



1 Reconciliation of observation- and inventory- based methane emissions for eight large
2 global emitters

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39

40 **Abstract**

41

42 Monitoring the spatial distribution and trends in surface greenhouse gas (GHG) fluxes, as well as flux
43 attribution to natural and anthropogenic processes, is essential to track progress under the Paris Agreement and
44 to inform its Global Stocktake. This study updates earlier syntheses (Petrescu et al., 2020, 2021, 2023) and
45 provides a consolidated synthesis of CH₄ emissions using bottom-up (BU) and top-down (TD) approaches for
46 the European Union (EU) and seven additional countries with large anthropogenic and/or natural emissions
47 (USA, Brazil, China, India, Indonesia, Russia, and the Democratic Republic of Congo (DR Congo)). The work
48 utilizes updated National GHG Inventories (NGHGs) reported by Annex I Parties under the United Nations
49 Framework Convention on Climate Change (UNFCCC) in 2023 and the latest available Biennial Update Reports
50 (BURs) reported by non-Annex I Parties. The NGHGs are considered in an integrated analysis that also relies
51 on independent flux estimates from global inventory datasets, process-based models, inverse modeling and, when
52 available, respective uncertainties. Whenever possible, it extends the period to 2021. Comparing NGHGs with
53 other approaches reveals that differences in the emission sources that are included in the estimate is a key source
54 of divergence between approaches. A key system boundary difference is whether both anthropogenic and natural
55 fluxes are included and, if they are, how fluxes belonging to these two sources are grouped/partitioned.
56 Additionally, the natural fluxes are sensitive to the prior geospatial distribution of emissions in atmospheric
57 inversions. Over the studied period, the total CH₄ emissions in the EU, USA, and Russia show a steady decreasing
58 trend since 1990, while for the non-EU emitters analyzed in this study, Brazil, China, India, Indonesia, and DR
59 Congo, CH₄ emissions have generally increased.

60 In the **EU**, the anthropogenic BU approaches are reporting relatively similar mean emissions over 2015
61 to 2020 of 18.5 ± 2.7 Tg CH₄ yr⁻¹ for EDGAR v7.0, 16 Tg CH₄ yr⁻¹ for GAINS and 19 Tg CH₄ yr⁻¹ for FAOSTAT,
62 with the NGHGI estimates of 15 ± 1.8 Tg CH₄ yr⁻¹. Inversions give higher emission estimates as they include
63 natural emissions. Over the same period, the three high-resolution regional inversions report a mean emission of
64 21 (19-25) Tg CH₄ yr⁻¹, while the mean of six coarser-resolution global inversions results in emission estimates
65 of 24 (23-25) Tg CH₄ yr⁻¹. The magnitude of BU natural emissions (peatland and mineral soils, lakes and
66 reservoirs, geological and biomass burning) accounts for 6.6 Tg CH₄ yr⁻¹ (Petrescu et al., 2023a) and explains
67 the differences between the TD inversions and the BU estimates of anthropogenic emissions (including
68 NGHGs). For the other Annex I Parties in this study (**USA and Russia**), over 2015 to 2020, the mean of the
69 four anthropogenic BU approaches reports 18.5 (13-27.9) Tg CH₄ yr⁻¹ for Russia and 29.1 (23.5- Tg CH₄ yr⁻¹ for
70 the USA, against total TD mean estimates of 37 (30-43) Tg CH₄ yr⁻¹ and 43.4 (42-48) Tg CH₄ yr⁻¹, respectively.
71 The averaged BU and TD natural emissions account for 16.2 Tg CH₄ yr⁻¹ for Russia and 14.6 Tg CH₄ yr⁻¹ for the
72 USA, partly explaining the gap between the BU anthropogenic and total TD emissions.

73 For the **non-Annex I Parties**, anthropogenic CH₄ estimates from UNFCCC BURs show large
74 differences with the other global inventory-based estimates and even more with atmospheric-based ones. This



75 poses an important potential challenge to monitoring the progress of the global CH₄ pledge and the Global
76 Stocktake, not only from the availability of data but also its accuracy.

77 By systematically comparing the BU with TD methods, this study provides recommendations for more
78 robust comparisons of available data sources and hopes to steadily engage more Parties in using observational
79 methods to complement their UNFCCC inventories, as well as considering their natural emissions. With
80 anticipated improvements in atmospheric modeling and observations, as well as modeling of natural fluxes,
81 future development needs to resolve knowledge gaps in both BU and TD approaches and to better quantify
82 remaining uncertainty. Consequently, TD methods may emerge as a powerful tool for verifying emission
83 inventories for CH₄, and other GHGs and informing international climate policy. The referenced datasets related
84 to figures are available at <https://doi.org/10.5281/zenodo.10276087> (Petrescu et al., 2023b).

85 1. Introduction

86
87 In 2021, the NOAA Global Monitoring Laboratory (GML) reported the largest annual increase in
88 atmospheric CH₄ mixing ratios since records began in 1983, with a 17 parts per billion (ppb) value (NOAA
89 (https://gml.noaa.gov/ccgg/trends_ch4/). In 2022, atmospheric CH₄ concentrations averaged 1912 ppb, 162 %
90 higher than pre-industrial levels. A similar, abnormally large growth rate of 14.8 ppb yr⁻¹ was detected from
91 total column mixing ratio measurements (XCH₄) by the Greenhouse Gases Observing Satellite (GOSAT) (Peng
92 et al., 2022). The drivers of the recent growth are most likely driven primarily by biogenic emissions (Basu et
93 al., 2022; Lan, et al., 2021a; Lanet al., 2021b; Lan et al., 2022; Nisbet et al., 2016, 2019), with smaller
94 contributions from increased fossil fuel emissions and a reduced atmospheric sink (Nisbet et al., 2023). These
95 processes drove the near record increase in atmospheric CH₄ growth in 2020, and furthermore outweighed the
96 slight observed decrease in anthropogenic CH₄ emissions accumulated from March–June 2020 as impact of the
97 COVID-19 slowdown (e.g. China) which might be small relative to the long-term positive trend in emissions.
98 (McNorton et al., 2022, Peng et al., 2022, Qu et al. 2022).

99 CH₄ in the atmosphere has many different sources, of both natural and anthropogenic origin. The natural
100 sources of CH₄ are dominated by wetlands, while anthropogenic emissions principally come from agricultural
101 activities (livestock and rice farming), waste management (landfills and water treatment plants) and the
102 production, transportation, and use of fossil fuels. Most of the agricultural sources are distributed sources, while
103 the energy-related industrial sources of CH₄ are a mix of large point sources, of which some are detectable by
104 satellite. Smaller point and distributed sources of fugitive emissions (e.g., leaks in pipelines and compression
105 stations) are more challenging to identify (Rutherford et al., 2021; Omara et al., 2022). While anthropogenic CH₄
106 emissions from fossil fuels, agriculture, and waste can be reduced by mitigation actions, increased natural
107 emissions lead to different challenges. It has been suggested that fluctuations in natural sources - dominated by
108 wetlands and open water bodies - were the main reasons for some of the atmospheric CH₄ anomalies observed
109 during the last decades (Rocher-Ros et al., 2023; Zhang et al., 2023; Nisbet et al., 2023; Lunt et al., 2019). Nisbet
110 et al., 2023 review recent studies, including those which quantified the observed methane growth in the last years.
111 Using a global inverse analysis of GOSAT satellite observations, it has been shown that increases between 2019-
112 2020 were in the range of 22-32 Tg CH₄ yr⁻¹ and were attributed to biogenic sources, half of which took place in
113 East Africa, as well as Canada (Qu et al., 2022 and Basu et al., 2022).



114 The contribution of CH₄ to global warming has been estimated to be about 0.5°C relative to the period
115 1850–1900 (IPCC, 2021) (Stavert et al., 2022). Methane has a relatively short perturbation lifetime (averaging
116 12.4 years, Balcombe et al., 2018) and a high global warming potential (86 and 34 for 20- and 100-years times
117 horizons respectively, compared to that of CO₂ emissions, IPCC, 2021, Table 7.15). Given the short lifetime, a
118 decline in CH₄ emissions will rapidly reduce the global warming contribution from CH₄ and help mitigate the
119 impact of climate change at decadal time scale (Cain et al., 2021). However, efforts to reduce CH₄ emissions
120 require a thorough understanding of the dominant CH₄ sources and sinks and their temporal and regional
121 distribution and trends (Stavert et al., 2022).

122 The Paris Agreement, a milestone of the UNFCCC to combat climate change and adapt to its effects,
123 entered into force on November 4, 2016. It asks each signatory to define and communicate its planned climate
124 actions, known as Nationally Determined Contributions (NDCs), and to report their progress towards their
125 targets. Next to commitments adopted by countries at COP26, the Global Methane Pledge (GMP) was launched.
126 The goal of the GMP is to cut anthropogenic CH₄ emissions by at least 30 % by 2030 with respect to 2020 levels,
127 and is seen as the fastest way to reduce near-term warming and is necessary to keep a 1.5°C temperature limit
128 within reach. Achieving this goal will drive significant energy security, food security, health, and development
129 gains. About 150 countries joined this pledge and about fifty already developed national CH₄ action plans or are
130 in the process of doing so. As agriculture and waste are the main anthropogenic sources for CH₄ emissions, a
131 GMP Food and Agriculture and Waste pathway was launched at COP27, foreseeing actions that increase
132 agricultural productivity, reduce emissions from dairy, food loss and waste by supporting small farmers and
133 increase innovation (<https://www.state.gov/global-methane-pledge-from-moment-to-momentum/>).

134 This paper updates previous studies (e.g., Petrescu et al., 2020, 2021, 2023) and aims to inform and attract
135 attention to the use of the results for diverse climate stakeholder needs beyond research. It deepens the analysis
136 on sources in the EU and in seven countries that have large anthropogenic and/or natural CH₄ emissions (USA,
137 Brazil, China, India, Indonesia, Russia and the Democratic Rep. of Congo). It examines both Annex I (EU, USA
138 and Russia) and non-Annex I estimates from observation-based BU process-based models and inversions-based
139 TD approaches (using satellite observations) by identifying and explaining differences with official inventory
140 reports submitted by parties to the UNFCCC. The seven countries were chosen based on location and the
141 importance / magnitude of their anthropogenic and natural emissions. By using multiple methodologies,
142 uncertainties can be estimated by looking at the range in both emissions and trends. Starting in 2024, the non-
143 Annex I Parties must - given they have sufficient capacities - report formal inventories under the Paris
144 Agreement's Enhanced Transparency Framework following the same guidelines and rules as the Annex I
145 countries. Furthermore, they will undergo more stringent reviews than those that previously looked at the BURs
146 and NDCs. This will also allow strengthening the robustness of such comparison exercises when using
147 independent atmospheric observations in estimating trends and patterns for regional and national CH₄ emissions
148 (IPCC, 2006).

149 **2. Methods and data**

150 **2.1. Verification practices in official UNFCCC NGHGs**

151



152 Quality assurance/quality control (QA/QC) is a key component of NGHGs development. Verification
153 is an additional step and refers specifically to methods that are external to the inventory and apply independent
154 data. There are two main methods of verification: 1) independent inventory-based estimates, 2) observation-
155 based emission estimates.

156 A challenge with comparisons against *independent inventory-based estimates* is that none are truly
157 independent as they may rely on, for example, the same activity data reported by a country (Andrew 2020).
158 Experience has shown that performing detailed comparisons (Petrescu et al., 2021, 2023) can help clarify
159 differences in system boundaries or even identify errors (Andrew 2020). Improving independent emission
160 inventories also has value, as these are often used in global studies where common methods across all countries
161 are desired.

162 *Observation-based estimates* use observations of atmospheric concentrations and prior fluxes that are
163 then coupled to a transport model. These methods are more complex, computationally expensive, but make use
164 of observational information that is independent from emission inventories.

165 Since most developed countries have reported UNFCCC inventories for decades and these have been
166 continually reviewed and refined, the focus of this work is on observation-based estimates. As an increasing
167 number of developing countries begin more detailed and frequent reporting, comparisons with independent TD
168 approaches will be an important method of verification for those countries.

169 The 2019 refinement of the 2006 IPCC guidelines highlight notable advances in the application of
170 inverse models of atmospheric transport for estimating emissions at the national scale. Building on this progress,
171 they extend the guidance on the use of atmospheric measurements for verification (IPCC, 2019). There are
172 several countries that currently use atmospheric measurements for verification of parts of their inventories.
173 Australia (Luhar et al 2020, Australian NIR, 2023) and New Zealand (Geddes et al., 2021) have estimated
174 regional CH₄ emissions to help better understand the methods and their potential. Germany performs various
175 cross validation checks with available data (German NIR, 2023), some of which are based on observations. The
176 UK and Switzerland (Annex 6 CHE NIR, 2023) have developed more comprehensive methods based on
177 inversion modeling, covering several GHGs in addition to CH₄. Building on modeling experience, the country
178 reporting confirms that most potential lies in using observations to verify fluorinated gases (Annex 6 UK NIR,
179 2023), but the large uncertainty in CH₄ emissions gives the potential for verification if a sufficient observation
180 network is used in inversion modeling (Bergamaschi et al., 2018, Thompson et al., 2014).

181 While inversions of CH₄ fluxes are associated with significant uncertainty, so are NGHGI estimates of
182 anthropogenic CH₄ emissions. Furthermore, inversions can provide information on subannual and subnational
183 variations in time and space that may indicate differences in source sector emission estimates. In geographic
184 areas with sufficiently dense ground-based observation networks, the inversions will have more value.

185

186 **2.2. Anthropogenic CH₄ emissions from the NGHGs**

187

188 Annex I countries report their annual GHG emissions to the UNFCCC in the so-called Common
189 Reporting Format (CRFs) data tables and National Inventory Reports (NIRs). Here, anthropogenic CH₄
190 emissions from the five UNFCCC sectors (incl. LULUCF) are grouped together. As part of the LULUCF sector,
191 we also have the CH₄ emissions from wetlands, which according to the are defined as managed "where the water



192 table is artificially changed (i.e. lowered or raised) or those created through human activity (e.g. damming a
193 river) and that do not fall into Forest Land, Cropland, or Grassland categories (IPCC, 2014). Reporting CH₄
194 emissions from managed wetlands are not mandatory, but if done, parties are encouraged to make use of the
195 2013 IPCC Wetlands supplement (IPCC, 2014). In the EU, if Member States report these emissions, they report
196 not only restored (rewetted) wetlands but also emissions from drained organic and mineral soils (e.g. peatlands,
197 ditches etc.). These are not large by magnitude but are large by area in the Nordic countries. According to NGHGI
198 data, in 2021, managed wetlands in the EU for which emissions were reported under the LULUCF (CRF Table
199 4(II) and Summary 1.As2 accessible for each EU¹) summed up to 0.21 Tg CH₄ yr⁻¹. Furthermore, the NGHGIs do
200 not include any lateral fluxes from inland waters but do include biomass burning anthropogenic emissions
201 reported under the LULUCF sector.

202 The presented uncertainties in the CH₄ emission levels of the individual countries and the EU are not
203 always reported in a complete and harmonized format, and therefore were calculated applying gap-filling and
204 harmonization procedures that are used to compile the EU GHG inventory reported under UNFCCC (EU NIR,
205 2023) (see SI and Appendix A1.1 in Petrescu et al., 2023a). The EU uncertainty analysis reported in the bloc's
206 National Inventory Report (NIR) is based on country-level, Approach 1 uncertainty estimates (IPCC, 2006, Vol.
207 1, Chap. 3) that are reported by EU Member States, previously under Article 7(1)(p) of Regulation (EU) 525/2013
208 and since 2023 under Article 26(3) and Annex V(Part 1)(m) of the Governance Regulation (EU) 2018/1999.

209 Non-Annex I countries report their updated NGHGIs to the UNFCCC, including a national inventory
210 report and information on mitigation actions, needs and support received in the so-called Biennial Update Reports
211 (BURs). In this study, Brazil, China, Indonesia, India and the Democratic Rep. of Congo (DR Congo) were
212 investigated. For Brazil, information from its fourth biennial update report (4th BUR) (Brazil, 2020) that give
213 both total and sectoral split emission values for years 1994, 2000, 2010, 2012, 2015 and 2016, were used. For
214 China, information from its second biennial update report (2nd BUR) Tables 2-10, 2-13, 2-14, 2-15, and 2-16
215 (China, 2019) were used. The information was available for both total and sectoral split emission values for 1994,
216 2005, 2010 and 2014. Uncertainties for 2014 are available in Table 2-12. Indonesia submitted its third biennial
217 update report (3rd BUR) in 2021 (Indonesia, 2021). Indonesian total CH₄ emissions time series per sector as
218 reported by the 2nd UNFCCC BUR (2001-2016) and revised 3rd BUR (2000 and 2019, Table 2). For 2017 and
219 2018, agricultural CH₄ emissions were detailed by the 3rd BUR. Data uncertainty for 2019 activity and EFs are
220 the same as reported in the 2nd BUR (2018). The result of the uncertainty analysis showed that the overall
221 uncertainty of Indonesia's National GHG inventory with AFOLU (including peat fires) for 2000 and 2019 were
222 approximately 20.0% and 19.9%, respectively. A smaller uncertainty, 10.4 % for 2000 and 13.8 % for 2019,
223 occurred when the FOLU (including forest fires), was excluded from the analysis. This shows that Indonesian
224 emission inventories are highly uncertain when forest fires are included in the analysis. The DR Congo submitted
225 its first BUR in 2022. However, we only used total values reported for 2000-2018 (Table 12 Congo, 2022). India
226 has submitted three BURs and information on sectoral CH₄ emissions are in each of them only for one year. We

¹<https://unfccc.int/process-and-meetings/transparency-and-reporting/reporting-and-review-under-the-convention/greenhouse-gas-inventories-annex-i-parties/national-inventory-submissions-2019>



227 compiled information for 2010 from the first BUR (India, 2016), for 2014 from the second BUR (India, 2018)
 228 and for 2016 from the third and latest BUR (India, 2021).

229 **2.3. Other CH₄ data sources and estimation approaches**
 230

231 The CH₄ emissions in the EU and non-Annex I countries used in the atmospheric inversions and
 232 anthropogenic and natural emissions estimates from various BU approaches and inventories (i.e., UNFCCC
 233 CRFs and BURs) covering specific products, sectors and activities are summarized in Table 1. The data and the
 234 detailed description of most products (Tables S1 and S2, Supplementary Information) span the period from 1990
 235 to 2021, with some of the data only available for shorter time periods. The estimates are available both from
 236 peer-reviewed literature and from unpublished research results from the VERIFY and CoCO₂ projects
 237 (Supplementary Information, SI) and in this work they are compared with NGHGIs reported in 2023 (time series
 238 for all (Annex I) or some years (non-Annex I) of the 1990-2021 period). The BU anthropogenic sources are from
 239 UNFCCC NGHGIs and three global inventory datasets/models: EDGARv7.0, FAOSTAT and GAINS. In this
 240 synthesis, data from FAOSTAT (Tubiello et al., 2022; FAO, 2023) is included for all economic sectors: Energy,
 241 IPPU, Waste and Other, and are sourced from the PRIMAP-hist v2.4 dataset (Gütschow et al., 2022) to build
 242 emissions indicators on agrifood systems and on the entire economy. Emission totals from the agrifood domain
 243 are computed following the Tier 1 methods of the Intergovernmental Panel on Climate Change (IPCC)
 244 Guidelines for NGHGIs. Agrifood systems emissions in FAOSTAT are largely based on FAO crop, livestock
 245 and land use statistics (Tubiello et al., 2022; FAO, 2023). They are complemented with activity data from the
 246 UN Statistical Division (UNSD), the International Energy Agency (IEA) and with geospatial information on
 247 drained organic soils and biomass fires (Conchedda and Tubiello, 2020; Prosperi et al., 2020).

248 The analysis focuses on both total and sectoral or partitioned information from both BU and TD
 249 estimates. As detailed in Table 1, not all inversions distinguish between sources, however in the following
 250 sections we discuss comparability between BU and TD for both total and partitioned results.

251 *Table 1: Sectors included in this study and data sources providing estimates for these sectors. CAMS stands for*
 252 *Copernicus Atmosphere Monitoring Service. References to data products are found in Table 2 Petrescu et al.,*
 253 *2023a and Table S1 and S2, SI.*

Anthropogenic (BU) ² CH ₄	Natural (BU) ³ CH ₄	Regional TD CH ₄	Global TD CH ₄
1. Energy: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT-PRIMAP	Wetlands EU: JSBACH-HIMMELI Global: LPJ-GUESS	No partitions – total emissions FLEXkF_v2023	Totals and partitioned emissions:
2. Industrial Products and Products in Use (IPPU): UNFCCC NGHGI (CRFs)			

² For consistency with the NGHGI, here we refer to the five reporting sectors as defined by the UNFCCC and the Paris Agreement decision (18/CMP.1), the IPCC Guidelines (IPCC, 2006), and their Refinement (IPCC, 2019a), with the only exception that the latest IPCC Refinement groups together Agriculture and LULUCF sectors in one sector (Agriculture, Forestry and Other land Use - AFOLU).

³ The term **natural** refers here to unmanaged natural CH₄ emissions (peatlands, mineral soils, geological, inland waters and biomass burning) not reported under the anthropogenic UNFCCC LULUCF sector.



and BURs), EDGAR v7.0, FAOSTAT-PRIMAP	Peatlands, mineral soils: EU: JSBACH-HIMMELI	CIF-FLEXPARTv10.4	MIROC4-ACTM (control and OH varying runs)
3. Agriculture: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT	Global: LPJ-GUESS	CIF-CHIMERE	CAMsv21r1 (NOAA and NOAA_GOSAT runs)
4. LULUCF: UNFCCC NGHGI (CRFs and BURs) and FAOSTAT	Inland waters fluxes EU: lakes, rivers and reservoirs (RECCAP2) Global: lakes and reservoirs ORNL DAAC		TM5-4DVAR (TROPOMI)
5. Waste: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT-PRIMAP	Geological fluxes updated activity (see SI) Biomass burning (GFEDv4.1s)		CTE-CH ₄ (GCP2021) CEOS (GOSAT) GEOS-Chem CTM (TROPOMI) for USA only

254 note: Not all models have a version id. Those that have, are used in previous syntheses (Petrescu et al., 2021 and 2023a).

255 We define as natural sources, all sources which do not belong to the anthropogenic partition (Table 2).
 256 The BU natural components for the EU were computed as the sum of the VERIFY products (biomass burning,
 257 inland waters and undisturbed peatlands plus mineral soils as described in Petrescu et al., 2021 and 2023) and
 258 geological emissions (Etiopie et al. 2019) updated for the VERIFY project. For the seven non-EU emitters, the
 259 BU natural fluxes are the sum of wetland emissions (LPJ-GUESS), lake and reservoir emissions (ORNL DAAC),
 260 biomass burning emissions (GFED4.1s) and geological emissions (updated activity, SI). The TD natural global
 261 estimates were calculated as the sum of all natural partitions reported by the inversions. Adjustments were made
 262 to have a consistent comparison between partitions, adding the missing ones from the BU estimates (Table 4).
 263 The error bar on the TD natural represents the range of the min/max between inversion estimates.

264 The total regional TD estimates (for EU) and their uncertainties were calculated as the mean and
 265 min/max range between FLEXkF_v2023, CIF-FLEXPART and CIF-CHIMERE inversions (see Priors table in
 266 Petrescu et al., 2023b). For the USA, we considered as a regional estimate the optimized emissions from the
 267 GEOS-Chem CTM (based on TROPOMI data for 2019) from Nesser et al., 2023, with the range from the eight
 268 members of the inversion ensemble shown as uncertainty (Table 2 in Nesser et al., 2023).

269 For all countries, the total global TD inversion estimates (time series) and uncertainties were calculated
 270 over the period 2015-2021 using the mean and min/max between CTE-GCP2021, MIROC4-ACTM both runs,
 271 CAMS v21r1 (both runs), and TM5-4DVAR (TROPOMI based). CEOS (GOSAT) provided an estimate only
 272 for 2019.

273 The units used in this paper are metric tons (t) [1kt = 10⁹ g; 1Mt (Tg) = 10¹² g] of CH₄. The referenced
 274 data for replicability purposes are available for download at <https://doi.org/10.5281/zenodo.10276087> (Petrescu
 275 et al., 2023b). Upon request, the computer code for plotting figures in the same style and layout can be provided.
 276 Throughout the paper and mostly for the complex figures, the following ISO3 country codes are used: BRA



277 (Brazil), CHN (China), IDN (Indonesia), IND (India), RUS (Russia) and COD (DR Congo). As before in the
278 text, the European Union consists of 27 MS, excludes the UK and is further abbreviated as EU.

279 3. Results

280

281 3.1. NGHGI official reported estimates (UNFCCC)

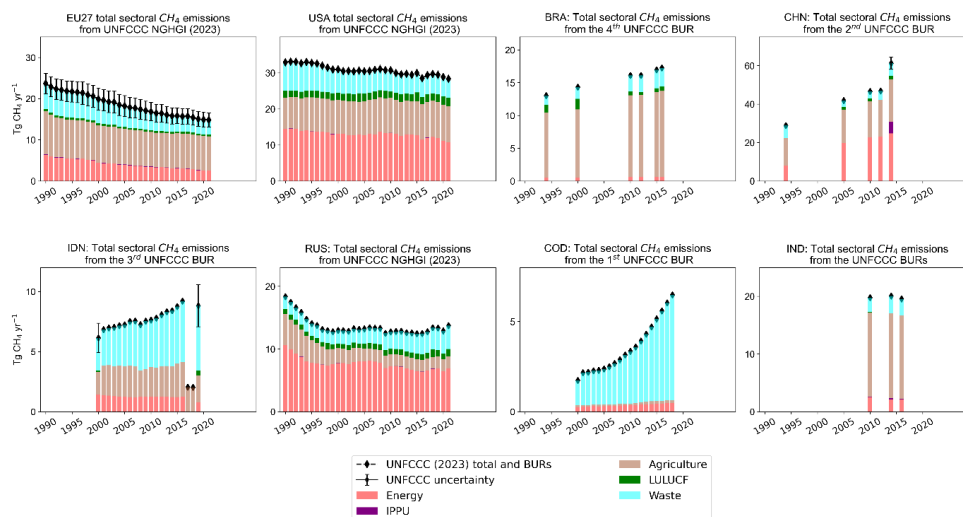
282

283 Figure 1 presents anthropogenic CH₄ emissions reported to the UNFCCC in 2023 from the NGHGI
284 CRFs (EU, USA and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR
285 Congo (1st in 2022) and India (all three BURs). The following section provides additional details for all the
286 countries.

287 For the EU, the total CH₄ emissions in 2021 account for 14.8 ± 1.8 Tg CH₄ yr⁻¹ and represent 12.8 % of
288 the total EU greenhouse gas emissions (in CO₂e, GWP 100 years, IPCC AR5⁴). CH₄ emissions are predominantly
289 from agriculture (Figure 1, brown), a sector which in 2021 accounted for 8.3 Tg CH₄ yr⁻¹ \pm 0.8 Tg CH₄ yr⁻¹ or 56
290 % of the total EU CH₄ emissions (incl. LULUCF). Anthropogenic NGHGI CH₄ emissions from the LULUCF
291 sector are very small for the EU e.g., 0.5 Tg CH₄ yr⁻¹ or 3 % in 2021, including emissions from biomass burning.
292 The EU data from Figure 1 shows steadily decreasing trends for all sectors with respect to the 1990 CH₄ levels.
293 The reduction in total CH₄ emissions in 2021 with respect to 1990 is 8.9 Tg CH₄ yr⁻¹ (37 %) at an average yearly
294 rate of -1%.

295 Between 1990-2021, the reported USA CH₄ emissions show a small decrease of 4.6 Tg CH₄ yr⁻¹, more
296 pronounced for the last two years (2020-2021), with an average reduction of -0.5 % per year (Fig. 1 black dotted
297 line). In the USA, the largest share of emissions comes from the Energy sector, for which, next to IPPU and
298 Waste, the highest reduction shares were registered (42%, 34 % and 26 %, respectively) while emissions from
299 Agriculture and LULUCF increases of 16 % and 23 %, respectively, were registered. After a notable decrease of
300 1.5 Tg registered in 2016 compared to 2015, emissions picked up again and had a second decreasing trend in
301 2020 and 2021, possibly due to the COVID pandemic. Overall, reported data indicates that reductions in the
302 USA CH₄ emissions have been slower than that in the EU.

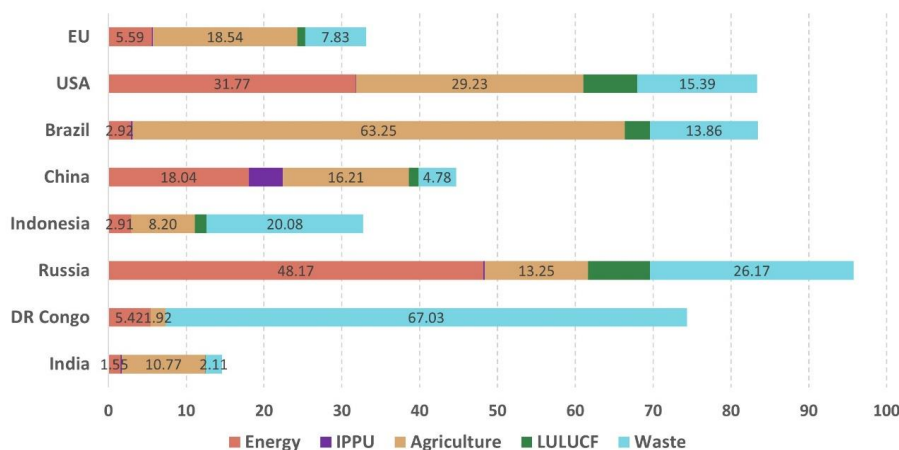
⁴ IPCC AR4 GWP 100 values are still used by the Member States in their NGHGI reporting to the UNFCCC.



303

304 *Figure 1: Total and sectoral CH₄ emissions (incl. LULUCF) from the UNFCCC NGHGI (2023) CRFs (EU, USA*
 305 *and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR Congo (1st in 2022)*
 306 *and India (all three BURs: 2016, 2018 and 2021). The relative error on the UNFCCC value represents the*
 307 *NGHGI (2023) reported uncertainties computed with the error propagation method (95% confidence interval)*
 308 *and gap-filled to provide respective estimates for each year. Information on Indonesian sectoral CH₄ emissions*
 309 *in 2017 and 2018 are only available for Agriculture. In 2014, China reported uncertainty as well (min 5.2 %*
 310 *and max 5.3 %). The overall uncertainty of Indonesia's National GHG inventory with AFOLU (including peat*
 311 *fires) for 2000 and 2019 were approximately 20.0% and 19.9%, respectively.*

312 The trend in total CH₄ emissions in *Brazil* is increasing strongly, with 32.5 % more emissions in 2016
 313 compared to 1994, registering a maximum annual growth rate of +22 % in 2010 compared to 2000, and a
 314 minimum annual increase rate of +1% in 2016 compared to 2015. The agricultural sector (76 % of the total) was
 315 the main driver of the growth, followed by the waste sector (16 % of the total). There are only small CH₄
 316 emissions from the Energy sector (some oil and gas activities). The Brazilian agricultural CH₄ emissions are the
 317 highest compared to all other countries on a per capita basis (see Figure 2).



318

319 *Figure 2: Per capita emissions (kg) intensity per sector based (IPCC, 2006) on reported emissions and*
 320 *population data from the last reported NIRs as follows: 2021 for the EU, USA and Russia, 2016 for Brazil and*
 321 *India, 2014 for China, 2019 for Indonesia and 2018 for DR Congo.*

322 *China's total CH₄ emissions are much larger than the emissions reported by many developed countries*
 323 *or the entire EU (see Figure 1), but on a per capita basis it is only in fifth place (Figure 2), with the highest*
 324 *contribution from the Energy sector (third place after Russia and the USA). The rapid growth of China's coal*
 325 *demand has important implications for CH₄ emissions from coal mining or coal mine methane (CMM) emissions*
 326 *(Gao et al., 2020). The Energy sector is the largest component of Chinese emissions (40 % in 2015), followed*
 327 *by Agriculture (36 %). The second Chinese UNFCCC BUR CH₄ data shows increasing trends in total CH₄*
 328 *emissions with an increase reported in 2014 of 113 % compared to 1994, corresponding to 32 Tg CH₄. The*
 329 *Energy and Agriculture sectors have increased by 214 % and 54 % in 2014 compared to 1994.*

330 *Indonesia's 3rd BUR data (2000 and 2019) show increasing trends in total CH₄ emissions. The time*
 331 *series 2001-2006 belongs to the 2nd BUR submitted in 2018. In 2019, Indonesian CH₄ emissions have increased*
 332 *by +44 % compared to 2000, corresponding to 2.6 Tg CH₄ yr⁻¹, an average yearly increase of 3 %, and the sector*
 333 *which contributes the most to this increase is the Waste sector, which nearly doubled its emissions in 2019*
 334 *compared to 2000, which is also seen in the per capita contribution (Figure 2). According to Qonitan et al., 2021,*
 335 *the major solid waste source in Indonesia is the household sector, which contributed 44-75% to total waste*
 336 *generated. The composition of municipal waste consists of 43.78% of food waste, 16.05% of paper, and 14.08%*
 337 *of plastics. CH₄ emissions from the other sectors remained nearly constant.*

338 *Russian CH₄ emissions have decreased by -25 % from 1990 to 2021, but most of this decrease happened*
 339 *during the collapse of the former Soviet Union. Since 2000, CH₄ emissions have remained rather low. The decline*
 340 *seen between 1990-2000 is primarily due to the Agricultural sector (-52 %) and Energy (-27 %). At the same*
 341 *time, the Waste sector started to increase its emissions (6 %). Between 2001-2021, the CH₄ emissions from the*
 342 *Agriculture and Energy sectors continue to decrease (by 17 % and 11 %, respectively), while the emissions from*



343 the Waste sector register an additional 76 % increase. IPPU emissions increased by 85 % but remain negligible
 344 compared to other sectors. Since the 2000s, also LULUCF emissions have increased by 53 %.

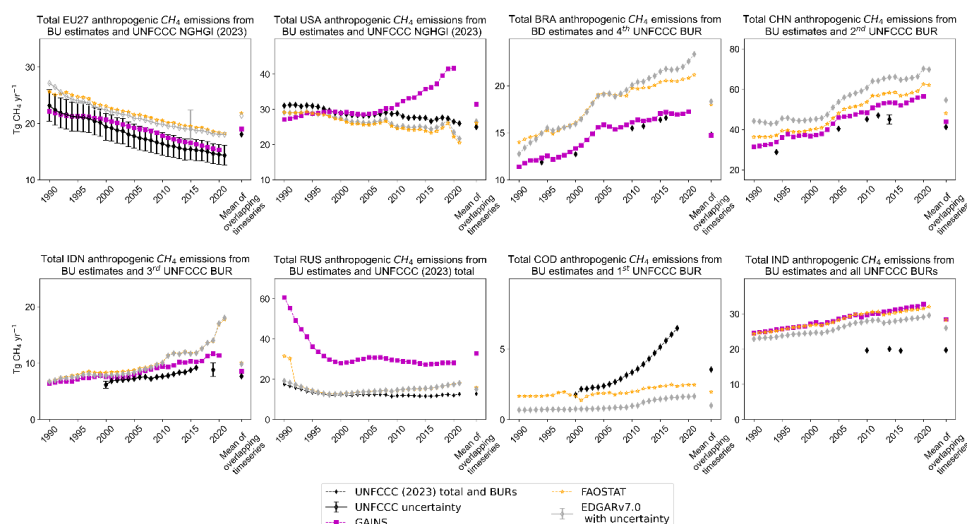
345 For its first BUR, *DR Congo* submitted emissions from Energy, AFOLU (Agriculture plus LULUCF)
 346 and Waste for 2000-2018. Since 2000, the DR Congo total CH₄ emissions have increased by a factor of four.
 347 Most of the CH₄ emissions are reported for the Waste sector, and account for 90 % of the total emissions. The
 348 high percentage of waste emissions in DR Congo is seen as well in the per capita Figure 2. Assè-Wassa Sama
 349 and Berenger, 2023 report confirm that between 2000 and 2021, CH₄ emissions, which in 2021 represent in DR
 350 Congo ~97% of total waste generated emissions, grew at a rate of 4 % yr⁻¹, compared with 2.7 % yr⁻¹ for total
 351 emissions. This increase was driven by the increase in emissions caused by solid waste disposal (+6.2 %). The
 352 CH₄ waste emissions come mainly from the treatment and discharge of wastewater (69 % in 2021, compared
 353 with 80 % in 2000), followed by the elimination of solid waste (31 % in 2021, compared with 20 % in 2000).
 354 The weight of emissions caused by the elimination of solid waste in the sector's total emissions has nevertheless
 355 increased by 11 percentage points between 2000 and 2021 (Assè-Wassa Sama and Berenger, 2023).

356 Each of *India's* BURs provide detailed information on sectoral CH₄ emissions only for one year. Most
 357 of the emissions in India belong to the Agriculture sector, amounting to almost 15 Tg CH₄ yr⁻¹ (in 2016),
 358 representing 74 % of the total anthropogenic emissions. From only three years of reported data, there is no clear
 359 notable trend.

360 3.2. NGHGI compared to other bottom-up estimates

361

362 Figure 3 shows UNFCCC (CRFs and BURs) estimates from EU and seven non-EU countries
 363 compared to global bottom-up inventories.



364

365 *Figure 3: Total anthropogenic CH₄ emissions (excl. LULUCF) from bottom-up (BU) inventories as: UNFCCC*
 366 *NGHGIs (2023) of CRFs (EU, USA and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia*
 367 *(3rd in 2021), DR Congo (1st in 2022), India (all three BURs: 2016, 2018 and 2021) and three other global*



368 *datasets: EDGARv7.0, GAINS (no IPPU) and FAOSTAT (PRIMAP based, except for AFOLU). The relative error*
369 *on the UNFCCC value represents the NGHGI (2023) reported uncertainties computed with the error*
370 *propagation method (95% confidence interval) and gap-filled to provide respective estimates for each year.*
371 *China and Indonesia report uncertainties, for 2014 and 2000 and 2019 respectively (BUR). Total COD*
372 *UNFCCC BUR emissions do not include IPPU. The EDGARv7.0 uncertainty is only for 2015 and was*
373 *calculated according to Solazzo et al., 2021 for EDGARv5.0. The mean of overlapping time series was calculated*
374 *for 1990-last available year as following: 2021 for UNFCCC NGHGI (2023), EDGARv7.0 and FAOSTAT and*
375 *2020 for GAINS.*

376

377 From Figure 3, it is notable that, except for the EU and USA which show decreasing trends in emissions
378 from all data sets (USA except for GAINS), all the other countries show increasing trends. The match between
379 UNFCCC reported emissions and all the data sources is satisfactory, with few exceptions.

380 For the EU, the difference between the UNFCCC NGHGI 1990-2020 average and the other three data
381 sets is less than 5 %. As previously discussed, the inventory-based data sources are consistent with each other
382 for capturing recent CH₄ emission reductions, but they are not independent because they use similar methodology
383 with different versions of the same AD (Petrescu et al., 2020, Figure 4).

384 For the USA, GAINS reports high emissions after 2010, with strong growth. This divergence is largely
385 found in the Energy sector, resulting from the EFs used for conventional gas production as well as for
386 unconventional shale gas extraction, which has increased rapidly since 2006 due to the development of hydraulic
387 fracturing technology (Supplementary Figure S6-1 in Höglund-Isaksson et al., 2020). The high share of
388 emissions from unconventional shale gas can be explained by the GAINS EFs which, in the absence of published
389 factors, are derived from the residual emissions after having subtracted estimated emissions for oil production
390 and conventional gas production from the total upstream emission estimated by Alvarez et al., (2018, Table 1)
391 As Alvarez et al. 2018 do not specify emission factors by type of gas produced, GAINsv4 splits it based on
392 activity data from other references (IEA-WEO, 2018 and EIA, 2019). On the other hand, the NGHGI EF seems
393 to be too low, and this is reflected by the low oil and gas emissions reported by the USEPA, 2017 in 2015,
394 compared to Alvarez et al., 2018 (Supplementary Table S6-3, Höglund-Isaksson et al., 2020). For the USA, total
395 gas production increased by 47 % between 2006 and 2017. Revisions for the agricultural livestock emissions
396 concern updates of AD and reported EFs to statistics from FAOSTAT (2018) and CRFs UNFCCC (2016; 2018),
397 and a review of available technical abatement options for CH₄.

398 For Brazil, UNFCCC and GAINS report emissions of similar magnitudes and trends. The EDGARv7.0
399 and FAOSTAT report on average around 23 % more emissions for the 1990-2021 period, but closely follow the
400 NGHGI trends. The similarity between trends could be explained by the use of the same EFs following Tier-1
401 IPCC 2006 Guidelines and UNFCCC NIRs (Janssens-Maenhout et al., 2019), while the higher emissions could
402 appear when using different AD information.

403 For China the inventory estimates agree with the BUR reported data, with EDGARv7.0 showing the
404 highest estimates. According to both GAINS and EDGARv7.0, the primary drivers for growth in Chinese CH₄



405 emissions are due to a mix of sources, mainly from the IPCC 2006 sector 1.B.1, fugitive emissions from solid
406 fuels activity linked to increased coal mining.

407 In Indonesia the three global datasets agree well up until 2010. From 2010, the inventories show a
408 continued increase in emissions, while the UNFCCC BUR emissions suggest a decline. EDGARv7.0 reports a
409 large increase in emissions from fugitive emissions from solid fuels (coal mining) (IPCC 2006, sector 1.B.1.) at
410 an increased average rate of 19 % per year and has increased by a factor of 152 until 2021 compared to 1990
411 (Figure 3).

412 For Russia, GAINS emissions are much higher than NGHGIs and the other two data sets due to the
413 revisions of the assumptions on the average composition of the associated gas generated from oil production
414 based on information provided in Huang et al. (2015). The higher emissions in GAINsv4 might be caused by a
415 greater source from venting of associated gas instead of flaring. GAINsv4 estimates a decline in global CH₄
416 emissions in the first half of the 1990s, primarily a consequence of the collapse of the Soviet Union and the
417 associated general decline in production levels in agriculture and fossil fuels (see regional emission illustrations
418 in figures S2–1 of the SI). In addition, as described by Evans and Roshchanka (2014) and assumed in Höglund-
419 Isaksson (2017), venting of associated petroleum gas declined significantly in Russia due to an increase in flaring.
420 It is unclear why this happened, but a possible explanation could be that the privatization of oil production in
421 this period meant that the new private owners were less willing to take the security risks of venting and invested
422 in flaring devices to avoid potential production disruptions. This hypothesis is however yet to be confirmed
423 (Höglund-Isaksson et al., 2020). FAOSTAT data for the Russian Federation starts in 1992, since the country did
424 not exist before this date. The former USSR statistics were used prior to 1992 without adjustments and this is the
425 cause of the 1990 and 1991 outliers in time series.. The slightly increasing trend observed in EDGARv7.0 and
426 FAOSTAT are set by emissions from the Energy sector.

427 For DR Congo estimates from GAINS are not available because they only report aggregated emissions
428 from a few African regions. Both FAOSTAT (PRIMAP based) and EDGARv7.0 estimates show similar slowly
429 increasing trends, potentially indicating the use of similar prior statistics (EFs). For non-AFOLU sectors the
430 PRIMAP-hist third party data priority scenario used in FAOSTAT also uses EDGAR data as an input data source
431 explaining similarities in these sectors. On the other hand, UNFCCC BUR data reports a strong increase in
432 emissions, which is due to a rapid growth of CH₄ emissions from the Waste sector, by a factor of four until 2018
433 compared to 2000. This increase happened at an average yearly rate of +8 %, with an initial sharp increase of
434 +30 % between 2000 and 2001. As previously discussed (section 3.1.) we believe that DR Congo BUR reported
435 Waste emissions are improbable and further investigation is needed.

436 For India, all bottom-up global inventories show similar trends and magnitudes of anthropogenic CH₄
437 emissions. The emissions of CH₄ averaged across EDGARv7.0, GAINS and FAOSTAT are 67% (2010), 68 %
438 (2014) and 65 % (2016) higher than the Indian BURs. All three BU inventories show an averaged steady increase
439 of 1 % yr⁻¹ between 1990-2020.

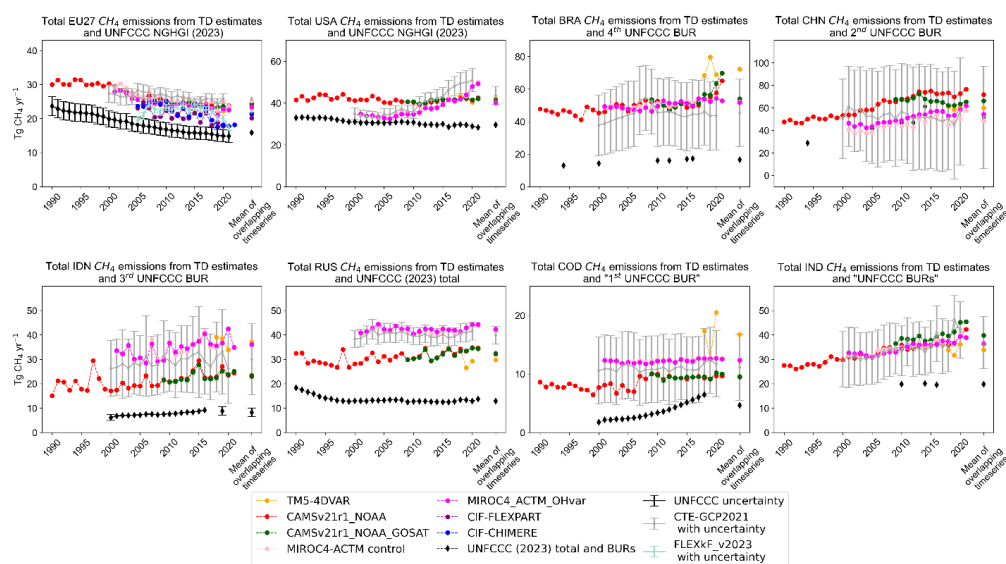
440 3.3. NGHGIs compared to TD atmospheric-based CH₄ estimates

441

442 Figure 4 presents the total TD estimates versus UNFCCC official reported emissions for the EU and the



443 seven non-EU emitters. The mean column on the right of each chart represents the mean of the overlapping time
 444 series (2009-last available year, except for TROPOMI, which was available only for 2018-2020). For China, the
 445 last BUR is available for 2014, and therefore we used that value. The inversions show total CH₄ emissions,
 446 including both anthropogenic and natural sources. We present here the total TD estimate against the
 447 anthropogenic NGHGI, emphasizing that the difference between BU and TD estimates might be due to the
 448 natural emissions.
 449



450
 451 *Figure 4: Total anthropogenic CH₄ emissions (incl. LULUCF) from UNFCCC NGHGI (2023) CRFs (EU, USA*
 452 *and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR Congo (1st in 2022),*
 453 *India (all three BURs: 2016, 2018 and 2021) and total TD estimates as following: for EU regional inversions*
 454 *(FLEKxv2023, CIF-FLEXPART and CIF-CHIMERE) and global inversions (TM5-4DVAR,*
 455 *CAMSv21r1_NOAA, CAMSv21r1_NOAA_GOSAT, CTE-GCP2021 and MIROC4-ACTM both runs) products.*
 456 *The relative error on the UNFCCC value represents the NGHGI (2023) reported uncertainties computed with*
 457 *the error propagation method (95% confidence interval) and gap-filled to provide respective estimates for each*
 458 *year. China reports uncertainties for 2014 (min 5.2 %, max 5.3 %) and Indonesia reports for 2000 and 2019 ,*
 459 *20 % and 19.9 % respectively. Total COD UNFCCC BUR emissions do not include IPPU. The last available*
 460 *years are CIF-CHIMERE (2022), TM5-4DVAR, CIF-FLEXPART and CTE-GCP2021 (2020) and*
 461 *FLEKxv2023, MIROC4-ACTM both runs, UNFCCC CRFs, and CAMSv21r1 both runs (2021). The mean of*
 462 *overlapping time series was calculated for 2009-2021, except for TM5-4DVAR (2018-2020).*

463
 464 Chandra et al., 2021 identify a few main sectors that triggered increases and decreases in the
 465 anthropogenic CH₄ emissions of different countries. The first is Energy, fugitive emissions from the oil and gas
 466 industry which helped to stabilize CH₄ concentration in the 1990s (decreased emissions), then they contributed
 467 to the renewed CH₄ growth since the late 2000s (increased emissions). The other major sectors that drove changes



468 in the CH₄ growth rate arose from Agriculture (increase in emissions from enteric fermentation and manure
469 management) and from Waste. The increase in emissions from enteric fermentation and manure management is
470 caused primarily by increased animal numbers (AD), and in addition by the greater intensity of ruminant farming
471 as estimated by the FAO and the emission inventories (e.g. EDGAR) which might take into account productivity
472 increases (Crippa et al., 2020; Wolf et al., 2017; FAOSTAT, 2018) while inventory emissions from Waste can
473 account for up to 43 % of the linear increase in emissions for the rest of the world.

474 In the EU, the averaged 2009-2021 total CH₄ emissions from global inversions are in the range of 23-
475 26 Tg CH₄ yr⁻¹, in line with previous estimates published in Petrescu et al. (2021, 2023). As this is the total flux,
476 while the UNFCCC NGHGI (2023) report only anthropogenic emissions (15.8 ± 1.8 Tg CH₄ yr⁻¹), the difference
477 can at least in part be explained by the sum of the natural emissions (~7 Tg CH₄ yr⁻¹). There is good agreement
478 in trends, but with inversions showing higher estimates. Tendentially, for the same period, regional inversions
479 report emissions between 20-22 Tg CH₄ yr⁻¹, lower than that of global inversions. Regional inversions use more
480 regional observations (e.g. ICOS, not just NOAA), have higher spatial resolution, and may thus better resolve
481 the transport. However, they may also have problems with the regional boundary conditions.

482 For the USA, averaged over the period 2009-2021, inversions indicate total CH₄ emissions in the range
483 of 40 - 44 Tg CH₄ yr⁻¹. In contrast, the UNFCCC NGHGIs (2023) report for the same period anthropogenic total
484 emissions of only 29 Tg CH₄ yr⁻¹. Regarding trends, those observed in the TD products are slightly increasing,
485 except for CAMS which shows no trend (Figure 4). The striking discrepancy between the trends from CAMS
486 and those from MIROC4-ACTM and CTE-GCP2021 are most likely caused by the increasing oil and gas
487 emissions from the Eastern USA (Permian Basin). The same increasing trend is also captured by GAINS (Figure
488 3). In their runs, both MIROC4-ACTM and CTE-GCP2021 use oil and gas priors from GAINS, while CAMS
489 uses priors from EDGAR (Figure 3). In SI, we discuss further differences in having CTE-GCP2021 run with
490 both EDGAR and GAINS oil & gas prior estimates.

491 For Brazil, inversions yield an average (range) total CH₄ emissions of 55 (42-72) Tg CH₄ yr⁻¹, with
492 TM5-4DVAR reporting the highest estimate. The UNFCCC estimate of anthropogenic emissions is 16.6 Tg CH₄
493 yr⁻¹. The two CAMS inversions report an increased trend during 2017-2021, with 15 Tg CH₄ higher emissions
494 in 2021 than in 2017. There is also a peak in the TROPOMI observation in 2019, and the TM5-4DVAR model
495 attributed this to biomass burning events, identified in the reported partitions (see
496 <https://doi.org/10.5281/zenodo.10276087> data figure).

497 For China, approximately 80 % of the CH₄ emission increase (21.5 Tg yr⁻¹) during 2000 – 2015 was
498 from fugitive emissions from coal (mines), consistent with what GAINS and EDGAR reports (Figure 3). The
499 TD estimates mostly agree, except for CAMS inversions which find 10 to 20 Tg CH₄ yr⁻¹ higher emission than
500 the other inversions. Both MIROC4-ACTM runs (control and OH inter-annual variability (IAV) varying run;
501 Patra et al., 2021) are in line with the BURs. Trend wise, all inversions agree on increased emissions after 2019.
502 Nevertheless, CAMS_GOSAT_NOAA registers a decrease after 2013 not seen in the other inversion trends.

503 For Indonesia, most TD results agree on the trend and show a slight increase in emissions. Similar trend
504 is also seen by the BURs. However, the CAMS inversions result in about 10 Tg CH₄ yr⁻¹ lower emissions than
505 the other inversions (MIROC4-ACTM and CTE-GCP2021). Regarding the East Asian estimates,
506 MIROC4_ACTM inversion simulates higher fluxes compared to the other inversions. Only recently they found



507 that annual total East Asian emissions have lowered more significantly than in Patra et al. (2016) or Chandra et
 508 al. (2021).

509 For Russia, the estimates from CAMS are both in the same range as the BU GAINS estimate (see Figure
 510 2) from 2000 onwards (between 30-40 Tg CH₄ yr⁻¹) but does not show such a strong decrease as GAINS from
 511 1990 to 2000, Fig. 2), while MIROC4-ACTM and CTE-GCP2021 are about 10 Tg CH₄ yr⁻¹ higher than CAMS).

512 For DR Congo, except for the TM5-4DVAR which also reports the highest emissions (wetlands - see
 513 partitions, Fig.5a), inversions do not show any trend. MIROC4-ACTM runs report about 5 Tg CH₄ yr⁻¹ higher
 514 emissions than both CAMS runs. The two high values in 2018 and 2020 seen by the TROPOMI satellite are
 515 triggered by high emissions from wetlands reported in the TM5-4DVAR partitions.

516 For India, the TD estimates of total emissions agree well on increased trends and magnitudes. In
 517 contrast, UNFCCC reporting does not show a trend, but too little reported data from BURs is available, therefore
 518 a plausible conclusion cannot be drawn.

519

520 3.4. Reconciliation and sectoral attribution of CH₄ emissions

521 3.4.1. Sectoral attribution of CH₄ emissions in TD products

522

523 Table 2 shows the partitions as originally reported by some of the inversions, which we name here
 524 “unharmonized partitions”. A straightforward, direct comparison of the fluxes is not possible because of the
 525 different ways each inversion allocates and groups the natural/anthropogenic fluxes. For example, not all
 526 inversions report soil fluxes as done by MIROC4-ACTM and CTE-GCP2021 (together with wetlands), or report
 527 the biomass burning fluxes separately from anthropogenic emissions (MIROC4-ACTM and TM5-4DVAR).
 528 Also the termites, oceans and geological fluxes are sometimes reported separately (MIROC4-ACTM) or grouped
 529 in “Other” (CTE-GCP2021, TM5-4DVAR). Regarding the anthropogenic emissions, TM5-4DVAR reports them
 530 as other, providing a separate partition for rice.

531

532 *Table 2: Unharmonized partitions originally reported by inverse products:*

Inversion	Anthropogenic	Rice	Soils	Wetlands	Ocean	Termites	Geological	Biomass burning	Other
CAMSv21r1 (both runs)	Yes (in Other)	Yes	No	Yes	Yes (in Other)	Yes (in Other)	No	Yes	Yes
MIROC4-ACTM (control and OH var)	Yes ((Agr, Waste, Oil/Gas, Biofuel, coal)	Yes (in Agr.)	Yes	Yes	Yes	Yes	Yes	Yes	Yes (separated)
CTE-GCP2021*	Yes (Agr, waste, fossil fuel, biofuel, biomass burning)	Yes (in Agr.)	Yes (BIO)		Yes (In Other)	Yes (In Other)	Yes (In Other)	In anthr.	Yes (Ocean, Termites, Geological)
CEOS (GOSAT)	Yes (Livestock, rice, waste, coal, oil, fire)	In anthr.	No	Yes	No	No	Yes (seeps)	In anthr.(but separate)	only seeps
TM5-4DVAR (TROPOMI)	Yes (in Other)	Yes	No	Yes	Yes (in other)	Yes (in other)	Yes (In Other)	Yes	Yes**



Inversion	Anthropogenic	Rice	Soils	Wetlands	Ocean	Termites	Geological	Biomass burning	Other
GEOS-Chem CTM (TROPOMI for USA)	Yes (Livestock, Oil Gas, Landfills, Wastewater, Other anthro (rice))	In. other anthr.	No	Yes	Yes (In Other)	Yes (In Other)	Yes (In Other)	Yes (In Other)	Yes***

533 *CTE-GCP2021 partitions refer to anthropogenic, bio and other.

534 ** In TMS-4DVAR (similar to the CAMSv20 set-up and CAMSv21r1), the "Other" partition includes anthropogenic sources
 535 except for the rice paddies. It also includes the small fluxes from termites, oceans, soil sink, geological etc.). More details on
 536 priors are found in Petrescu et al., 2023b, Priors table.

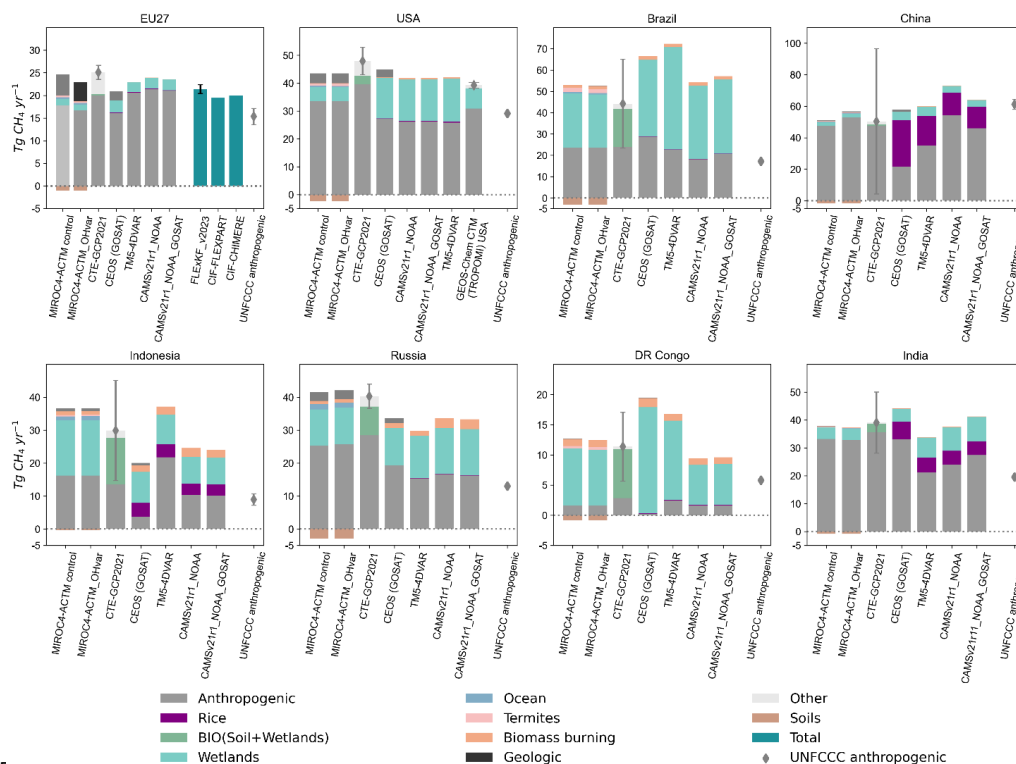
537 ***Named Other biogenic

538

539 Figure 5 shows the UNFCCC NGHGI anthropogenic total reported estimate (diamond) next to all TD
 540 estimates. All global inversions report total and disaggregated partitions, while the regional inversions report
 541 only the total emissions (green column).

542

UNFCCC anthropogenic and unharmonized CH_4 emissions from partitions reported by TD estimates (average 2015-last available year)



544 Figure 5: Total (green) and disaggregated anthropogenic and natural CH_4 emissions from TD estimates
 545 compared to UNFCCC NGHGI (2023) anthropogenic emissions (incl. LULUCF) (diamond) for the EU and



546 seven global emitters outside the EU (USA, Brazil, China, Indonesia, Russia, DR Congo and India). The
 547 UNFCCC anthropogenic value represents the sum of all five IPCC sectors (Energy, IPPU, Agriculture,
 548 LULUCF and Waste). The partitions reported by the TD global inversions are detailed in Table 2. The relative
 549 error on the UNFCCC CRF value represents the NGHGI (2023) reported uncertainties computed with the error
 550 propagation method (95% confidence interval) and gap-filled to provide respective estimates for each year (see
 551 Petrescu et al., 2023a, Appendix). China value and uncertainties (min 5.2 %, max 5.3 %) are for 2014 only and
 552 Indonesia uncertainties for 2019, 19.9 %. For the USA CEOS (GOSAT) we used the Nesser et al., 2023 total
 553 uncertainty of min 1.1 and max 1 Tg yr⁻¹. CTE-GCP2021 provides uncertainties for each partition, but here the
 554 uncertainty of the total flux is shown. FLEXkF_v2023 reports the relative uncertainty (%) of the posterior
 555 emissions. The plotted data represents the average between 2015 and last available year as follows: CIF-
 556 CHIMERE (2022), TM5-4DVAR, CIF-FLEXPART and CTE-GCP2021 (2020) and FLEXkF_v2023, MIROC4-
 557 ACTM both runs, UNFCCC CRFs, and CAMSv21r1 both runs (2021). GEOS-Chem CTM (TROPOMI) USA
 558 reports only for 2019 (Nesser et al., 2023).

559 We note that CTE-GCP2021 reports the net natural land-biosphere flux “BIO flux” (soil+wetlands),
 560 while other inversions report wetlands and soil separately, thus the BIO flux will be smaller than the wetlands,
 561 because mineral soils are a CH₄ sink. Regarding the allocation of different fluxes to partitions, sometimes Rice
 562 emissions are part of the Agriculture component (anthropogenic partition) (MIROC4-ACTM, CTE-GCP2021)
 563 while CEOS (GOSAT) and GEOS-Chem CTM (USA TROPOMI) report separate partitions for Rice in
 564 Anthropogenic emissions. Same for the biomass burning - CTE-GCP2021 and CEO report it as part of
 565 Anthropogenic emissions, while GEOS-Chem CTM as part of Others. The rest of the inversions report it
 566 separately; this different allocation makes comparisons for these two sources challenging.

567
 568 To facilitate comparisons between all TD products, we aggregated and harmonized the partitions in
 569 three main categories, as summarized in Table 3 and Figure 6. The dark green columns in Figure 6 show the total
 570 flux for regional EU inversions which did not report partitions.

571

572 *Table 3: Harmonized partitions from inverse products:*

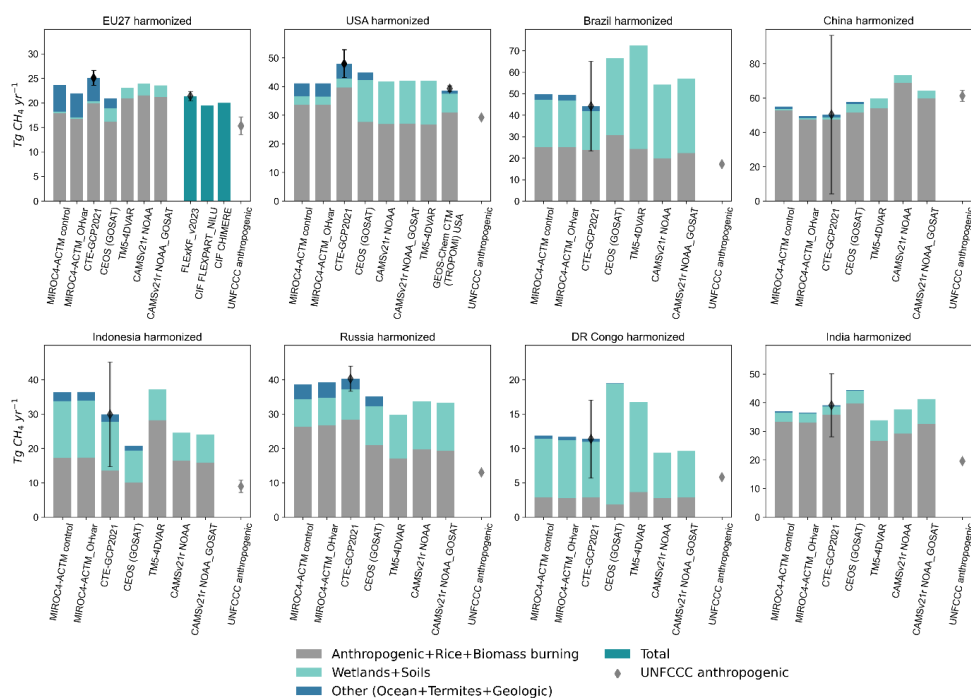
Inversions	Anthropogenic + Rice + Biomass burning			Soils + Wetlands		Other (Ocean + Termites + Geological)		
	Anthropogenic	Rice	Biomass burning	Soils	Wetlands	Ocean	Termites	Geological
CAMSv21r1 (both runs)	= Other	Yes	Yes	No	Yes	Yes	Yes	Yes
MIROC4-ACTM (control and OH var)	Yes ((Agr (livestock + rice), Waste, Oil/Gas, Biofuel, coal)	In Agr.	Yes, summed to anthr.	Yes	Yes	Yes	Yes	Yes
CTE-GCP2021*	Yes (Agr (rice is in), waste, fossil fuel, biofuel, biomass burning)	in Agr.	In anthr.	Yes (BIO)		Yes (Other)		
CEOS (GOSAT)	Yes (Livestock, rice, waste, coal, oil, fire)	In anthr.	In anthr.	No	Yes	No	No	Yes



TM5-4DVAR (TROPOMI)	Others + Rice+ BB	In anthr.	Yes, summed to anthr.	In Other	Yes	Yes	Yes	Yes
GEOS-Chem CTM (TROPOMI) USA	Yes	In anthr.	In other biogenic	No	Yes	Yes	Yes	Yes

573 *CTE-GCP2021 partitions refer to Anthropogenic, Bio and Other. Other fluxes are imposed
 574

UNFCCC anthropogenic and harmonized CH_4 emissions from partitions reported by TD estimates (average 2015-last available year)



575
 576 *Figure 6: Total (green) and disaggregated anthropogenic and natural CH_4 emissions from TD estimates*
 577 *compared to UNFCCC NGHGI (2023) anthropogenic emissions (incl. LULUCF) for the EU and seven global*
 578 *emitters (USA, Brazil, China, Indonesia, Russia and DR Congo). The UNFCCC anthropogenic value represents*
 579 *the sum of all five IPCC sectors (Energy, IPPU, Agriculture, LULUCF and Waste). The partitions reported by*
 580 *the TD global inversions are harmonized and detailed in Table 3. The relative error on the UNFCCC CRF value*
 581 *represents the NGHGI (2023) reported uncertainties computed with the error propagation method (95%*
 582 *confidence interval) and gap-filled to provide respective estimates for each year (see Petrescu et al., 2023a,*
 583 *Appendix). In 2014, China UNFCCC value and reported uncertainties (min 5.2 % and max 5.3 %) are for 2014*
 584 *while Indonesia reported uncertainties for 2019, 19.9 %. India UNFCCC value is for 2016. CTE-GCP2021*
 585 *provides uncertainties for each partition, but here we plotted the uncertainty of the total flux. FLEXkF_v2023*
 586 *reports the relative uncertainty (%) of the posterior emissions. The plotted data represents the average between*
 587 *2015 and last available reported year as follows: CIF-CHIMERE (2022), UNFCCC CRFs, TM5-4DVAR, CIF-*



588 *FLEXPART and CTE-GCP2021 (2020) and FLEXkf_v2023, MIROC4-ACTM both runs, and CAMSv21r1 both*
 589 *runs (2021). GEOS-Chem CTM (TROPOMI) USA reports only for 2019 (Nesser et al., 2023).*

590 **3.4.2. Reconciliation of BU and TD CH₄ estimates**
 591

592 Figure 7 summarizes the total CH₄ fluxes for the EU and the seven global emitters as following: BU
 593 anthropogenic sources disaggregated per sectors, BU natural emissions, TD natural emissions from regional and
 594 global inversions, and total emissions from global TD estimates (see 2.3 and SI for description of all data
 595 products).

596 The way data is currently reported by the inversions, is inconsistent regarding the comparison between
 597 BU natural and TD natural sources. TD products differ in the sources they report (Table 2) or they allocate them
 598 to different categories. We consider natural the following sources: biomass burning, soils, oceans and termites
 599 (often reported by inversions under category “Other”), wetlands, geological and lakes & reservoirs (or
 600 freshwaters). Due to lack of information, biomass burning emissions were considered among the natural sources,
 601 recognizing that in regions like tropical forests, some of these events are influenced by human intervention. To
 602 make the products from Figure 7 comparable, we added the missing BU information from TD, and vice-versa,
 603 presented in hashed pattern. In this way, comparison between BU and TD natural emission estimates is consistent
 604 regarding the “apples to apples” comparison, but became “apples of different flavors” (see Table 4):

605

606 *Table 4: BU and TD natural partitions as presented in Figure 7:*

Product name	TD natural partitions		
	reported	missing*	added from
TM5-4DVAR (TROPOMI)	BB and wetlands	oceans, termites, soils, geological, lakes and reservoirs	MIROC4-ACTM (termites, oceans and soils), DAAC lakes and reservoirs, geological, updated for this study (SI)
CEOS (GOSAT)	Fires (BB), Seeps and wetlands	termites, oceans, soils, lakes and reservoirs	MIROC4-ACTM (termites, oceans and soils), DAAC lakes and reservoirs
MIROC4-ACTM control	BB, wetlands, oceans, termites, soils, geological	lakes and reservoirs	DAAC lakes and reservoirs
MIROC4-ACTM_OHvar	BB, wetlands, oceans, termites, soils, geological	lakes and reservoirs	DAAC lakes and reservoirs
CAMSv21r1_NOAA	BB, wetlands, “Others” include anthropogenic and was not used	termites, oceans, soils, lakes and reservoirs, geological	MIROC4-ACTM (termites, oceans and soils), DAAC lakes and reservoirs, geological, updated for this study (SI)
CAMSv21r1_NOAA_GOSAT	BB, wetlands, “Others” include anthropogenic and was not used	termites, oceans, soils, lakes and reservoirs, geological	MIROC4-ACTM (termites, oceans and soils), DAAC lakes and reservoirs, geological, updated for this study (SI)
CTE-GCP2021	soils + wetlands (BIO), termites and oceans	BB	BB from GFEDv4.1s
Product name	BU natural partitions		
	available	missing**	added from



GFEDv4.1s DAAC LPJ-GUESS Geological	GFEDv4.1s DAAC LPJ-GUESS Geological updated in this study (SI)	soils termites oceans	MIROC4-ACTM
--	--	-----------------------------	-------------

607 note: in TD products termites, oceans emissions are imposed from existing literature

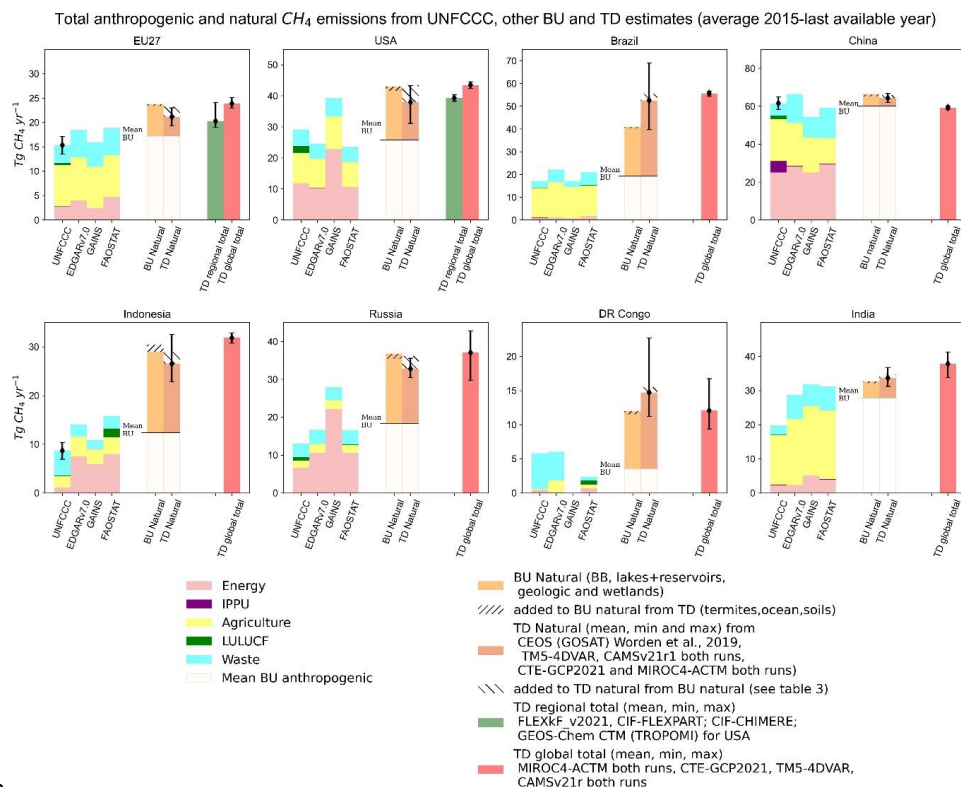
608 * presented as hatched pattern in the figure “\\”

609 ** presented as hatched pattern in the figure “///”

610

611 We note from Figure 7 that in all Annex-I countries (EU, USA, Russia) and China, TD and BU natural
 612 emissions are consistent with each other, after including the missing sources, as detailed in Table 4. For Brazil
 613 and DR Congo, the gap between the two natural components is highly significant, while less for Indonesia and
 614 India. We hypothesize that mapping of the wetlands extent might cause these inconsistencies.

615 For an easier visual comparison and reconciliation between BU and TD estimates, we added in white
 616 the mean of the BU anthropogenic estimates, underneath the BU and TD natural estimates. We note that for most
 617 countries, the sum of the anthropogenic and natural components matches those of the TD global total estimates.
 618 This gives confidence that, to a certain extent and albeit inconsistencies between products, BU anthropogenic
 619 emission estimates are accurate and consistent with the observation-based estimates and can be used to reconcile
 620 with the atmospheric-based estimates.



62

622 *Figure 7: Total anthropogenic and natural CH₄ emissions from BU and TD estimates presented as average of*
 623 *2015-last available year for EU and seven global emitters (USA, Brazil, China, Indonesia, Russia, DR Congo*
 624 *and India). The BU anthropogenic estimates belong to: UNFCCC NGHGI (2023) CRFs and BURs (incl.*
 625 *LULUCF) as totals and sectoral shares, EDGARv7.0, GAINS and FAOSTAT-PRIMAP. The relative error on the*
 626 *UNFCCC CRF value represents the NGHGI (2023) reported uncertainties computed with the error propagation*
 627 *method (95% confidence interval) and gap-filled to provide respective estimates for each year (see Petrescu et*
 628 *al., 2023a, Appendix). In 2014, China reported an uncertainty of min 5.2% - max 5.3%. The BU Natural*
 629 *emissions for the EU are the sum of the VERIFY products (biomass burning, inland waters, geological and*
 630 *peatlands plus mineral soils as described in Petrescu et al., 2021 and 2023a, Appendix A2.1). For the seven non-*
 631 *EU emitters, the BU Natural fluxes are the sum of wetland emissions (LPJ-GUESS), lakes and reservoirs fluxes*
 632 *(ORNL DAAC, Johnson et al., 2022), geological (updated activity in SI) and biomass burning emissions*
 633 *(GFED4.1s). The TD natural global estimates are presented in Table 1. The uncertainty on the TD natural*
 634 *emissions is the min/max of all estimates. To both BU and TD estimates missing information was added (see*
 635 *Table 4). The natural emissions have been plotted starting at the mean of the BU anthropogenic estimates, to*
 636 *retain comparability across the natural emission estimates, but also compare with the total TD estimates. The*
 637 *total regional TD estimates (for EU) belong to the mean and min/max of FELXkF_v2023, CIF-FLEXPART and*
 638 *CIF-CHIMERE and for USA GEOS-Chem CTM (TROPOMI) for the year 2019 (Nesser et al., 2023). The total*
 639 *global TD inversions represent the average of the 2015-last available year of the mean and min/max of CTE-*



640 *GCP2021, MIROC4-ACTM both runs, CAMS v21r both runs and TM5-4DVAR. The last available years are*
641 *2022 for CIF-CHIMERE, 2021 for EDGARv7.0, FAOSTAT, MIROC4-ACTM both runs, UNFCCC CRFs, and*
642 *CAMSv21r1 both runs, and 2020 for CIF-FLEXPART and CTE-GCP2021. TM5-4DVAR partitioned data is only*
643 *available between 2018-2020.*

644 However, this should be interpreted with caution because in Europe, natural emission priors come from
645 regional ecosystem model simulations, where drained peatland, drainage ditches areas, and pristine areas are
646 lumped together. Therefore, if both LULUCF sector and natural BU emissions are included in the total budget
647 estimation, there is some overlap and possible double-counting. Especially, ecosystem model estimates of ‘soil
648 sink’ or ‘inundated soil emissions’ may be overlapping with NGHGI managed peatland forest soil category (or
649 agricultural soils). The separation of emissions into different categories requires further clarification together
650 with inventory makers. Furthermore, it should be assessed which emissions should be called natural and which
651 anthropogenic (e.g., LULUCF, Agriculture) by inversions.

652

653 **4. Discussion and recommendations on reconciliation procedures**

654

655 Figures 3, 4a, b, 5 and 6 visually compare inversions and inventory estimates. A valid comparison
656 should include consistent types of information (tiers, priors, methodology) and show a full uncertainty analysis
657 to determine whether differences between estimates are statistically significant. However, very few datasets
658 provide this necessary uncertainty information that is critical for interpreting comparisons. Furthermore, from
659 the perspective of verification, some adjustment of estimates (additions/subtractions) may be required to reflect
660 variations in which anthropogenic and natural sources and sinks are included in (or excluded from) the respective
661 estimates. The following paragraphs discuss the potential challenges and opportunities for comparing NGHGI
662 against external BU and TD estimates from the perspective of national agencies responsible for inventory
663 compilation and reporting.

664 The two most common issues were identified by the series of previous syntheses and other scientific
665 literature (Petrescu et al., 2021, 2023; McGrath et al., 2023, Andrew 2020; Grassi et al., 2018). The first is the
666 **geographical scope and boundaries** of inverse modelling versus inventory estimates. Independent of which
667 GHG are analyzed, a general system boundary issue is masking of gridded results to the country level, where it
668 is important to know how modelling groups have defined emissions in each grid cell and to ensure the mask
669 correctly captures the territorial perspective (e.g., country and economic zones), in line with how official
670 NGHGIs are reported (EEA, 2013). Given that NGHGIs do not report gridded emissions, comparison with
671 gridded inversions may introduce some under/over estimation of total emissions within the country border.
672 Particularly for inversion models with a coarser grid, aggregation of the grid cells to country levels will not
673 necessarily give a perfect match. This problem lessens with larger regions, or regions with long coastlines, and
674 is one reason why VERIFY (<https://verify.lsce.ipsl.fr/>) aggregates many smaller countries together to larger
675 regions. The second issue linked to boundaries is the **structural system boundaries**, an inconsistency with great
676 implications in comparing the inventory- with inversions-based estimates for **source attribution**, e.g.,
677 anthropogenic vs. natural. Most emission inventories aim at estimating anthropogenic emissions, while most
678 inversions see both anthropogenic and natural emissions. Thus, methods are needed to separate the anthropogenic



679 flux from the total flux (Deng et al. 2022, and above section 3.4), similarly to what was done in Andrew 2020
680 who refers to system boundaries when comparing the various inventory datasets focusing on what sources are
681 included in the respective estimates. This is a particularly important issue for CH₄ where, globally, natural
682 emissions are of similar magnitude as anthropogenic emissions, with larger variations at regional scales, mainly
683 due to seasonality (i.e. wetlands). Further, climate change may modify natural emissions in ways that models
684 can't yet resolve, for example, a warmer climate may increase natural emissions of CH₄ (Yvon-Durocher., et al.,
685 2014).

686 Another identified source of inconsistencies between inventory estimates and inversions-based
687 estimates are **uncontrolled events**. CH₄ from the fossil fuel industry can contribute to high releases to the
688 atmosphere over a short period of time, given the large number of uncontrolled emission point sources in oil and
689 gas (O&G) and coal production areas worldwide (Jackson et al., 2020). Such processes include leakage from
690 landfills, spontaneous events from oil and gas production activities, so-called uncontrolled gas well blasts etc. In
691 recent years, there is a high interest in quantifying emissions from these events (Jacob et al., 2016) and a series
692 of top-down studies using satellite imagery quantifies these sparse but important events, which are difficult to
693 include in the national inventories leading to a potential underestimate of emissions. However, they can be
694 identified and quantified to ultimately be included in the inventories, as done for the USA by Massakkers et al.,
695 2016 and 2022. Recently, under the CoCO₂ project (<https://coco2-project.eu/>) a hot-spot satellite detection
696 interactive map ([Published studies on hot spot detection \(CO₂, CH₄\) - uMap \(openstreetmap.fr\)](#)) was released as
697 a user-centric interface featuring published studies on hot-spot detection between 2010 to 2021. It allows for
698 advanced filtering by year, gas, activity, geographical zone, and country.

699 Varying **temporal and spatial resolutions** are also an important factor to consider. Inventory-based
700 estimates use annual temporal resolutions and are not spatially distributed yet are compiled at a high source-
701 sector resolution. Inversion models can produce estimates at a potentially fine grid scale (kilometers) and fine
702 temporal resolution (hours). The optimized net fluxes and/or partitioned anthropogenic and natural flux totals
703 may be too coarse to identify biases at the source-sector level used in inventories. A region (e.g. city or region
704 within a country) would be interested in annual emissions, but the availability of more detail could help identify
705 and manage 'events' (acute pipeline leaks) from individual facilities (power plant or industrial site), and this may
706 help inventory agencies verify emissions from these facilities. The fine spatial resolution allows aggregation to
707 city- or region-level, matching as close as possible to jurisdiction boundaries. Uncertainties in inversion estimates
708 reduce when aggregating in time and space since the grid-cell errors are often negatively correlated. Thus, from
709 TD, we can be more confident in estimates for large regions than in estimates for small regions or for a single
710 grid cell. This problem lessens with denser observational coverage.

711 When comparing inventory- and inversion-based emissions, there are difficulties in analyzing **trends**
712 due to different time scale **variability**. Inventory-based approaches report emissions at the annual level, but often
713 do not consider intra-annual (seasonal) variations, important for the microbial sources. Further, the Paris
714 Agreement is set around five-yearly Global Stocktakes, which indicates a desire to average trends, prioritizing
715 the multi-annual trend over IAV, canceling out extremes from both weather and socio-economic fluctuations.
716 Inversion models, on the other hand, include variations over a wide range of timescales, but in particular for IAV
717 (e.g. OH and weather) that remains challenging to assess. For an effective comparison, inversion-based estimates



718 need to have IAVs statistically removed to make comparisons with NGHGs easier (e.g., 5-year or 10-year
719 averages or trend analysis). Additionally, averages of ensembles of inversions may mask underlying differences
720 and trends in individual inversions. More broadly, methods to identify if a difference between two independent
721 datasets is statistically significant (levels of trends) exist but are inhibited by: a) information on inventory
722 uncertainty; and b) computation cost to generate inversion emission uncertainties. A suite of approaches to
723 resolve uncertainties in both BU and TD methods, and to enable comparisons between approaches, will be
724 important for the design and usefulness of future integrated monitoring systems. Some attempts have been initiated
725 (e.g. CIF (Berchet et al., 2021) in the VERIFY H2020 project and the CH₄ inversion intercomparison funded by
726 WMO).

727 From the point of view of the observation information, a more valid comparison between inversions is
728 made when all inversions use the same **priors**. In this context, we define as priors input data in the form of
729 atmospheric observations (e.g. satellite retrievals, ground-based observation networks (ICOS)) and/or bottom-
730 up emissions datasets (e.g. EDGAR, GAINS) used as input parameters to the inverse models. When combined
731 with observation data, the inversion system produces a posterior estimate of emissions, which can then be
732 compared to the prior estimate, preferably incorporating a full uncertainty analysis. The posterior emissions
733 depend to some extent on the prior that was used; the extent of this dependency is determined by the number of
734 observations used in the inversion, by how the observations relate to the emissions (governed by atmospheric
735 transport) and by the uncertainties assigned on the prior emissions and the observations. Thus, the inversions
736 would be more robust with better quantified uncertainties for the prior emissions. Whereas the comparison on an
737 inversion with NGHGs or other inversions would be made more robust by having more information on how
738 dependent the posterior estimate is on the prior. This stresses the need for more systematic measurements of
739 fluxes necessary to produce adequate prior data (Bastviken et al., 2022) and synthesized atmospheric
740 observations with their uncertainties to robustly constrain the inversions.

741 It is not generally clear how inventory uncertainties can be compared to inversion uncertainties;
742 however, it is important that both methods provide comprehensive **uncertainty estimates**. The “uncertainty
743 reduction” characterizes the improvement brought by the inversion over the prior emission estimate. The prior
744 emissions used as input into an inversion model should have uncertainty statistics, and a full inversion analysis
745 will include uncertainties on the posterior estimate, with the reduction in uncertainty between the two estimates
746 of particular interest. The inventory-based emission estimate will additionally have uncertainty estimates, though
747 some argue these statistics are not sufficiently robust for verification purposes (National Academies of Sciences,
748 Engineering, and Medicine 2022). There are often offsets in inversion models, because of systematic
749 inconsistencies between observations and chemistry-transport models, which may make trends more robust than
750 instantaneous estimates. Strongly dependent on the use of the information, uncertainties might be given different
751 weights, e.g. in a policy context, the uncertainty on the emission trend may be more important than emission
752 level uncertainties. However, this is harder to estimate as it requires knowledge of correlations in emission
753 estimates over time. Given the above listed challenges in quantifying uncertainties, research projects such as
754 VERIFY and global initiatives like the Global Carbon Project (GCP), make use of multi-model analysis
755 (**ensembles**). This way the UNFCCC emission inventory is compared against, for example, 13 land surface
756 models and 22 inversion models (Saunois et al., 2020; Deng et al. 2022). From a scientific perspective, the model



757 ensemble is often considered a more robust estimate of the mean and uncertainty, as transport model uncertainty
758 can be accounted for. From an inventory perspective, individual model comparisons may be more efficient, as
759 various input variables or processes can be compared directly to the inventory. Doing this for each model
760 becomes time consuming. Currently, most inventory comparisons in UNFCCC National Inventory Reports (UK,
761 Switzerland) use single model comparisons.

762 Finally, there are no **standardized procedures** for inverse analysis systems; therefore, there is room
763 for additional progress and refinement of emission estimates and uncertainties derived from atmospheric
764 observation and inverse models. The Community Inversion Framework (CIF; Berchet et al., 2021) is a good
765 example and moves in this direction. However, improvements are still needed to ensure common formatting and
766 presentation of the results, in addition to the use of common language and terminology.

767 Based on the above discussions, improvements/additions in the following areas could enhance the
768 information that national inventory agencies could gain from comparing NGHGs with other BU and TD
769 estimates:

- 770 1. Generate spatially distributed data of NGHGs by the respective inventory agencies.
- 771 2. A better quantification of uncontrolled spontaneous events of point source estimates to complement
772 NGHGs (Maasackers et al., 2016 and 2022 for the EPA in the USA).
- 773 3. Denser atmospheric observation networks to feed into inversions and reduce uncertainties at grid-cell levels.
- 774 4. Routine provision to quantify uncertainties from inversion models.
- 775 5. Develop methods to compare estimates and ensembles with inventory estimates.
- 776 6. Increase quality in systematic measurements of fluxes to produce better priors with better uncertainties.
- 777 7. More accurate transport models to increase robustness in emission estimates.
- 778 8. Clear prescribed inventory methods to estimate and assess uncertainties, particularly statistical significance.
- 779 9. Accurate model outputs (EFs) to be used in direct comparisons with inventory data.
- 780 10. Accurate estimates of natural emissions with better spatial distribution, to subtract from total inversion
781 estimates and compare with anthropogenic NGHGs.
- 782 11. Develop a common language, terminology and data formats to compare inversions with NGHGI formats.

783

784 **5. Data availability**

785

786 Data files reported in this work which were used for calculations and figures are available for public
787 download at <https://doi.org/10.5281/zenodo.10276087> (Petrescu et al., 2023b). The data are reachable with one
788 click (without the need for entering login and password), with a second click to download the data, consistent
789 with the two click access principle for data published in ESSD (Carlson and Oda, 2018). The data and the DOI
790 number are subject to future updates and only refer to this version of the paper. The raw gridded data is available
791 upon request, directly from the data providers, as detailed in the Supplementary Information, Table S2.

792 **6. Conclusions**

793



794 We analyzed data from both anthropogenic and natural CH₄ fluxes, from both BU and TD observation-
795 based estimates (Table 1). BU estimates show that, for most non-Annex I countries, the largest CH₄ emissions
796 are from the Energy sector, followed by emissions from the Agricultural sector (Figure 7), primarily from
797 ruminant livestock (enteric fermentation). The inversions (Figure 5) attribute most of the fluxes to the
798 anthropogenic emissions (gray) including rice (purple) (EU, China) while in tropical countries emissions are
799 attributed mostly to natural processes (wetlands).

800 The EU and the seven large emitters analyzed here contribute an anthropogenic emission of 173 Tg CH₄
801 yr⁻¹ (sum of last UNFCCC reported year, Figure 1,2), representing roughly half of the total global anthropogenic
802 emissions (386 Tg CH₄ yr⁻¹) reported by EDGARv7.0 in 2021, while the average of the anthropogenic component
803 from the atmospheric global inversions (MIROC booth runs, CTE-GCP2021, CEOS and CAMS booth runs) is
804 181 Tg CH₄ yr⁻¹ (Figure 5). Regarding comparability issues (Petrescu et al., 2023b, Matrix products table),
805 comparisons between UNFCCC and BU products (Figure 3) highlight substantial deviations, likely mainly due
806 to assumptions regarding gas/oil emissions (e.g., GAINS for Russia and the USA).

807 We highlighted the challenge of reconciling BU and TD estimates, due to different priors used in the
808 simulations (Petrescu et al., 2023b, Priors Table). It is also challenging to compare different TD products due to
809 the source attributions and allocation of fluxes to different sectors and activities within sectors (Table 2 and
810 Figure 5).

811 The comparison between UNFCCC and the TD estimates (Figure 4) agrees largely with the findings of
812 Deng et al. (2022) who applied different methodologies to calculate natural emissions. This approach produced
813 different results and highlighted the need for better estimation of natural fluxes from ecosystem-based models,
814 similar to our findings (Brazil and DR Congo, Figure 7). For the moment, it is difficult to discuss and draw
815 conclusions regarding emissions trends seen by inversions. From Figure 4, only UNFCCC shows clear patterns
816 during the last three decades (e.g. declined in the USA, and EU (regulations) and Russia (dissolution of the Soviet
817 Union) because are driven mainly by the anthropogenic component which is better constrained in bottom-up
818 inventories, while inversions include the natural component as well. Therefore, as described above in section 4,
819 comparability issues such as IAV and seasonal variability might strongly influence trends.

820 Given that, in most cases, the UNFCCC BURs reports are incomplete for the non-Annex I parties
821 (China, Indonesia, DR Congo) it is important to acknowledge that the TD estimates might become a useful way
822 to complement inventories and play a role in the validation of the BU estimates. In most cases, the gap between
823 the anthropogenic BU fluxes from inventories and total TD fluxes can be largely explained by the natural fluxes
824 (Figure 7).

825 There is still a pressing need for reporting of uncertainties in both prior and posterior emissions, even if
826 some TD inversions do report it (CTE-GCP2021 and FLEXkF_v2023, Figure 4) as the standard deviation of
827 ensemble members. The use of a variety of priors across different inversion systems can also inhibit
828 comparability with inventories and between inversions. Generally, inversions are still ill-constrained by
829 observations (only 60 sites globally plus satellites) and the prior flux uncertainty for each of the 54 regions is
830 large. Therefore, the monthly results could be more ill-constrained than the annual totals.

831 Even if comparisons between CH₄ inversion estimates and NGHGs are currently uncertain because of
832 the spread in the inversion results, TD inversions inferred from atmospheric observations represent independent



833 data against which inventory totals and trends can be compared, considering the recommendations listed at the
834 end of section 4.

835

836 **7. Appendix**

837

838 All the information regarding models/methods descriptions is available in the Supplementary Information (SI)
839 file. Appendices A1 and A2 in Petrescu et al., 2023a contain detailed information about Table 1 products. Further
840 information on new products together with references and contact details are found in Tables S1 and S2 in SI.

841

842 The tables with priors used by all the products and the matrix highlighting the comparability issues identified in
843 section 4 are found in the Zenodo data repository, Petrescu et al., 2023b.

844

845 **Supplementary Information (link)**

846

847

848 **Author contributions**

849

850 AMRP designed research and led the discussions; AMRP wrote the initial draft of the paper and edited all the
851 following versions; GPP drafted the initial version of section 4, edited the final version of this manuscript and
852 advised on the context; PP processed all the original EU data submitted to the VERIFY portal; RLT, SH, BM,
853 DaB, RL, PKP, AT, RMA, LHI, FNT, GC and JG edited and gave consistent comments and suggestions to the
854 manuscript; all co-authors are data providers and contributed to subsequent versions of the manuscript by
855 providing specific comments and information related to their data in the main text, providing as well product
856 descriptions for the Supplementary Information file.

857 **Competing interests**

858 At least one of the (co-)authors is a member of the editorial board of Earth System Science Data

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860

861

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