Wu, L., Wang, Z., Feng, H., Yu, Q., and He, J.: PMF-LP: the first 10 m plastic-mulched farmland
1070 distribution map (2019-2021) in the Loess Plateau of China generated using training sample generation and classifier
1071 transfer method, Zenodo [data set], <https://doi.org/10.5281/zenodo.13369426>, 2024.
- 1072 Zhao, Y., Mao, X., Li, S., Huang, X., Che, J., and Ma, C.: A review of plastic film mulching on water, heat, nitrogen balance,
1073 and crop growth in farmland in China, *Agronomy-Basel*, 13, 2515, <https://doi.org/10.3390/agronomy13102515>, 2023.
- 1074 Zheng, W., Wang, R., Cao, Y., Jin, N., Feng, H., and He, J.: Remote sensing recognition of plastic-film-mulched farmlands on
1075 Loess Plateau based on Google Earth Engine (in Chinese), *Transactions of the Chinese Society for Agricultural*
1076 *Machinery*, 53, 224-234, <https://doi.org/10.6041/j.issn.1000-1298.2022.01.025>, 2022.
- 1077 Zhong, L., Hu, L., and Zhou, H.: Deep learning based multi-temporal crop classification, *Remote Sens. Environ.*, 221, 430-443,
1078 <https://doi.org/10.1016/j.rse.2018.11.032>, 2019.
- 1079 Zhou, C., Huang, J., Xiao, Y., Du, M., and Li, S.: A novel approach: Coupling prior knowledge and deep learning methods for
1080 large-scale plastic greenhouse extraction using Sentinel-1/2 data, *Int. J. Appl. Earth Obs. Geoinf.*, 132, 104073,
1081 <https://doi.org/10.1016/j.jag.2024.104073>, 2024.
- 1082 Zhou, Q., Zhu, Z., Xian, G., and Li, C.: A novel regression method for harmonic analysis of time series, *ISPRS J. Photogramm.*
1083 *Remote Sens.*, 185, 48-61, <https://doi.org/10.1016/j.isprsjprs.2022.01.006>, 2022.
- 1084 Zhou, W., Wei, H., Chen, Y., Zhang, X., Hu, J., Cai, Z., Yang, J., Hu, Q., Xiong, H., Yin, G., and Xu, B.: Monitoring intra-annual
1085 and interannual variability in spatial distribution of plastic-mulched citrus in cloudy and rainy areas using multisource
1086 remote sensing data, *Eur. J. Agron.*, 151, 126981, <https://doi.org/10.1016/j.eja.2023.126981>, 2023.
- 1087 Zou, Q., Ni, L., Zhang, T., and Wang, Q.: Deep learning based feature selection for remote sensing scene classification, *IEEE*
1088 *Geosci. Remote Sens. Lett.*, 12, 2321-2325, <https://doi.org/10.1109/LGRS.2015.2475299>, 2015.
- 1089