

1 Comparison of observation- and inventory- based methane emissions for eight large global
2 emitters

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40 **Abstract**

41

42 Monitoring the spatial distribution and trends in surface greenhouse gas (GHG) fluxes, as well as flux
43 attribution to natural and anthropogenic processes, is essential to track progress under the Paris Agreement and
44 to inform its Global Stocktake. This study updates earlier syntheses (Petrescu et al., 2020, 2021, 2023a) and
45 provides a consolidated synthesis of CH₄ emissions using bottom-up (BU) and top-down (TD) approaches for
46 the European Union (EU) and is expanded to include seven additional countries with large anthropogenic and/or
47 natural emissions (USA, Brazil, China, India, Indonesia, Russia, and the Democratic Republic of Congo (DR
48 Congo)). Our aim is to demonstrate the use of different emission estimates to help improve national GHG
49 emission inventories for a diverse geographical range of stakeholders.

50 We use updated National GHG Inventories (NGHGIs) reported by Annex I Parties under the United
51 Nations Framework Convention on Climate Change (UNFCCC) in 2023 and the latest available Biennial Update
52 Reports (BURs) reported by non-Annex I Parties. Comparing NGHGIs with other approaches highlights that
53 different system boundaries are a key source of divergence. A key system boundary difference is whether both
54 anthropogenic and natural fluxes are included and, if they are, how fluxes belonging to these two sources are
55 partitioned.

56 Over the studied period, the total CH₄ emission estimates in the EU, USA, and Russia show a steady
57 decreasing trend since 1990, while for the non-Annex I emitters analyzed in this study, Brazil, China, India,
58 Indonesia, and DR Congo, CH₄ emissions have generally increased. Quantitatively, in the **EU** the mean of 2015-
59 2020 anthropogenic UNFCCC NGHGIs (15 ± 1.8 Tg CH₄ yr⁻¹) and the mean of the BU CH₄ emissions (17.8 (16-
60 19) Tg CH₄ yr⁻¹) generally agree on the magnitude, while inversions show higher emission estimates (medians
61 of 21 (19-22) Tg CH₄ yr⁻¹ and 24 (22-25) Tg CH₄ yr⁻¹ for the three regional and six global inversions,
62 respectively), as they include natural emissions, which for the EU were quantified at 6.6 Tg CH₄ yr⁻¹ (Petrescu
63 et al., 2023a). Similarly, for the other Annex I Parties in this study (**USA and Russia**), the gap between the BU
64 anthropogenic and total TD emissions is partly explained by the natural emissions.

65 For the **non-Annex I Parties**, anthropogenic CH₄ estimates from UNFCCC BURs show large
66 differences with the other global inventory-based estimates and even more with atmospheric-based ones. This
67 poses an important potential challenge to monitoring the progress of the global CH₄ pledge and the Global
68 Stocktake. Our analysis provides a useful baseline to prepare for the influx of inventories from non-Annex I
69 Parties as regular reporting starts under the Enhanced Transparency Framework of the Paris Agreement.

70 By systematically comparing the BU with TD methods, this study provides recommendations for more
71 robust comparisons of available data sources and hopes to steadily engage more Parties in using observational
72 methods to complement their UNFCCC inventories, as well as considering their natural emissions. With
73 anticipated improvements in atmospheric modeling and observations, as well as modeling of natural fluxes,
74 future development needs to resolve knowledge gaps in both BU and TD approaches and to better quantify

75 remaining uncertainty. TD methods may emerge as a powerful tool to help improve NGHGs of CH₄ emissions,
76 but further confidence is needed in the comparability and robustness of the estimates.

77 The referenced datasets related to figures are available at <https://doi.org/10.5281/zenodo.12582667>
78 (Petrescu et al., 2023b).

79 **1. Introduction**

80

81 In 2021, the NOAA Global Monitoring Laboratory (GML) reported the largest annual increase in
82 atmospheric CH₄ mixing ratios since records began in 1983, at 17 parts per billion (ppb) (NOAA
83 (https://gml.noaa.gov/ccgg/trends_ch4/). In 2022, atmospheric CH₄ concentrations averaged 1912 ppb yr⁻¹, 162
84 % higher than pre-industrial levels. A similar, abnormally large growth rate of 14.8 ppb yr⁻¹ was detected from
85 total column mixing ratio measurements (XCH₄) by the Greenhouse Gases Observing Satellite (GOSAT) (Peng
86 et al., 2022).

87 CH₄ in the atmosphere has many different sources, of both natural and anthropogenic origin. The natural
88 sources of CH₄ are dominated by wetlands, while anthropogenic emissions principally come from agricultural
89 activities (livestock and rice farming), waste management (landfills and water treatment plants) and the
90 production, transportation, and use of fossil fuels. Most of the agricultural sources are distributed sources, while
91 the energy-related industrial sources of CH₄ are a mix of large point sources, of which some are detectable by
92 satellite (Jacob et al., 2022) and smaller point and distributed sources of fugitive emissions (e.g., leaks in
93 pipelines and compression stations), which are more challenging to identify (Rutherford et al., 2021; Omara et
94 al., 2022).

95 While anthropogenic CH₄ emissions from fossil fuels, agriculture, and waste can be reduced by
96 mitigation actions, increased natural emissions lead to different challenges. It has been suggested that the drivers
97 of the recent growth are most likely driven primarily by biogenic emissions (Basu et al., 2022; Lan, et al., 2021a;
98 Lan et al., 2021b; Lan et al., 2022; Nisbet et al., 2016, 2019), with smaller contributions from increased fossil
99 fuel emissions and a reduced atmospheric sink (Nisbet et al., 2023). Fluctuations in natural sources - dominated
100 by wetlands and open water bodies - were the main reasons for some of the atmospheric CH₄ anomalies observed
101 during the last decades (Rocher-Ros et al., 2023; Zhang et al., 2023; Nisbet et al., 2023; Lunt et al., 2019). Nisbet
102 et al., 2023 review recent studies, including those which quantified the observed methane growth in the last years.
103 Using a global inverse analysis of GOSAT satellite observations, increases in the range of 22-32 Tg CH₄ yr⁻¹
104 were detected between 2019 and 2020 and were attributed to biogenic sources, half of which took place in East
105 Africa (~ 15 Tg yr⁻¹), and some were observed in Canada and Alaska (4.8 Tg yr⁻¹) (Qu et al., 2022 and Basu et
106 al., 2022).

107 Chandra et al., 2021 identified a few main sectors that triggered increases and decreases in the
108 anthropogenic CH₄ emissions of different countries. The first is energy, with its fugitive emissions from the oil
109 and gas industry whose decline in emissions helped stabilize CH₄ concentration in the 1990s, before they
110 contributed to the renewed CH₄ growth since the late 2000s (increased emissions). The other major sectors that
111 drove changes in the CH₄ growth rate were agriculture (increase in emissions from enteric fermentation and
112 manure management) and waste. The increase in emissions from enteric fermentation and manure management

113 was caused primarily by increased animal numbers, and in addition by the greater intensity of ruminant farming
114 as estimated by the FAO and the emission inventories (e.g. EDGAR) which might take into account productivity
115 increases (Crippa et al., 2020; Wolf et al., 2017; FAOSTAT, 2018) while inventory emissions from waste can
116 account for up to 43 % of the linear increase in emissions for the rest of the world.

117 The contribution of CH₄ to global warming has been estimated to be about 0.5°C relative to the period
118 1850–1900 (IPCC, 2021) (Stavert et al., 2022). Methane has a relatively short perturbation lifetime (averaging
119 12.4 years, Balcombe et al., 2018) and a high global warming potential (86 and 34 for 20- and 100-years times
120 horizons respectively, compared to that of CO₂ emissions, IPCC, 2021, Table 7.15). Given the short lifetime, a
121 decline in CH₄ emissions will rapidly reduce the global warming contribution from CH₄ and help mitigate the
122 impact of climate change at decadal time scale (Cain et al., 2021). However, efforts to reduce CH₄ emissions
123 require a thorough understanding of the dominant CH₄ sources and sinks and their temporal and regional
124 distribution and trends (Stavert et al., 2022).

125 The Paris Agreement, a milestone of the UNFCCC to combat climate change and adapt to its effects,
126 entered into force on November 4, 2016. It asks each signatory to define and communicate its planned climate
127 actions, known as Nationally Determined Contributions (NDCs), and to report their progress towards their
128 targets. Next to commitments adopted by countries at COP26, the Global Methane Pledge (GMP) was launched.
129 The goal of the GMP is to cut anthropogenic CH₄ emissions by at least 30 % by 2030 with respect to 2020 levels
130 and is seen as the fastest way to reduce near-term warming and is necessary to keep a 1.5°C temperature limit
131 within reach. Achieving this goal will drive significant gains, through specific energy and agriculture defined
132 pathways including innovative actions, national targeted policies, and green climate funds to help smallholder
133 farmers (<https://www.state.gov/global-methane-pledge-from-moment-to-momentum/>). About 150 countries
134 joined this pledge and about fifty have already developed national CH₄ action plans or are doing so. As
135 agriculture and waste are the main anthropogenic sources for CH₄ emissions, a GMP Food and agriculture
136 pathway and a GMP waste pathway were launched at COP27, foreseeing actions that increase agricultural
137 productivity, while reducing emissions from dairy, food loss and waste by supporting small farmers and
138 innovation (<https://www.state.gov/global-methane-pledge-from-moment-to-momentum/>).

139 Starting in 2024, non-Annex I Parties to the UNFCCC must - given they have sufficient capacities -
140 report formal inventories under the Paris Agreement’s Enhanced Transparency Framework following the same
141 guidelines and rules as the Annex I countries (Perugini et al., 2021). Furthermore, they will undergo more
142 stringent reviews than those that previously looked at the Biennial Update Reports (BURs) and NDCs. This will
143 also allow strengthening the robustness of such comparison exercises when using independent atmospheric
144 observations in estimating trends and patterns for regional and national CH₄ emissions (IPCC, 2006). The influx
145 of new inventories will place additional demands on independent emission estimates to help improve and inform
146 National Greenhouse Gas Inventories (NGHGs), particularly in countries with low capacity.

147 With increased focus on CH₄ in climate policy, there is a demand to ensure that planned emission
148 reductions are realized. Further, as non-Annex I countries begin regular reporting of emission estimates under
149 the Enhanced Transparency Framework, there is a need to help countries improve their GHG emission estimates.
150 This has created an active field of research comparing NGHGI and independent estimates. Our analysis builds
151 on the three-year EU funded project CoCO₂, which had as main objective the building of prototype systems for

152 a European Monitoring and Verification Support capacity for anthropogenic CO₂ (and CH₄) emissions
153 (CO₂MVS). In this context, one of the results of the CoCO₂ project was the production of a Blueprint for a
154 decision support system to be used in an eventual CO₂MVS, aiming at informing and attracting attention of
155 diverse climate stakeholders on the use of the results needed beyond research. Therefore, the objectives of this
156 study reflect those of the Blueprint and focus on user engagement. It builds on dialogues with a broad community
157 of users (e.g. scientists, inventory agencies, policy makers), considering their opinions and needs when it comes
158 to comparisons between independent approaches. Furthermore, this study expands beyond the EU to include
159 seven countries that have large anthropogenic and/or natural CH₄ emissions (USA, Brazil, China, India,
160 Indonesia, Russia and the Democratic Rep. of Congo). It examines both Annex I (EU, USA and Russia) and non-
161 Annex I estimates from observation-based BU process-based models and inversions-based TD approaches (using
162 satellite observations) by identifying and explaining differences with official inventory reports submitted by
163 parties to the UNFCCC. The seven countries were chosen based on location and the importance / magnitude of
164 their anthropogenic and natural emissions. By using multiple methodologies, uncertainties can be estimated by
165 looking at the range in both emissions and trends.

166

167 **2. Methods and data**

168 In this work we focus on comparing BU and TD emission estimates. The ‘reconciliation process’
169 described in this work is the action of making one dataset comparable with another to assess their consistency.
170 In this respect, we attempt to obtain consistent results from both BU and TD estimates, through harmonization
171 of the results, concepts and definitions. After the reconciliation process, the estimates do not necessarily agree,
172 representing uncertainties in the different methods and datasets. We now describe the key data and methods used
173 in our analysis.

174 **2.1. Verification practices in official UNFCCC NGHGs**

175

176 Quality assurance/quality control (QA/QC) is a key component of NGHGs development. Verification
177 is an additional step and refers specifically to methods that are external to the inventory and apply independent
178 data. There are two main methods of verification: 1) independent inventory-based estimates, 2) observation-
179 based emission estimates.

180 A challenge with comparisons against *independent inventory-based estimates* is that none of them is
181 truly independent as they may rely on, for example, the same activity data reported by a country (Andrew 2020).
182 Experience has shown that performing detailed comparisons (Petrescu et al., 2021, 2023a, Lauerwald et al., 2024)
183 can help clarify differences in system boundaries or even identify errors (Andrew 2020). Improving independent
184 emission inventories also has value, as these are often used in global studies where common methods across all
185 countries are desired.

186 *Observation-based estimates* use observations of atmospheric concentrations and prior fluxes that are
187 then coupled to a transport model. These methods are more complex and computationally expensive and can
188 make use of both direct observations and emission inventories.

189 The 2019 refinement of the 2006 IPCC guidelines highlighted notable advances in the application of
190 inverse models of atmospheric transport for estimating emissions at the national scale. Building on this progress,

191 they extend the guidance on the use of atmospheric measurements for verification (IPCC, 2019). There are
192 several countries that currently use atmospheric measurements for verification of parts of their inventories.
193 Australia (Luhar et al 2020, AUS NIR, 2023) and New Zealand (Geddes et al., 2021) have estimated regional
194 CH₄ emissions to help better understand the methods and their potential. Germany performs various cross-
195 validation checks with available data (German NIR, 2023), some of which are based on observations. The UK
196 and Switzerland (Annex 6 CHE NIR, 2023) have developed more comprehensive methods based on inversion
197 modeling, covering several GHGs in addition to CH₄. Building on modeling experience, the country reporting
198 confirms that most potential lies in using observations to verify fluorinated gases (Annex 6 UK NIR, 2023), but
199 the large uncertainty in CH₄ emissions gives the potential for verification if a sufficient observation network is
200 used in inversion modeling (Bergamaschi et al., 2018, Thompson et al., 2014).

201 While inversions of CH₄ fluxes are associated with significant uncertainty, so are NGHGI estimates of
202 anthropogenic CH₄ emissions. Furthermore, inversions can provide information on subannual and subnational
203 variations in time and space that may indicate differences in source sector emission estimates. In geographic
204 areas with sufficiently dense ground-based observation networks, the inversions will have more value.
205

206 **2.2. Anthropogenic CH₄ emissions from the NGHGIs**

207

208 Annex I countries report their annual GHG emissions to the UNFCCC in the so-called Common
209 Reporting Format (CRFs) data tables and National Inventory Reports (NIRs). Here, anthropogenic CH₄
210 emissions from the five UNFCCC sectors, incl. Land Use, Land Use Change and Forestry (LULUCF) are
211 grouped together. As part of the LULUCF sector, we also have the CH₄ emissions from wetlands, which
212 according to the IPCC guidelines are defined as managed "where the water table is artificially changed (i.e.
213 lowered or raised) or those created through human activity (e.g. damming a river) and that do not fall into Forest
214 Land, Cropland, or Grassland categories (IPCC, 2014)". Reporting CH₄ emissions from managed wetlands is not
215 mandatory, but if done, parties are encouraged to make use of the 2013 IPCC Wetlands supplement (IPCC,
216 2014). In the EU, if Member States report these emissions, they report not only restored (rewetted) wetlands but
217 also emissions from drained organic and mineral soils (e.g. peatlands, ditches, etc.). These are not large by
218 magnitude but are large by area in the Nordic countries. According to NGHGI data, in 2021, managed wetlands
219 in the EU, for which emissions were reported under the LULUCF (CRF Table 4(II) and Summary 1.As2
220 accessible for each EU country), summed up to 0.21 Tg CH₄ yr⁻¹, in comparison to total emissions of ~15 Tg
221 CH₄ yr⁻¹. Furthermore, the NGHGIs do not include any lateral fluxes from inland waters but do include biomass
222 burning anthropogenic emissions reported under the LULUCF sector.

223 The presented uncertainties in the CH₄ emission levels of the individual countries and the EU are not
224 always reported in a complete and harmonized format, and therefore were calculated applying gap-filling and
225 harmonization procedures that are used to compile the EU GHG inventory reported under UNFCCC (EU NIR,
226 2023) (see SI and Appendix A1.1 in Petrescu et al., 2023a). The EU uncertainty analysis reported in the bloc's
227 National Inventory Report (NIR) is based on country-level, Approach 1 uncertainty estimates (IPCC, 2006, Vol.
228 1, Chap. 3) that are reported by EU Member States, previously under Article 7(1)(p) of Regulation (EU) 525/2013
229 and since 2023 under Article 26(3) and Annex V(Part 1)(m) of the Governance Regulation (EU) 2018/1999.

230 Non-Annex I countries report their updated NGHGs to the UNFCCC, including a national inventory
231 report and information on mitigation actions, needs and support received in Biennial Update Reports (BURs). In
232 this study, Brazil, China, Indonesia, India and the Democratic Rep. of Congo (DR Congo) were investigated. For
233 Brazil, information from its fourth BUR (Brazil, 2020) was used, giving both total and sectoral split emission
234 values for years 1994, 2000, 2010, 2012, 2015 and 2016. For China, information from its second BUR Tables 2-
235 10, 2-13, 2-14, 2-15, and 2-16 was used (China, 2019). The information was available for both total and sectoral
236 split emission values for 1994, 2005, 2010 and 2014. Uncertainties for 2014 are available in Table 2-12.
237 Indonesia submitted its third BUR in 2021 (Indonesia, 2021). Indonesian total sectoral CH₄ emissions time series
238 as reported by the 2nd BUR (2001-2016) were revised in the 3rd BUR (2000 and 2019, Table 2). For 2017 and
239 2018, only the agricultural CH₄ emissions were detailed by the 3rd BUR (Fig. 2-24), but the total emissions for
240 these two years were not reported. Uncertainty for 2019 activity data and emission factors (EFs) are the same as
241 reported in the 2nd BUR (2018). The overall uncertainty of Indonesia's National GHG inventory with agriculture
242 and LULUCF (including peat fires) for 2000 and 2019 were approximately 20.0% and 19.9%, respectively. A
243 much smaller uncertainty, 10.4 % for 2000 and 13.8 % for 2019, occurred when the forestry and land use sector
244 (including forest fires), was excluded from the analysis, pointing to the high uncertainty of emissions from forest
245 fires in Indonesia. The DR Congo submitted its first BUR in 2022, and we used timeseries reported for 2000-
246 2018 (Table 12 Congo, 2022). India has submitted three BURs and information on sectoral CH₄ emissions are
247 in each of them only for one year. We compiled information for 2010 from the first BUR (India, 2016), for 2014
248 from the second BUR (India, 2018) and for 2016 from the third and latest BUR (India, 2021).

249 **2.3. Other CH₄ data sources and estimation approaches**

250

251 The CH₄ emissions in the EU and non-Annex I countries used in the atmospheric inversions and
252 anthropogenic and natural emissions estimates from various BU approaches and inventories (i.e., UNFCCC
253 CRFs and BURs) covering specific products, sectors and activities are summarized in Table 1. The data and the
254 detailed description of most products (Tables S1 and S2, Supplementary Information) span the period from 1990
255 to 2021, with some of the data only available for shorter periods. The estimates are available both from peer-
256 reviewed literature and from unpublished research results from the VERIFY and CoCO₂ projects (Supplementary
257 Information, SI) and in this work they are compared with NGHGs reported in 2023 (time series for all (Annex
258 I) or some years (non-Annex I) of the 1990-2021 period). The BU anthropogenic sources are from UNFCCC
259 NGHGs and three global inventory datasets/models: EDGARv7.0, FAOSTAT/PRIMAP-hist 2.4 and GAINS.
260 In this synthesis, FAOSTAT (Tubiello et al., 2022; FAO, 2023) data includes estimates for all economic sectors:
261 Energy, Industrial Processes and Products Use (IPPU), Waste and Other, which are sourced from the PRIMAP-
262 hist v2.4 dataset (Gütschow et al., 2022) to build emissions indicators on agrifood systems and on the entire
263 economy. Emission totals from the agrifood domain are computed following the Tier 1 methods of the
264 Intergovernmental Panel on Climate Change (IPCC) Guidelines for NGHGs. Agrifood systems emissions in
265 FAOSTAT are largely based on FAO crop, livestock and land-use statistics (Tubiello et al., 2022; FAO, 2023).
266 They are complemented with activity data from the UN Statistics Division (UNSD), the International Energy
267 Agency (IEA) and with geospatial information on drained organic soils and biomass fires (Conchedda and
268 Tubiello, 2020; Prospero et al., 2020).

269 The analysis focuses on both total and sectoral or partitioned information from both BU and TD
 270 estimates. As detailed in Table 1, not all inversions distinguish between sources but in the following sections we
 271 discuss comparability between BU and TD for both total and partitioned results.

272 *Table 1: Sectors included in this study and data sources providing estimates for these sectors. CAMS stands for*
 273 *Copernicus Atmosphere Monitoring Service. References to data products are found in Table 2 Petrescu et al.,*
 274 *2023a and Table S1 and S2, SI.*

| Anthropogenic (BU) ¹ CH ₄ | Natural (BU) ² CH ₄ | Regional TD CH ₄ | Global TD CH ₄ |
|--|--|--|---|
| 1. Energy: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT/PRIMAP-hist 2.4 | Wetlands EU: JSBACH-HIMMELI Global: LPJ-GUESS | No partitions – total emissions FLEXkF_v2023 | Totals and partitioned emissions: |
| 2. Industrial Products and Products in Use (IPPU): UNFCCC NGHGI (CRFs and BURs), EDGAR v7.0, FAOSTAT/PRIMAP-hist 2.4 | Peatlands, mineral soils: EU: JSBACH-HIMMELI Global: LPJ-GUESS | CIF-FLEXPARTv10.4 CIF-CHIMERE | MIROC4-ACTM (control and OH varying runs) CAMSv21r1 (NOAA and NOAA_GOSAT runs) |
| 3. Agriculture: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT | Inland waters fluxes EU: lakes, rivers and reservoirs (RECCAP2) | | TM5-4DVAR (TROPOMI) |
| 4. LULUCF: UNFCCC NGHGI (CRFs and BURs) and FAOSTAT | Global: lakes and reservoirs ORNL DAAC | | CTE-CH ₄ (GCP2021) |
| 5. Waste: UNFCCC NGHGI (CRFs and BURs), GAINS, EDGAR v7.0, FAOSTAT/PRIMAP-hist 2.4 | Geological fluxes updated activity (see SI) Biomass burning (GFEDv4.1s) | | CEOS (GOSAT) GEOS-Chem CTM (TROPOMI) for USA only |

275 note: Not all models have a version id. Those that have, are used in previous syntheses (Petrescu et al., 2021 and 2023a).

276 We define natural sources as all sources which do not belong to the anthropogenic partition: wetlands,
 277 geological, termites, ocean, inland waters, soils and biomass burning (Table 2). The BU natural components for
 278 the EU were computed as the sum of the VERIFY products (biomass burning, inland waters and undisturbed
 279 peatlands plus mineral soils (as described in Petrescu et al., 2021 and 2023) and geological emissions (Etiope et
 280 al. 2019 updated for the VERIFY project). For the seven non-EU emitters, the BU natural fluxes are the sum of
 281 wetland emissions (LPJ-GUESS), lake and reservoir emissions (ORNL DAAC), biomass burning emissions

¹ For consistency with the NGHGI, here we refer to the five reporting sectors as defined by the UNFCCC and the Paris Agreement decision (18/CMP.1), the IPCC Guidelines (IPCC, 2006), and their Refinement (IPCC, 2019a), with the only exception that the latest IPCC Refinement groups together Agriculture and LULUCF sectors in one sector (Agriculture, Forestry and Other land Use - AFOLU).

² The term **natural** refers here to unmanaged natural CH₄ emissions (peatlands, mineral soils, geological, inland waters and biomass burning) not reported under the anthropogenic UNFCCC LULUCF sector.

282 (GFED4.1s) and geological emissions (updated activity, SI). The TD natural global estimates were calculated as
283 the sum of all natural partitions reported by the inversions. Adjustments were made to have a consistent
284 comparison between partitions, adding the missing ones from the BU estimates (Table 4). The error bar on the
285 TD natural represents the range of the min/max between inversion estimates.

286 The total regional TD estimates (for EU) and their uncertainties were calculated as the mean and
287 min/max range between FLEXkF_v2023, CIF-FLEXPART and CIF-CHIMERE inversions (see Priors table in
288 Petrescu et al., 2023b). For the USA, we considered the optimized emissions from the GEOS-Chem CTM (based
289 on TROPOMI data for 2019) from Nesser et al. (2023), with the range from the eight members of the inversion
290 ensemble shown as uncertainty (Table 2 in Nesser et al., 2023).

291 For all countries, the total global TD inversion estimates (time series) and uncertainties were calculated
292 over the period 2015-2021 using the mean and min/max between CTE-GCP2021, MIROC4-ACTM both runs,
293 CAMS v21r1 (both runs), and TM5-4DVAR (TROPOMI based). CEOS (GOSAT) provided an estimate only
294 for 2019.

295 The units used in this paper are metric tons (t) [$1\text{kt} = 10^9\text{ g}$; $1\text{Mt (Tg)} = 10^{12}\text{ g}$] of CH_4 . The referenced
296 data for replicability purposes are available for download at <https://doi.org/10.5281/zenodo.12582667> (Petrescu
297 et al., 2023b). Upon request, the computer code for plotting figures in the same style and layout can be provided.
298 Throughout the paper and mostly for the complex figures, the following ISO3 country codes are used: USA
299 (United States of America), BRA (Brazil), CHN (China), IDN (Indonesia), RUS (Russia), COD (DR Congo) and
300 IND (India). Next to these we also refer to CHE (Switzerland) and AUS (Australia). The European Union consists
301 of 27 Member States, excludes the United Kingdom (UK) and is further abbreviated as EU. All abbreviations
302 are summarized in the SI, Table S5.

303 **3. Results**

304

305 **3.1. NGHGI official reported estimates (UNFCCC)**

306

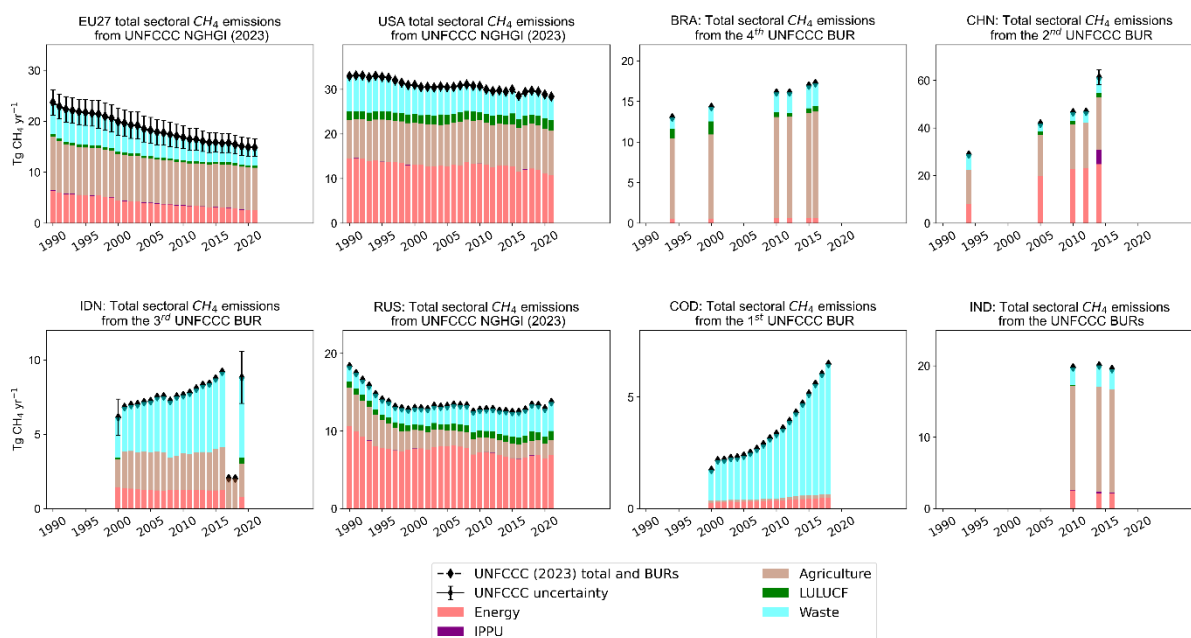
307 Figure 1 presents anthropogenic CH_4 emissions reported to the UNFCCC in 2023 from the NGHGI
308 CRFs (EU, USA and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR
309 Congo (1st in 2022) and India (all three BURs). The following section provides additional details for all the
310 countries.

311 For the EU, the total anthropogenic CH_4 emissions in 2021 amount to $14.8 \pm 1.8\text{ Tg CH}_4\text{ yr}^{-1}$ and
312 represent 12.8 % of the total EU greenhouse gas emissions (in CO_2 equivalents, GWP 100 years, IPCC AR5³).
313 CH_4 emissions are predominantly from agriculture (Figure 1, brown), which accounted for $8.3\text{ Tg CH}_4\text{ yr}^{-1} \pm 0.8$
314 $\text{Tg CH}_4\text{ yr}^{-1}$ or 56 % of the total EU CH_4 emissions in 2021 (incl. LULUCF). Anthropogenic CH_4 emissions from
315 the LULUCF sector are very small for the EU: $0.5\text{ Tg CH}_4\text{ yr}^{-1}$ or 3 % in 2021, including emissions from biomass
316 burning. The EU data from Figure 1 shows steadily decreasing trends for all sectors with respect to the 1990

³ IPCC AR4 GWP 100 values are still used by the Member States in their NGHGI reporting to the UNFCCC.

317 levels. The reduction in total CH₄ emissions in 2021 with respect to 1990 is 8.9 Tg CH₄ yr⁻¹ (37 %) at an average
 318 yearly rate of -1%.

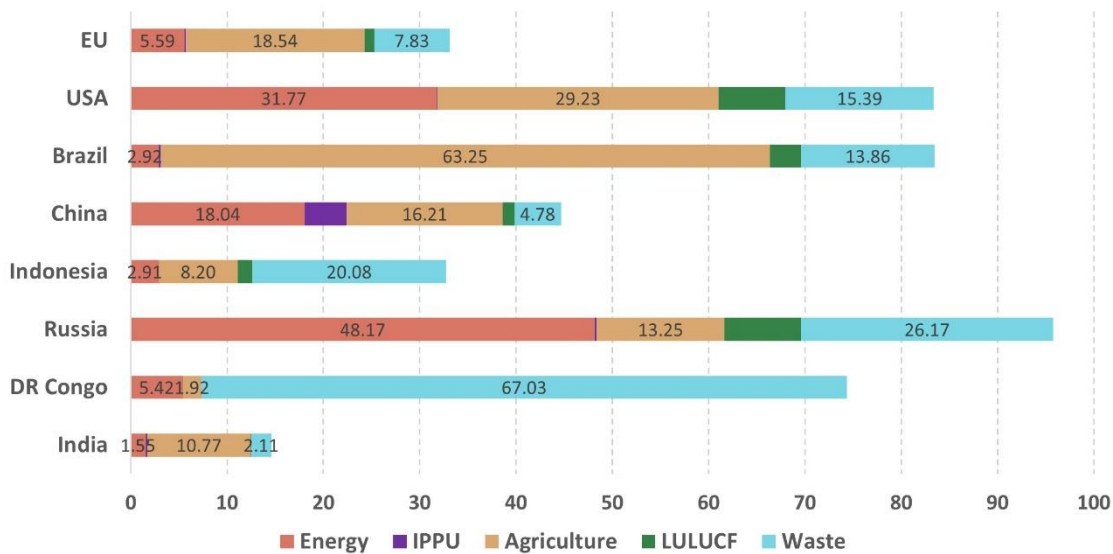
319 In 2021, the USA reported anthropogenic CH₄ emissions of 28.3 Tg and, compared to 1990, the reported
 320 USA CH₄ emissions show a small decrease of 4.6 Tg CH₄ yr⁻¹, more pronounced for the last two years (2020-
 321 2021), with an average reduction rate of -0.5 % per year (Fig. 1 black dotted line). In the USA, the largest share
 322 of emissions comes from the energy sector (38%), and next to IPPU and waste, had the highest reductions since
 323 1990 (42%, 34% and 26%, respectively). Emissions from agriculture (35%, the second largest sector) and
 324 LULUCF increased 16 % and 23%, respectively. CH₄ emissions have been slowly declining since 1990 but had
 325 a notable decrease of 1.5 Tg CH₄ yr⁻¹ in 2016 compared to 2015, before increasing again and had a second
 326 decreasing trend in 2020 and 2021, possibly due to the COVID pandemic. Overall, reported data indicates that
 327 reductions in the USA CH₄ emissions have declined more slowly than that in the EU. The EU also has much
 328 lower CH₄ emissions than the US on a per capita basis (Figure 2).



329
 330 *Figure 1: Total and sectoral CH₄ emissions (incl. LULUCF) from the UNFCCC NGHGI (2023) CRFs (EU, USA*
 331 *and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR Congo (1st in 2022)*
 332 *and India (all three BURs: 2016, 2018 and 2021). The relative error on the UNFCCC value represents the*
 333 *NGHGI (2023) reported uncertainties computed with the error propagation method (95% confidence interval)*
 334 *and gap-filled to provide respective estimates for each year. Information on Indonesian sectoral CH₄ emissions*
 335 *in 2017 and 2018 are only available for agriculture. The overall uncertainty of Indonesia's National GHG*
 336 *inventory with AFOLU (including peat fires) for 2000 and 2019 were approximately 20.0% and 19.9%,*
 337 *respectively. In 2014, China reported uncertainty as well (min 5.2 % and max 5.3 %).*

338 The trend in total anthropogenic CH₄ emissions in *Brazil* is strongly increasing, with 32.5 % more
 339 emissions in 2016 compared to 1994. Given that the Brazilian BUR inventory does not include data between
 340 2001 and 2019, it is difficult to discuss the yearly growth rates. We can only note that the agricultural sector (76
 341 % of the total) was the main driver of the growth, followed by the waste sector (16 % of the total). There are

342 only small CH₄ emissions from the energy sector (some oil and gas activities). The Brazilian agricultural CH₄
 343 emissions are the highest of the eight countries on a per capita basis (see Figure 2).



344

345 *Figure 2: Per capita emissions (kg) intensity per sector based (IPCC, 2006) on reported emissions and*
 346 *population data from the last reported NIRs as follows: 2021 for the EU, USA and Russia, 2016 for Brazil and*
 347 *India, 2014 for China, 2019 for Indonesia and 2018 for DR Congo.*

348 *China's total anthropogenic CH₄ emissions are much larger than the emissions reported by many*
 349 *developed countries or the entire EU (see Figure 1), but on a per capita basis it is only fifth of the eight countries*
 350 *considered (Figure 2). China's CH₄ emissions have grown 114% from 1995 to 2014, when they reached 32 Tg*
 351 *CH₄. The highest contributions to China's CH₄ emissions are from energy (40%) and agriculture*

352 *(36%). The rapid growth of China's coal demand has important implications for CH₄ emissions from*
 353 *coal mines (Gao et al., 2020). The energy and agriculture sectors have respectively increased by 214 % and 54*
 354 *% in 2014 compared to 1994.*

355 *Indonesia's 3rd BUR data (2000 and 2019) show increasing trends in total anthropogenic CH₄ emissions.*
 356 *The time series 2001-2006 derives from the 2nd BUR submitted in 2018. In 2019, Indonesian CH₄ emissions had*
 357 *increased by +44 % compared to 2000, corresponding to 2.6 Tg CH₄ yr⁻¹, an average yearly increase of 3 %, and*
 358 *the sector which contributes the most to this increase is the waste sector, which nearly doubled its emissions in*
 359 *2019 compared to 2000. According to Qonitan et al., 2021, the major solid waste source in Indonesia is the*
 360 *household sector, which contributed 44-75% to total waste generated. The composition of municipal waste*
 361 *consists of 44% food waste, 16% paper, and 14% plastics. CH₄ emissions from the other sectors remained nearly*
 362 *constant. For 2017 and 2018 the Indonesian 3rd BUR does not report total emissions other than agricultural*
 363 *emissions, which were taken from the report Figure 2-24. The last data point (2019) shows lower total emissions*
 364 *because it belongs to the revised versions of the (3rd) BUR while the previous data points 2000-2016 belong to*
 365 *the 2nd BUR.*

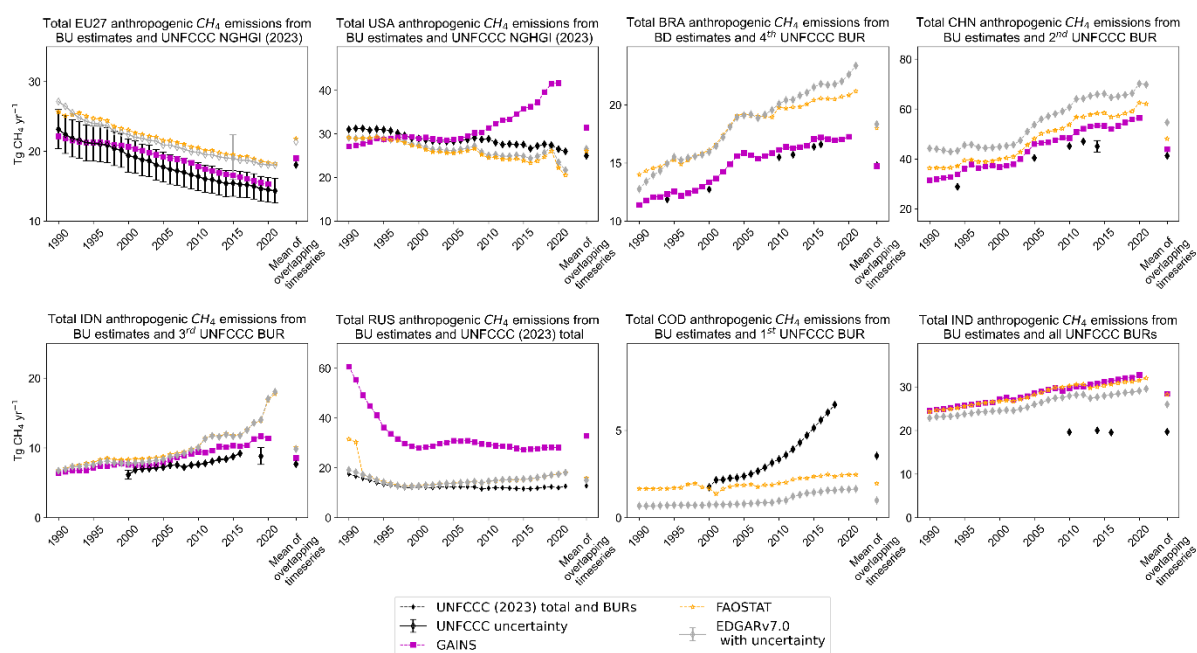
366 *Russia's* anthropogenic CH₄ emissions have decreased by -25 % from 1990 to 2021, but most of this
367 decrease happened during the dissolution of the Soviet Union. Since 2000, CH₄ emissions have increased
368 slightly, but remain lower than pre-2000 levels. The decline seen between 1990 and 2000 is primarily due to the
369 agricultural sector (-52 %) and energy (-27 %). At the same time, the waste sector started to increase its emissions
370 (6 %). Between 2001 and 2021, the CH₄ emissions from the agriculture and energy sectors continue to decrease
371 (by 17 % and 11 %, respectively), while the emissions from the waste sector register an additional 76 % increase.
372 IPPU emissions increased by 85 % but remain negligible compared to other sectors. Since the 2000s, LULUCF
373 emissions have also increased, by 53 %.

374 For its first BUR, *DR Congo* submitted emissions from energy, AFOLU (agriculture plus LULUCF)
375 and waste for 2000-2018. Since 2000, the DR Congo total anthropogenic CH₄ emissions have increased by a
376 factor of four. Most of the CH₄ emissions are reported for the waste sector, and account for 90 % of the total
377 emissions. The high percentage of waste emissions in DR Congo is also seen in the per capita emissions (Figure
378 2). Assè-Wassa Sama and Berenger (2023) confirm that between 2000 and 2021, CH₄ emissions, which in 2021
379 represent in DR Congo ~97% of total waste generated emissions, grew at a rate of 4 % yr⁻¹, compared with 2.7
380 % yr⁻¹ for total emissions. This increase was driven by the increase in emissions caused by solid waste disposal
381 (+6.2 %). The CH₄ waste emissions come mainly from the treatment and discharge of wastewater (69 % in 2021,
382 compared with 80 % in 2000), followed by the elimination of solid waste (31 % in 2021, compared with 20 %
383 in 2000). The weight of emissions caused by the elimination of solid waste in the sector's total emissions has
384 nevertheless increased by 11 percentage points between 2000 and 2021 (Assè-Wassa Sama and Berenger, 2023).

385 Each of *India's* BURs provide detailed information on sectoral CH₄ emissions only for one year. Most
386 of the emissions in India belong to the agriculture sector, amounting to almost 15 Tg CH₄ yr⁻¹ (in 2016),
387 representing 74 % of the total anthropogenic emissions. However, with only three years of reported data
388 available, no clear or notable trend is observed.

389 **3.2. NGHGI compared to other bottom-up estimates** 390

391 Figure 3 shows UNFCCC (CRFs and BURs) estimates from EU and seven non-EU countries compared
392 to global bottom-up inventories. The EU and USA show decreasing trends in emissions from all data sets (except
393 for GAINS in the USA), while all the other countries show increasing trends in all datasets. The match between
394 UNFCCC reported emissions and all other data sources is satisfactory, with a few notable exceptions.



395

396 *Figure 3: Total anthropogenic CH₄ emissions (excl. LULUCF) from bottom-up (BU) inventories as: UNFCCC*
 397 *NGHGIs (2023) of CRFs (EU, USA and Russia) and BURs (Brazil (4th in 2021), China (2nd in 2019), Indonesia*
 398 *(3rd in 2021), DR Congo (1st in 2022), India (all three BURs:2016, 2018 and 2021) and three other global*
 399 *datasets: EDGARv7.0, GAINS (no IPPU) and FAOSTAT (PRIMAP based, except for AFOLU). The relative error*
 400 *on the UNFCCC value represents the NGHGI (2023) reported uncertainties computed with the error*
 401 *propagation method (95% confidence interval) and gap-filled to provide respective estimates for each year.*
 402 *China and Indonesia report uncertainties, for 2014 and 2000 and 2019 respectively (BUR). Total COD*
 403 *UNFCCC BUR emissions do not include IPPU. The EDGARv7.0 uncertainty is only for 2015 and was*
 404 *calculated according to Solazzo et al., 2021 for EDGARv5.0. The mean of overlapping time series was calculated*
 405 *for 1990-last available year as following: 2021 for UNFCCC NGHGI (2023), EDGARv7.0 and FAOSTAT and*
 406 *2020 for GAINS.*

407 For the EU, the difference between the UNFCCC NGHGI 1990-2020 average and the other three data
 408 sets is less than 5%. As previously discussed, the inventory-based data sources are consistent with each other
 409 for capturing recent CH₄ emission reductions, but they are not independent because they use similar methodology
 410 with different versions of the same activity data (AD) (Petrescu et al., 2020, Figure 4).

411 For the USA, GAINS reports high emissions after 2010, with strong growth. This divergence is largely
 412 found in the energy sector, resulting from the EFs used for conventional gas production as well as for
 413 unconventional shale gas extraction, which has increased rapidly since 2006 due to the development of hydraulic
 414 fracturing technology (Supplementary Figure S6-1 in Höglund-Isaksson et al., 2020). The high share of
 415 emissions from unconventional shale gas can be explained by the GAINS EFs which, in the absence of published
 416 factors, are derived from the residual emissions after having subtracted estimated emissions for oil production
 417 and conventional gas production from the total upstream emission estimated by Alvarez et al., (2018, Table 1)
 418 As Alvarez et al. 2018 do not specify EFs by type of gas produced, GAINSv4 splits it based on activity data from
 419 other references, International Energy Agency-World Energy Outlook (IEA-WEO, 2018) and Energy
 420 Information Administration (EIA, 2019). On the other hand, the NGHGI EF seems to be too low, and this is

421 reflected by the low oil and gas emissions reported by the United States Environmental Protection Agency
422 (USEPA 2017) for 2015, compared to Alvarez et al., 2018 (Supplementary Table S6-3, Höglund-Isaksson et al.,
423 2020). For the USA, total gas production increased by 47 % between 2006 and 2017. Revisions for the
424 agricultural livestock emissions concern updates of AD and reported EFs to statistics from FAOSTAT (2018)
425 and CRFs (UNFCCC 2016; 2018), and a review of available technical abatement options for CH₄.

426 For *Brazil*, UNFCCC and GAINS report emissions of similar magnitudes and trends. The EDGARv7.0
427 and FAOSTAT report on average around 23 % more emissions for the 1990-2021 period, but closely follow the
428 NGHGI trends. The similarity between trends could be explained by the use of the same EFs following Tier-1
429 IPCC 2006 Guidelines and UNFCCC NIRs (Janssens-Maenhout et al., 2019), while the higher emissions could
430 appear when using different AD information.

431 For *China* the inventory estimates agree reasonably with the BUR reported data, with EDGARv7.0
432 showing the highest estimates. According to both GAINS and EDGARv7.0, the primary drivers for growth in
433 Chinese CH₄ emissions are due to a mix of sources, mainly from the IPCC 2006 sector 1.B.1, fugitive emissions
434 from solid fuels activity linked to increased coal mining.

435 In *Indonesia* the three global datasets agree well up until 2010. From 2010, the third-party datasets show
436 a continued increase in emissions, while the UNFCCC BUR emissions suggest a decline. EDGARv7.0 reports a
437 large increase in emissions from fugitive emissions from solid fuels (coal mining) (IPCC 2006, sector 1.B.1.) at
438 an increased average rate of 19 % per year and has increased by a factor of 152 until 2021 compared to 1990
439 (Figure 3).

440 For *Russia*, GAINS emissions are much higher than NGHGI and the other two data sets due to the
441 revisions of the assumptions on the average composition of the associated gas generated from oil production
442 based on information provided in Huang et al. (2015). The higher emissions in GAINsv4 might be caused by a
443 greater source from venting of associated gas instead of flaring. GAINsv4 estimates a decline in global CH₄
444 emissions in the first half of the 1990s, primarily a consequence of the dissolution of the Soviet Union and the
445 associated general decline in production levels in agriculture and fossil fuels (see regional emission illustrations
446 in figures S2–1 of the SI). In addition, as described by Evans and Roshchanka (2014) and assumed in Höglund-
447 Isaksson (2017), venting of associated petroleum gas declined significantly in Russia due to an increase in flaring.
448 It is unclear why this happened, but a possible explanation could be that the privatization of oil production in
449 this period meant that the new private owners were less willing to take the security risks of venting and invested
450 in flaring devices to avoid potential production disruptions. This hypothesis is however yet to be confirmed
451 (Höglund-Isaksson et al., 2020). FAOSTAT data for the Russian Federation starts in 1992, but the former USSR
452 statistics were used prior to 1992 without adjustments and this is the cause of the 1990 and 1991 outliers in time
453 series. The slightly increasing trend observed in EDGARv7.0 and FAOSTAT are set by emissions from the
454 energy sector.

455 For *DR Congo* estimates from GAINS are not available because they only report aggregated emissions
456 from a few African regions. Both FAOSTAT (PRIMAP based) and EDGARv7.0 estimates show similar slowly
457 increasing trends, potentially indicating the use of similar prior statistics (EFs). For non-AFOLU sectors the
458 PRIMAP-hist third party data priority scenario used in FAOSTAT also uses EDGAR data as an input data source

459 explaining similarities in these sectors. On the other hand, UNFCCC BUR data reports a strong increase in
460 emissions, which is due to a rapid growth of CH₄ emissions from the waste sector, by a factor of four until 2018
461 compared to 2000. This increase happened at an average yearly rate of +8 %, with an initial sharp increase of
462 +30 % between 2000 and 2001. As previously discussed, (section 3.1.) we believe that DR Congo BUR reported
463 waste emissions are improbable and further investigation is needed.

464 For *India*, all bottom-up global inventories show similar trends and magnitudes of anthropogenic CH₄
465 emissions. The emissions of CH₄ averaged across EDGARv7.0, GAINS and FAOSTAT are 67% (2010), 68 %
466 (2014) and 65 % (2016) higher than the Indian BURs. All three BU inventories show an averaged steady increase
467 of 1 % yr⁻¹ between 1990-2020.

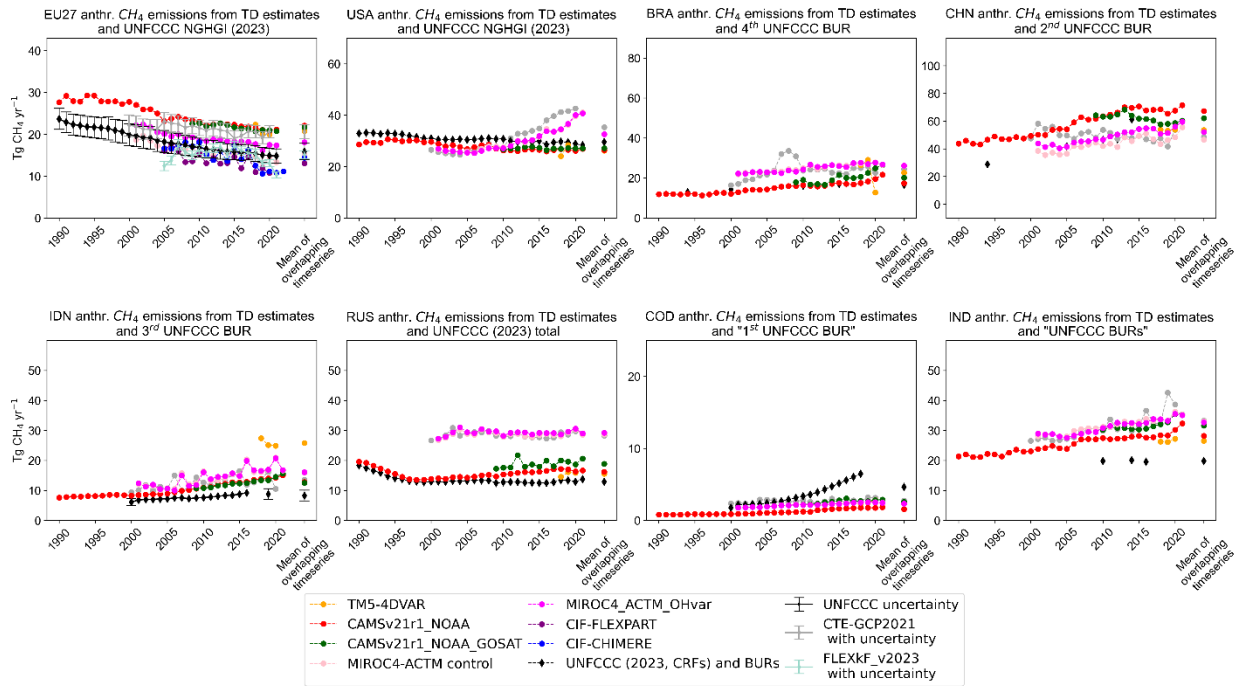
468 **3.3. NGHGs compared to TD atmospheric-based CH₄ estimates**

469

470 In Figure 4 we compare the reported TD anthropogenic estimates, after removing natural emissions,
471 with the UNFCCC official reported emissions for the EU and the seven non-EU emitters. The mean column on
472 the right of each chart represents the mean of the overlapping time series (2009-last available year, except for
473 TROPOMI, which was available only for 2018-2020). For the EU, the three regional inversions report total
474 emissions because they do not separate anthropogenic from natural emissions. Therefore, we subtracted from the
475 total the natural emissions as calculated in Petrescu et al., 2023a which amount to 6.6 Tg CH₄ yr⁻¹ and are the
476 sum of biomass burning, wetlands, geological and inland water CH₄ emissions. For the global inversions, the
477 anthropogenic estimates were calculated by subtracting from the total fluxes the reported natural partitions as
478 follows: for the two CAMS inversions and TM5-4DVAR (TROPOMI based) the sum of biomass burning and
479 wetlands, for MIROC4-ACTM runs the natural is the represented by the sum of the biomass burning, geologic,
480 ocean, termites, soils and wetlands, for CTE-GCP2021 the sum of the biologic (wetlands + soils) and other
481 (ocean, termites, geological). Because not all inversions report the same partitions, we consider this a coarse
482 comparison, and we detail the harmonization of the natural emissions in the next section (Table 4 and Figure 7).
483 For China, the last BUR is available for 2014, and therefore we used that value.

484

485



486

487 *Figure 4: Anthropogenic CH₄ emissions from UNFCCC NGHGI (2023) CRFs (EU, USA and Russia)*
 488 *(Brazil (4th in 2021), China (2nd in 2019), Indonesia (3rd in 2021), DR Congo (1st in 2022), India (all three BURs:*
 489 *2016, 2018 and 2021) and TD estimates as following: for EU anthropogenic regional inversions*
 490 *(FLEXkF_v2023, CIF-FLEXPART and CIF-CHIMERE) and for global inversions anthropogenic estimates*
 491 *calculated as total TD minus natural TD reported partitions (TM5-4DVAR, CAMSv21r1_NOAA,*
 492 *CAMSv21r1_NOAA_GOSAT, CTE-GCP2021 and MIROC4-ACTM both runs) products. The relative error on the*
 493 *UNFCCC value represents the NGHGI (2023) reported uncertainties computed with the error propagation*
 494 *method (95% confidence interval) and gap-filled to provide respective estimates for each year. China reports*
 495 *uncertainties for 2014 (min 5.2 %, max 5.3 %) and Indonesia reports for 2000 and 2019, 20 % and 19.9 %*
 496 *respectively. Total COD UNFCCC BUR emissions do not include IPPU. The last available years are CIF-*
 497 *CHIMERE (2022), TM5-4DVAR, CIF-FLEXPART and CTE-GCP2021 (2020) and FLEXkF_v2023, MIROC4-*
 498 *ACTM both runs, UNFCCC CRFs, and CAMSv21r1 both runs (2021). The mean of overlapping time series was*
 499 *calculated for 2009-2021, except for TM5-4DVAR (2018-2020).*

500

501 In the EU, the average anthropogenic CH₄ emissions from global inversions for 2009-2021 were 19 Tg
 502 CH₄ yr⁻¹ with a min-max range of 15-23 Tg CH₄ yr⁻¹, in line with previous estimates published in Petrescu et al.
 503 (2021, 2023a) and the recent RECCAP2 European GHG budgets study of Lauerwald et al., 2024. This is
 504 consistent with the UNFCCC NGHGI (2023) which report for the same period anthropogenic emissions of 15.8
 505 ± 1.8 Tg CH₄ yr⁻¹, noting the uncertainty ranges of both estimates overlap. There is good agreement in trends,
 506 but with inversions showing a larger year to year variability. The regional inversions, for the same period, report
 507 averaged emissions of 14 Tg CH₄ yr⁻¹ with a min-max range of 13-20 Tg CH₄ yr⁻¹. We note that the regional
 508 inversions tend to report slightly lower emissions than the global inversions, closer to the UNFCCC estimates.
 509 One reason could be that regional inversions use better-constrained regional observations (e.g. ICOS, not just

510 NOAA), have higher spatial resolution, and may thus better resolve the transport. However, they may still have
511 problems with the regional boundary conditions.

512 For the *USA*, averaged over the period 2009-2021, inversions indicate anthropogenic CH₄ emissions of
513 30 Tg CH₄ yr⁻¹ with min-max range of 26-35 Tg CH₄ yr⁻¹, well in line with the UNFCCC NGHGs (2023) which
514 for the same period report anthropogenic total emissions of 29 Tg CH₄ yr⁻¹. The trends observed in TD products
515 are slightly increasing after 2010, except for CAMS which shows no trend (Figure 4). The striking discrepancy
516 between the trends from CAMS and those from MIROC4-ACTM and CTE-GCP2021 are most likely caused by
517 the increasing oil and gas emissions from the Eastern USA (Permian Basin). The same increasing trend is also
518 captured by GAINS (Figure 3). In their runs, both MIROC4-ACTM and CTE-GCP2021 use oil and gas priors
519 from GAINS, while CAMS uses priors from EDGAR (Figure 3). We discuss further differences in having CTE-
520 GCP2021 run with both EDGAR and GAINS oil & gas prior estimates in the SI.

521 For *Brazil*, inversions yield an average (range) of anthropogenic CH₄ emissions of 23 (17-27) Tg CH₄
522 yr⁻¹, slightly higher than the UNFCCC estimate of 16.6 Tg CH₄ yr⁻¹. The two CAMS inversions have trends
523 which match the trend of the UNFCCC reports estimates.

524 For *China*, approximately 80 % of the CH₄ emission increase (21.5 Tg yr⁻¹) during 2000 – 2015 was
525 from fugitive emissions from coal (mines), consistent with what GAINS and EDGAR report (Figure 3). The TD
526 estimates mostly agree with the BURs, except for CAMS inversions which show 10 to 20 Tg CH₄ yr⁻¹ higher
527 emission than the other inversions. Both MIROC4-ACTM runs (control and OH inter-annual variability (IAV)
528 varying run; Patra et al., 2021) are in line with the BURs. Trend wise, all inversions agree on increased emissions
529 after 2019, with a slight decrease after 2013 which picks up again after 2018 seen in all inversion trends.

530 For *Indonesia*, most TD results agree on the trend and show a slight increase in emissions. A similar
531 trend is also seen by the BURs. However, the CAMS inversions show linear increased trends while the other
532 inversions have a more variable trend. Regarding the East Asian estimates, MIROC4_ACTM inversion simulates
533 higher fluxes compared to the other inversions. Only recently they found that annual total East Asian emissions
534 have lowered more significantly than in Patra et al. (2016) or Chandra et al. (2021), therefore new runs with
535 updated input set-ups are currently being investigated.

536 For *Russia*, the estimates from MIROC4-ACTM runs and CTE-GCP2021 are both in the same range as
537 the BU GAINS estimate (see Figure 2) from 2000 onwards (between 30-40 Tg CH₄ yr⁻¹) but does not show such
538 a strong decrease as GAINS from 1990 to 2000), while CAMS runs report about 10 Tg CH₄ yr⁻¹ lower emissions
539 than the other two inversions, matching the UNFCCC estimates. The reason for higher estimates reported by
540 CTE-GCP2021 and MIROC4-ACTM is most likely the use of oil and gas priors from GAINS.

541 For *DR Congo*, inversions show the same slightly increasing trend, similar to that of UNFCCC BURs,
542 without the abrupt increase after 2010. The inversions appear to confirm the overreported growth in emissions
543 from waste.

544 For *India*, all the TD anthropogenic estimates agree well on increased trends and magnitudes. In
545 contrast, UNFCCC reporting does not show any trend, but given the insufficient data from BURs, a plausible
546 conclusion cannot be drawn.

547

548

549 **3.4. Sectoral attribution of CH₄ emissions in TD products**

550

551 In some cases, inversions can be used to partition emissions to different sources. Table 2 shows the
 552 partitions as originally reported by some of the inversions, which we name here “unharmonized partitions”. A
 553 straightforward, direct comparison of the fluxes is not possible because of the different ways each inversion
 554 allocates and groups the natural/anthropogenic fluxes. For example, not all inversions report soil fluxes as done
 555 by MIROC4-ACTM and CTE-GCP2021 (together with wetlands) or report the biomass burning fluxes separately
 556 from anthropogenic emissions (MIROC4-ACTM and TM5-4DVAR). Rice is also sometimes allocated to natural
 557 emissions. Termites, oceans and geological fluxes are sometimes reported separately (MIROC4-ACTM) or
 558 grouped in “Other” (CTE-GCP2021, TM5-4DVAR). Regarding the anthropogenic emissions, TM5-4DVAR
 559 reports them as other, providing a separate partition for rice. Figure 5 shows the UNFCCC NGHGI anthropogenic
 560 total reported estimate (diamond) next to all TD estimates. All global inversions report total and disaggregated
 561 partitions, while the regional inversions report only the total emissions (green column).

562

563 *Table 2: Unharmonized partitions originally reported by inverse products:*

| Inversion | Anthropogenic | Rice | Soils | Wetlands | Ocean | Termites | Geological | Biomass burning | Other |
|---|--|-----------------|-----------|----------|----------------|----------------|----------------|-------------------------|-----------------------------------|
| CAMSv21r1 (both runs) | Yes (as Other) | Yes | No | Yes | Yes (in Other) | Yes (in Other) | No | Yes | Yes** |
| MIROC4-ACTM (control and OH var) | Yes ((agr, waste, oil/gas, biofuel, coal) | Yes (in agr.) | Yes | Yes | Yes | Yes | Yes | Yes | Yes (separated) |
| CTE-GCP2021* | Yes (agr, waste, fossil fuel, biofuel, biomass burning) | Yes (in agr.) | Yes (BIO) | | Yes (as Other) | Yes (as Other) | Yes (as Other) | as anthr. | Yes (Ocean, Termites, Geological) |
| CEOS (GOSAT) | Yes (livestock, rice, waste, coal, oil, fire) | as anthr. | No | Yes | No | No | Yes (seeps) | as anthr.(but separate) | only seeps |
| TM5-4DVAR (TROPOMI) | Yes (as Other) | Yes | No | Yes | Yes (as other) | Yes (as other) | Yes (as Other) | Yes | Yes** |
| GEOS-Chem CTM (TROPOMI for USA) | Yes (livestock, oil/gas, landfills, wastewater, Other anthro. (rice) | as other anthr. | No | Yes | Yes (as Other) | Yes (as Other) | Yes (as Other) | Yes (as Other) | Yes*** |

564 *CTE-GCP2021 partitions refer to anthropogenic, bio and other.

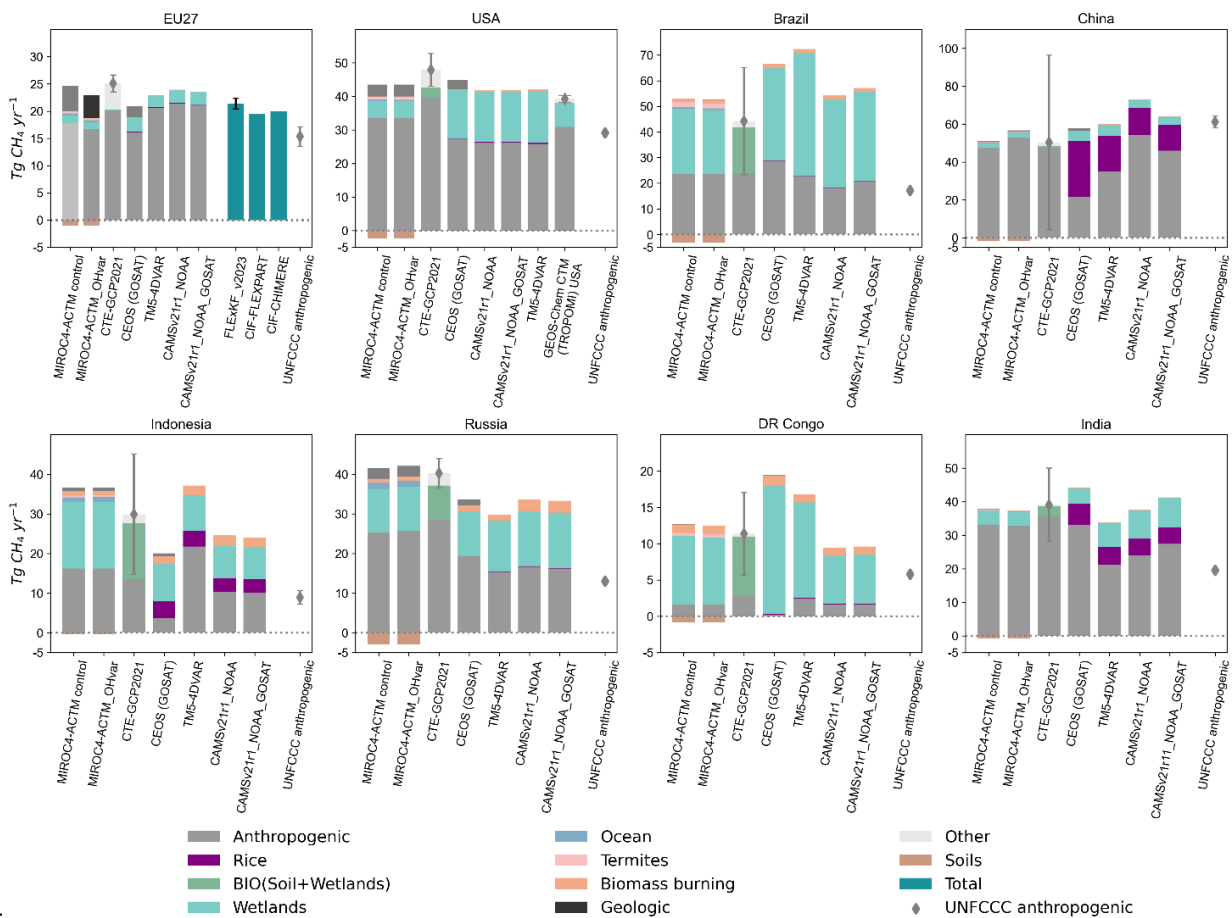
565 ** In TM5-4DVAR (similar to the CAMSv20 set-up and CAMSv21r1), the "Other" partition includes anthropogenic sources
 566 except for the rice paddies. It also includes the small fluxes from termites, oceans, soil sink, geological etc.). More details on
 567 priors are found in Petrescu et al., 2023b, Priors table.

568 ***Named Other biogenic

569

570

UNFCCC anthropogenic and unharmonized CH_4 emissions from partitions reported by TD estimates (average 2015-last available year)



572 Figure 5: Total (green) and disaggregated anthropogenic and natural CH_4 emissions from TD estimates
 573 compared to UNFCCC NGHGI (2023) anthropogenic emissions (incl. LULUCF) (diamond) for the EU and
 574 seven global emitters outside the EU (USA, Brazil, China, Indonesia, Russia, DR Congo and India). The
 575 UNFCCC anthropogenic value represents the sum of all five IPCC sectors (Energy, IPPU, Agriculture,
 576 LULUCF and Waste). The partitions reported by the TD global inversions are detailed in Table 2. The relative
 577 error on the UNFCCC CRF value represents the NGHGI (2023) reported uncertainties computed with the error
 578 propagation method (95% confidence interval) and gap-filled to provide respective estimates for each year (see
 579 Petrescu et al., 2023a, Appendix). China value and uncertainties (min 5.2 %, max 5.3 %) are for 2014 only and
 580 Indonesia uncertainties for 2019, 19.9 %. For the USA CEOS (GOSAT) we used the Nesser et al., 2023 total
 581 uncertainty of min 1.1 and max 1 $Tg\ yr^{-1}$. CTE-GCP2021 provides uncertainties for each partition, but here the
 582 uncertainty of the total flux is shown. FLEXkF_v2023 reports the relative uncertainty (%) of the posterior
 583 emissions. The plotted data represents the average between 2015 and last available year as follows: CIF-
 584 CHIMERE (2022), TMS-4DVAR, CIF-FLEXPART and CTE-GCP2021 (2020) and FLEXkF_v2023, MIROC4-
 585 ACTM both runs, UNFCCC CRFs, and CAMSv21r1 both runs (2021). GEOS-Chem CTM (TROPOMI) USA
 586 reports only for 2019 (Nesser et al., 2023).

587 Since the different models define sectors differently, also whether they are natural or anthropogenic,
 588 harmonization is required to make them comparable. CTE-GCP2021 reports the net natural land-biosphere flux

589 “BIO flux” (soil+wetlands), while other inversions report wetlands and soil separately. Rice emissions are
590 sometimes a part of the agriculture component (anthropogenic partition) (MIROC4-ACTM, CTE-GCP2021)
591 while CEOS (GOSAT) and GEOS-Chem CTM (USA TROPOMI) report separate partitions for rice in
592 anthropogenic emissions, while CAMS reports rice separate from anthropogenic and natural. Same for the
593 biomass burning - CTE-GCP2021 and CEO report it as part of anthropogenic emissions, while GEOS-Chem
594 CTM as part of Others. The rest of the inversions report it separately; this different allocation makes comparisons
595 for these two sources challenging. To facilitate comparisons between all TD products, we aggregated and
596 harmonized the partitions in three main categories, as summarized in Table 3 and Figure 6. The dark green
597 columns in Figure 6 show the total flux for regional EU inversions which did not report partitions.

598

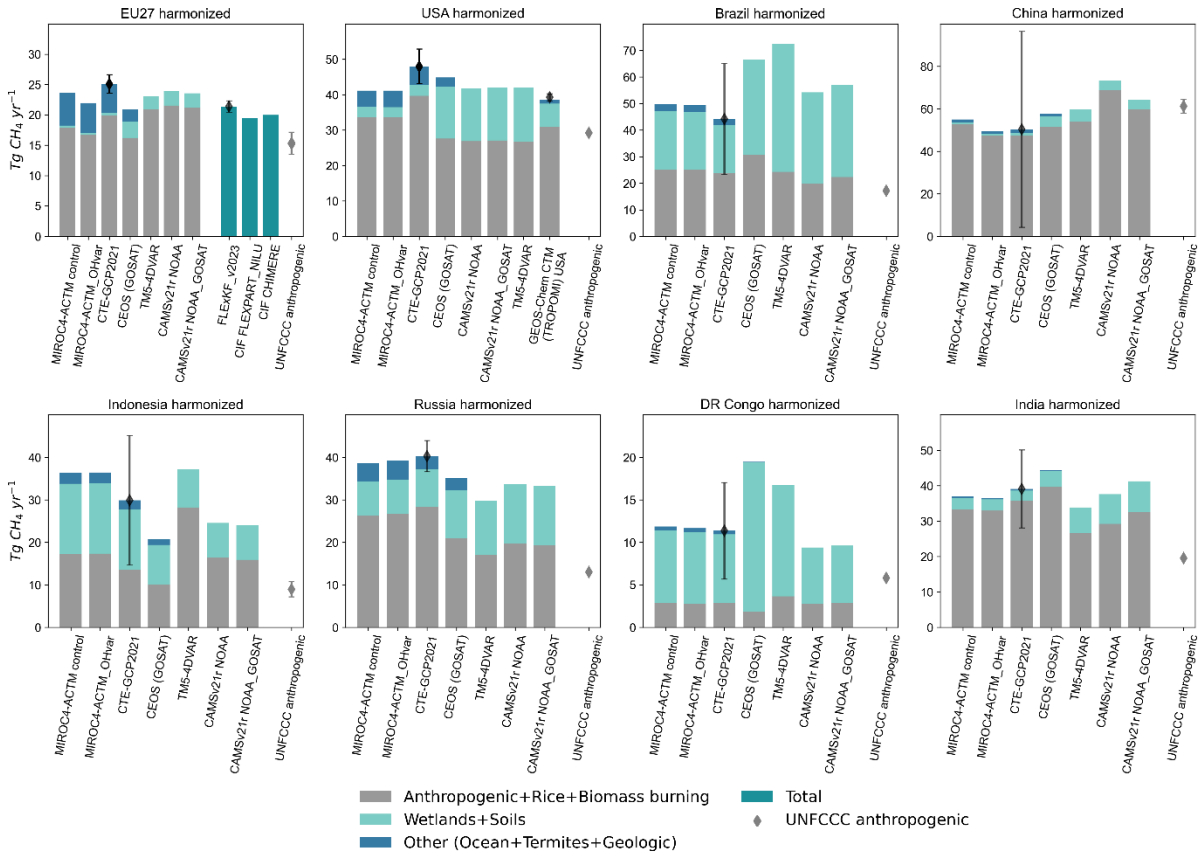
599 *Table 3: Harmonized partitions from inverse products:*

| Inversions | Anthropogenic + Rice + Biomass burning | | | Soils + Wetlands | | Other (Ocean + Termites + Geological) | | |
|----------------------------------|--|-----------|-----------------------|------------------|----------|---------------------------------------|----------|------------|
| | Anthropogenic | Rice | Biomass burning | Soils | Wetlands | Ocean | Termites | Geological |
| CAMsv21r1 (both runs) | = Other | Yes | Yes | No | Yes | Yes | Yes | Yes |
| MIROC4-ACTM (control and OH var) | Yes ((agr (livestock + rice), waste, oil/gas, biofuel, coal) | In agr. | Yes, summed to anthr. | Yes | Yes | Yes | Yes | Yes |
| CTE-GCP2021* | Yes (agr (rice is in), waste, fossil fuel, biofuel, biomass burning) | in agr. | In anthr. | Yes (BIO) | | Yes (Other) | | |
| CEOS (GOSAT) | Yes (livestock, rice, waste, coal, oil, fire) | In anthr. | In anthr. | No | Yes | No | No | Yes |
| TM5-4DVAR (TROPOMI) | Others + rice+ BB | In anthr. | Yes, summed to anthr. | In Other | Yes | Yes | Yes | Yes |
| GEOS-Chem CTM (TROPOMI) USA | Yes | In anthr. | In other biogenic | No | Yes | Yes | Yes | Yes |

600 *CTE-GCP2021 partitions refer to Anthropogenic, Bio and Other. Other fluxes are imposed

601

UNFCCC anthropogenic and harmonized CH_4 emissions from partitions reported by TD estimates (average 2015-last available year)



602

603 *Figure 6: Total (green) and disaggregated anthropogenic and natural CH_4 emissions from TD estimates*
 604 *compared to UNFCCC NGHGI (2023) anthropogenic emissions (incl. LULUCF) for the EU and seven global*
 605 *emitters (USA, Brazil, China, Indonesia, Russia and DR Congo). The UNFCCC anthropogenic value represents*
 606 *the sum of all five IPCC sectors (Energy, IPPU, Agriculture, LULUCF and Waste). The partitions reported by*
 607 *the TD global inversions are harmonized and detailed in Table 3. The relative error on the UNFCCC CRF value*
 608 *represents the NGHGI (2023) reported uncertainties computed with the error propagation method (95%*
 609 *confidence interval) and gap-filled to provide respective estimates for each year (see Petrescu et al., 2023a,*
 610 *Appendix). In 2014, China UNFCCC value and reported uncertainties (min 5.2 % and max 5.3 %) are for 2014*
 611 *while Indonesia reported uncertainties for 2019, 19.9 %. India UNFCCC value is for 2016. CTE-GCP2021*
 612 *provides uncertainties for each partition, but here we plotted the uncertainty of the total flux. FLEXkF_y2023*
 613 *reports the relative uncertainty (%) of the posterior emissions. The plotted data represents the average between*
 614 *2015 and last available reported year as follows: CIF-CHIMERE (2022), UNFCCC CRFs, TMS-4DVAR, CIF-*
 615 *FLEXPART and CTE-GCP2021 (2020) and FLEXkF_y2023, MIROC4-ACTM both runs, and CAMSv21r1 both*
 616 *runs (2021). GEOS-Chem CTM (TROPOMI) USA reports only for 2019 (Nesser et al., 2023).*

617 3.5. Comparison of BU and TD CH_4 estimates

618

619 Figure 7 summarizes the total CH_4 fluxes for the EU and the seven global emitters as following: BU
 620 anthropogenic sources disaggregated per sector, BU natural emissions, TD natural emissions from regional and

621 global inversions, and total emissions from global TD estimates (see 2.3 and SI for description of all data
 622 products). This figure brings all the estimates together to demonstrate the reconciliation process.

623 Inversions currently report in a way that makes comparison between BU natural and TD natural sources
 624 difficult. TD products differ in the sources they report (Table 2) or they allocate them to different categories. We
 625 consider natural the following sources: biomass burning, soils, oceans and termites (often reported by inversions
 626 under category “Other”), wetlands, geological and lakes & reservoirs (or freshwaters). Due to lack of
 627 information, biomass burning emissions were considered among the natural sources, recognizing that in regions
 628 like tropical forests, some of these events are influenced by human intervention. To make the products from
 629 Figure 7 comparable, we added the missing BU information from TD, and vice-versa, presented in hatched
 630 pattern. In this way, comparison between BU and TD natural emission estimates is consistent regarding the
 631 “apples to apples” comparison, but became “apples of different flavors” (see Table 4):

632
 633 *Table 4: BU and TD natural partitions as presented in Figure 7:*

| Product name | TD natural partitions | | | |
|--|---|---|-----------------------------|---|
| | Reported | Missing* (not in priors) | Not reported** | Missing was added from: |
| CAMsv21r1_NOAA | BB, wetlands, “Others” include anthropogenic and was not used | lakes and reservoirs, geological | termites, oceans, soil sink | DAAC lakes and reservoirs, geological, updated in this study (see SI) |
| CAMsv21r1_NOAA_GOSAT | BB, wetlands, “Others” include anthropogenic and was not used | lakes and reservoirs, geological | termites, oceans, soil sink | DAAC lakes and reservoirs, geological, updated in this study (SI) |
| MIROC4-ACTM control | BB, wetlands, oceans, termites, soils, geological | lakes and reservoirs | | DAAC lakes and reservoirs |
| MIROC4-ACTM_OHvar | BB, wetlands, oceans, termites, soils, geological | lakes and reservoirs | | DAAC lakes and reservoirs |
| CTE-GCP2021 | soils + wetlands (BIO), termites and oceans | lakes and reservoirs | BB, geologic | DAAC lakes and reservoirs |
| CEOS (GOSAT) | Fires (BB), Seeps and wetlands | termites, oceans, soils, lakes and reservoirs | | MIROC4-ACTM (termites, oceans and soils), DAAC lakes and reservoirs |
| TM5-4DVAR (TROPOMI) | BB and wetlands | lakes and reservoirs, geological, | termites, oceans, soil sink | DAAC lakes and reservoirs, geological, updated in this study (see SI) |
| Product name | BU natural partitions | | | |
| | | Reported | Not reported** | Added from |
| Biomass burning Lakes and reservoirs Wetlands | | GFEDv4.1s DAAC LPJ-GUESS | soils termites oceans | MIROC4-ACTM |

| | | | |
|-------------------|---|--|--|
| Geological | Geological emissions updated in this study (SI) | | |
|-------------------|---|--|--|

634 note: in TD products termites, oceans emissions are imposed from existing literature

635 * missing = not in the priors, presented as hatched pattern in the figure “\\”

636 **Not reported = data not available, presented as hatched pattern in the figure “//”

637

638 For an easier visual comparison and reconciliation between BU and TD estimates, we added the mean
639 of the BU anthropogenic estimates (off-white), underneath the BU and TD natural estimates. To note that for
640 some countries (e.g. Russia, DR Congo) this area might look like subtracted from the BU natural estimates, but
641 this is due to the sign convention used in this study (sink = negative and source = positive). In most cases, the
642 missing soil sink emissions are represented as a downward area.

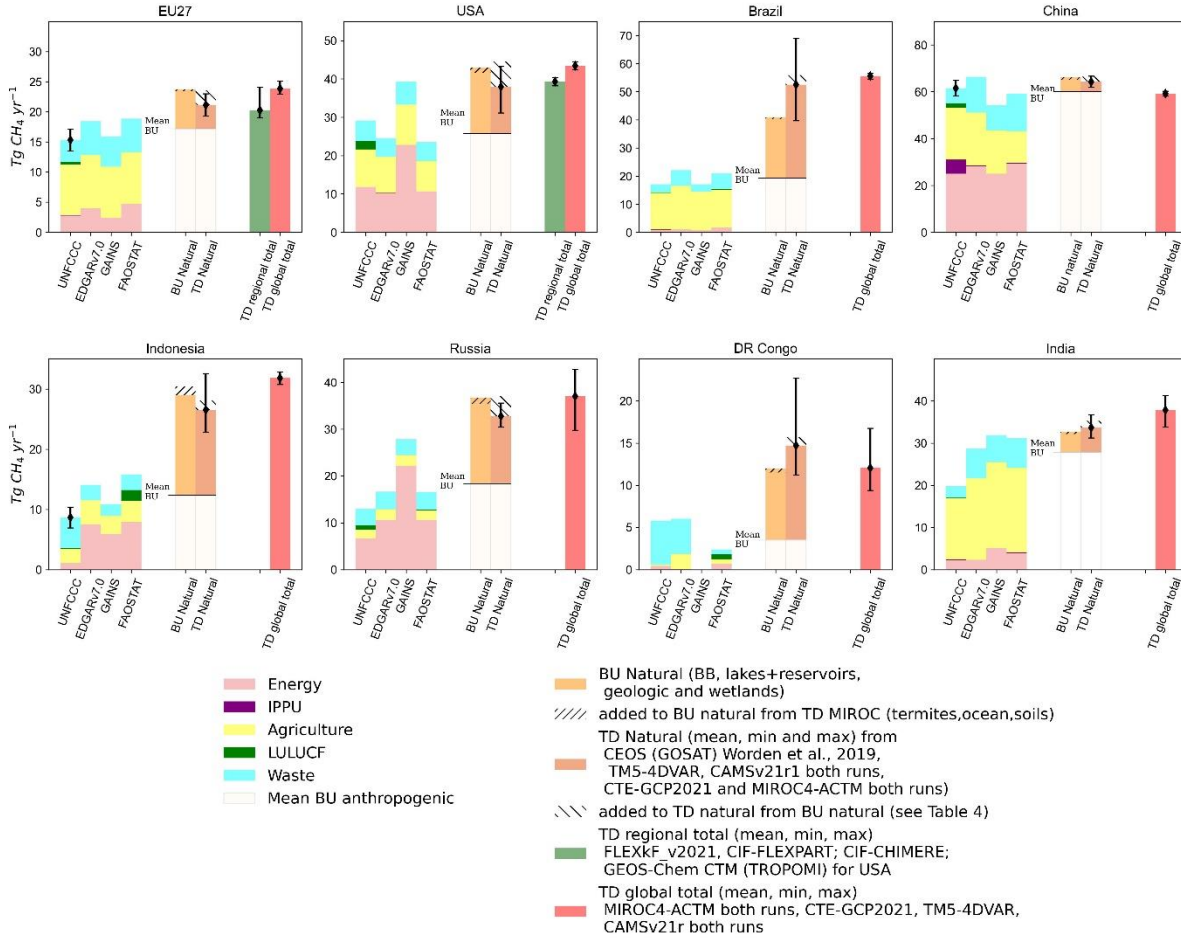
643 We note that for most countries, the sum of the anthropogenic and natural components matches those
644 of the TD global total estimates. This gives confidence that, to a certain extent and albeit with inconsistencies
645 between products, BU anthropogenic emission estimates are accurate and consistent with the observation-based
646 estimates and can be used to reconcile with the atmospheric-based estimates. We note from Figure 7 that in all
647 Annex I countries (EU, USA, Russia) and China, TD and BU natural emissions are consistent with each other,
648 after including the missing sources, as detailed in Table 4. For Brazil and DR Congo, the gap between the two
649 natural components is highly significant, while less for Indonesia and India. We hypothesize that mapping of the
650 wetland extent might cause these inconsistencies.

651

652

653

Total anthropogenic and natural CH_4 emissions from UNFCCC, other BU and TD estimates (average 2015-last available year)



65.

655 *Figure 7: Total anthropogenic and natural CH_4 emissions from BU and TD estimates presented as average of*
 656 *2015-last available year for EU and seven global emitters (USA, Brazil, China, Indonesia, Russia, DR Congo*
 657 *and India). The BU anthropogenic estimates belong to: UNFCCC NGHGI (2023) CRFs and BURs (incl.*
 658 *LULUCF) as totals and sectoral shares, EDGARv7.0, GAINS and FAOSTAT/PRIMAP-hist. The relative error*
 659 *on the UNFCCC CRF value represents the NGHGI (2023) reported uncertainties computed with the error*
 660 *propagation method (95% confidence interval) and gap-filled to provide respective estimates for each year (see*
 661 *Petrescu et al., 2023a, Appendix). In 2014, China reported an uncertainty of min 5.2% - max 5.3%. The BU*
 662 *Natural emissions for the EU are the sum of the VERIFY products (biomass burning, inland waters, geological*
 663 *and peatlands plus mineral soils as described in Petrescu et al., 2021 and 2023a, Appendix A2.1). For the seven*
 664 *non-EU emitters, the BU Natural fluxes are the sum of wetland emissions (LPJ-GUESS), lakes and reservoirs*
 665 *fluxes (ORNL DAAC, Johnson et al., 2022), geological (updated activity in SI) and biomass burning emissions*
 666 *(GFED4.1s). The TD natural global estimates are presented in Table 1. The uncertainty on the TD natural*
 667 *emissions is the min/max of all estimates. To both BU and TD estimates, missing (as not reported or not*
 668 *included in the priors) was added (see Table 4). The natural emissions have been plotted starting at the mean of*
 669 *the BU anthropogenic estimates, to retain comparability across the natural emission estimates, but also compare*
 670 *with the total TD estimates. The total regional TD estimates (for EU) belong to the mean and min/max of*
 671 *FELXkF_v2023, CIF-FLEXPART and CIF-CHIMERE and for USA GEOS-Chem CTM (TROPOMI) for the year*
 672 *2019 (Nesser et al., 2023). The total global TD inversions represent the average of the 2015-last available year*

673 *of the mean and min/max of CTE-GCP2021, MIROC4-ACTM both runs, CAMS v21r both runs and TM5-4DVAR.*
674 *The last available years are 2022 for CIF-CHIMERE, 2021 for EDGARv7.0, FAOSTAT, MIROC4-ACTM both*
675 *runs, UNFCCC CRFs, and CAMSv21r1 both runs, and 2020 for CIF-FLEXPART and CTE-GCP2021. TM5-*
676 *4DVAR partitioned data is only available between 2018 and 2020.*

677 However, Figure 7 should be interpreted with caution because in Europe, natural emission priors come
678 from regional ecosystem model simulations, where drained peatland, drainage ditches areas, and pristine areas
679 are lumped together. Therefore, if both LULUCF sector and natural BU emissions are included in the total budget
680 estimation, there is some overlap and possible double counting. Especially, ecosystem model estimates of ‘soil
681 sink’ or ‘inundated soil emissions’ may overlap with NGHGI managed peatland forest soil category (or
682 agricultural soils). The separation of emissions into different categories requires further clarification together
683 with inventory makers. Furthermore, it should be assessed which emissions should be called natural and which
684 anthropogenic (e.g., LULUCF, agriculture) by inversions.

685

686 **4. Challenges comparing bottom-up and top-down estimates**

687

688 An off-the-shelf comparison of BU and TD estimates is not possible, with a variety of adjustments
689 needed for comparability, often without the necessary data. Broadly speaking, inversions have not necessarily
690 been designed to compare directly to NGHGIs. A valid comparison should have consistent system boundaries
691 and perform a full uncertainty analysis to determine whether differences between estimates are statistically
692 significant given the constraining observational data.

693 The two most common issues limiting comparability are geographic scope and system boundaries
694 (Petrescu et al., 2021, 2023; McGrath et al., 2023, Andrew 2020; Grassi et al., 2018). The **geographical scope**
695 of inverse modelling versus inventory estimates should be controllable, but it can be challenging for small
696 countries or coarse inversions. Inversions are generally performed on a spatial grid and require aggregation, in
697 line with how official NGHGIs are reported (EEA, 2013). Inconsistent **system boundaries** have implications in
698 comparing the inventory- with inversions-based estimates for **source attribution**, e.g., anthropogenic vs. natural.
699 Most emission inventories aim at estimating anthropogenic emissions, while most inversions estimate both
700 anthropogenic and natural emissions. This is a particularly important issue for CH₄ where, globally, natural
701 emissions are of similar magnitude as anthropogenic emissions, with larger variations at regional scales, mainly
702 due to seasonality (i.e. wetlands). Thus, methods are needed to separate the anthropogenic flux from the total
703 flux (Deng et al. 2022, and above section 3.4). Similar issues arise with fossil CO₂ (Andrew 2020) as different
704 datasets can report different emission sources. Standardization procedures, such as The Community Inversion
705 Framework (CIF; Berchet et al., 2021), may help resolve some of these issues.

706 When comparing inventory- and inversion-based emissions, there are difficulties in analyzing **trends**
707 due to different time scale **variability**. Inventory-based approaches report emissions at the annual level, but often
708 do not consider interannual variations. Further, the Paris Agreement is set around five-yearly Global Stocktakes,
709 which indicates a desire to average trends, prioritizing the multi-annual trend over IAV, canceling out extremes
710 from both weather and socio-economic fluctuations. Inversion models, on the other hand, include variations over
711 a wide range of timescales, but in particular for IAV (e.g. OH and weather) that remains challenging to assess.

712 For an effective comparison, inversion-based estimates need to have IAVs statistically removed to make
713 comparisons with NGHGI easier (e.g., 5-year or 10-year averages or trend analysis). Additionally, averages of
714 ensembles of inversions may mask underlying differences and trends in individual inversions. Many research
715 projects make use of multi-model ensembles (Saunio et al., 2020; Deng et al. 2022, Lauerwald et al., 2024,
716 Zhang et al., 2024). From a scientific perspective, the model ensemble is often considered a more robust estimate
717 of the mean and uncertainty, as often individual estimates make errors due to some limitations and biases, while
718 in an ensemble, these errors are averaged out. From an inventory perspective, individual model comparisons may
719 be more efficient, as various input variables or processes can be compared directly to the inventory. Currently,
720 most inventory comparisons in UNFCCC NGHGI (e.g. UK, Switzerland) use single-model comparisons.

721 A strength of inversions is that they provide high temporal and spatial resolutions, which are not directly
722 capitalized when comparing with inventories. CH₄ from the fossil-fuel industry can contribute to large releases
723 to the atmosphere over a short period of time, given the large number of uncontrolled emission point sources in
724 oil and gas (O&G) and coal production areas worldwide (Jackson et al., 2020). Such processes include leakage
725 from landfills, spontaneous events from oil and gas production activities, so-called uncontrolled gas well blasts
726 etc. (Jacob et al., 2016, 2022). These uncontrolled events are difficult to include in the national inventories
727 leading to a potential underestimate of emissions (Massakkers et al., 2016, 2022). Recently, under the CoCO₂
728 project (<https://coco2-project.eu/>) a hot-spot satellite detection interactive map (Published studies on hot spot
729 detection (CO₂, CH₄) - uMap (openstreetmap.fr)) was released as a user-centric interface featuring published
730 studies on hot-spot detection between 2010 and 2021. It allows for advanced filtering by year, gas, activity,
731 geographical zone, and country.

732 A key challenge when comparing inversions with NGHGI is ensuring independence from the assumed
733 prior emissions. A more valid comparison between inversions is made when all inversions use the same **priors**.
734 In this context, we define as priors input data in the form of atmospheric observations (e.g. satellite retrievals,
735 ground-based observation networks (ICOS)) and/or bottom-up emissions datasets (e.g. EDGAR, GAINS) used
736 as input parameters to the inverse models. A key issue is the prior emission estimate. Theoretically, a constant
737 emission prior could be used, but this would require a dense observational network. Because of sparse
738 observations, inversion modelers assess how far observations have shifted the prior emissions to the posterior
739 emissions, preferably incorporating a full uncertainty analysis. The posterior emissions depend to a varying
740 extent on the prior that was used; the extent of this dependency is determined by the number of observations
741 used in the inversion, by how the observations relate to the emissions (governed by atmospheric transport) and
742 by the uncertainties assigned on the prior emissions and the observations. Thus, better quantified uncertainties
743 for the prior emissions would lead to more robust inversions. Whereas the comparison of an inversion with
744 NGHGI or other inversions would be made more robust by having more information on how dependent the
745 posterior estimate is on the prior. This stresses the need for more systematic in-situ data necessary to produce
746 adequate prior data (Bastviken et al., 2022) and synthesized atmospheric observations with their uncertainties to
747 robustly constrain the inversions.

748 It is not generally clear how inventory uncertainties can be compared to inversion uncertainties; however, it is
749 important that both methods provide comprehensive **uncertainty estimates**. The prior emissions used as input
750 into an inversion model should have robust uncertainty estimates, particularly with correlations in space and

751 time. This allows a full inversion system to better characterize how observations reduce uncertainty when
752 estimating the posterior estimate. Very few inversions routinely report this information. The inventory-based
753 emission estimate will additionally have uncertainty estimates, though these statistics may not be sufficiently
754 robust for verification purposes (National Academies of Sciences, Engineering, and Medicine 2022). There are
755 often offsets in inversion models, because of systematic inconsistencies between observations and chemistry-
756 transport models, which may make trends more robust than instantaneous estimates. Though, estimating
757 uncertainty in trends also requires understanding the correlation structure in time.

758 A key challenge for comparisons between NGHGI and independent estimates is to understand the
759 reasons for differences. In the case of BU comparisons, obtaining sufficient activity data and emission factors
760 should enable an accurate reconciliation of different estimates. However, in practice, it is often not possible to
761 obtain the necessary data. For inversions it is more complex. Often a close collaboration may be needed between
762 the inversion modeler and NGHGI team (e.g. UK NIR). If an inversion indicated a different trend in agricultural
763 CH₄ emissions, it is necessary to track down if this is a real difference or artifact of the inversion system. After
764 this, the spatial and temporal data in the inversion could be useful to the NGHGI team to locate what is causing
765 the difference. Many of the comparisons we show in this article ultimately remain comparisons, with detailed
766 reconciliations likely requiring intensive country-level case studies.

767 **5. Data availability**

768

769 Data files reported in this work which were used for calculations and figures are available for public
770 download at <https://doi.org/10.5281/zenodo.12582667> (Petrescu et al., 2023b). The data are reachable with one
771 click (without the need for entering login and password), with a second click to download the data, consistent
772 with the two click access principle for data published in ESSD (Carlson and Oda, 2018). The data and the DOI
773 number are subject to future updates and only refer to this version of the paper. The raw gridded data is available
774 upon request, directly from the data providers, as detailed in the Supplementary Information, Table S2.

775 **6. Conclusions**

776

777 We analyzed data from both anthropogenic and natural CH₄ fluxes, from both BU and TD observation-
778 based estimates (Table 1). BU estimates show that the largest sectors depend on the country: agriculture (EU,
779 Brazil, India), energy (US, China, Russia), and waste (Indonesia, DR Congo). The inversions attribute most of
780 the fluxes to the anthropogenic emissions, with tropical countries having a larger share of natural emissions
781 (wetlands). The EU and the seven other large emitters analyzed here contribute an anthropogenic emission of
782 173 Tg CH₄ yr⁻¹ (sum of last UNFCCC reported year, Figure 1,2), representing roughly half of the total global
783 anthropogenic emissions (386 Tg CH₄ yr⁻¹) reported by EDGARv7.0 in 2021. For comparison, the average of
784 the anthropogenic component from the atmospheric global inversions (MIROC booth runs, CTE-GCP2021,
785 CEOS and CAMS booth runs) is 181 Tg CH₄ yr⁻¹ (Figure 5).

786 We performed comparisons to the UNFCCC NGHGIs using the BU and TD data. Comparisons between
787 UNFCCC and BU products (Figure 3) reveal some deviations, particularly related to assumptions on gas/oil
788 emissions (e.g., GAINS for Russia and the USA) and waste (e.g., Indonesia, DR Congo). It is more challenging

789 to compare BU and TD estimates, due to different attribution to source activities (Table 2 and Figure 5) and
790 different priors used in the simulations (Petrescu et al., 2023b, Priors Table). The comparison between UNFCCC
791 and the TD estimates (Figure 4) agrees largely with the findings of Deng et al. (2022) who applied different
792 methodologies to calculate natural emissions. In most cases, the gap between the anthropogenic BU fluxes from
793 inventories and total TD fluxes can be largely explained by the natural fluxes (Figure 7). It is difficult to draw
794 definitive conclusions on emissions trends seen by inversions, as the adjustments for natural emissions and IAV
795 and seasonal variability might strongly influence trends. Despite this, given that, in most cases, the UNFCCC
796 BURs reports are incomplete for the non-Annex I parties (China, Indonesia, DR Congo) it is important to
797 acknowledge that the TD estimates might become a useful way to complement inventories and play a role in the
798 validation of the BU estimates.

799 There is still a pressing need for reporting of uncertainties in both prior and posterior emissions, even if
800 some TD inversions do report it as the standard deviation of ensemble members (CTE-GCP2021 and
801 FLEXkF_v2023, Figure 4). The use of a variety of priors across different inversion systems can also inhibit
802 comparability with inventories and between inversions. Generally, inversions are still ill-constrained by
803 observations (only 60 sites globally plus satellites) and the prior flux uncertainty for each of the 54 regions is
804 large. Therefore, the monthly results could be more ill-constrained than the annual totals. Even if comparisons
805 between CH₄ inversion estimates and NGHGs are currently uncertain because of the spread in the inversion
806 results, TD inversions inferred from atmospheric observations represent partly independent data against which
807 inventory totals and trends can be compared, considering the most encountered issues discussed in section 4.

808

809 **7. Appendix**

810

811 All the information regarding models/methods descriptions is available in the Supplementary Information (SI)
812 file. Appendices A1 and A2 in Petrescu et al., 2023a contain detailed information about Table 1 products. Further
813 information on new products together with references and contact details are found in Tables S1 and S2 in SI.

814

815 The tables with priors used by all the products and the matrix highlighting the comparability issues identified in
816 section 4 are found in the Zenodo data repository, Petrescu et al., 2023b.

817

818 **Supplementary Information (link)**

819

820

821 **Author contributions**

822

823 AMRP designed research and led the discussions; AMRP wrote the initial draft of the paper and edited all the
824 following versions; GPP drafted the initial version of section 4, edited the final version of this manuscript,
825 contributed to the revised version and advised on the context; PP processed all the original EU data submitted to
826 the VERIFY portal; RLT, SH, BM, DaB, RL, PKP, AT, RMA, LHI, FNT, GC and JG edited and gave consistent
827 comments and suggestions to the initial manuscript; DaB, RL and RMA provided input to the final revised
828 version; all co-authors are data providers and contributed to subsequent versions of the manuscript by providing

829 specific comments and information related to their data in the main text, providing as well product descriptions
830 for the Supplementary Information file.

831 **Competing interests**

832 At least one of the (co-)authors is a member of the editorial board of Earth System Science Data.

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859

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