

1 Comparison of observation- and inventory- based methane emissions for eight large global
2 emitters

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41

42 **Abstract**

43

44 Monitoring the spatial distribution and trends in surface greenhouse gas (GHG) fluxes, as well as flux
45 attribution to natural and anthropogenic processes, is essential to track progress under the Paris Agreement and
46 to inform its Global Stocktake. This study updates earlier syntheses (Petrescu et al., 2020, 2021, 2023) and
47 provides a consolidated synthesis of CH₄ emissions using bottom-up (BU) and top-down (TD) approaches for
48 the European Union (EU) and is expanded to include seven additional countries with large anthropogenic and/or
49 natural emissions (USA, Brazil, China, India, Indonesia, Russia, and the Democratic Republic of Congo (DR
50 Congo)). Our aim is to demonstrate the use of different emission estimates to help improve national GHG
51 emission inventories for a diverse geographical range of stakeholders.

52 We use updated National GHG Inventories (NGHGs) reported by Annex I Parties under the United
53 Nations Framework Convention on Climate Change (UNFCCC) in 2023 and the latest available Biennial Update
54 Reports (BURs) reported by non-Annex I Parties. Comparing NGHGs with other approaches highlights that
55 different system boundaries are a key source of divergence. A key system boundary difference is whether both
56 anthropogenic and natural fluxes are included and, if they are, how fluxes belonging to these two sources are
57 partitioned.

58 Over the studied period, the total CH₄ emission estimates in the EU, USA, and Russia show a steady
59 decreasing trend since 1990, while for the non-Annex I emitters analyzed in this study, Brazil, China, India,
60 Indonesia, and DR Congo, CH₄ emissions have generally increased. Quantitatively, in the EU the mean of 2015-
61 2020 anthropogenic UNFCCC NGHGs (15 ± 1.8 Tg CH₄ yr⁻¹) and the mean of the BU CH₄ emissions (17.8 (16-
62 19) Tg CH₄ yr⁻¹) generally agree on the magnitude, while inversions show higher emission estimates (medians
63 of 21 (19-22) Tg CH₄ yr⁻¹ and 24 (22-25) Tg CH₄ yr⁻¹ for the three regional and six global inversions,
64 respectively), as they include natural emissions, which for the EU were quantified at 6.6 Tg CH₄ yr⁻¹ (Petrescu
65 et al., 2023). Similarly, for the other Annex I Parties in this study (**USA and Russia**), the gap between the BU
66 anthropogenic and total TD emissions is partly explained by the natural emissions.

67 For the **non-Annex I Parties**, anthropogenic CH₄ estimates from UNFCCC BURs show large
68 differences with the other global inventory-based estimates and even more with atmospheric-based ones. This
69 poses an important potential challenge to monitoring the progress of the global CH₄ pledge and the Global
70 Stocktake. Our analysis provides a useful baseline to prepare for the influx of inventories from non-Annex I
71 Parties as regular reporting starts under the Enhanced Transparency Framework of the Paris Agreement.

72 By systematically comparing the BU with TD methods, this study provides recommendations for more
73 robust comparisons of available data sources and hopes to steadily engage more Parties in using observational
74 methods to complement their UNFCCC inventories, as well as considering their natural emissions. With

75 anticipated improvements in atmospheric modeling and observations, as well as modeling of natural fluxes,
76 future development needs to resolve knowledge gaps in both BU and TD approaches and to better quantify
77 remaining uncertainty. TD methods may emerge as a powerful tool to help improve NGHGs of CH₄ emissions,
78 but further confidence is needed in the comparability and robustness of the estimates.

79 The referenced datasets related to figures are available at <https://doi.org/10.5281/zenodo.12818506>
80 (Petrescu et al., 2024).

81 **1. Introduction**

82

83 In 2021, the NOAA Global Monitoring Laboratory (GML) reported the largest annual increase in
84 atmospheric CH₄ mixing ratios since records began in 1983, at 17 parts per billion (ppb) (NOAA
85 (https://gml.noaa.gov/ccgg/trends_ch4/). In 2022, atmospheric CH₄ concentrations averaged 1912 ppb yr⁻¹, 162
86 % higher than pre-industrial levels. A similar, abnormally large growth rate of 14.8 ppb yr⁻¹ was detected from
87 total column mixing ratio measurements (XCH₄) by the Greenhouse Gases Observing Satellite (GOSAT) (Peng
88 et al., 2022).

89 CH₄ in the atmosphere has many different sources, of both natural and anthropogenic origin. The natural
90 sources of CH₄ are dominated by wetlands, while anthropogenic emissions principally come from agricultural
91 activities (livestock and rice farming), waste management (landfills and water treatment plants) and the
92 production, transportation, and use of fossil fuels. Most of the agricultural sources are distributed sources, while
93 the energy-related industrial sources of CH₄ are a mix of large point sources, of which some are detectable by
94 satellite (Jacob et al., 2022) and smaller point and distributed sources of fugitive emissions (e.g., leaks in
95 pipelines and compression stations), which are more challenging to identify (Rutherford et al., 2021; Omara et
96 al., 2022).

97 While anthropogenic CH₄ emissions from fossil fuels, agriculture, and waste can be reduced by
98 mitigation actions, increased natural emissions lead to different challenges. It has been suggested that the drivers
99 of the recent growth are most likely driven primarily by biogenic emissions (Basu et al., 2022; Lan, et al., 2021a;
100 Lan et al., 2021b; Lan et al., 2022; Nisbet et al., 2016, 2019), with smaller contributions from increased fossil
101 fuel emissions and a reduced atmospheric sink (Nisbet et al., 2023). Fluctuations in natural sources - dominated
102 by wetlands and open water bodies - were the main reasons for some of the atmospheric CH₄ anomalies observed
103 during the last decades (Rocher-Ros et al., 2023; Zhang et al., 2023; Nisbet et al., 2023; Lunt et al., 2019). Nisbet
104 et al., 2023 review recent studies, including those which quantified the observed methane growth in the last years.
105 Using a global inverse analysis of GOSAT satellite observations, increases in the range of 22-32 Tg CH₄ yr⁻¹
106 were detected between 2019 and 2020 and were attributed to biogenic sources, half of which took place in East
107 Africa (~ 15 Tg yr⁻¹), and some were observed in Canada and Alaska (4.8 Tg yr⁻¹) (Qu et al., 2022 and Basu et
108 al., 2022).

109 Chandra et al., 2021 identified a few main sectors that triggered increases and decreases in the
110 anthropogenic CH₄ emissions of different countries. The first is energy, with its fugitive emissions from the oil
111 and gas industry whose decline in emissions helped stabilize CH₄ concentration in the 1990s, before they
112 contributed to the renewed CH₄ growth since the late 2000s (increased emissions). The other major sectors that

113 drove changes in the CH₄ growth rate were agriculture (increase in emissions from enteric fermentation and
114 manure management) and waste. The increase in emissions from enteric fermentation and manure management
115 was caused primarily by increased animal numbers, and in addition by the greater intensity of ruminant farming
116 as estimated by the FAO and the emission inventories (e.g. EDGAR) which might take into account productivity
117 increases (Crippa et al., 2020; Wolf et al., 2017; FAOSTAT, 2018) while inventory emissions from waste can
118 account for up to 43 % of the linear increase in emissions for the rest of the world.

119 The contribution of CH₄ to global warming has been estimated to be about 0.5°C relative to the period
120 1850–1900 (IPCC, 2021) (Stavert et al., 2022). Methane has a relatively short perturbation lifetime (averaging
121 12.4 years, Balcombe et al., 2018) and a high global warming potential (86 and 34 for 20- and 100-years times
122 horizons respectively, compared to that of CO₂ emissions, IPCC, 2021, Table 7.15). Given the short lifetime, a
123 decline in CH₄ emissions will rapidly reduce the global warming contribution from CH₄ and help mitigate the
124 impact of climate change at decadal time scale (Cain et al., 2021). However, efforts to reduce CH₄ emissions
125 require a thorough understanding of the dominant CH₄ sources and sinks and their temporal and regional
126 distribution and trends (Stavert et al., 2022).

127 The Paris Agreement, a milestone of the UNFCCC to combat climate change and adapt to its effects,
128 entered into force on November 4, 2016. It asks each signatory to define and communicate its planned climate
129 actions, known as Nationally Determined Contributions (NDCs), and to report their progress towards their
130 targets. Next to commitments adopted by countries at COP26, the Global Methane Pledge (GMP) was launched.
131 The goal of the GMP is to cut anthropogenic CH₄ emissions by at least 30 % by 2030 with respect to 2020 levels
132 and is seen as the fastest way to reduce near-term warming and is necessary to keep a 1.5°C temperature limit
133 within reach. Achieving this goal will drive significant gains, through specific energy and agriculture defined
134 pathways including innovative actions, national targeted policies, and green climate funds to help smallholder
135 farmers (<https://www.state.gov/global-methane-pledge-from-moment-to-momentum/>). About 150 countries
136 joined this pledge and about fifty have already developed national CH₄ action plans or are doing so. As
137 agriculture and waste are the main anthropogenic sources for CH₄ emissions, a GMP Food and agriculture
138 pathway and a GMP waste pathway were launched at COP27, foreseeing actions that increase agricultural
139 productivity, while reducing emissions from dairy, food loss and waste by supporting small farmers and
140 innovation (<https://www.state.gov/global-methane-pledge-from-moment-to-momentum/>).

141 Starting in 2024, non-Annex I Parties to the UNFCCC must - given they have sufficient capacities -
142 report formal inventories under the Paris Agreement’s Enhanced Transparency Framework following the same
143 guidelines and rules as the Annex I countries (Perugini et al., 2021). Furthermore, they will undergo more
144 stringent reviews than those that previously looked at the Biennial Update Reports (BURs) and NDCs. This will
145 also allow strengthening the robustness of such comparison exercises when using independent atmospheric
146 observations in estimating trends and patterns for regional and national CH₄ emissions (IPCC, 2006). The influx
147 of new inventories will place additional demands on independent emission estimates to help improve and inform
148 National Greenhouse Gas Inventories (NGHGs), particularly in countries with low capacity.

149 With increased focus on CH₄ in climate policy, there is a demand to ensure that planned emission
150 reductions are realized. Further, as non-Annex I countries begin regular reporting of emission estimates under
151 the Enhanced Transparency Framework, there is a need to help countries improve their GHG emission estimates.

152 This has created an active field of research comparing NGHGI and independent estimates. Our analysis builds
153 on the three-year EU funded project CoCO₂, which had as main objective the building of prototype systems for
154 a European Monitoring and Verification Support capacity for anthropogenic CO₂ (and CH₄) emissions
155 (CO₂MVS). In this context, one of the results of the CoCO₂ project was the production of a Blueprint for a
156 decision support system to be used in an eventual CO₂MVS, aiming at informing and attracting attention of
157 diverse climate stakeholders on the use of the results needed beyond research. Therefore, the objectives of this
158 study reflect those of the Blueprint and focus on user engagement. It builds on dialogues with a broad community
159 of users (e.g. scientists, inventory agencies, policy makers), considering their opinions and needs when it comes
160 to comparisons between independent approaches. Furthermore, this study expands beyond the EU to include
161 seven countries that have large anthropogenic and/or natural CH₄ emissions (USA, Brazil, China, India,
162 Indonesia, Russia and the Democratic Rep. of Congo). It examines both Annex I (EU, USA and Russia) and non-
163 Annex I estimates from observation-based BU process-based models and inversions-based TD approaches (using
164 satellite observations) by identifying and explaining differences with official inventory reports submitted by
165 parties to the UNFCCC. The seven countries were chosen based on location and the importance / magnitude of
166 their anthropogenic and natural emissions. By using multiple methodologies, uncertainties can be estimated by
167 looking at the range in both emissions and trends.

168

169 **2. Methods and data**

170 In this work we focus on comparing BU and TD emission estimates. The ‘reconciliation process’
171 described in this work is the action of making one dataset comparable with another to assess their consistency.
172 In this respect, we attempt to obtain consistent results from both BU and TD estimates, through harmonization
173 of the results, concepts and definitions. After the reconciliation process, the estimates do not necessarily agree,
174 representing uncertainties in the different methods and datasets. We now describe the key data and methods used
175 in our analysis.

176 **2.1. Verification practices in official UNFCCC NGHGIs**

177

178 Quality assurance/quality control (QA/QC) is a key component of NGHGIs development. Verification
179 is an additional step and refers specifically to methods that are external to the inventory and apply independent
180 data. There are two main methods of verification: 1) independent inventory-based estimates, 2) observation-
181 based emission estimates.

182 A challenge with comparisons against *independent inventory-based estimates* is that none of them is
183 truly independent as they may rely on, for example, the same activity data reported by a country (Andrew 2020).
184 Experience has shown that performing detailed comparisons (Petrescu et al., 2021, 2023, Lauerwald et al., 2024)
185 can help clarify differences in system boundaries or even identify errors (Andrew 2020). Improving independent
186 emission inventories also has value, as these are often used in global studies where common methods across all
187 countries are desired.

188 *Observation-based estimates* use observations of atmospheric concentrations and prior fluxes that are
189 then coupled to a transport model. These methods are more complex and computationally expensive and can
190 make use of both direct observations and emission inventories.

191 The 2019 refinement of the 2006 IPCC guidelines highlighted notable advances in the application of
192 inverse models of atmospheric transport for estimating emissions at the national scale. Building on this progress,
193 they extend the guidance on the use of atmospheric measurements for verification (IPCC, 2019). There are
194 several countries that currently use atmospheric measurements for verification of parts of their inventories.
195 Australia (Luhar et al 2020, AUS NIR, 2023) and New Zealand (Geddes et al., 2021) have estimated regional
196 CH₄ emissions to help better understand the methods and their potential. Germany performs various cross-
197 validation checks with available data (German NIR, 2023), some of which are based on observations. The UK
198 and Switzerland (Annex 6 CHE NIR, 2023) have developed more comprehensive methods based on inversion
199 modeling, covering several GHGs in addition to CH₄. Building on modeling experience, the country reporting
200 confirms that most potential lies in using observations to verify fluorinated gases (Annex 6 UK NIR, 2023), but
201 the large uncertainty in CH₄ emissions gives the potential for verification if a sufficient observation network is
202 used in inversion modeling (Bergamaschi et al., 2018, Thompson et al., 2014).

203 While inversions of CH₄ fluxes are associated with significant uncertainty, so are NGHGI estimates of
204 anthropogenic CH₄ emissions. Furthermore, inversions can provide information on subannual and subnational
205 variations in time and space that may indicate differences in source sector emission estimates. In geographic
206 areas with sufficiently dense ground-based observation networks, the inversions will have more value.

207

208 **2.2. Anthropogenic CH₄ emissions from the NGHGIs**

209

210 Annex I countries report their annual GHG emissions to the UNFCCC in the so-called Common
211 Reporting Format (CRFs) data tables and National Inventory Reports (NIRs). Here, anthropogenic CH₄
212 emissions from the five UNFCCC sectors, incl. Land Use, Land Use Change and Forestry (LULUCF) are
213 grouped together. As part of the LULUCF sector, we also have the CH₄ emissions from wetlands, which
214 according to the IPCC guidelines are defined as managed "where the water table is artificially changed (i.e.
215 lowered or raised) or those created through human activity (e.g. damming a river) and that do not fall into Forest
216 Land, Cropland, or Grassland categories (IPCC, 2014)". Reporting CH₄ emissions from managed wetlands is not
217 mandatory, but if done, parties are encouraged to make use of the 2013 IPCC Wetlands supplement (IPCC,
218 2014). In the EU, if Member States report these emissions, they report not only restored (rewetted) wetlands but
219 also emissions from drained organic and mineral soils (e.g. peatlands, ditches, etc.). These are not large by
220 magnitude but are large by area in the Nordic countries. According to NGHGI data, in 2021, managed wetlands
221 in the EU, for which emissions were reported under the LULUCF (CRF Table 4(II) and Summary 1.As2
222 accessible for each EU country), summed up to 0.21 Tg CH₄ yr⁻¹, in comparison to total emissions of ~15 Tg
223 CH₄ yr⁻¹. Furthermore, the NGHGIs do not include any lateral fluxes from inland waters but do include biomass
224 burning anthropogenic emissions reported under the LULUCF sector.

225 The presented uncertainties in the CH₄ emission levels of the individual countries and the EU are not
226 always reported in a complete and harmonized format, and therefore were calculated applying gap-filling and
227 harmonization procedures that are used to compile the EU GHG inventory reported under UNFCCC (EU NIR,
228 2023) (see SI and Appendix A1.1 in Petrescu et al., 2023). The EU uncertainty analysis reported in the bloc's
229 National Inventory Report (NIR) is based on country-level, Approach 1 uncertainty estimates (IPCC, 2006, Vol.

230 1, Chap. 3) that are reported by EU Member States, previously under Article 7(1)(p) of Regulation (EU) 525/2013
231 and since 2023 under Article 26(3) and Annex V(Part 1)(m) of the Governance Regulation (EU) 2018/1999.

232 Non-Annex I countries report their updated NGHGs to the UNFCCC, including a national inventory
233 report and information on mitigation actions, needs and support received in Biennial Update Reports (BURs). In
234 this study, Brazil, China, Indonesia, India and the Democratic Rep. of Congo (DR Congo) were investigated. For
235 Brazil, information from its fourth BUR (Brazil, 2020) was used, giving both total and sectoral split emission
236 values for years 1994, 2000, 2010, 2012, 2015 and 2016. For China, information from its second BUR Tables 2-
237 10, 2-13, 2-14, 2-15, and 2-16 was used (China, 2019). The information was available for both total and sectoral
238 split emission values for 1994, 2005, 2010 and 2014. Uncertainties for 2014 are available in Table 2-12.
239 Indonesia submitted its third BUR in 2021 (Indonesia, 2021). Indonesian total sectoral CH₄ emissions time series
240 as reported by the 2nd BUR (2001-2016) were revised in the 3rd BUR (2000 and 2019, Table 2). For 2017 and
241 2018, only the agricultural CH₄ emissions were detailed by the 3rd BUR (Fig. 2-24), but the total emissions for
242 these two years were not reported. Uncertainty for 2019 activity data and emission factors (EFs) are the same as
243 reported in the 2nd BUR (2018). The overall uncertainty of Indonesia's National GHG inventory with agriculture
244 and LULUCF (including peat fires) for 2000 and 2019 were approximately 20.0% and 19.9%, respectively. A
245 much smaller uncertainty, 10.4 % for 2000 and 13.8 % for 2019, occurred when the forestry and land use sector
246 (including forest fires), was excluded from the analysis, pointing to the high uncertainty of emissions from forest
247 fires in Indonesia. The DR Congo submitted its first BUR in 2022, and we used timeseries reported for 2000-
248 2018 (Table 12 Congo, 2022). India has submitted three BURs and information on sectoral CH₄ emissions are
249 in each of them only for one year. We compiled information for 2010 from the first BUR (India, 2016), for 2014
250 from the second BUR (India, 2018) and for 2016 from the third and latest BUR (India, 2021).

251 **2.3. Other CH₄ data sources and estimation approaches**

252

253 The CH₄ emissions in the EU and non-Annex I countries used in the atmospheric inversions and
254 anthropogenic and natural emissions estimates from various BU approaches and inventories (i.e., UNFCCC
255 CRFs and BURs) covering specific products, sectors and activities are summarized in Table 1. The data and the
256 detailed description of most products (Tables S1 and S2, Supplementary Information) span the period from 1990
257 to 2021, with some of the data only available for shorter periods. The estimates are available both from peer-
258 reviewed literature and from unpublished research results from the VERIFY and CoCO₂ projects (Supplementary
259 Information, SI) and in this work they are compared with NGHGs reported in 2023 (time series for all (Annex
260 I) or some years (non-Annex I) of the 1990-2021 period). The BU anthropogenic sources are from UNFCCC
261 NGHGs and three global inventory datasets/models: EDGARv7.0, FAOSTAT/PRIMAP-hist 2.4, GAINS and
262 the TNO_CoCO₂_PED18-21 priors emissions datasets for 2010-2018 and 2021. In this synthesis, FAOSTAT
263 (Tubiello et al., 2022; FAO, 2023) data includes estimates for all economic sectors: Energy, Industrial Processes
264 and Products Use (IPPU), Waste and Other, which are sourced from the PRIMAP-hist v2.4 dataset (Gütschow
265 et al., 2022) to build emissions indicators on agrifood systems and on the entire economy. Emission totals from
266 the agrifood domain are computed following the Tier 1 methods of the Intergovernmental Panel on Climate
267 Change (IPCC) Guidelines for NGHGs. Agrifood systems emissions in FAOSTAT are largely based on FAO
268 crop, livestock and land-use statistics (Tubiello et al., 2022; FAO, 2023). They are complemented with activity

269 data from the UN Statistics Division (UNSD), the International Energy Agency (IEA) and with geospatial
 270 information on drained organic soils and biomass fires (Conchedda and Tubiello, 2020; Prosperi et al., 2020).
 271 The TNO CoCO₂ PED datasets for 2010-2018 and 2021 are based on the UNFCCC reported data in 2020 and
 272 2023, respectively for the EU27 countries, on the DACCIWAv.2 dataset (Keita et al., 2021) for the African
 273 continent and the CAMS-GLOB-ANT v5.3 dataset (Soulie et al., 2024) for all other countries (no data for
 274 COD). The methodology is detailed in the CoCO₂ deliverables D2.1 Prior Emission Dataset (PED) 2016 | CoCO₂:
 275 Prototype system for a Copernicus CO₂ service (coco2-project.eu) and D2.2 Prior Emissions data 2021 | CoCO₂:
 276 Prototype system for a Copernicus CO₂ service (coco2-project.eu).

277 The analysis focuses on both total and sectoral or partitioned information from both BU and TD
 278 estimates. As detailed in Table 1, not all inversions distinguish between sources but in the following sections we
 279 discuss comparability between BU and TD for both total and partitioned results.

280 *Table 1: Sectors included in this study and data sources providing estimates for these sectors. CAMS stands for*
 281 *Copernicus Atmosphere Monitoring Service. References to data products are found in Table 2 Petrescu et al.,*
 282 *2023 and Table S1 and S2, SI.*

Anthropogenic (BU) ¹ CH ₄	Natural (BU) ² CH ₄	Regional TD CH ₄	Global TD CH ₄
1. Energy: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT/PRIMAP-hist 2.4, TNO_CoCO ₂ _PED18-21	Wetlands EU: JSBACH-HIMMELI Global: LPJ-GUESS	No partitions – total emissions FLEXkF_v2023 CIF-FLEXPARTv10.4	Totals and partitioned emissions: MIROC4-ACTM (control and OH varying runs)
2. Industrial Products and Products in Use (IPPU): UNFCCC NGHGI (CRFs and BURs), EDGAR v7.0, FAOSTAT/PRIMAP-hist 2.4, TNO_CoCO ₂ _PED18-21	Peatlands, mineral soils: EU: JSBACH-HIMMELI Global: LPJ-GUESS	CIF-CHIMERE	CAMSv21r1 (NOAA and NOAA_GOSAT runs)
3. Agriculture: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT, TNO_CoCO ₂ _PED18-21	Inland waters fluxes EU: lakes, rivers and reservoirs (RECCAP2) Global: lakes and reservoirs ORNL DAAC		TM5-4DVAR (TROPOMI)
4. LULUCF: UNFCCC NGHGI (CRFs and BURs) and FAOSTAT	Geological fluxes updated activity (see SI)		CTE-CH ₄ (GCP2021)
5. Waste: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT/PRIMAP-hist			CEOS (GOSAT)
			GEOS-Chem CTM (TROPOMI) for USA only

¹ For consistency with the NGHGI, here we refer to the five reporting sectors as defined by the UNFCCC and the Paris Agreement decision (18/CMP.1), the IPCC Guidelines (IPCC, 2006), and their Refinement (IPCC, 2019a), with the only exception that the latest IPCC Refinement groups together Agriculture and LULUCF sectors in one sector (Agriculture, Forestry and Other land Use - AFOLU).

² The term **natural** refers here to unmanaged natural CH₄ emissions (peatlands, mineral soils, geological, inland waters and biomass burning) not reported under the anthropogenic UNFCCC LULUCF sector.

2.4, TNO_CoCO ₂ _PED18-21	Biomass burning (GFEDv4.1s)		
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283 note: Not all models have a version id. Those that have, are used in previous syntheses (Petrescu et al., 2021 and 2023).

284 We define natural sources as all sources which do not belong to the anthropogenic partition: wetlands,
 285 geological, termites, ocean, inland waters, soils and biomass burning (Table 2). The BU natural components for
 286 the EU were computed as the sum of the VERIFY products (biomass burning, inland waters and undisturbed
 287 peatlands plus mineral soils (as described in Petrescu et al., 2021 and 2023) and geological emissions (Etiope et
 288 al. 2019 updated for the VERIFY project). For the seven non-EU emitters, the BU natural fluxes are the sum of
 289 wetland emissions (LPJ-GUESS), lake and reservoir emissions (ORNL DAAC), biomass burning emissions
 290 (GFED4.1s) and geological emissions (updated activity, SI). The TD natural global estimates were calculated as
 291 the sum of all natural partitions reported by the inversions. Adjustments were made to have a consistent
 292 comparison between partitions, adding the missing ones from the BU estimates (Table 4). The error bar on the
 293 TD natural represents the range of the min/max between inversion estimates.

294 The total regional TD estimates (for EU) and their uncertainties were calculated as the mean and
 295 min/max range between FLEXkF_v2023, CIF-FLEXPART and CIF-CHIMERE inversions (see Priors table in
 296 Petrescu et al., 2024). For the USA, we considered the optimized emissions from the GEOS-Chem CTM (based
 297 on TROPOMI data for 2019) from Nesser et al. (2023), with the range from the eight members of the inversion
 298 ensemble shown as uncertainty (Table 2 in Nesser et al., 2023).

299 For all countries, the total global TD inversion estimates (time series) and uncertainties were calculated
 300 over the period 2015-2021 using the mean and min/max between CTE-GCP2021, MIROC4-ACTM both runs,
 301 CAMS v21r1 (both runs), and TM5-4DVAR (TROPOMI based). CEOS (GOSAT) provided an estimate only
 302 for 2019.

303 The units used in this paper are metric tons (t) [$1\text{kt} = 10^9\text{ g}$; $1\text{Mt (Tg)} = 10^{12}\text{ g}$] of CH₄. The referenced
 304 data for replicability purposes are available for download at <https://doi.org/10.5281/zenodo.12818506> (Petrescu
 305 et al., 2024). Upon request, the computer code for plotting figures in the same style and layout can be provided.
 306 Throughout the paper and mostly for the complex figures, the following ISO3 country codes are used: USA
 307 (United States of America), BRA (Brazil), CHN (China), IDN (Indonesia), RUS (Russia), COD (DR Congo) and
 308 IND (India). Next to these we also refer to CHE (Switzerland) and AUS (Australia). The European Union consists
 309 of 27 Member States, excludes the United Kingdom (UK) and is further abbreviated as EU. All abbreviations
 310 are summarized in the SI, Table S5.

311 **3. Results**

312

313 **3.1. NGHGI official reported estimates (UNFCCC)**

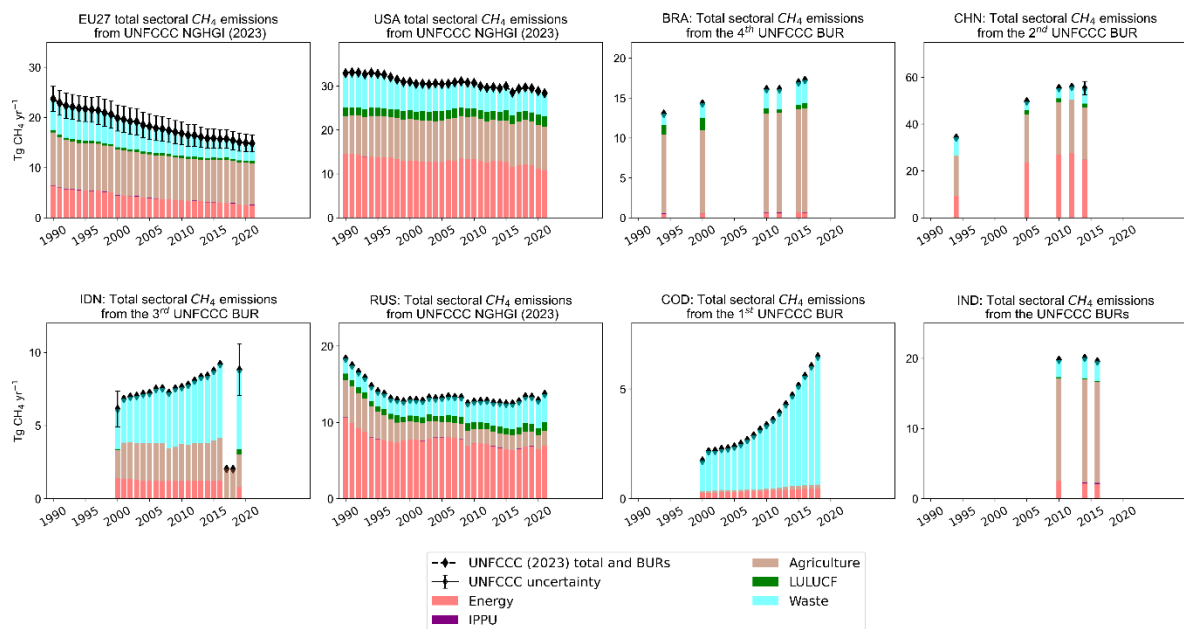
314

315 Figure 1 presents anthropogenic CH₄ emissions reported to the UNFCCC in 2023 from the NGHGI
 316 CRFs (EU, USA and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR
 317 Congo (1st in 2022) and India (all three BURs). The following section provides additional details for all the
 318 countries.

319 For the *EU*, the total anthropogenic CH₄ emissions in 2021 amount to 14.8 ± 1.8 Tg CH₄ yr⁻¹ and
 320 represent 12.8 % of the total EU greenhouse gas emissions (in CO₂ equivalents, GWP 100 years, IPCC AR5³).
 321 CH₄ emissions are predominantly from agriculture (Figure 1, brown), which accounted for 8.3 Tg CH₄ yr⁻¹ ± 0.8
 322 Tg CH₄ yr⁻¹ or 56 % of the total EU CH₄ emissions in 2021 (incl. LULUCF). Anthropogenic CH₄ emissions from
 323 the LULUCF sector are very small for the EU: 0.5 Tg CH₄ yr⁻¹ or 3 % in 2021, including emissions from biomass
 324 burning. The EU data from Figure 1 shows steadily decreasing trends for all sectors with respect to the 1990
 325 levels. The reduction in total CH₄ emissions in 2021 with respect to 1990 is 8.9 Tg CH₄ yr⁻¹ (37 %) at an average
 326 yearly rate of -1%.

327 In 2021, the *USA* reported anthropogenic CH₄ emissions of 28.3 Tg and, compared to 1990, the reported
 328 USA CH₄ emissions show a small decrease of 4.6 Tg CH₄ yr⁻¹, more pronounced for the last two years (2020-
 329 2021), with an average reduction rate of -0.5 % per year (Fig. 1 black dotted line). In the USA, the largest share
 330 of emissions comes from the energy sector (38%), and next to IPPU and waste, had the highest reductions since
 331 1990 (42%, 34% and 26%, respectively). Emissions from agriculture (35%, the second largest sector) and
 332 LULUCF increased 16 % and 23%, respectively. CH₄ emissions have been slowly declining since 1990 but had
 333 a notable decrease of 1.5 Tg CH₄ yr⁻¹ in 2016 compared to 2015, before increasing again and had a second
 334 decreasing trend in 2020 and 2021, possibly due to the COVID pandemic. Overall, reported data indicates that
 335 reductions in the USA CH₄ emissions have declined more slowly than that in the EU. The EU also has much
 336 lower CH₄ emissions than the US on a per capita basis (Figure 2).

337



338

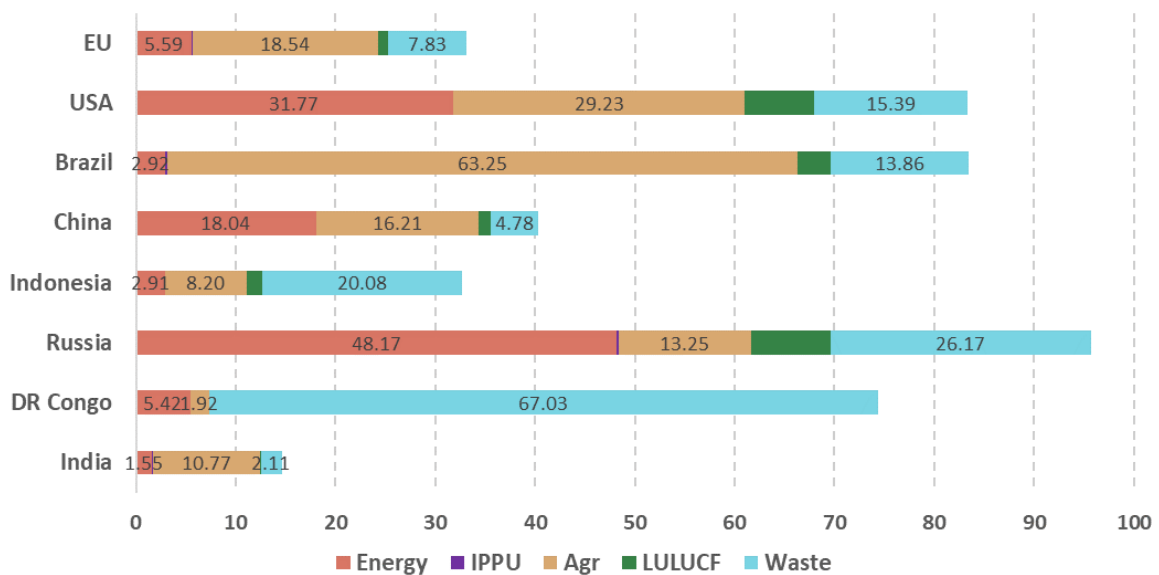
339 *Figure 1: Total and sectoral CH₄ emissions (incl. LULUCF) from the UNFCCC NGHGI (2023) CRFs (EU, USA*
 340 *and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR Congo (1st in 2022)*
 341 *and India (all three BURs: 2016, 2018 and 2021). The relative error on the UNFCCC value represents the*

³ IPCC AR4 GWP 100 values are still used by the Member States in their NGHGI reporting to the UNFCCC.

342 *NGHGI (2023) reported uncertainties computed with the error propagation method (95% confidence interval)*
 343 *and gap-filled to provide respective estimates for each year. Information on Indonesian sectoral CH₄ emissions*
 344 *in 2017 and 2018 are only available for Agriculture. The overall uncertainty of Indonesia's National GHG*
 345 *inventory with AFOLU (including peat fires) for 2000 and 2019 were approximately 20.0% and 19.9%,*
 346 *respectively. In 2014, China reported uncertainty as well (min 5.2 % and max 5.3 %).*

347 The trend in total anthropogenic CH₄ emissions in *Brazil* is strongly increasing, with 32.5 % more
 348 emissions in 2016 compared to 1994. Given that the Brazilian BUR inventory does not include data between
 349 2001 and 2019, it is difficult to discuss the yearly growth rates. We can only note that the agricultural sector (76
 350 % of the total) was the main driver of the growth, followed by the waste sector (16 % of the total). There are
 351 only small CH₄ emissions from the Energy sector (some oil and gas activities). The Brazilian agricultural CH₄
 352 emissions are the highest of the eight countries on a per capita basis (see Figure 2).

353



354

355 *Figure 2: Per capita emissions (kg) intensity per sector based (IPCC, 2006) on reported emissions and*
 356 *population data from the last reported NIRs as follows: 2021 for the EU, USA and Russia, 2016 for Brazil and*
 357 *India, 2014 for China, 2019 for Indonesia and 2018 for DR Congo.*

358 *China's* total anthropogenic CH₄ emissions are much larger than the emissions reported by many
 359 developed countries or the entire EU (see Figure 1), but on a per capita basis it is only fifth of the eight countries
 360 considered (Figure 2). China's CH₄ emissions have grown 61 % from 1994 to 2014, when they reached 55 Tg
 361 CH₄. The highest contributions to China's CH₄ emissions in 2014 are from energy (45 %) and agriculture (40
 362 %). The rapid growth of China's coal demand has important implications for CH₄ emissions from coal mines
 363 (Gao et al., 2020). The energy and agriculture sectors have respectively increased by 166 % and 30 % in 2014
 364 compared to 1994.

365 *Indonesia's* 3rd BUR data (2000 and 2019) show increasing trends in total anthropogenic CH₄ emissions.
366 The time series 2001-2006 derives from the 2nd BUR submitted in 2018. In 2019, Indonesian CH₄ emissions had
367 increased by +44 % compared to 2000, corresponding to 2.6 Tg CH₄ yr⁻¹, an average yearly increase of 3 %, and
368 the sector which contributes the most to this increase is the waste sector, which nearly doubled its emissions in
369 2019 compared to 2000. According to Qonitan et al., 2021, the major solid waste source in Indonesia is the
370 household sector, which contributed 44-75% to total waste generated. The composition of municipal waste
371 consists of 44% food waste, 16% paper, and 14% plastics. CH₄ emissions from the other sectors remained nearly
372 constant. For 2017 and 2018 the Indonesian 3rd BUR does not report total emissions other than agricultural
373 emissions, which were taken from the report Figure 2-24. The last data point (2019) shows lower total emissions
374 because it belongs to the revised versions of the (3rd) BUR while the previous data points 2000-2016 belong to
375 the 2nd BUR.

376 *Russia's* anthropogenic CH₄ emissions have decreased by -25 % from 1990 to 2021, but most of this
377 decrease happened during the dissolution of the Soviet Union. Since 2000, CH₄ emissions have increased
378 slightly, but remain lower than pre-2000 levels. The decline seen between 1990 and 2000 is primarily due to the
379 agricultural sector (-52 %) and energy (-27 %). At the same time, the waste sector started to increase its emissions
380 (6 %). Between 2001 and 2021, the CH₄ emissions from the agriculture and energy sectors continue to decrease
381 (by 17 % and 11 %, respectively), while the emissions from the waste sector register an additional 76 % increase.
382 IPPU emissions increased by 85 % but remain negligible compared to other sectors. Since the 2000s, LULUCF
383 emissions have also increased, by 53 %.

384 For its first BUR, *DR Congo* submitted emissions from energy, AFOLU (agriculture plus LULUCF)
385 and waste for 2000-2018. Since 2000, the DR Congo total anthropogenic CH₄ emissions have increased by a
386 factor of four. Most of the CH₄ emissions are reported for the waste sector, and account for 90 % of the total
387 emissions. The high percentage of waste emissions in DR Congo is also seen in the per capita emissions (Figure
388 2). Assè-Wassa Sama and Berenger (2023) confirm that between 2000 and 2021, CH₄ emissions, which in 2021
389 represent in DR Congo ~97% of total waste generated emissions, grew at a rate of 4 % yr⁻¹, compared with 2.7
390 % yr⁻¹ for total emissions. This increase was driven by the increase in emissions caused by solid waste disposal
391 (+6.2 %). The CH₄ waste emissions come mainly from the treatment and discharge of wastewater (69 % in 2021,
392 compared with 80 % in 2000), followed by the elimination of solid waste (31 % in 2021, compared with 20 %
393 in 2000). The weight of emissions caused by the elimination of solid waste in the sector's total emissions has
394 nevertheless increased by 11 percentage points between 2000 and 2021 (Assè-Wassa Sama and Berenger, 2023).

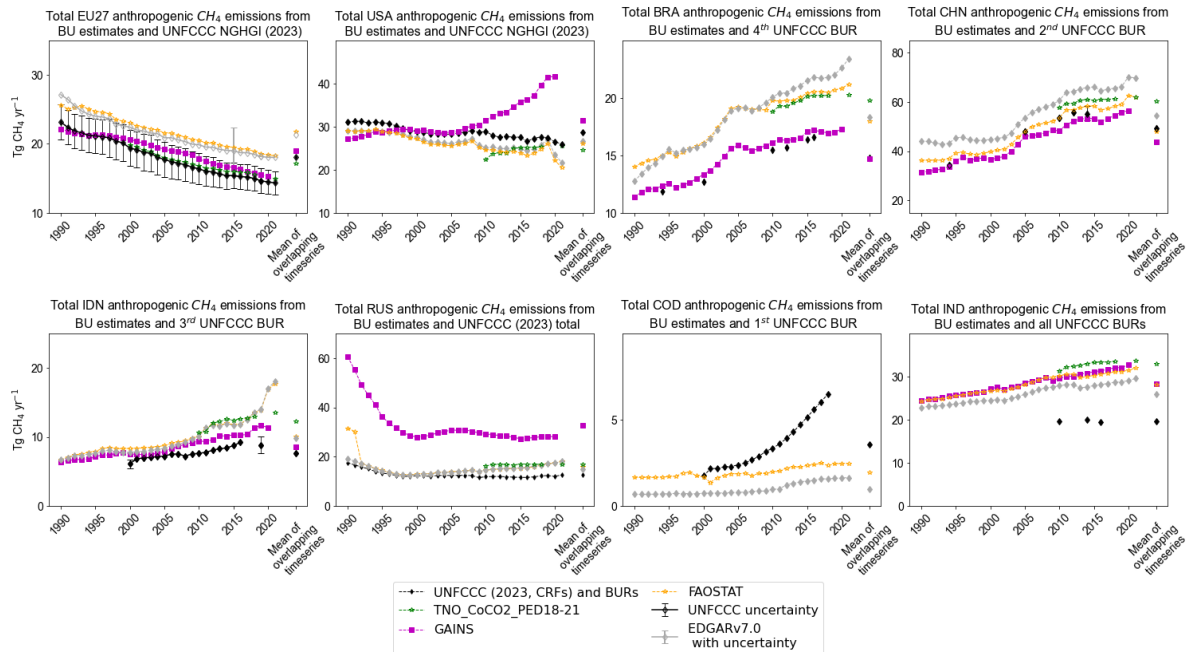
395 Each of *India's* BURs provide detailed information on sectoral CH₄ emissions only for one year. Most
396 of the emissions in India belong to the agriculture sector, amounting to almost 15 Tg CH₄ yr⁻¹ (in 2016),
397 representing 74 % of the total anthropogenic emissions. However, with only three years of reported data
398 available, no clear or notable trend is observed.

399 **3.2. NGHGI compared to other bottom-up estimates**

400

401 Figure 3 shows UNFCCC (CRFs and BURs) estimates from EU and seven non-EU countries compared
402 to global bottom-up inventories. The EU and USA show decreasing trends in emissions from all data sets (except

403 for GAINS in the USA), while all the other countries show increasing trends in all datasets. The match between
 404 UNFCCC reported emissions and all other data sources is satisfactory, with a few notable exceptions. To note
 405 that the TNO_CoCO₂_PED18-21 and the FAOSTAT/PRIMAP-hist have very similar trends for all countries
 406 except EU, as both FAOSTAT/PRIMAP-hist and CAMS-GLOB-ANT (used in TNO_CoCO₂_PED18-21 dataset
 407 for countries outside EU and Africa) are partly based on EDGAR.
 408



409

410

411 *Figure 3: Total anthropogenic CH₄ emissions (excl. LULUCF) from bottom-up (BU) inventories as: UNFCCC*
 412 *NGHGs (2023) of CRFs (EU, USA and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia*
 413 *(3rd in 2021), DR Congo (1st in 2022), India (all three BURs:2016, 2018 and 2021) and three other global*
 414 *datasets: EDGARv7.0, GAINS (no IPPU), FAOSTAT (PRIMAP-hist based, except for AFOLU) and*
 415 *TNO_CoCO₂_PED18-21. The relative error on the UNFCCC value represents the NGHGI (2023) reported*
 416 *uncertainties computed with the error propagation method (95% confidence interval) and gap-filled to provide*
 417 *respective estimates for each year. China and Indonesia report uncertainties, for 2014 and 2000 and 2019*
 418 *respectively (BUR). Total COD UNFCCC BUR emissions do not include IPPU. The EDGARv7.0 uncertainty*
 419 *is only for 2015 and was calculated according to Solazzo et al., 2021 for EDGARv5.0. The mean of overlapping*
 420 *time series was calculated for 1990-last available year as following: 2021 for UNFCCC NGHGI (2023),*
 421 *EDGARv7.0, FAOSTAT and TNO_CoCO₂_PED18-21 and 2020 for GAINS.*

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For the EU, the difference between the UNFCCC NGHGI 1990-2020 average and the other three data
 sets, is less than 5 %. TNO_CoCO₂_PED18-21 data for EU27 are gap filled data based on the UNFCCC country
 reported numbers, therefore it follows closely the trend of the NGHGI data. The difference between EDGAR
 and FAOSTAT and the other datasets originates from country-specific emission factors being used in the other
 three inventories for the EU, especially for fossil fuel production. As previously discussed, the inventory-based
 data sources are consistent with each other for capturing recent CH₄ emission reductions, but they are not

428 independent because they use similar methodology with different versions of the same activity data (AD)
429 (Petrescu et al., 2020, Figure 4).

430 For the *USA*, GAINS reports high emissions after 2010, with strong growth. This divergence is largely
431 found in the Energy sector, resulting from the EFs used for conventional gas production as well as for
432 unconventional shale gas extraction, which has increased rapidly since 2006 due to the development of hydraulic
433 fracturing technology (Supplementary Figure S6-1 in Höglund-Isaksson et al., 2020). The high share of
434 emissions from unconventional shale gas can be explained by the GAINS EFs which, in the absence of published
435 factors, are derived from the residual emissions after having subtracted estimated emissions for oil production
436 and conventional gas production from the total upstream emission estimated by Alvarez et al., (2018, Table 1)
437 As Alvarez et al. 2018 do not specify EFs by type of gas produced, GAINsv4 splits it based on activity data from
438 other references, International Energy Agency-World Energy Outlook (IEA-WEO, 2018) and Energy
439 Information Administration (EIA, 2019). On the other hand, the NGHGI EF seems to be too low, and this is
440 reflected by the low oil and gas emissions reported by the United States Environmental Protection Agency
441 (USEPA 2017) for 2015, compared to Alvarez et al., 2018 (Supplementary Table S6-3, Höglund-Isaksson et al.,
442 2020). For the USA, total gas production increased by 47 % between 2006 and 2017. Revisions for the
443 agricultural livestock emissions concern updates of AD and reported EFs to statistics from FAOSTAT (2018)
444 and CRFs (UNFCCC 2016; 2018), and a review of available technical abatement options for CH₄.

445 For *Brazil*, UNFCCC and GAINS report emissions of similar magnitudes and trends. The EDGARv7.0
446 and FAOSTAT report on average around 23 % more emissions for the 1990-2021 period, but closely follow the
447 NGHGI trends. The similarity between trends could be explained by the use of the same EFs following Tier-1
448 IPCC 2006 Guidelines and UNFCCC NIRs (Janssens-Maenhout et al., 2019), while the higher emissions could
449 appear when using different AD information.

450 For *China* the inventory estimates agree reasonably with the BUR reported data, with EDGARv7.0
451 showing the highest estimates. According to both GAINS and EDGARv7.0, the primary drivers for growth in
452 Chinese CH₄ emissions are due to a mix of sources, mainly from the IPCC 2006 sector 1.B.1, fugitive emissions
453 from solid fuels activity linked to increased coal mining.

454 In *Indonesia* the three global datasets agree well up until 2010. From 2010, the third-party datasets show
455 a continued increase in emissions, while the UNFCCC BUR emissions suggest a decline. EDGARv7.0 reports a
456 large increase in emissions from fugitive emissions from solid fuels (coal mining) (IPCC 2006, sector 1.B.1.) at
457 an increased average rate of 19 % per year and has increased by a factor of 152 until 2021 compared to 1990
458 (Figure 3).

459 For *Russia*, GAINS emissions are much higher than NGHGI and the other two data sets due to the
460 revisions of the assumptions on the average composition of the associated gas generated from oil production
461 based on information provided in Huang et al. (2015). The higher emissions in GAINsv4 might be caused by a
462 greater source from venting of associated gas instead of flaring. GAINsv4 estimates a decline in global CH₄
463 emissions in the first half of the 1990s, primarily a consequence of the dissolution of the Soviet Union and the
464 associated general decline in production levels in agriculture and fossil fuels (see regional emission illustrations
465 in figures S2-1 of the SI). In addition, as described by Evans and Roshchanka (2014) and assumed in Höglund-

466 Isaksson (2017), venting of associated petroleum gas declined significantly in Russia due to an increase in flaring.
467 It is unclear why this happened, but a possible explanation could be that the privatization of oil production in
468 this period meant that the new private owners were less willing to take the security risks of venting and invested
469 in flaring devices to avoid potential production disruptions. This hypothesis is however yet to be confirmed
470 (Höglund-Isaksson et al., 2020). FAOSTAT data for the Russian Federation starts in 1992, but the former USSR
471 statistics were used prior to 1992 without adjustments and this is the cause of the 1990 and 1991 outliers in time
472 series. The slightly increasing trend observed in EDGARv7.0 and FAOSTAT are set by emissions from the
473 Energy sector.

474 For *DR Congo* estimates from GAINS are not available because they only report aggregated emissions
475 from a few African regions. Both FAOSTAT (PRIMAP based) and EDGARv7.0 estimates show similar slowly
476 increasing trends, potentially indicating the use of similar prior statistics (EFs). For non-AFOLU sectors the
477 PRIMAP-hist third party data priority scenario used in FAOSTAT also uses EDGAR data as an input data source
478 explaining similarities in these sectors. On the other hand, UNFCCC BUR data reports a strong increase in
479 emissions, which is due to a rapid growth of CH₄ emissions from the Waste sector, by a factor of four until 2018
480 compared to 2000. This increase happened at an average yearly rate of +8 %, with an initial sharp increase of
481 +30 % between 2000 and 2001. As previously discussed, (section 3.1.) we believe that DR Congo BUR reported
482 waste emissions are improbable and further investigation is needed.

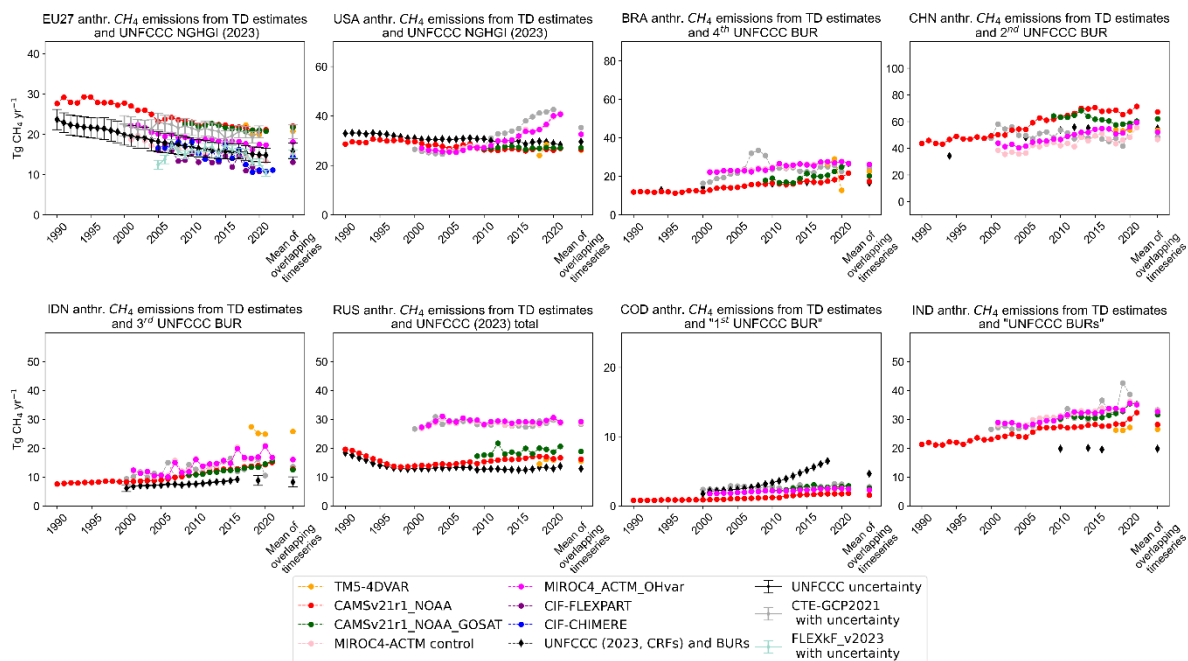
483 For *India*, all bottom-up global inventories show similar trends and magnitudes of anthropogenic CH₄
484 emissions. The emissions of CH₄ averaged across EDGARv7.0, GAINS and FAOSTAT are 67% (2010), 68 %
485 (2014) and 65 % (2016) higher than the Indian BURs. All three BU inventories show an averaged steady increase
486 of 1 % yr⁻¹ between 1990-2020.

487 **3.3. NGHGs compared to TD atmospheric-based CH₄ estimates**

488

489 In Figure 4 we compare the reported TD anthropogenic estimates, after removing natural emissions,
490 with the UNFCCC official reported emissions for the EU and the seven non-EU emitters. The mean column on
491 the right of each chart represents the mean of the overlapping time series (2009-last available year, except for
492 TROPOMI, which was available only for 2018-2020). For the EU, the three regional inversions report total
493 emissions because they do not separate anthropogenic from natural emissions. Therefore, we subtracted from the
494 total the natural emissions as calculated in Petrescu et al., 2023 which amount to 6.6 Tg CH₄ yr⁻¹ and are the sum
495 of biomass burning, wetlands, geological and inland water CH₄ emissions. For the global inversions, the
496 anthropogenic estimates were calculated by subtracting from the total fluxes the reported natural partitions as
497 follows: for the two CAMS inversions and TM5-4DVAR (TROPOMI based) the sum of biomass burning and
498 wetlands, for MIROC4-ACTM runs the natural is the represented by the sum of the biomass burning, geologic,
499 ocean, termites, soils and wetlands, for CTE-GCP2021 the sum of the biologic (wetlands + soils) and other
500 (ocean, termites, geological). Because not all inversions report the same partitions, we consider this a coarse
501 comparison, and we detail the harmonization of the natural emissions in the next section (Table 4 and Figure 7).
502 For China, the last BUR is available for 2014, and therefore we used that value.

503
504



506

507 *Figure 4: Anthropogenic CH₄ emissions from UNFCCC NGHGI (2023) CRFs (EU, USA and Russia) and BURs*
 508 *(Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR Congo (1st in 2022), India (all three BURs:*
 509 *2016, 2018 and 2021) and TD estimates as following: for EU anthropogenic regional inversions*
 510 *(FLEXkF_v2023, CIF-FLEXPART and CIF-CHIMERE) and for global inversions anthropogenic estimates*
 511 *calculated as total TD minus natural TD reported partitions (TM5-4DVAR, CAMSv21r1_NOAA,*
 512 *CAMSv21r1_NOAA_GOSAT, CTE-GCP2021 and MIROC4-ACTM both runs) products. The relative error on the*
 513 *UNFCCC value represents the NGHGI (2023) reported uncertainties computed with the error propagation*
 514 *method (95% confidence interval) and gap-filled to provide respective estimates for each year. China reports*
 515 *uncertainties for 2014 (min 5.2 %, max 5.3 %) and Indonesia reports for 2000 and 2019, 20 % and 19.9 %*
 516 *respectively. Total COD UNFCCC BUR emissions do not include IPPU. The last available years are CIF-*
 517 *CHIMERE (2022), TM5-4DVAR, CIF-FLEXPART and CTE-GCP2021 (2020) and FLEXkF_v2023, MIROC4-*
 518 *ACTM OHvar and control, UNFCCC CRFs, and CAMSv21r1_NOAA and NOAA_GOSAT runs (2021). The*
 519 *mean of overlapping time series was calculated for 2009-2021, except for TM5-4DVAR (2018-2020).*

520

521 In the EU, the average anthropogenic CH₄ emissions from global inversions for 2009-2021 were 19 Tg
 522 CH₄ yr⁻¹ with a min-max range of 15-23 Tg CH₄ yr⁻¹, in line with previous estimates published in Petrescu et al.
 523 (2021, 2023) and the recent RECCAP2 European GHG budgets study of Lauerwald et al., 2024. This is consistent
 524 with the UNFCCC NGHGI (2023) which report for the same period anthropogenic emissions of 15.8 ± 1.8 Tg
 525 CH₄ yr⁻¹, noting the uncertainty ranges of both estimates overlap. There is good agreement in trends, but with
 526 inversions showing a larger year to year variability. The regional inversions, for the same period, report averaged
 527 emissions of 14 Tg CH₄ yr⁻¹ with a min-max range of 13-20 Tg CH₄ yr⁻¹. We note that the regional inversions
 528 tend to report slightly lower emissions than the global inversions, closer to the UNFCCC estimates. One reason
 529 could be that regional inversions use better-constrained regional observations (e.g. ICOS, not just NOAA), have

530 higher spatial resolution, and may thus better resolve the transport. However, they may still have problems with
531 the regional boundary conditions.

532 For the *USA*, averaged over the period 2009-2021, inversions indicate anthropogenic CH₄ emissions of
533 30 Tg CH₄ yr⁻¹ with min-max range of 26-35 Tg CH₄ yr⁻¹, well in line with the UNFCCC NGHGs (2023) which
534 for the same period report anthropogenic total emissions of 29 Tg CH₄ yr⁻¹. The trends observed in TD products
535 are slightly increasing after 2010, except for CAMS which shows no trend (Figure 4). The striking discrepancy
536 between the trends from CAMS and those from MIROC4-ACTM and CTE-GCP2021 are most likely caused by
537 the increasing oil and gas emissions from the Eastern USA (Permian Basin). The same increasing trend is also
538 captured by GAINS (Figure 3). In their runs, both MIROC4-ACTM and CTE-GCP2021 use oil and gas priors
539 from GAINS, while CAMS uses priors from EDGAR (Figure 3). We discuss further differences in having CTE-
540 GCP2021 run with both EDGAR and GAINS oil & gas prior estimates in the SI.

541 For *Brazil*, inversions yield an average (range) of anthropogenic CH₄ emissions of 23 (17-27) Tg CH₄
542 yr⁻¹, slightly higher than the UNFCCC estimate of 16.6 Tg CH₄ yr⁻¹. The two CAMS inversions have trends
543 which match the trend of the UNFCCC reports estimates.

544 For *China*, approximately 80 % of the CH₄ emission increase (21.5 Tg yr⁻¹) during 2000 – 2015 was
545 from fugitive emissions from coal (mines), consistent with what GAINS and EDGAR report (Figure 3). The TD
546 estimates mostly agree with the BURs, except for CAMS inversions which show 10 to 20 Tg CH₄ yr⁻¹ higher
547 emission than the other inversions. Both MIROC4-ACTM runs (control and OH inter-annual variability (IAV)
548 varying run; Patra et al., 2021) are in line with the BURs. Trend wise, all inversions agree on a slight decrease
549 after 2013 and show increased emissions after 2018.

550 For *Indonesia*, most TD results agree on the trend and show a slight increase in emissions. A similar
551 trend is also seen by the BURs. However, the CAMS inversions show linear increased trends while the other
552 inversions have a more variable trend. Regarding the East Asian estimates, MIROC4_ACTM inversion simulates
553 higher fluxes compared to the other inversions. Only recently they found that annual total East Asian emissions
554 have lowered more significantly than in Patra et al. (2016) or Chandra et al. (2021), therefore new runs with
555 updated input set-ups are currently being investigated.

556 For *Russia*, the estimates from the two MIROC4-ACTM runs and CTE-GCP2021 are both in the same
557 range as the BU GAINS estimate (see Figure 2) from 2000 onwards (between 30-40 Tg CH₄ yr⁻¹) but does not
558 show such a strong decrease as GAINS from 1990 to 2000), while CAMS runs report about 10 Tg CH₄ yr⁻¹ lower
559 emissions than the other two inversions, matching the UNFCCC estimates. The reason for higher estimates
560 reported by CTE-GCP2021 and MIROC4-ACTM is most likely the use of oil and gas priors from GAINS.

561 For *DR Congo*, inversions show the same slightly increasing trend, similar to that of UNFCCC BURs,
562 without the abrupt increase after 2010. The inversions appear to confirm the overreported growth in emissions
563 from waste.

564 For *India*, all the TD anthropogenic estimates agree well on increased trends and magnitudes. In
565 contrast, UNFCCC reporting does not show any trend, but given the insufficient data from BURs, a plausible
566 conclusion cannot be drawn.

567

568 ***3.4. Sectoral attribution of CH₄ emissions in TD products***

569

570 In some cases, inversions can be used to partition emissions to different sources. Table 2 shows the
571 partitions as originally reported by some of the inversions, which we name here “unharmonized partitions”. A
572 straightforward, direct comparison of the fluxes is not possible because of the different ways each inversion
573 allocates and groups the natural/anthropogenic fluxes. For example, not all inversions report soil fluxes as done
574 by MIROC4-ACTM and CTE-GCP2021 (together with wetlands) or report the biomass burning fluxes separately
575 from anthropogenic emissions (MIROC4-ACTM and TM5-4DVAR). Rice is also sometimes allocated to natural
576 emissions. Termites, oceans and geological fluxes are sometimes reported separately (MIROC4-ACTM) or
577 grouped in “Other” (CTE-GCP2021, TM5-4DVAR). Regarding the anthropogenic emissions, TM5-4DVAR
578 reports them as other, providing a separate partition for rice. Figure 5 shows the UNFCCC NGHGI anthropogenic
579 total reported estimate (diamond) next to all TD estimates. All global inversions report total and disaggregated
580 partitions, while the regional inversions report only the total emissions (green column).

581

582 *Table 2: Unharmonized partitions originally reported by inverse products:*

Inversion	Anthropogenic	Rice	Soils	Wetlands	Ocean	Termites	Geological	Biomass burning	Other
CAMSv21r1_N OAA and NOAA_GOSAT runs	Yes (as Other)	Yes	No	Yes	Yes (in Other)	Yes (in Other)	No	Yes	Yes**
MIROC4- ACTM (control and OH var)	Yes ((Agr, Waste, Oil/Gas, Biofuel, coal)	Yes (in Agr.)	Yes	Yes	Yes	Yes	Yes	Yes	Yes (separated)
CTE-GCP2021*	Yes (Agr, waste, fossil fuel, biofuel, biomass burning)	Yes (in Agr.)	Yes (BIO)		Yes (as Other)	Yes (as Other)	Yes (as Other)	as anthr.	Yes (Ocean, Termites, Geological)
CEOS (GOSAT)	Yes (Livestock, rice, waste, coal, oil, fire)	as anthr.	No	Yes	No	No	Yes (seeps)	as anthr.(but separate)	only seeps
TM5-4DVAR (TROPOMI)	Yes (as Other)	Yes	No	Yes	Yes (as other)	Yes (as other)	Yes (as Other)	Yes	Yes**
GEOS-Chem CTM (TROPOMI for USA)	Yes (Livestock, Oil Gas, Landfills, Wastewater, Other anthro (rice)	as other anthr.	No	Yes	Yes (as Other)	Yes (as Other)	Yes (as Other)	Yes (as Other)	Yes***

583 *CTE-GCP2021 partitions refer to anthropogenic, bio and other.

584 ** In TM5-4DVAR (similar to the CAMSv20 set-up and CAMSv21r1), the "Other" partition includes anthropogenic sources
585 except for the rice paddies. It also includes the small fluxes from termites, oceans, soil sink, geological etc.). More details on
586 priors are found in Petrescu et al., 2024, Priors table.

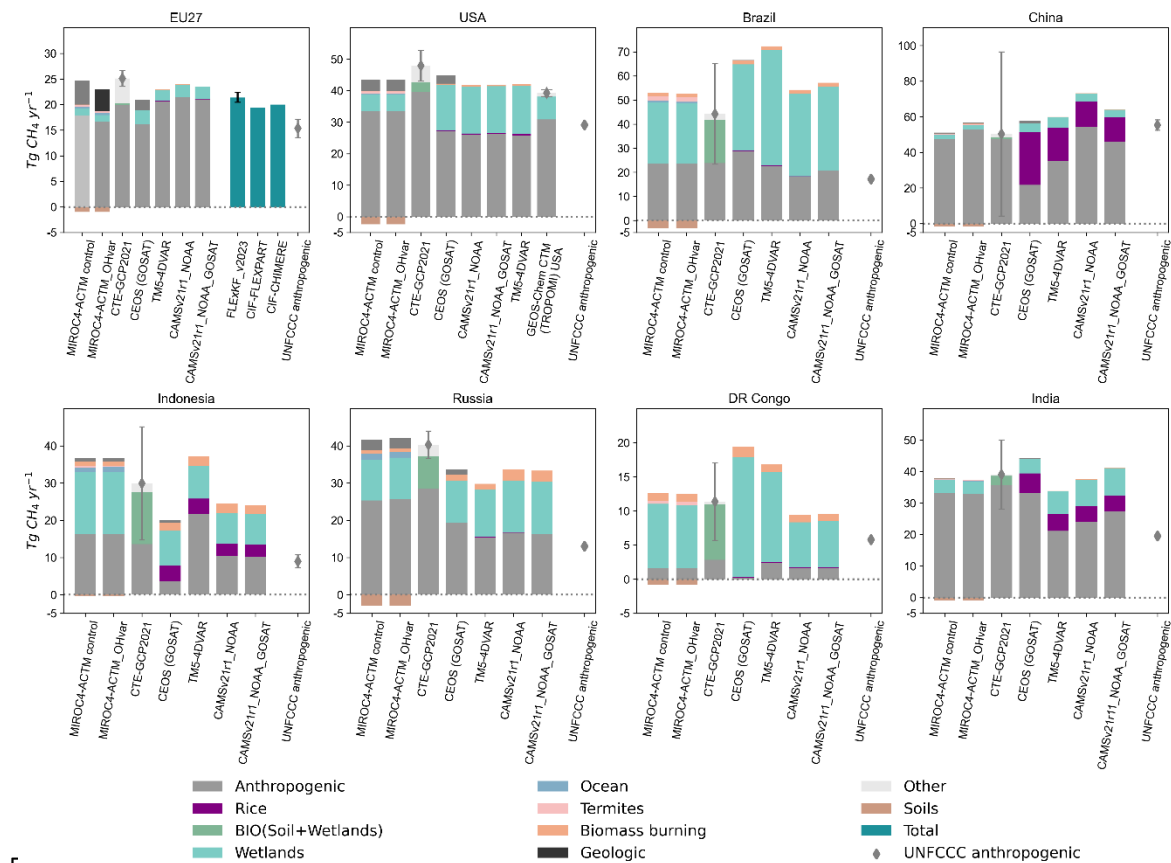
587 ***Named Other biogenic

588

589

590

UNFCCC anthropogenic and unharmonized CH_4 emissions from partitions reported by TD estimates (average 2015-last available year)



592 *Figure 5: Total (green) and disaggregated anthropogenic and natural CH_4 emissions from TD estimates*
 593 *compared to UNFCCC NGHGI (2023) anthropogenic emissions (incl. LULUCF) (diamond) for the EU and*
 594 *seven global emitters outside the EU (USA, Brazil, China, Indonesia, Russia, DR Congo and India). The*
 595 *UNFCCC anthropogenic value represents the sum of all five IPCC sectors (Energy, IPPU, Agriculture,*
 596 *LULUCF and Waste). The partitions reported by the TD global inversions are detailed in Table 2. The relative*
 597 *error on the UNFCCC CRF value represents the NGHGI (2023) reported uncertainties computed with the error*
 598 *propagation method (95% confidence interval) and gap-filled to provide respective estimates for each year (see*
 599 *Petrescu et al., 2023, Appendix). China value and uncertainties (min 5.2 %, max 5.3 %) are for 2014 only and*
 600 *Indonesia uncertainties for 2019, 19.9 %. For the USA CEOS (GOSAT) we used the Nessar et al., 2023 total*
 601 *uncertainty of min 1.1 and max 1 $Tg\ yr^{-1}$. CTE-GCP2021 provides uncertainties for each partition, but here the*
 602 *uncertainty of the total flux is shown. FLEXkF_y2023 reports the relative uncertainty (%) of the posterior*
 603 *emissions. The plotted data represents the average between 2015 and last available year as follows: CIF-*
 604 *CHIMERE (2022), TM5-4DVAR, CIF-FLEXPART and CTE-GCP2021 (2020) and FLEXkF_y2023, MIROC4-*
 605 *ACTM_OHvar and control, UNFCCC CRFs, and CAMSv21r1_NOAA and NOAA_GOSAT runs (2021).*
 606 *GEOS-Chem CTM (TROPOMI) USA reports only for 2019 (Nessar et al., 2023).*

607 Since the different models define sectors differently, also whether they are natural or anthropogenic,
 608 harmonization is required to make them comparable. CTE-GCP2021 reports the net natural land-biosphere flux
 609 “BIO flux” (soil+wetlands), while other inversions report wetlands and soil separately. Rice emissions are
 610 sometimes a part of the agriculture component (anthropogenic partition) (MIROC4-ACTM, CTE-GCP2021)

611 while CEOS (GOSAT) and GEOS-Chem CTM (USA TROPOMI) report separate partitions for rice in
612 anthropogenic emissions, while CAMS reports rice separate from anthropogenic and natural. Same for the
613 biomass burning - CTE-GCP2021 and CEO report it as part of anthropogenic emissions, while GEOS-Chem
614 CTM as part of Others. The rest of the inversions report it separately; this different allocation makes comparisons
615 for these two sources challenging. To facilitate comparisons between all TD products, we aggregated and
616 harmonized the partitions in three main categories, as summarized in Table 3 and Figure 6. The dark green
617 columns in Figure 6 show the total flux for regional EU inversions which did not report partitions.

618

619 *Table 3: Harmonized partitions from inverse products:*

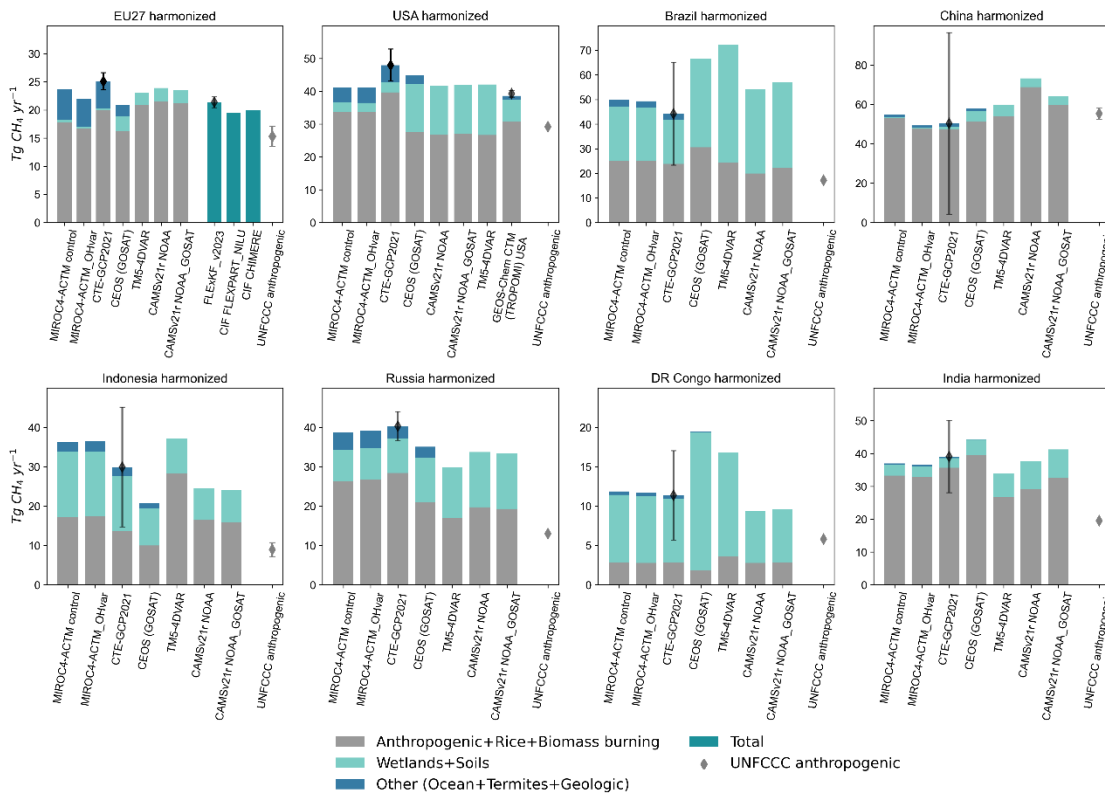
Inversions	Anthropogenic + Rice + Biomass burning			Soils + Wetlands		Other (Ocean + Termites + Geological)		
	Anthropogenic	Rice	Biomass burning	Soils	Wetlands	Ocean	Termites	Geological
CAMsv21r1_NOA A and NOAA_GOSAT runs	= Other	Yes	Yes	No	Yes	Yes	Yes	Yes
MIROC4-ACTM (control and OH var)	Yes ((Agr (livestock + rice), Waste, Oil/Gas, Biofuel, coal)	In Agr.	Yes, summed to anthr.	Yes	Yes	Yes	Yes	Yes
CTE-GCP2021*	Yes (Agr (rice is in), waste, fossil fuel, biofuel, biomass burning)	in Agr.	In anthr.	Yes (BIO)		Yes (Other)		
CEOS (GOSAT)	Yes (Livestock, rice, waste, coal, oil, fire)	In anthr.	In anthr.	No	Yes	No	No	Yes
TM5-4DVAR (TROPOMI)	Others + Rice+ BB	In anthr.	Yes, summed to anthr.	In Other	Yes	Yes	Yes	Yes
GEOS-Chem CTM (TROPOMI) USA	Yes	In anthr.	In other biogenic	No	Yes	Yes	Yes	Yes

620 *CTE-GCP2021 partitions refer to Anthropogenic, Bio and Other. Other fluxes are imposed

621

622

UNFCCC anthropogenic and harmonized CH_4 emissions from partitions reported by TD estimates (average 2015-last available year)



623

624 *Figure 6: Total (green) and disaggregated anthropogenic and natural CH_4 emissions from TD estimates*
 625 *compared to UNFCCC NGHGI (2023) anthropogenic emissions (incl. LULUCF) for the EU and seven global*
 626 *emitters (USA, Brazil, China, Indonesia, Russia and DR Congo). The UNFCCC anthropogenic value represents*
 627 *the sum of all five IPCC sectors (Energy, IPPU, Agriculture, LULUCF and Waste). The partitions reported by*
 628 *the TD global inversions are harmonized and detailed in Table 3. The relative error on the UNFCCC CRF value*
 629 *represents the NGHGI (2023) reported uncertainties computed with the error propagation method (95%*
 630 *confidence interval) and gap-filled to provide respective estimates for each year (see Petrescu et al., 2023,*
 631 *Appendix). In 2014, China UNFCCC value and reported uncertainties (min 5.2 % and max 5.3 %) are for 2014*
 632 *while Indonesia reported uncertainties for 2019, 19.9 %. India UNFCCC value is for 2016. CTE-GCP2021*
 633 *provides uncertainties for each partition, but here we plotted the uncertainty of the total flux. FLEXkF_v2023*
 634 *reports the relative uncertainty (%) of the posterior emissions. The plotted data represents the average between*
 635 *2015 and last available reported year as follows: CIF-CHIMERE (2022), UNFCCC CRFs, TM5-4DVAR, CIF-*
 636 *FLEXPART and CTE-GCP2021 (2020) and FLEXkF_v2023, MIROC4-ACTM both runs, and CAMSv21r1 both*
 637 *runs (2021). GEOS-Chem CTM (TROPOMI) USA reports only for 2019 (Nesser et al., 2023).*

638 3.5. Comparison of BU and TD CH_4 estimates

639

640 Figure 7 summarizes the total CH_4 fluxes for the EU and the seven global emitters as following: BU
 641 anthropogenic sources disaggregated per sector, BU natural emissions, TD natural emissions from regional and
 642 global inversions, and total emissions from global TD estimates (see 2.3 and SI for description of all data
 643 products). This figure brings all the estimates together to demonstrate the reconciliation process.

644 Inversions currently report in a way that makes comparison between BU natural and TD natural sources
645 difficult. TD products differ in the sources they report (Table 2) or they allocate them to different categories. We
646 consider natural the following sources: biomass burning, soils, oceans and termites (often reported by inversions
647 under category “Other”), wetlands, geological and lakes & reservoirs (or freshwaters). Due to lack of
648 information, biomass burning emissions were considered among the natural sources, recognizing that in regions
649 like tropical forests, some of these events are influenced by human intervention. To make the products from
650 Figure 7 comparable, we added the missing BU information from TD, and vice-versa, presented in hatched
651 pattern. In this way, comparison between BU and TD natural emission estimates is consistent regarding the
652 “apples to apples” comparison, but became “apples of different flavors” (see Table 4):
653

654 *Table 4: BU and TD natural partitions as presented in Figure 7:*

Product name	TD natural partitions			
	Reported	Missing* (not in priors)	Not reported**	Missing was added from:
CAMsv21r1_NOAA	BB, wetlands, “Others” include anthropogenic and was not used	lakes and reservoirs, geological	termites, oceans, soil sink	DAAC lakes and reservoirs, geological, updated in this study (see SI)
CAMsv21r1_NOAA_GOSAT	BB, wetlands, “Others” include anthropogenic and was not used	lakes and reservoirs, geological	termites, oceans, soil sink	DAAC lakes and reservoirs, geological, updated in this study (SI)
MIROC4-ACTM control	BB, wetlands, oceans, termites, soils, geological	lakes and reservoirs		DAAC lakes and reservoirs
MIROC4-ACTM_OHvar	BB, wetlands, oceans, termites, soils, geological	lakes and reservoirs		DAAC lakes and reservoirs
CTE-GCP2021	soils + wetlands (BIO), termites and oceans	lakes and reservoirs	BB, geologic	DAAC lakes and reservoirs
CEOS (GOSAT)	Fires (BB), Seeps and wetlands	termites, oceans, soils, lakes and reservoirs		MIROC4-ACTM (termites, oceans and soils), DAAC lakes and reservoirs
TM5-4DVAR (TROPOMI)	BB and wetlands	lakes and reservoirs, geological,	termites, oceans, soil sink	DAAC lakes and reservoirs, geological, updated in this study (see SI)
Product name	BU natural partitions			
		Reported	Not reported**	Added from
Biomass burning Lakes and reservoirs Wetlands Geological		GFEDv4.1s DAAC LPJ-GUESS Geological emissions updated in this study (SI)	soils termites oceans	MIROC4-ACTM

655 note: in TD products termites, oceans emissions are imposed from existing literature
656 * missing = not in the priors, presented as hatched pattern in the figure “\\”
657 **Not reported = data not available, presented as hatched pattern in the figure “//”

658

659 For an easier visual comparison and reconciliation between BU and TD estimates, we added the mean
660 of the BU anthropogenic estimates (off-white), underneath the BU and TD natural estimates. To note that for
661 some countries (e.g. Russia, DR Congo) this area might look like subtracted from the BU natural estimates, but
662 this is due to the sign convention used in this study (sink = negative and source = positive). In most cases, the
663 missing soil sink emissions are represented as a downward area.

664 We note that for most countries, the sum of the anthropogenic and natural components matches those
665 of the TD global total estimates. This gives confidence that, to a certain extent and albeit with inconsistencies
666 between products, BU anthropogenic emission estimates are accurate and consistent with the observation-based
667 estimates and can be used to reconcile with the atmospheric-based estimates. We note from Figure 7 that in all
668 Annex I countries (EU, USA, Russia) and China, TD and BU natural emissions are consistent with each other,
669 after including the missing sources, as detailed in Table 4. For Brazil and DR Congo, the gap between the two
670 natural components is highly significant, while less for Indonesia and India. We hypothesize that mapping of the
671 wetland extent might cause these inconsistencies.

672

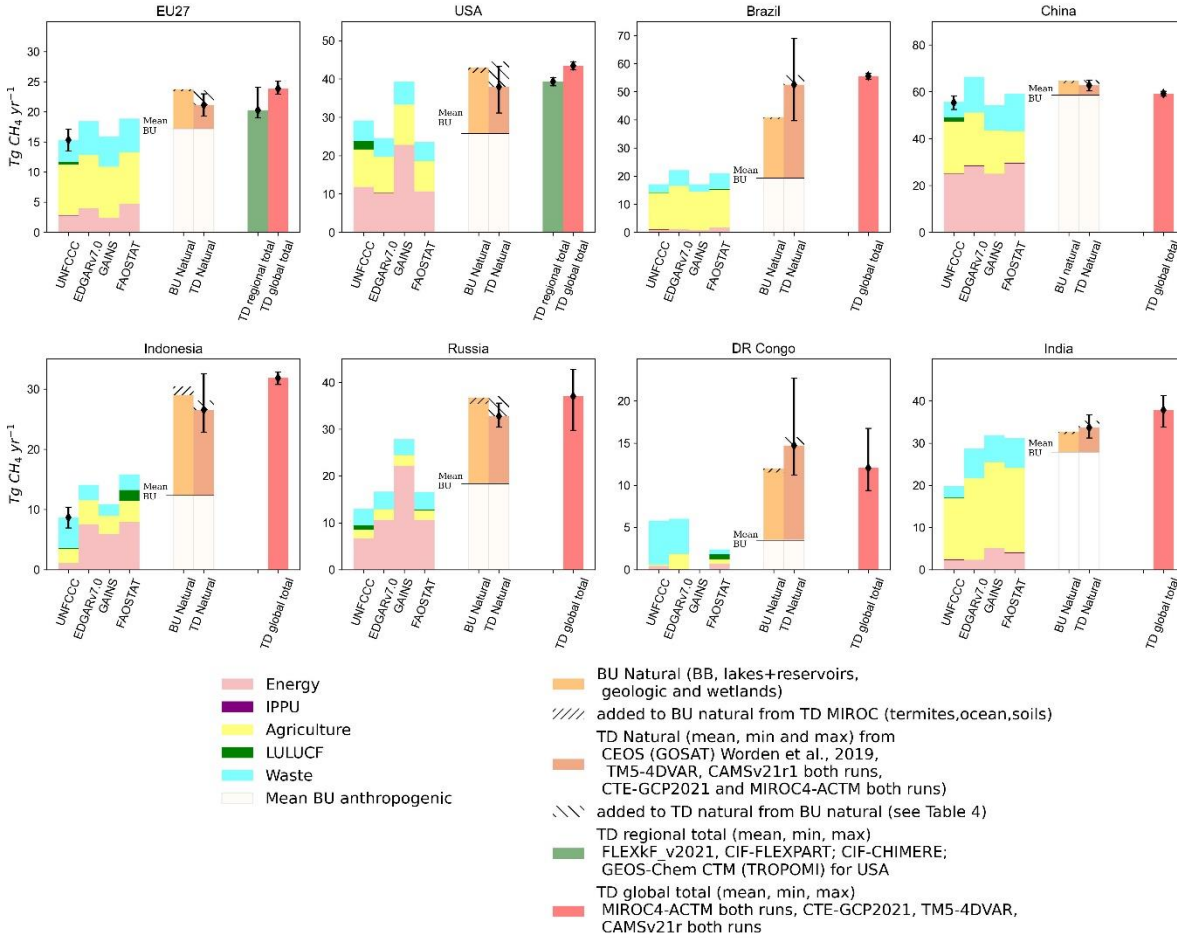
673

674

675

676

Total anthropogenic and natural CH_4 emissions from UNFCCC, other BU and TD estimates (average 2015-last available year)



67

678 *Figure 7: Total anthropogenic and natural CH_4 emissions from BU and TD estimates presented as average of*
 679 *2015-last available year for EU and seven global emitters (USA, Brazil, China, Indonesia, Russia, DR Congo*
 680 *and India). The BU anthropogenic estimates belong to: UNFCCC NGHGI (2023) CRFs and BURs (incl.*
 681 *LULUCF) as sectoral shares, EDGARv7.0, GAINS and FAOSTAT/PRIMAP-hist. The relative error on the*
 682 *UNFCCC CRF value represents the NGHGI (2023) reported uncertainties computed with the error propagation*
 683 *method (95% confidence interval) and gap-filled to provide respective estimates for each year (see Petrescu et*
 684 *al., 2023, Appendix). In 2014, China reported an uncertainty of min 5.2% - max 5.3%. The BU Natural emissions*
 685 *for the EU are the sum of the VERIFY products (biomass burning, inland waters, geological and peatlands plus*
 686 *mineral soils as described in Petrescu et al., 2021 and 2023, Appendix A2.1). For the seven non-EU emitters,*
 687 *the BU Natural fluxes are the sum of wetland emissions (LPJ-GUESS), lakes and reservoirs fluxes (ORNL DAAC,*
 688 *Johnson et al., 2022), geological (updated activity in SI) and biomass burning emissions (GFED4.1s). The TD*
 689 *natural global estimates are presented in Table 1. The uncertainty on the TD natural emissions is the min/max*
 690 *of all estimates. To both BU and TD estimates, missing (as not reported or not included in the priors) was added*
 691 *(see Table 4). The natural emissions have been plotted starting at the mean of the BU anthropogenic estimates,*
 692 *to retain comparability across the natural emission estimates, but also compare with the total TD estimates. The*
 693 *total regional TD estimates (for EU) belong to the mean and min/max of FELXkF_v2023, CIF-FLEXPART and*
 694 *CIF-CHIMERE and for USA GEOS-Chem CTM (TROPOMI) for the year 2019 (Nesser et al., 2023). The total*
 695 *global TD inversions represent the average of the 2015-last available year of the mean and min/max of CTE-*

696 *GCP2021, MIROC4-ACTM both runs, CAMS v21r both runs and TM5-4DVAR. The last available years are*
697 *2022 for CIF-CHIMERE, 2021 for EDGARv7.0, FAOSTAT, MIROC4-ACTM both runs, UNFCCC CRFs, and*
698 *CAMSv21r1 both runs, and 2020 for CIF-FLEXPART and CTE-GCP2021. TM5-4DVAR partitioned data is only*
699 *available between 2018 and 2020.*

700 However, Figure 7 should be interpreted with caution because in Europe, natural emission priors come
701 from regional ecosystem model simulations, where drained peatland, drainage ditches areas, and pristine areas
702 are lumped together. Therefore, if both LULUCF sector and natural BU emissions are included in the total budget
703 estimation, there is some overlap and possible double counting. Especially, ecosystem model estimates of ‘soil
704 sink’ or ‘inundated soil emissions’ may overlap with NGHGI managed peatland forest soil category (or
705 agricultural soils). The separation of emissions into different categories requires further clarification together
706 with inventory makers. Furthermore, it should be assessed which emissions should be called natural and which
707 anthropogenic (e.g., LULUCF, Agriculture) by inversions.

708

709 **4. Challenges comparing bottom-up and top-down estimates**

710

711 An off-the-shelf comparison of BU and TD estimates is not possible, with a variety of adjustments
712 needed for comparability, often without the necessary data. Broadly speaking, inversions have not necessarily
713 been designed to compare directly to NGHGIs. A valid comparison should have consistent system boundaries
714 and perform a full uncertainty analysis to determine whether differences between estimates are statistically
715 significant given the constraining observational data.

716 The two most common issues limiting comparability are geographic scope and system boundaries
717 (Petrescu et al., 2021, 2023; McGrath et al., 2023, Andrew 2020; Grassi et al., 2018). The **geographical scope**
718 of inverse modelling versus inventory estimates should be controllable, but it can be challenging for small
719 countries or coarse inversions. Inversions are generally performed on a spatial grid and require aggregation, in
720 line with how official NGHGIs are reported (EEA, 2013). Inconsistent **system boundaries** have implications in
721 comparing the inventory- with inversions-based estimates for **source attribution**, e.g., anthropogenic vs. natural.
722 Most emission inventories aim at estimating anthropogenic emissions, while most inversions estimate both
723 anthropogenic and natural emissions. This is a particularly important issue for CH₄ where, globally, natural
724 emissions are of similar magnitude as anthropogenic emissions, with larger variations at regional scales, mainly
725 due to seasonality (i.e. wetlands). Thus, methods are needed to separate the anthropogenic flux from the total
726 flux (Deng et al. 2022, and above section 3.4). Similar issues arise with fossil CO₂ (Andrew 2020) as different
727 datasets can report different emission sources. Standardization procedures, such as The Community Inversion
728 Framework (CIF; Berchet et al., 2021), may help resolve some of these issues.

729 When comparing inventory- and inversion-based emissions, there are difficulties in analyzing **trends**
730 due to different time scale **variability**. Inventory-based approaches report emissions at the annual level, but often
731 do not consider interannual variations. Further, the Paris Agreement is set around five-yearly Global Stocktakes,
732 which indicates a desire to average trends, prioritizing the multi-annual trend over IAV, canceling out extremes
733 from both weather and socio-economic fluctuations. Inversion models, on the other hand, include variations over
734 a wide range of timescales, but in particular for IAV (e.g. OH and weather) that remains challenging to assess.

735 For an effective comparison, inversion-based estimates need to have IAVs statistically removed to make
736 comparisons with NGHGs easier (e.g., 5-year or 10-year averages or trend analysis). Additionally, averages of
737 ensembles of inversions may mask underlying differences and trends in individual inversions. Many research
738 projects make use of multi-model ensembles (Saunio et al., 2020; Deng et al. 2022, Lauerwald et al., 2024,
739 Zhang et al., 2024). From a scientific perspective, the model ensemble is often considered a more robust estimate
740 of the mean and uncertainty, as often individual estimates make errors due to some limitations and biases, while
741 in an ensemble, these errors are averaged out. From an inventory perspective, individual model comparisons may
742 be more efficient, as various input variables or processes can be compared directly to the inventory. Currently,
743 most inventory comparisons in UNFCCC NGHGI (e.g. UK, Switzerland) use single-model comparisons.

744 A strength of inversions is that they provide high temporal and spatial resolutions, which are not directly
745 capitalized when comparing with inventories. CH₄ from the fossil-fuel industry can contribute to large releases
746 to the atmosphere over a short period of time, given the large number of uncontrolled emission point sources in
747 oil and gas (O&G) and coal production areas worldwide (Jackson et al., 2020). Such processes include leakage
748 from landfills, spontaneous events from oil and gas production activities, so-called uncontrolled gas well blasts
749 etc. (Jacob et al., 2016, 2022). These uncontrolled events are difficult to include in the national inventories
750 leading to a potential underestimate of emissions (Massakkers et al., 2016, 2022). Recently, under the CoCO₂
751 project (<https://coco2-project.eu/>) a hot-spot satellite detection interactive map (Published studies on hot spot
752 detection (CO₂, CH₄) - uMap (openstreetmap.fr)) was released as a user-centric interface featuring published
753 studies on hot-spot detection between 2010 and 2021. It allows for advanced filtering by year, gas, activity,
754 geographical zone, and country.

755 A key challenge when comparing inversions with NGHGs is ensuring independence from the assumed
756 prior emissions. A more valid comparison between inversions is made when all inversions use the same **priors**.
757 In this context, we define as priors input data in the form of atmospheric observations (e.g. satellite retrievals,
758 ground-based observation networks (ICOS)) and/or bottom-up emissions datasets (e.g. EDGAR, GAINS) used
759 as input parameters to the inverse models. A key issue is the prior emission estimate. Theoretically, a constant
760 emission prior could be used, but this would require a dense observational network. Because of sparse
761 observations, inversion modelers assess how far observations have shifted the prior emissions to the posterior
762 emissions, preferably incorporating a full uncertainty analysis. The posterior emissions depend to a varying
763 extent on the prior that was used; the extent of this dependency is determined by the number of observations
764 used in the inversion, by how the observations relate to the emissions (governed by atmospheric transport) and
765 by the uncertainties assigned on the prior emissions and the observations. Thus, better quantified uncertainties
766 for the prior emissions would lead to more robust inversions. This stresses the need for more systematic in-situ
767 data necessary to produce adequate prior data (Bastviken et al., 2022) and synthesized atmospheric observations
768 with their uncertainties to robustly constrain the inversions.

769 It is not generally clear how inventory uncertainties can be compared to inversion uncertainties; however, it is
770 important that both methods provide comprehensive **uncertainty estimates**. The prior emissions used as input
771 into an inversion model should have robust uncertainty estimates, particularly with correlations in space and
772 time. This allows a full inversion system to better characterize how observations reduce uncertainty when
773 estimating the posterior estimate. Very few inversions routinely report this information. The inventory-based

774 emission estimate will additionally have uncertainty estimates, though these statistics may not be sufficiently
775 robust for verification purposes (National Academies of Sciences, Engineering, and Medicine 2022). There are
776 often offsets in inversion models, because of systematic inconsistencies between observations and chemistry-
777 transport models, which may make trends more robust than instantaneous estimates. Though, estimating
778 uncertainty in trends also requires understanding the correlation structure in time.

779 A key challenge for comparisons between NGHGI and independent estimates is to understand the
780 reasons for differences. In the case of BU comparisons, obtaining sufficient activity data and emission factors
781 should enable an accurate reconciliation of different estimates. However, in practice, it is often not possible to
782 obtain the necessary data. For inversions it is more complex. Often a close collaboration may be needed between
783 the inversion modeler and NGHGI team (e.g. UK NIR). If an inversion indicated a different trend in agricultural
784 CH₄ emissions, it is necessary to track down if this is a real difference or artifact of the inversion system. After
785 this, the spatial and temporal data in the inversion could be useful to the NGHGI team to locate what is causing
786 the difference. Many of the comparisons we show in this article ultimately remain comparisons, with detailed
787 reconciliations likely requiring intensive country-level case studies.

788 **5. Data availability**

789

790 Data files reported in this work which were used for calculations and figures are available for public
791 download at <https://doi.org/10.5281/zenodo.12818506> (Petrescu et al., 2024). The data are reachable with one
792 click (without the need for entering login and password), with a second click to download the data, consistent
793 with the two click access principle for data published in ESSD (Carlson and Oda, 2018). The data and the DOI
794 number are subject to future updates and only refer to this version of the paper. The raw gridded data is available
795 upon request, directly from the data providers, as detailed in the Supplementary Information, Table S2.

796 **6. Conclusions**

797

798 We analyzed data from both anthropogenic and natural CH₄ fluxes, from both BU and TD observation-
799 based estimates (Table 1). BU estimates show that the largest sectors depend on the country: Agriculture (EU,
800 Brazil, India), Energy (US, China, Russia), and Waste (Indonesia, DR Congo). The inversions attribute most of
801 the fluxes to the anthropogenic emissions, with tropical countries having a larger share of natural emissions
802 (wetlands). The EU and the seven other large emitters analyzed here contribute an anthropogenic emission of
803 173 Tg CH₄ yr⁻¹ (sum of last UNFCCC reported year, Figure 1,2), representing roughly half of the total global
804 anthropogenic emissions (386 Tg CH₄ yr⁻¹) reported by EDGARv7.0 in 2021. For comparison, the average of
805 the anthropogenic component from the atmospheric global inversions (MIROC booth runs, CTE-GCP2021,
806 CEOS and CAMS booth runs) is 181 Tg CH₄ yr⁻¹ (Figure 5).

807 We performed comparisons to the UNFCCC NGHGIs using the BU and TD data. Comparisons between
808 UNFCCC and BU products (Figure 3) reveal some deviations, particularly related to assumptions on gas/oil
809 emissions (e.g., GAINS for Russia and the USA) and waste (e.g., Indonesia, DR Congo). It is more challenging
810 to compare BU and TD estimates, due to different attribution to source activities (Table 2 and Figure 5) and
811 different priors used in the simulations (Petrescu et al., 2024, Priors Table). The comparison between UNFCCC

812 and the TD estimates (Figure 4) agrees largely with the findings of Deng et al. (2022) who applied different
813 methodologies to calculate natural emissions. In most cases, the gap between the anthropogenic BU fluxes from
814 inventories and total TD fluxes can be largely explained by the natural fluxes (Figure 7). It is difficult to draw
815 definitive conclusions on emissions trends seen by inversions, as the adjustments for natural emissions and IAV
816 and seasonal variability might strongly influence trends. Despite this, given that, in most cases, the UNFCCC
817 BURs reports are incomplete for the non-Annex I parties (China, Indonesia, DR Congo) it is important to
818 acknowledge that the TD estimates might become a useful way to complement inventories and play a role in the
819 validation of the BU estimates.

820 There is still a pressing need for reporting of uncertainties in both prior and posterior emissions, even if
821 some TD inversions do report it as the standard deviation of ensemble members (CTE-GCP2021 and
822 FLEXkF_v2023, Figure 4). The use of a variety of priors across different inversion systems can also inhibit
823 comparability with inventories and between inversions. Generally, inversions are still ill-constrained by
824 observations (only 60 sites globally plus satellites) and the prior flux uncertainty for each of the 54 regions is
825 large. Therefore, the monthly results could be more ill-constrained than the annual totals. Even if comparisons
826 between CH₄ inversion estimates and NGHGIs are currently uncertain because of the spread in the inversion
827 results, TD inversions inferred from atmospheric observations represent partly independent data against which
828 inventory totals and trends can be compared, considering the most encountered issues discussed in section 4.

829

830 **7. Appendix**

831

832 All the information regarding models/methods descriptions is available in the Supplementary Information (SI)
833 file. Appendices A1 and A2 in Petrescu et al., 2023 contain detailed information about Table 1 products. Further
834 information on new products together with references and contact details are found in Tables S1 and S2 in SI.

835

836 The tables with priors used by all the products and the matrix highlighting the comparability issues identified in
837 section 4 are found in the Zenodo data repository, Petrescu et al., 2024.

838

839 **Supplementary Information (link)**

840

841

842 **Author contributions**

843

844 AMRP designed research and led the discussions; AMRP wrote the initial draft of the paper and edited all the
845 following versions; GPP drafted the initial version of section 4, edited the final version of this manuscript,
846 contributed to the revised version and advised on the context; PP processed all the original EU data submitted to
847 the VERIFY portal; RLT, SH, BM, DaB, RL, PKP, AT, RMA, LHI, FNT, GC and JG edited and gave consistent
848 comments and suggestions to the initial manuscript; DaB, RL and RMA provided input to the final revised
849 version; all co-authors are data providers and contributed to subsequent versions of the manuscript by providing
850 specific comments and information related to their data in the main text, providing as well product descriptions
851 for the Supplementary Information file.

852 **Competing interests**

853 At least one of the (co-)authors is a member of the editorial board of Earth System Science Data.

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883 AVENGERS.

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885 **8. References**

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