

1 **High-resolution physicochemical dataset of atmospheric** 2 **aerosols over the Tibetan Plateau and its surroundings**

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31 **Abstract**

32 Atmospheric aerosol in the Tibetan Plateau (TP) and its surroundings has attracted
33 significant scientific interest in recent decades due to its notable impacts on regionally
34 climatic and cryospheric changes, ecological and environmental securities, and the
35 hydrological cycle. However, our understanding on the atmospheric aerosol in this
36 remote region is highly limited by the scarcity of available dataset, owing to the
37 extremely harsh natural conditions. This challenge has been mitigated in recent decades
38 by establishing field observatories at typical sites within the TP and its surroundings. A
39 continuous project initiated in 2015 aims to explore the properties and sources of
40 atmospheric aerosols, as well as their regional differences, through multiple short-term
41 intensive observations across this vast region utilizing a suite of high-time-resolution
42 online instruments. This paper presents a systematic hourly scaled dataset of aerosol
43 physicochemical and optical properties at eight sites across the TP and its surroundings
44 derived from the project. It includes size-resolved chemical compositions of submicron
45 aerosols, high-resolution mass spectra and sources of organic aerosols, size
46 distributions of particle number concentrations, particle light scattering and absorption
47 coefficients, particle light absorptions attributed to different carbonaceous substances
48 including black carbon and brown carbon, and number concentrations of cloud
49 condensation nuclei. In brief, atmospheric aerosols in these remote sites were all well-
50 mixed and highly aged, reflecting their dominated regional transport sources. However,
51 the southern TP region exhibited high contributions of carbonaceous organic aerosols,
52 neutralized bulk submicron aerosols, and a relatively higher light absorption capacity,
53 whereas in the northern TP region, secondary inorganic species were the main
54 contributors to the overall acidic submicron aerosols. Beyond providing insights into
55 the regional differences in aerosol sources and properties across the TP and its
56 surroundings, the datasets will also benefit simulations of aerosol radiative forcing and
57 evaluations of interactions among different Earth system components in numerical
58 models in this region. The datasets are accessible through the National Cryosphere
59 Desert Data Center, Chinese Academy of Sciences
60 (<https://doi.org/10.12072/ncdc.NIEER.db2200.2022>; Xu, 2022).

61 **1 Introduction**

62 Tibetan Plateau (TP), with an average elevation exceeding 4,000 m above sea level
63 (a.s.l.) and spanning a surface area of approximately 2.5 million square kilometers,
64 stands as the highest plateau on the Earth. Its high-altitude mountain ranges, integral to
65 one of the world’s most crucial cryospheric regions, have earned the TP the monikers
66 “roof of the world”, “Third Pole”, and “Asian Water Tower” (Qiu, 2008; Yao et al.,
67 2019). The TP and its surroundings play a pivotal role in influencing the global and
68 regional climate systems, hydrological cycles, and cryospheric changes through its vast
69 and complex topography and its function as a significant heat source (Duan and Wu,
70 2005; Yao et al., 2012; Chen et al., 2021). In recent decades, a major concern has
71 focusing on the significant climatic warming and the rapid changes in the cryosphere
72 of this region (Kang et al., 2010), which exhibits a warming trend that exceeding that
73 of the Northern Hemisphere (0.34 vs. 0.29 °C/decade) (You et al., 2021; Zhou and
74 Zhang, 2021).

75 Atmospheric aerosols, as one of the most complex and critical components in the
76 atmosphere, significantly influence climatic warming and cryospheric changes in the
77 TP regions. They exert crucial direct and indirect effects on atmospheric energy budget
78 and the albedos of snow and ice surfaces, impacting Earth’s climate system (Xu et al.,
79 2009; Kang et al., 2019b). Notably, light-absorbing carbonaceous aerosols (CAs) such
80 as black carbon (BC) and brown carbon (BrC) directly absorb solar radiation, warming
81 the atmosphere and contributing to positive forcing on Earth’s energy budget
82 (Ramanathan et al., 2007; Kopacz et al., 2011). For instance, Li et al. (2018)
83 demonstrated that BC causes significantly greater albedo reduction (~46%) and
84 instantaneous radiative forcing ($7\text{--}64\text{ W m}^{-2}$) on aged snow surfaces of a TP glacier
85 than mineral dust. Moreover, aerosol particles over the TP significantly affect ice cloud
86 properties and cloud development through their semi-direct effects (Liu et al., 2019).
87 Given the vast and remote nature of the TP, coupled with its complex topography,
88 meteorology, and harsh environment, *in-situ* observation of atmospheric aerosols poses
89 substantial challenges. Consequently, numerical model simulations based on reanalysis
90 data have emerged as a predominant and crucial method in recent decades. Notable
91 studies include Lau et al. (2006), who assessed the impact of atmospheric aerosols on
92 the Asian summer monsoon intensification using the NASA finite-volume general
93 circulation model; Kopacz et al. (2011), who explored the origin and radiative forcing

94 of BC in the TP and Himalayas with the GEOS-Chem global chemical transport model;
95 and Liu et al. (2015), who examined the transport of summer dust and anthropogenic
96 aerosols using a three-dimensional aerosol transport–radiation model with satellite data
97 inputs. Despite the significant insights gained from these simulations, *in-situ*
98 observations of atmospheric aerosols in the TP are increasingly recognized as critical
99 for evaluating and enhancing model accuracy in this remote area. The absence of *in-*
100 *situ* aerosol data to refine models introduces considerable uncertainty into the results.
101 Furthermore, while model simulations primarily focus on spatial distribution across
102 broad regions, they often overlook temporal variations or the inherent evolution
103 mechanisms at high temporal resolution, which *in-situ* observations could illuminate.

104 Recent advancements in observational techniques and instrumentation have enabled
105 numerous *in-situ* measurement within the TP and its surroundings, aiming to delineate
106 the physical, chemical, and optical properties of aerosols, along with their potential
107 sources, transport pathways, and regional distributions. A detailed compilation of direct
108 ambient aerosol measurements in the TP, employing a variety of observational methods
109 and instruments, is presented in Table S1 in the supplementary material. Notably, off-
110 line atmospheric filter sampling has emerged as a key *in-situ* aerosol collection method
111 in the TP, favored for its low-cost and feasibility under the region’s harsh conditions
112 and logistical constraints. This method has effectively captured the composition, size,
113 light absorption properties, sources, and variations of ambient aerosols, including
114 diverse CAs components such as BC, BrC, organic carbon (OC), water-soluble OC
115 (WSOC), humic-like substances (HULS), and polycyclic aromatic hydrocarbons (PAHs)
116 in the remote TP region (Cao et al., 2009; Zhao et al., 2013; Xu et al., 2014a; Zhang et
117 al., 2014; Cong et al., 2015; Wan et al., 2015; Xu et al., 2015; Kang et al., 2016; Li et
118 al., 2016b; Xu et al., 2020). Furthermore, off-line filter sampling has been instrumental
119 in mapping the regional aerosol distribution across the TP, facilitated by its simplicity
120 and the ability to conduct simultaneous observations at multiple locations (Li et al.,
121 2016a; Chen et al., 2019; Kang et al., 2022). Despite these advancements, off-line filter
122 sampling studies in the TP’s remote regions remain insufficient for current needs. The
123 gathered data is generally fragmented and unsystematic, characterized by low temporal
124 resolution, limited aerosol property parameters, and sparse data points. Typically, these
125 studies have been localized to specific sites, utilizing single instruments with temporal
126 resolutions ranging from days to weeks over brief periods. Such low temporal

127 resolution limits the accurate understanding of aerosols' temporal evolution and
128 underlying mechanisms, especially during rapid, short-term events. Additionally,
129 integrating and comparing data across different research groups is challenging due to
130 variations in research focuses, measured parameters, sampling methodologies,
131 laboratory filter processing, and data analysis techniques. Moreover, despite the relative
132 ease of off-line sampling, extensive areas of the TP still lack such observational efforts.
133 To date, comprehensive research focusing on multiple aerosol parameters through real-
134 time online consecutive measurements (with high temporal resolutions from minute to
135 hour scales) at multiple sites remains a rarity in the TP.

136 Research on atmospheric aerosols in China has achieved significant advancements over
137 the past decade. The Aerodyne high-resolution time-of-flight aerosol mass spectrometer
138 (HR-ToF-AMS) has become a widely used online instrument, enabling numerous
139 studies to characterize the real-time, size-resolved chemical compositions and sources
140 of submicron aerosols (Li et al., 2017; Zhou et al., 2020). However, its application in
141 the TP has been limited, a situation attributed to the instrument's demanding operational
142 requirements and the area's challenging observation conditions. Since 2015, our
143 research group has initiated a continuous and systematic observation project, aiming to
144 investigate the regional differences on aerosol sources and properties across the TP.
145 This effort involves annual deployments of the HR-ToF-AMS alongside other high-
146 resolution, real-time online instruments at various sites. Remarkably, our dataset
147 represents the first and only collection that extensively covers a wide range of aerosol
148 parameters (including physical, chemical, and optical properties and their diverse
149 sources) across multiple geographic environments (e.g., urban, remote, high-altitude
150 mountain, grassland, and subtropical forest) within the TP and its surroundings. This is
151 achieved through real-time and high-time-resolution online observations, providing a
152 comprehensive resource for understanding regional aerosol variations and serving as
153 essential input for future aerosol radiative forcing simulations and Earth system
154 interaction assessments. The structure of this paper is as follows: Sections 2 and 3
155 describe the observation sites, instrumental deployments, and data processing. Section
156 4 presents the high-time-resolution aerosol data, encompassing physical, chemical, and
157 optical properties as well as source information. The limitations and the uniqueness of
158 our dataset are discussed in Section 5.

159 **2 Observation site descriptions**

160 Between 2015 and 2022, intensive observations of atmospheric aerosol chemistry were
161 carried out at eight sites across the TP and its surroundings. These sites comprise seven
162 remote sites (QOMS, Motuo, NamCo, Ngari, Waliguan, LHG, and Bayanbulak) and
163 one urban site (Lhasa), the latter serving as a contrast for comparative analysis. Figure
164 1 illustrates the geographical locations of these sites along with photographs during
165 each observation. Table 1 details the specifics of each site including the sampling
166 periods and the instruments deployed during each field campaign. The subsequent
167 sections provide brief descriptions these sites, arranged geographically from the
168 southern to the northern parts of the TP.

169 **2.1 QOMS**

170 The Qomolangma Station for Atmospheric and Environmental Observation and
171 Research, Chinese Academy of Sciences (86.56°E, 28.21°N; 4276 m a.s.l.; abbreviated
172 as QOMS in this study, with similar abbreviations for other sites hereafter), is situated
173 in the Rongbuk valley basin on the northern slope of Mt. Everest. The climate in the
174 northern slope of Mt. Everest has obvious seasonal variation, heavily influenced by the
175 Indian monsoon system (Bonasoni et al., 2010; Cong et al., 2015). During the pre-
176 monsoon season (typically March to May), dominant westerlies facilitates the long-
177 range transport of atmospheric pollutants from South Asia, making QOMS an ideal
178 high-altitude observatory on the south edge of the TP for examining transboundary
179 pollutant transport into the plateau's interior. In the summer monsoon season
180 (June–August), prevailing southerly winds bring warm and wet airflow from the Indian
181 Ocean, leading to increased humidity and precipitation in the plateau.

182 **2.2 Motuo**

183 Motuo County, located in the lower reaches of the Yarlung Tsangpo River on the
184 southern slopes of the eastern Himalayas and Gangrigab Mountains in the southeast
185 edge of the TP. The county, sited halfway up a mountain, enjoys a subtropical humid
186 climate characterized by relatively high temperatures and abundant rainfall. With a
187 small population of ~15,000, Motuo County remains one of the TP's most pristine
188 regions. The sampling site in Motuo (29.30°N, 95.32°E; 1305 m a.s.l.) was located atop
189 a hill overlooking the Yarlung Tsangpo Grand Canyon. This vantage point makes it an
190 ideal location for directly monitoring the transboundary transport of atmospheric
191 pollutants and moisture from Southeast Asia and the Indian Ocean into the TP.

192 **2.3 NamCo**

193 The Nam Co Station for Multisphere Observation and Research, Chinese Academy of
194 Sciences (NamCo; 90.95°E, 30.77°N; 4730 m a.s.l.) is a high-altitude observatory in the
195 south-central part of the TP. Situated at the southeast shore of Nam Co Lake, the station
196 is surrounded by a pristine and isolated area. The region experiences a typical semi-arid
197 plateau monsoon climate, characterized by increased precipitation during the summer
198 monsoon season. NamCo is a pivotal inland site within the TP, predominantly
199 influenced by air masses from the south and west.

200 **2.4 Ngari**

201 The Ngari Station for Desert Environment Observation and Research, Chinese
202 Academy of Science (Ngari; 79.70°E, 33.39°N; 4270 m a.s.l.) is located in the Rutog
203 County within the Ngari Prefecture of the Tibet Autonomous Region, China. This area
204 is at the southwestern edge of the TP and is characterized by its semi-arid climate, sparse
205 vegetation, and intense solar radiation. As a key member of “high-cold region
206 observation and research network for land surface processes & environment of China”,
207 the Ngari station plays a crucial role in monitoring climate, hydrological, atmospheric,
208 and ecological environmental changes in the TP’s western territories. Additionally, it
209 contributes to understanding the interactions between the Indian monsoon system and
210 the westerlies.

211 **2.5 Waliguan**

212 The Waliguan Baseline Observatory (Waliguan; 100.9°E, 36.28°N; 3816 m a.s.l.) is one
213 of the twenty-nine baseline stations of Global Atmosphere Watch (GAW) under the
214 World Meteorological Organization (WMO). Situated at the top of Mt. Waliguan, which
215 rises approximately 600 m, the observatory is located in a pristine region with minimal
216 human activity impact. Waliguan is represented as a key observatory on the northeastern
217 edge of the TP, predominantly influenced by air masses from the northeast during the
218 summer. This location is strategic for studying the transport and impact of air pollutants
219 from industrial regions in northwestern China to the TP in this study.

220 **2.6 LHG**

221 The Qilian Observation and Research Station of Cryosphere and Ecologic Environment,
222 Chinese Academy of Sciences (LHG; 39.50°N, 96.51°E; 4180 m a.s.l.), is located
223 approximately 1km from the terminus of Laohugou Glacier No.12. This glacier is

224 among the largest mountain glacier in the northern slope of the western Qilian
225 Mountains. LHG serves as another notable station in the northeastern TP, distinctly
226 remote from human settlements. The climate in this region is predominantly arid and
227 continental, influenced by the East-Asian monsoon in summer and the Westerlies in
228 winter. A pronounced mountain-valley breeze during summer facilitates the upward
229 transport of air masses from lower altitudes, making LHG an ideal site for background
230 air mass sampling and for investigating the transport and potential impacts of air
231 pollutants from surrounding areas.

232 **2.7 Bayanbulak**

233 The Bayanbulak National Basic Meteorological Station (Bayanbulak; 84.35° N, 42.83°
234 E; 2454 m a.s.l.) is located in the Bayanbulak grassland, northwest of Hejing County in
235 the Xinjiang Uygur Autonomous Region, China. Bayanbulak lies within an
236 intermontane basin in the central Tianshan Mountains, surrounded by numerous snow
237 mountains with altitudes more than 3000 m. The climate in Bayanbulak grassland is
238 characterized by a typical temperate continental mountain climate, with an annual
239 average precipitation ranging between 200 to 300 mm. Bayanbulak town experiences
240 minimal human activities and traffic, maintaining its pristine environment.

241 **2.8 Lhasa**

242 Lhasa (29.65°N, 91.03°E; 3650 m a.s.l.) is the capital of the Tibet Autonomous Region,
243 China, and located in the south-central part of the TP. The city lies in a broad river
244 valley, surrounded by mountains that reach up to 5500 m, with the Lhasa River passing
245 through the city from west to east. Our observation site is located in Binhe Park,
246 adjacent to the Lhasa River. Notably, the Norbulingka scenic area, one of the main
247 activity centers for local Tibetans celebrating their religious festivals such as the Sho
248 Dun festival, is located ~1 km to the northwest of the sampling site. Moreover, the
249 Potala Palace, the center of Tibetan Buddhism, is ~1.8 km to the northeast. Given
250 Lhasa's unique energy structure and the distinct living habits of the residents,
251 comparative observations are carried out at this urban site. These studies aim to
252 investigate the primary aerosol properties and sources, particularly from various
253 residential combustion activities.

254 **3 Online sampling, instrumental setup, and data processing**

255 **3.1 Online real-time aerosol sampling over the TP**

256 Atmospheric aerosol observations were conducted at each site using a suite of real-time,
257 high-resolution instruments. This instruments typically included a HR-ToF-AMS
258 (Aerodyne Research Inc., Billerica, MA, USA) to determine the chemical composition
259 (organic aerosol (OA), nitrate, sulfate, ammonium, and chloride) of non-refractory
260 submicron aerosol (PM_{10}); A scanning mobility particle sizer (SMPS, model 3936, TSI
261 Inc., Shoreview, MN, USA) was used to measure the size distribution and number
262 concentration of submicron particles; A photoacoustic extinctions (PAX, DMT Inc.,
263 Boulder, CO, USA) was used for obtaining particle light absorption, scattering, and
264 extinction coefficients (B_{abs} , B_{scat} , and B_{ext}) along with single scattering albedo (SSA)
265 at 405 nm, as well as BC mass concentration; An Aethalometer (model AE33/AE31,
266 Magee Scientific Corp., Berkeley, CA, USA) was used for acquiring the B_{abs} across
267 seven wavelengths (370–950 nm); A cloud condensation nuclei (CCN) counter (model
268 CCN-100, DMT Inc., Boulder, CO, USA) measured CCN number concentrations at
269 various water vapor supersaturations (SS). The measurement uncertainties for each
270 instrument are difficult to quantify based on data from a single instrument. The
271 uncertainty showing below are referred from other studies: <30% for HR-ToF-AMS
272 (Jimenez et al., 2016) and SMPS, <40% for Aethalometer (Backman et al., 2017), <10%
273 for PAX (Selimovic et al., 2018), and <25% for CCNC (Rose et al., 2008). Details on
274 the specific instruments deployed and the sampling periods for each observation
275 campaign are summarized in Table 1.

276 The HR-ToF-AMS was the cornerstone instrument for atmospheric aerosol chemistry
277 observations across all field campaigns. The SMPS was deployed at QOMS, Motuo,
278 LHG, and Lhasa, whereas the PAX was employed at QOMS, Motuo, Ngari, Waliguan,
279 and Lhasa. The Aethalometer was utilized at QOMS, NamCo, and Waliguan, with the
280 CCN-100 operational at Motuo, Waliguan, and LHG. Field observations in the southern,
281 western, and central regions of the TP were predominantly conducted during the pre-
282 monsoon season, aimed to study the transboundary transport of pollutants from South
283 Asia into the TP, under the influence of Westerlies and the Indian monsoon. For instance,
284 observations took place from 12 April to 12 May at QOMS, Motuo from 26 March to
285 22 May, NamCo from 31 May to 1 July, and Ngari from 1 Jun to 5 July. On the other
286 hand, measurements in the remote regions of the northern TP and its surroundings were
287 carried out during the summer to track aerosol transport from surrounding polluted
288 areas, considering the effects of the Westerlies or the intensified East Asian monsoon.

289 Specifically, observations occurred from 1 to 31 July at Waliguan, from 4 to 29 August
290 at LHG, and from 29 August to 26 September at Bayanbulak. The observation at Lhasa
291 was conducted from 31 August to 26 September, focusing on capturing the peak
292 atmospheric oxidation capacity during the summer.

293 **3.2 Instrumental setup**

294 Despite the variations in instrumentation across different observation campaigns, the
295 core setup for sampling was largely consistent. Figure 1b illustrates the standard
296 sampling configuration for each campaign typically involved housing all instruments
297 within an air-conditioned trailer or room. The inlets were induced to the instruments
298 from the roof with a cyclone (model URG-2000-30EH, URG Corp., Chapel Hill, NC,
299 USA) in the front of the inlet to eliminate particles with an aerodynamic diameter (D_{va})
300 exceeding 2.5 μm . These fine particles were then passed through a Nafion dryer via
301 1/2-inch stainless steel tubing to dehumidify the airflows before being directed into the
302 instruments for real-time analysis. For detailed information on the setup and
303 methodology of the instruments, it can be found in our previous publications (Xu et al.,
304 2018; Zhang et al., 2018; Zhang et al., 2019; Zhao et al., 2022).

305 **3.3 Instrumental operation and data processing**

306 The measurement principles, operation procedures, calibration methods, and data
307 analysis for these instruments used in this study are thoroughly detailed in Texts S1–S4
308 of supplementary material. Here, we highlight only key descriptions and critical setting
309 as follows: (1) The HR-ToF-AMS was operated at V-mode during most of the eight
310 field campaigns to accommodate the relatively low signal-to-noise ratios due to low
311 aerosol mass loading in the TP regions. (2) Particle size observations were not
312 conducted during the NamCo and LHG campaigns due to a chopper malfunction in the
313 HR-ToF-AMS. (3) Different relative ionization efficiency (RIE) values were used for
314 ammonium and sulfate according to the ionization efficiency calibrations of HR-ToF-
315 AMS in different campaigns. (4) Different size parameters were achieved according to
316 the particle sizing calibrations in different campaigns. (5) Elemental ratios of OA such
317 as oxygen-to-carbon (O/C), hydrogen-to-carbon (H/C), organic matter-to-organic
318 carbon (OM/OC), and nitrogen-to-carbon (N/C), were determined using the improved
319 method (Canagaratna et al., 2015). (6) A default collection efficiency (CE) value of 0.5
320 were employed to the HR-ToF-AMS measurements during the QOMS, NamCo, Ngari,

321 Waliguan, and Lhasa campaigns in consideration of their overall neutralized bulk
322 submicron aerosols, whereas a composition-dependent CE (Middlebrook et al., 2012)
323 value was adopted at Motuo, LHG, and Bayanbulak, where aerosols were slightly acidic.
324 (7) Source apportionment of OA during all observations were performed by the positive
325 matrix factorization (PMF) analysis. The details of the PMF solution determination for
326 each site are not presented here but can be referenced in our previous publication for
327 select campaigns (Xu et al., 2018; Zhang et al., 2018; 2019). (8) Only the chemical
328 compositions of non-refractory PM₁ are reported for the Bayanbulak campaign due to
329 the absence of BC observations. (9) The sample and sheath flow rates of SMPS were
330 set at 0.3 and 3.0 L min⁻¹, respectively, at both QOMS and Lhasa, covering a particle
331 size range between 14.6 and 661.2 nm in mobility diameter (D_m), whereas 0.5 and 5.0
332 L min⁻¹ at LHG and Motuo sites with a range of 10.9–495.8 nm in D_m . (10)
333 Aethalometer measurements were corrected for filter-based loading and multiple
334 scattering effects. A traditional absorption Ångström exponents (AAE) method (Zhang
335 et al., 2021) was adopted to apportion total B_{abs} into two parts of BC and BrC ($B_{abs,BC}$
336 and $B_{abs,BrC}$). (11) CCN number concentrations were measured at five different SS
337 values of 0.2%, 0.4%, 0.6%, 0.8%, and 1.0% at a 30-min cycle during the Motuo,
338 Waliguan, and LHG campaigns.

339 **4 Aerosol properties, sources, and radiative forcing over the TP**

340 **4.1 Mass loading and chemical composition of submicron aerosols**

341 Figure S1 presents the temporal variations of PM₁ chemical species (OA, nitrate, sulfate,
342 ammonium, chloride, and BC) observed across the eight observations in the TP and its
343 surroundings. The mass concentrations of PM₁ and its components exhibited distinct
344 variations, with a few periods of high mass loading observed throughout each
345 campaign's sampling period. Despite variations in sampling years (2015–2022),
346 seasons (March–September), and altitudes (1350–4730 m a.s.l.) across these sites, the
347 distinct PM₁ mass concentrations and chemical compositions clearly illustrate the
348 regional difference among these sites. On average, total PM₁ mass concentrations
349 across the eight campaigns ranged from 1.9 to 9.1 μg m⁻³ (Fig. 2 and Table 2). The
350 highest mass concentration was observed at Waliguan, driven by the transport of
351 anthropogenic aerosols and gaseous pollutants from urban centers in northwestern
352 China. In contrast, the lowest values were observed at NamCo and Bayanbulak,
353 reflecting their background and pristine environmental conditions. The average PM₁

354 mass level across the TP and its surroundings was comparable to those observed at other
355 high-altitude, coastal, forest, and remote background sites globally ($0.46\text{--}15.1\ \mu\text{g m}^{-3}$;
356 Table 3), yet remains significantly lower than those observed at densely urban
357 ($34.4\text{--}71.5\ \mu\text{g m}^{-3}$) and suburban ($21.4\text{--}44.9\ \mu\text{g m}^{-3}$) areas in other parts of China (Li
358 et al., 2017). This suggests the predominantly clean atmospheric conditions in the
359 remote and high-altitude regions of the TP.

360 The chemical compositions of PM_{10} also exhibited significant regional differences
361 across the TP (Fig. 2), highlighting varied aerosol sources in different areas of the TP.
362 At five sites (QOMS, Motuo, NamCo, Ngari, and Lhasa) located in the southern,
363 western, or central TP, OA and BC together accounted for as high as 64.9–85.7% of the
364 total PM_{10} mass (Table 2). This high contribution was largely attributed to the frequent
365 transport of biomass-burning emissions from South and Southeast Asia to the TP during
366 the pre-monsoon season (Bonasoni et al., 2010; Cong et al., 2015; Zhang et al., 2018),
367 along with significant local biomass burning from religious activities in Lhasa (Cui et
368 al., 2018; Zhao et al., 2022). In contrast, at the three northern sites (Waliguan, LHG,
369 and Bayanbulak), inorganic species (sulfate, nitrate, and ammonium; referred to as SNA)
370 accounted for more than 60% of the total PM_{10} . Sulfate was the most significant
371 component of SNA (38.1–46.0%), aligning with observations from another high-
372 altitude site in the northeastern TP (Menyuan; 28%) and various rural and remote sites
373 (19–64%) in East Asia (Du et al., 2015). The pronounced SNA contributions,
374 particularly sulfate, in the northern TP and its surroundings, were mainly related to the
375 regional transport of anthropogenic aerosols and gaseous precursor from nearby urban
376 areas as well as important in-cloud aqueous reactions during the transportation to the
377 mountains (Zhang et al., 2019).

378 **4.2 Bulk acidity, size distribution, and diurnal variation of submicron aerosols**

379 Particle phase acidity significantly influences aerosol physicochemical properties,
380 affecting hygroscopicity, toxicity, and heterogeneous reactions. The bulk acidity of
381 submicron aerosols was evaluated at each site following the method in Zhang et al.
382 (2007b) and Schueneman et al. (2021) using AMS measurements. A detailed
383 description of this method can be found in Text S5 of the supplementary material or our
384 previous publications (Zhang et al., 2018; Zhang et al., 2019). Notably, we observed
385 clear regional variations between the southern and northern TP, largely due to
386 differences in aerosol sources and composition (Fig. 3). Linear regression slopes at

387 QOMS, NamCo, Ngari, and Lhasa (located in the southern, western, or central TP) were
388 fitted to be 1.2, 1.11, 0.98, and 1.18, respectively, indicating the submicron aerosols at
389 these sites were generally neutralized, occasionally showing an excess of ammonium.
390 The result is consistent with previous findings on high ammonia availability from
391 agriculture emissions in the South Asia (Van Damme et al., 2015). Moreover, as
392 reported in our previous publications, atmospheric aerosols at QOMS and NamCo were
393 significantly influenced by biomass-burning emissions from South and Southeast Asia
394 during the pre-monsoon season (Xu et al., 2018; Zhang et al., 2018), and Lhasa
395 experienced intense biomass fuel burning during frequent religious festivals (Zhao et
396 al., 2022). In contrast, submicron particles at the remaining four sites, particularly LHG
397 and Bayanbulak in the north, were overall acidic, with regression slopes ranging from
398 0.73 to 0.86, where sulfate was a major PM₁ component (46.0% and 41.6%). Similar
399 findings of acidic submicron aerosol particles have also been observed at Menyuan and
400 LHG in the northern TP in previous studies (Du et al., 2015; Xu et al., 2015), mainly
401 related to the transport of enriched SNA species or their gaseous precursors from the
402 industrial areas in northwestern China.

403 The size distributions of non-refractory PM₁ chemical species, obtained from HR-ToF-
404 AMS measurement, provide valuable insights into aerosol sources, oxidation degrees,
405 mixing states, formation, transformation, and growth mechanisms as well as their
406 impacts on CCN activity. Typically, size distributions for SNA species and oxidized OA
407 peaked in the accumulation mode ($\sim 400\text{--}600$ nm in D_{va}), as a result of secondary
408 formation processes. In contrast, fresh organics from primary emission sources exhibit
409 smaller size (Zhang et al., 2005b; Aiken et al., 2009). In this study, we focus on organics
410 and the combined three SNA species to highlight regional variations in size distributions
411 across the TP. As shown in Figs. 4a and S2 and Table 2, the peak diameters of OA and
412 SNA size distributions varied significantly, from 584.4 and 634.5 nm at Ngari to a
413 smaller size of 228.1 and 250.0 nm at Lhasa, respectively. This variation suggests the
414 distinctly different sources and aging processes of atmospheric aerosols across the TP,
415 particularly between those high-altitude remote sites and urban sites. For instance, bulk
416 PM₁ at QOMS was reported to be internally well-mixed and aged, attributed to long-
417 range transport from biomass-burning emissions in South Asia (Zhang et al., 2018),
418 whereas local primary sources, including cooking, traffic exhausts, and biomass
419 burning, totally accounted for more than 60% of the total OA in urban Lhasa (Zhao et

420 al., 2022). The crucial influence of aerosol sources on size distributions is further
421 supported by the correlation between the mode size and O/C ratios of OA ($R^2 = 0.74$)
422 (Fig. 4a).

423 The diurnal variations of PM₁ chemical compositions are typically influenced by
424 multiple factors, including meteorological conditions (e.g., planetary boundary layer
425 (PBL) height, wind direction and speed, temperature, relative humidity), various
426 primary emission sources (e.g., vehicle exhausts during traffic rush hours, cooking
427 emissions, and coal combustion emissions for heating), and distinct formation
428 mechanisms (e.g., daytime photochemical oxidation, nighttime heterogeneous
429 reactions, and gas-particle partitioning of secondary species). A comprehensive
430 understanding of these diurnal variations is crucial for exploring the dynamic evolution
431 of aerosol compositions and identifying the primary drivers (source, meteorology, or
432 secondary formation) behind the variations in different chemical species.

433 Distinct diurnal patterns in the total PM₁ mass concentrations were observed across
434 different field campaigns (Fig. 4b). At remote sites of QOMS, LHG, NamCo, and
435 Waliguan, located in valleys or atop mountains, variations were largely governed by
436 mountain-valley wind circulation and change in PBL height. For instance, QOMS
437 exhibited a distinct diurnal pattern with continuously decreasing concentrations during
438 the daytime, but relatively higher concentrations at night. The minimum occurred at
439 around ~15:00 in this valley site, likely due to strong afternoon glacier winds and a
440 higher PBL (Zhang et al., 2018). Conversely, LHG and NamCo experienced lower PM₁
441 concentrations from night to early morning, with increase in the afternoon, attributed at
442 LHG to up-slope wind transport and at NamCo to aerosols descending from higher
443 layers and enhanced afternoon transport from the west (Xu et al., 2018). A complex
444 diurnal pattern of PM₁ was observed at Waliguan, influenced by diffusion conditions,
445 wind directions, and air mass sources, including afternoon air masses from the northeast,
446 which likely carried industrial pollutants (Zhang et al., 2019). At Motuo, the diurnal
447 pattern of PM₁ was relatively stable, with two weak peaks linked to local combustion
448 activities in the late morning and evening. Ngari exhibited relatively higher nighttime
449 loadings and lower daytime loadings, mainly due to the variations of PBL height.
450 Bayanbulak, on the other hand, had relatively low and stable PM₁ mass throughout the
451 entire day due to its background location. In contrast, urban Lhasa displayed two
452 pronounced peaks correlating with primary emissions during morning and evening rush

453 hours (Zhao et al., 2022). Although the diurnal pattern of PM_{10} were mainly shaped by
454 mountain-valley winds and PBL height in those remote sites, and primary emissions in
455 the urban site in this study, secondary formation processes, including photochemical
456 oxidation and aqueous-phase reactions were also played a key role in the formation of
457 inorganic and organic aerosol species. This could be evidenced by the afternoon peaks
458 of oxygenated OA (OOA) components observed across almost all the sites, which were
459 commonly formed by photo-chemical processes (Xu et al., 2018; Zhang et al., 2018;
460 Zhang et al., 2019; Zhao et al., 2022).

461 **4.3 High-resolution mass spectrum and elemental ratios of organic aerosol**

462 The high-resolution mass spectrum (HRMS) and elemental ratios of OA were
463 determined to identify the possible sources, formation and evolution mechanisms, and
464 oxidation states at each site. A direct comparison of the average O/C ratios from the
465 eight field campaigns was presented in Fig. 5a. It is apparent that the O/C ratios at the
466 remote sites of QOMS, Motuo, NamCo, Ngari, Waliguan, and LHG typically reached
467 or exceeded 1.0, indicating highly oxidized OA. In contrast, Bayanbulak exhibited a
468 lower O/C ratio of 0.69, and the urban site of Lhasa showed an even lower ratio of 0.44.
469 These variations in O/C ratios across sites primarily reflect differences in OA sources
470 and aging processes. Remote sites in the TP were generally received well-mixed and
471 aged OA due to long-range transport (Xu et al., 2018; Zhang et al., 2018; Zhang et al.,
472 2019). Meanwhile, local emissions from activities like cooking, traffic, and biomass
473 burning significantly contribute to the OA in urban Lhasa, resulting in a comparatively
474 low O/C ratio (Zhao et al., 2022). This pattern of higher O/C ratios at remote sites and
475 lower ratios at urban sites were observed in previous findings across China, such as O/C
476 ratios of 0.98, 1.11, and 1.16 observed at Mt. Wuzhi (Zhu et al., 2016), Mt. Yulong
477 (Zheng et al., 2017), and LHG (Xu et al., 2015), respectively, versus urban sites where
478 O/C ratios typically fell below 0.5 at most urban sites (Zhou et al., 2020). The Van
479 Krevelen diagram, which plots H/C versus O/C ratios to illustrate changes in OA
480 elemental composition due to atmospheric aging, shows an overall slope of -0.66 for
481 the bulk OA across all campaigns. This result is comparable to slopes of -0.58 and
482 -0.47 obtained in previous studies, further illustrating common pathways in OA aging
483 (Chen et al., 2015; Zhou et al., 2020).

484 The average HRMSs of OA between the remote site (Waliguan) and the urban site
485 (Lhasa) were directly compared to investigate the difference in ionic compositions (Fig.

486 5c). Waliguan was selected for comparison due to its representation of overall highly
487 aged OA, a characteristic shared with other remote sites (Fig. S3). The HRMS of OA
488 at Waliguan and Lhasa displayed significant differences. At Waliguan, the m/z 44,
489 predominantly composed of CO_2^+ and a key marker for OOA, was the most prominent
490 peak (18%) in the OA HRMS. The CO_2^+ and its related ions (CO^+ , H_2O^+ , HO^+ and
491 O^+) together contributed over 41% of the total OA signals. Additionally, two
492 oxygenated ion fragments ($\text{C}_x\text{H}_y\text{O}_1^+$ and $\text{C}_x\text{H}_y\text{O}_2^+$) accounted for as much as 66% of
493 the total OA signals (Fig. 5c), suggesting the highly oxygenated nature of OA at this
494 remote site. In contrast, the OA HRMS at Lhasa was remarkably similar to those
495 observed in urban environments, with significant contributions from four m/z values at
496 43, 55, 57, and 60. These ions are recognized as markers for less oxidized OA or
497 primary emissions related to traffic, cooking, and biomass burning (Zhang et al., 2005a;
498 Alfara et al., 2007; He et al., 2010), making up a significant contribution to the OA
499 signals in Lhasa. Specifically, non-oxygenated ion fragments (C_xH_y^+) contributed as
500 much as 64.5% of the total OA in Lhasa, whereas the oxygenated fragments accounted
501 for only 33.6%. This pattern of fresh ion fragments in the OA HRMS in Lhasa is
502 comparable to those measured at other urban cities, such as 56% and 59% in Lanzhou
503 (Xu et al., 2014b; Xu et al., 2016), 51.2% in Nanjing (Wang et al., 2016), and 51.2% in
504 New York (Sun et al., 2011).

505 **4.4 OA components from PMF source apportionment**

506 Source apportionments of OA were performed using PMF analysis on OA HRMS data
507 for each field campaign. Figure 6 presents the average mass contributions of OA
508 components from the selected 2–4 factor solutions across eight field campaigns, while
509 Figure S4 details the specific HRMS signatures for each OA component. In regions
510 with limited local emissions but significant influence from regional transport, such as
511 NamCo, LHG, and Bayanbulak, two secondary OA factors with different oxidation
512 degrees, namely a less oxidized OOA (LO-OOA) and a more oxidized OOA (MO-
513 OOA), were identified. For instance, during the NamCo campaign, the MO-OOA and
514 LO-OOA accounted for 59.0% and 41.0% of the total OA mass, with average O/C ratios
515 of 0.96 and 0.49, respectively. The Bayanbulak campaign exhibited a similar result,
516 with MO-OOA (average O/C of 1.12) and LO-OOA (average O/C of 0.55) accounting
517 for 66.3% and 33.7% of the OA mass, respectively. Contrastingly, the LHG campaign
518 revealed a different pattern, with 24.9% MO-OOA and 75.1% LO-OOA, albeit with

519 higher O/C ratios of 1.29 and 1.08, respectively. Note that the properties of each OOA
520 factor could be different across the locations in the TP despite the same name.
521 Additionally, biomass-burning-related OA (BBOA) was also a prevalent component in
522 the TP. At QOMS, the OA was composed by 42.4% MO-OOA, 43.9% BBOA, and 13.9%
523 nitrogen-containing OA (NOA), with average O/C ratios of 1.34, 0.85, and 1.10,
524 respectively. The high O/C ratio and significant contributions of BBOA and NOA at
525 QOMS were linked to biomass burning emissions transported from South Asia during
526 the pre-monsoon season (Cong et al., 2015; Zhang et al., 2018; Kang et al., 2019a). At
527 Waliguan, the OA was composed by 34.4% MO-OOA, 40.4% relatively aged BBOA
528 (agBBOA), 18.3% BBOA, and 6.9% hydrocarbon-like OA (HOA), with average O/C
529 ratios of 1.42, 1.02, 0.69, and 0.33, respectively. The agBBOA exhibited an enhanced
530 contribution to OA as the OA mass concentration increased, ranging from ~10% to 70%
531 when OA mass varied from $<1.0 \mu\text{g m}^{-3}$ to $7 \mu\text{g m}^{-3}$ (Zhang et al., 2019). High
532 contributions of BBOA at Waliguan were associated with regional transport of biomass
533 burning emissions from areas in the northeast (Zhang et al., 2019). At Ngari, the OA
534 was composed by 43.7% MO-OOA, 28.5% LO-OOA, and 27.8% BBOA, with average
535 O/C ratios of 1.43, 1.00, and 0.56, respectively. In contrast, the Motuo exhibited OA
536 components of 36.9% MO-OOA, 46.9% LO-OOA, and 16.2% BBOA, with O/C ratios
537 of 1.30, 1.11, and 0.25, respectively. The lower BBOA contribution and O/C ratio at
538 Motuo suggest a weaker local biomass burning emissions. At urban Lhasa, four OA
539 factors were identified including an OOA with O/C ratio of 0.54 and three primary OA
540 components, i.e., BBOA (O/C of 0.13), cooking-related OA (COA, O/C of 0.12), and
541 HOA (O/C of 0.11). These components were markedly different from those at the above
542 remote sites with the three primary OA components accounting for more than 60% of
543 the total OA, suggesting the abundant primary aerosol sources from the residential
544 activities. In addition, the BBOA contribution increased significantly (up to 36%)
545 during a major local festival in Lhasa, suggesting the crucial aerosol source from
546 biomass burning during religious activities in the city (Zhao et al., 2022).
547 In summary, our study identified diverse OA components with varying O/C ratios at
548 different sites, indicating the heterogeneity of sources and oxidation states of OA across
549 the TP regions.

550 **4.5 Number concentrations of submicron aerosols and cloud condensation nuclei**

551 The measurement of particle number size distribution (PNSD) was useful for studying

552 the formation and growth mechanisms of aerosol particles in the atmosphere. Figure 7a
553 shows the high-resolution temporal variations of the PNSDs during four field
554 campaigns (QOMS, Motuo, LHG, and Lhasa), revealing significant variability in
555 number concentrations and size distribution patterns across the different sites. On
556 average, the total number concentration was 709.3 and 3994.4 cm^{-3} at QOMS and
557 Lhasa, respectively, while it was 1639.2 and 1462.0 cm^{-3} at Motuo and LHG. Notably,
558 the variations in particle number concentrations were not consistent with mass
559 concentrations measured from the HR-ToF-AMS at the four sites (Table 2). For instance,
560 although the PM_{10} mass concentration at Lhasa was comparable to that at QOMS (4.7
561 versus 4.4 $\mu\text{g m}^{-3}$), the number concentration at Lhasa was more than five times higher
562 than that at QOMS. This inconsistency was mainly related to the difference on size
563 distribution at different sites. As mentioned above, submicron aerosols at QOMS were
564 predominantly secondary due to long-range transport from South Asia and
565 characterized by accumulation mode size. In contrast, Lhasa exhibited fresher aerosols,
566 emitted from local residential activities and characterized by Aitken mode size. The
567 variation in submicron aerosol sizes across the TP was further evidenced by the peak
568 diameters in the average mass and number size distributions (Figs. 4a and 7b). For
569 instance, the average OA mass size distribution peaked at 510.2 and 430.5 nm in D_{va}
570 for QOMS and Motuo, respectively. Meanwhile, the average number size distributions
571 at these two sites had peak at 109.4 and 131.0 nm in D_m . In contrast, Lhasa displayed
572 significantly smaller peak diameters of only 228.1 nm in D_{va} and 28.9 nm in D_m .

573 New particle formation (NPF) events were observed at several sites in our study.
574 Typically, an NPF event is characterized by a rapid burst in nucleation mode followed
575 by the subsequent growth into larger particles, as defined as banana-shaped temporal
576 developments in the PNSD (Dal Maso et al., 2005). Figure 7a displays the cases of
577 banana-shaped patterns in the PNSD, which were frequently observed at urban Lhasa.
578 Throughout the 27-day Lhasa campaign, a total of 10 NPF events were observed (Zhao
579 et al., 2022). In contrast, such banana-shaped pattern in the PNSD was relatively rare
580 at the other three remote sites (QOMS, Motuo, and LHG), which might be related to
581 their predominance of long-range transported aerosol with overall highly-aged states
582 and limited gaseous precursors.

583 Cloud condensation nuclei (CCN) is a distinct class of atmospheric aerosol particles
584 which could be activated as cloud droplets at a certain supersaturated water vapor

585 condition and played important roles in cloud formation, precipitation, climate change,
586 and regional hydrological cycle (Andreae and Rosenfeld, 2008). Across the TP field
587 campaigns, CCN measurements were conducted at three sites: Motuo, Waliguan, and
588 LHG. The temporal variations of CCN number concentrations at each SS exhibited a
589 similar trend with the total number concentration from the SMPS measurement and the
590 PM_{10} mass concentration from the HR-ToF-AMS measurement during each campaign.
591 On average, the CCN number concentrations at Motuo were 974.0, 1142.6, 1240.1,
592 1296.5, and 1337.9 cm^{-3} at SS level of 0.2%, 0.4%, 0.6%, 0.8%, and 1.0%, respectively.
593 At Waliguan, relatively comparable average values of 233.7, 857.8, 1138.7, 1313.1, and
594 1407.0 cm^{-3} were observed at corresponding SS levels. In contrast, LHG exhibited
595 significantly lower average CCN concentration of 120.5, 340.1, 417.8, 468.0, and 504.5
596 cm^{-3} at the same SS levels, respectively (Table 2). The lower CCN number
597 concentrations at LHG compared to Waliguan and Motuo were consistent with its lower
598 PM_{10} mass loading. Comparing with other regions, the CCN number concentrations at
599 the three TP sites were almost an order of magnitude lower than those observed in
600 polluted urban environments or from specific combustion emissions. For instance, CCN
601 concentrations reached 12963 cm^{-3} ($SS = 0.70\%$) in Wuqing, 9890 cm^{-3} ($SS = 0.86\%$)
602 in Beijing (Deng et al., 2011; Gunthe et al., 2011), 7913 cm^{-3} ($SS = 0.70\%$) at Panyu in
603 the Pearl River Delta, as well as 11565 cm^{-3} ($SS = 0.87\%$) and 10000 cm^{-3} ($SS = 0.80\%$)
604 during unique biomass burning plumes (Rose et al., 2010; Zhang et al., 2020). However,
605 the CCN values from our study were comparable to those measured at eight remote
606 marine sites in the South China Sea (228–2150 cm^{-3} at $SS = 0.87\%$) and the amazon
607 rain forest (941 cm^{-3} at $SS = 0.74\%$) (Pöhlker et al., 2016; Atwood et al., 2017). These
608 comparisons again highlight the overall clean atmospheric condition in the TP.

609 **4.6 Aerosol optical properties and light absorptions from BC and BrC**

610 In this study, the parameters of B_{scat} , B_{abs} , and SSA of fine particles at 405 nm were
611 observed at the field campaigns of QOMS, Motuo, Waliguan, Ngari, and Lhasa, to
612 explore the variations in aerosol optical properties across the TP. On average, the B_{scat}
613 and B_{abs} at 405 nm during the five campaigns were 121.9, 44.9, 36.3, 8.9, and 2.1 Mm^{-1}
614 and 10.8, 7.0, 4.1, 3.6, and 1.9 Mm^{-1} , respectively. These values yielded average SSA
615 values of 0.89, 0.83, 0.86, 0.67, and 0.52, correspondingly (Fig. 8a and Table 2). The
616 B_{scat} and B_{abs} values at the TP sites were significantly lower than those reported in
617 various urban areas in China, such as 459.5 and 47.2 Mm^{-1} at 630 nm in Beijing (Xie

618 et al., 2019), 272 and 31 Mm^{-1} at 532 nm in Xi'an (Zhu et al., 2015), and 418 and 91
619 Mm^{-1} at 540 nm in Guangzhou (Andreae et al., 2008), again suggesting the overall
620 clean atmospheric condition in the TP. Although the PM_{10} mass concentrations at QOMS
621 was comparable to or even lower than those at the other four sites, QOMS exhibited the
622 highest B_{scat} , B_{abs} , and SSA values. This discrepancy may be due to variations in aerosol
623 chemical compositions and their mass scattering and absorbing efficiencies. In contrast,
624 Lhasa exhibited a significantly lower SSA compared to the other four remote sites,
625 suggesting a prevalence of fresh aerosols in the urban environment. On the other hand,
626 aerosols at the four remote sites were highly aged, leading to significant photobleaching
627 of BrC chromophores and an obvious decrease in their light absorptivity.

628 Real-time online measurements of particle B_{abs} at seven wavelengths ranging from 370
629 to 950 nm were also conducted using an aethalometer at QOMS, NamCo, and Waliguan,
630 respectively, to explore regional variations in aerosol absorption properties across the
631 TP. Overall, the multi-wavelength B_{abs} decreased significantly with the increasing
632 wavelength during the three measurement campaigns, with fitted AAE values of 1.73,
633 1.28, and 1.12, respectively (Fig. 8b). The average B_{abs} at the shortest wavelength of
634 370 nm was 13.40, 3.25, and 2.66 Mm^{-1} at the three sites, respectively (Table 2).
635 Despite a relatively low PM_{10} mass concentrations at QOMS, the B_{abs} at 370 nm was
636 five times higher than that at Waliguan, mainly due to a higher contribution of light-
637 absorbing aerosol components in the southern TP. Specifically, OA and BC together
638 accounted for nearly 80% of the total PM_{10} at QOMS, whereas this contribution
639 decreased to only 37.5% at Waliguan. The obviously higher AAE at QOMS also
640 suggested a dominant light-absorbing contribution from BrC or the significant lensing
641 effect of coated BC (Zhang et al., 2021). The inserted plots in Fig. 8b illustrate
642 significant decreases in particle $B_{\text{abs,BC}}$ and $B_{\text{abs,BrC}}$ to total B_{abs} with increasing
643 wavelength, yet their contributions to total B_{abs} ($fB_{\text{abs,BC}}$ and $fB_{\text{abs,BrC}}$) varied inversely.
644 BC was the primary light-absorbing component across all the three sites, contributing
645 66.9%, 78.7%, and 77.6% to the total B_{abs} at 370 nm at QOMS, NamCo, and Waliguan
646 sites, respectively; its contribution increased apparently with longer wavelengths (Table
647 2). Conversely, BrC showed more significant contributions to total B_{abs} at shorter
648 wavelengths. For instance, the average $B_{\text{abs,BrC}}$ at 370 nm were 4.42, 0.69, and 0.60
649 Mm^{-1} at the three sites, respectively, ultimately contributing 33.1%, 21.3%, and 22.4%
650 to the total B_{abs} . The significantly higher values of total B_{abs} , $B_{\text{abs,BC}}$, $B_{\text{abs,BrC}}$, and

651 $f_{B_{abs}, BrC}$ in the southern TP region could be related to the important contributions of
652 light-absorbing CAs from transported biomass burning emissions (Xu et al., 2020,
653 2022).

654 **4.7 Estimation of aerosol radiative forcing in the different TP regions**

655 Atmospheric aerosols play a significant role in impacting Earth's climate systems
656 through affecting solar radiation and exerting a positive forcing on the energy budget
657 (Bond and Bergstrom, 2006). In this study, aerosol direct radiative forcings (DRF)
658 caused by BC, organic carbon (OC), and water-soluble ions (WSIs) are estimated,
659 respectively, using Santa Barbara DISORT (Discrete Ordinate Radiative Transfer)
660 Atmospheric Radiative Transfer (SBDART) model (Ricchiuzzi et al., 1998). A detailed
661 introduction and operation of this model are described in Text S6 in the supplementary
662 material. Since the model's performance is evaluated and calibrated by comparing the
663 values with measurements from Aethalometer and PAX results, the aerosol DRF
664 estimations are limited to QOMS, NamCo, and Waliguan. Furthermore, these three sites
665 are located in the southern, central, and northern regions of the TP, respectively, which
666 allows for an exploration of regional variations in aerosol DRF across the TP.

667 Figure 9 presents the results of DRFs attributed to BC, OC, and WSIs across the three
668 campaigns. BC demonstrated a pronounced warming effect at the top of the atmosphere,
669 with average DRF values of $+2.5 \pm 0.5$, $+2.1 \pm 0.1$, and $+1.9 \pm 0.1 \text{ W m}^{-2}$ during the
670 QOMS, Waliguan, and NamCo campaigns, respectively. In contrast, BC exhibited a
671 noticeable cooling effect at the earth's surface, with average DRF values of -4.7 ± 0.8 ,
672 -4.1 ± 0.2 , and $-3.7 \pm 0.1 \text{ W m}^{-2}$ across the three campaigns. The combination of these
673 two effects resulted in significantly high net atmospheric forcing by BC, amounting to
674 $+7.3 \pm 1.2$, $+6.2 \pm 0.3$, and $+5.6 \pm 0.2 \text{ W m}^{-2}$ for the three campaigns. These results
675 suggest the important radiative effect of BC in the TP, especially in the southern region
676 influenced by long-range transported biomass burning emission from South Asia. In
677 contrast, OC and WSIs exhibited cooling effects at both the top of the atmosphere and
678 the Earth's surface, characterized by negative and relatively low average DRFs.
679 Consequently, the net atmospheric forcings for OC and WSIs were notably lower
680 compared to BC across the three campaigns, with values of $+2.0 \pm 1.2$, $+0.7 \pm 0.2$, and
681 $+0.9 \pm 0.7 \text{ W m}^{-2}$ for OC, and $+1.9 \pm 0.8$, $+1.4 \pm 0.6$, and $+1.2 \pm 0.2 \text{ W m}^{-2}$ for WSIs
682 at QOMS, Waliguan and NamCo, respectively. Interestingly, at QOMS, the average
683 atmospheric DRF of OC accounted for 27.3% of that of BC, whereas at Waliguan and

684 NamCo, the fractions were only 11.1% and 15.7%, respectively. The higher
685 atmospheric DRF observed at QOMS suggests a dominant contribution from light-
686 absorbing BC and BrC aerosols compared to Waliguan and NamCo.

687 It was worth noting that the simulations of DRF effects in this study were only
688 conducted at three specific sites over limited periods. Future research should focus on
689 long-term comprehensive measurements and DRF simulations across the entire TP to
690 enhance our understanding of aerosol impacts on regional climate.

691 **4.8 Long-range transport of aerosols from surrounding areas**

692 To further understand the potential sources and specific transport pathways of aerosols
693 at each site, particularly for those remote sites, three- or five-days air mass back
694 trajectories were calculated during each measurement period at an ending height of 500
695 m above ground level every 6h using the Hybrid Single Particle Lagrangian Integrated
696 Trajectory (HYSPLIT) model (Draxler and Rolph, 2003). The cluster analysis on the
697 trajectories was based on the total spatial minimum variance method. Figure 10 displays
698 the average backward trajectory clusters across all eight field campaigns.

699 In general, distinct air mass sources were identified across the TP. The five sites (QOMS,
700 Motuo, Lhasa, NamCo, and Ngari) located in the southern or south-central part of the
701 TP generally showed air masses from south or southwest with different transport
702 distances and pathways during their measurement periods. For instance, during the
703 QOMS campaign, 38% of the air masses originated from the west, covering a long
704 distance, while another 40% originated from the southwest, covering a shorter distance.
705 In the Motuo campaign, two major clusters both originated from the southwest with
706 different distances (77% at shorter distance compared to only 13% at a longer distance).
707 Similarly, during the NamCo campaign, two major clusters with comparable
708 contributions (37% and 34%) and transport distances but different pathways were
709 identified from the south. In the Ngari campaign, air masses showed similar transport
710 distances, with 56% originating from southwest and 26% from the south. These air mass
711 clusters originating from the south of the TP generally traverse polluted regions in South
712 Asia such as the Indo-Gangetic Plain, Nepal, and Bangladesh, carrying significant
713 amounts of pollutants, particularly biomass-burning emissions into the inland of the TP.
714 In contrast, air masses at northern sites were primarily influenced by Westerlies and
715 East Asian monsoon during summer season. In the campaigns of Bayanbulak, air

716 masses originated from the west with different distances, i.e., 69% in relatively shorter
717 distance versus 18% in a longer distance. During the LHG campaign, the air masses
718 originated from the northwest of the site with 63% in longer distance but 27% in shorter
719 distance. For the Waliguan campaign, the air mass originated from two distinct
720 directions. Most of air masses (57%) came from the northeast of the site with a shorter
721 distance, while the remaining air masses originated from the west and northwest with
722 longer distances.

723 In summary, significant variations in air mass sources and transport pathways were
724 observed across the TP, particularly between southern and northern regions. These
725 differences play a crucial role in shaping the different physiochemical and optical
726 properties of aerosols across the TP regions.

727 **5 Dataset limitations and applications**

728 Our dataset was compiled from eight short-term intensive field observations across the
729 TP utilizing a suite of high-resolution online instruments. However, it is important to
730 note that our dataset does have certain limitations that proved to be quite challenging to
731 address in these remote regions.

732 The primary limitation stems from the short and inconsistent measurement periods
733 across different observational years and seasons at different sites, impeding robust
734 comparisons of aerosol properties across the TP. This limitation also hampers the ability
735 to ascertain long-term and seasonal characteristics. The harsh natural environments,
736 challenging weather conditions, limited logistical support and instruments, and
737 stringent instrumental requirements (such as the necessity for comprehensive field
738 stations with stable power supply) presented significant challenges during our field
739 observations in these remote TP regions. It is worth noting that online HR-ToF-AMS
740 observations, such as the one we conducted, are predominantly short-term and intensive
741 observations carried out worldwide due to the instability issues and its challenging
742 maintenance required for long-term measurements. The short-term intensive
743 measurement is enough to capture and characterize the dynamic evolution of aerosol
744 properties at a high-time-resolution (Jimenez et al., 2009; Li et al., 2017). Long-term
745 high-time-resolution observation utilizing HR-ToF-AMS have been rarely conducted
746 thus far, even in urban environment with relatively favorable observational conditions
747 and logistic support compared to our remote TP sites. Consequently, performing

748 continuous long-term observations or simultaneous comparison at multiple sites in
749 these high-altitude remote TP regions, without stable power supply, is exceedingly
750 challenging. Furthermore, assessing the representativeness of each observation for the
751 spatial scale is particularly challenging due to the limited number of observatories
752 across the TP. Actually, these observatories have been strategically established based
753 on the representation of specific geographic and climatic features.

754 In addition to the limitations above, our team has made significant efforts to conduct
755 this comprehensive observation project over the past decade, aiming to study the
756 regional differences in aerosol sources and properties across the TP. The dataset
757 generated from our project represents the first and exclusive high-time-resolution
758 dataset focusing on atmospheric aerosol physicochemical and optical properties,
759 covering most region of the TP. The applications of this dataset in atmospheric science
760 are multifaceted. Firstly, the high-time-resolution observations offer crucial advantages
761 in understanding the rapid evolution and diurnal variations of aerosol properties during
762 a short period or special event. Furthermore, these observations are invaluable for
763 model simulation and verification, providing a wealth of data points that can be utilized
764 for assessing aerosol loading, chemical composition, size distribution, and other
765 parameters essential for model accuracy and validation. Such advantages are not
766 achievable with traditional off-line samplings, which typically exhibit low time
767 resolutions ranging from days to weeks. Secondly, the eight sites encompassed in our
768 project effectively represent a wide range of the TP. This is particularly noteworthy
769 considering the limited availability of observatories on the TP. These sites facilitate
770 comparisons of aerosol sources and properties among different regions. Thirdly, our
771 observations encompass a wide range of aerosol physical, chemical, and optical
772 parameters, including aerosol mass loadings, chemical compositions, size distribution,
773 diurnal variations, number concentrations, light scattering and absorption coefficients,
774 and so on. This comprehensive dataset plays a crucial role in fostering a profound
775 understanding of aerosol properties in the TP.

776 Overall, it is noteworthy that our online observational aerosol datasets, focusing on
777 multiple parameters with at least hourly-scale resolution at various sites across the TP,
778 are the only ones reported to date.

779 **6 Data availability**

780 The high-resolution online measurement datasets, encompassing aerosol physical,
781 chemical, and optical properties over the Tibetan Plateau and its surroundings in our
782 observation project have been released and are now available for download from the
783 National Cryosphere Desert Data Center
784 (<https://doi.org/10.12072/ncdc.NIEER.db2200.2022>; Xu, 2022). These datasets are
785 provided in an Excel file comprising eight worksheets. The first sheet of the Excel file
786 contains a concise description of the dataset, including the dataset name, observation
787 stations, sampling periods, online instruments used, and corresponding references. The
788 remaining seven sheets present the high-resolution measurement data obtained from the
789 online instruments employed during the eight campaigns. These instruments include
790 HR-ToF-AMS, SMPS, PAX, aethalometer, and CCN-100.

791 **7 Conclusions**

792 This study presents a comprehensive dataset encompassing aerosol physicochemical
793 and optical properties, with a particular focus on high-resolution size-resolved chemical
794 characteristics and sources of submicron aerosols, conducted at eight different sites of
795 the TP and its surroundings. The datasets collected offer insights into temporal and
796 diurnal variations, size distribution of PM₁ chemical compositions, HRMS and
797 chemical components of OA, particle light scattering and absorption coefficients, and
798 CCN number concentrations at different supersaturations in different campaigns.

799 The datasets offer valuable insights into regional variations in aerosol properties and
800 sources. In the southern TP region, atmospheric aerosols were primarily influenced by
801 biomass burning emissions transported from South Asia, resulted in high mass
802 contributions (>70%) of CAs and overall neutralized PM₁, as well as an enhanced light
803 absorption capability of aerosols. In contrast, in the northern TP, secondary inorganic
804 species, particularly sulfate, contributed significantly to total PM₁ due to the regional
805 transport of anthropogenic aerosol and gaseous precursor emissions from urban areas
806 in northwestern China. Furthermore, in contrast to the well-mixed, highly-aged, and
807 regionally transported aerosols observed in the remote sites, atmospheric aerosols in
808 the urban Lhasa were mainly originated from local primary sources such as cooking,
809 traffic vehicle exhausts, and biofuel combustion during the residential activities.
810 Consequently, these aerosol particles were relatively fresh, characterized by small size
811 and low oxidation degree, but exhibited a high frequency of NPF origins.

812 **Appendix A: Main Abbreviations**

TP	Tibetan Plateau
HR-ToF-AMS	high-resolution time-of-flight aerosol mass spectrometer
SMPS	scanning mobility particle sizer
PAX	photo-acoustic extinctionsimeter
CCN	cloud condensation nuclei
SS	supersaturation
PM ₁	submicron aerosol
BC	black carbon
BrC	brown carbon
OA	organic aerosol
SNA	sulfate, nitrate, and ammonium
D_m	mobility diameter
D_{va}	aerodynamic diameter
CE	collection efficiency
HRMS	high-resolution mass spectrum
PBL	planetary boundary layer
O/C	oxygen-to-carbon ratio
H/C	hydrogen-to-carbon ratio
N/C	nitrogen-to-carbon ratio
OM/OC	organic matter-to-organic carbon ratio
PMF	positive matrix factorization
OOA	oxygenated organic aerosol
LO-OOA	less oxidized oxygenated organic aerosol
MO-OOA	more oxidized oxygenated organic aerosol
BBOA	biomass-burning-related organic aerosol
agBBOA	aged biomass-burning-related organic aerosol
NOA	nitrogen-containing organic aerosol
HOA	traffic-related hydrocarbon-like organic aerosol
COA	cooking-related organic aerosol
PNSD	particle number size distribution
NPF	new particle formation
B_{scat}	light scattering coefficient
B_{abs}	light absorption coefficient
B_{ext}	light extinction coefficient
SSA	single scattering albedo
AAE	absorption Ångström exponents
$B_{abs,BC}$	light absorption coefficient from BC
$B_{abs,BrC}$	light absorption coefficient from BrC
OC	organic carbon
WSIs	water-soluble ions
DRF	direct radiative forcing

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 815 LZ, MZ, JS, JShi, YL, CX, YT, KL, XG, and QZ conducted the field measurements,
 816 JX, XZ, WZ, and YT analyzed the data. All authors reviewed and commented on the
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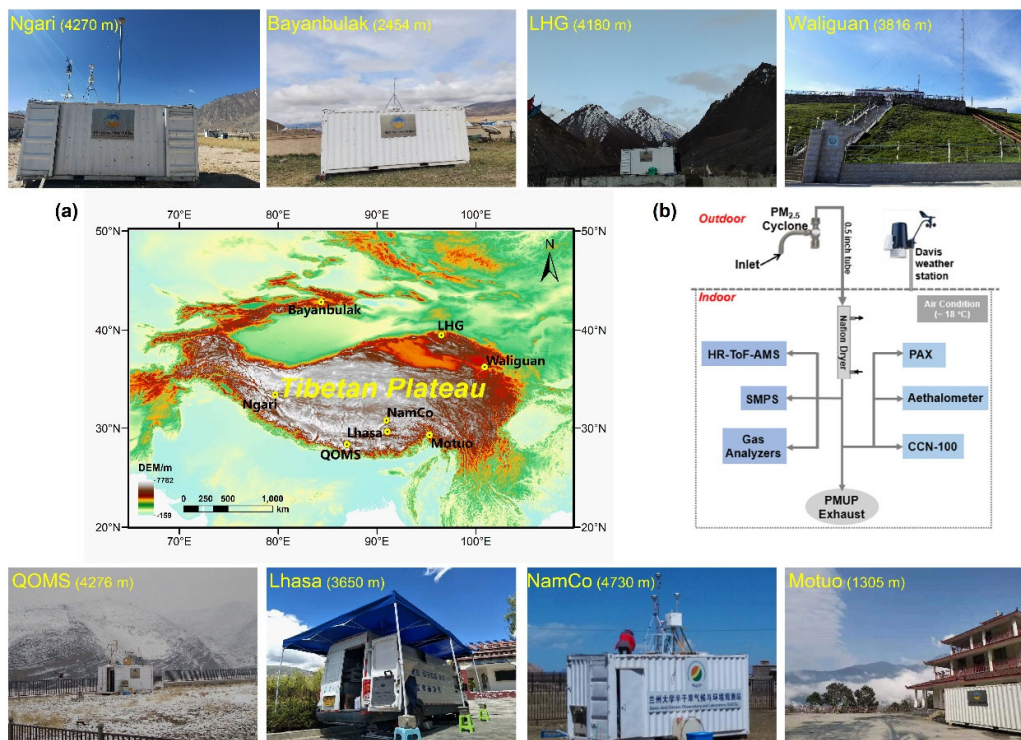
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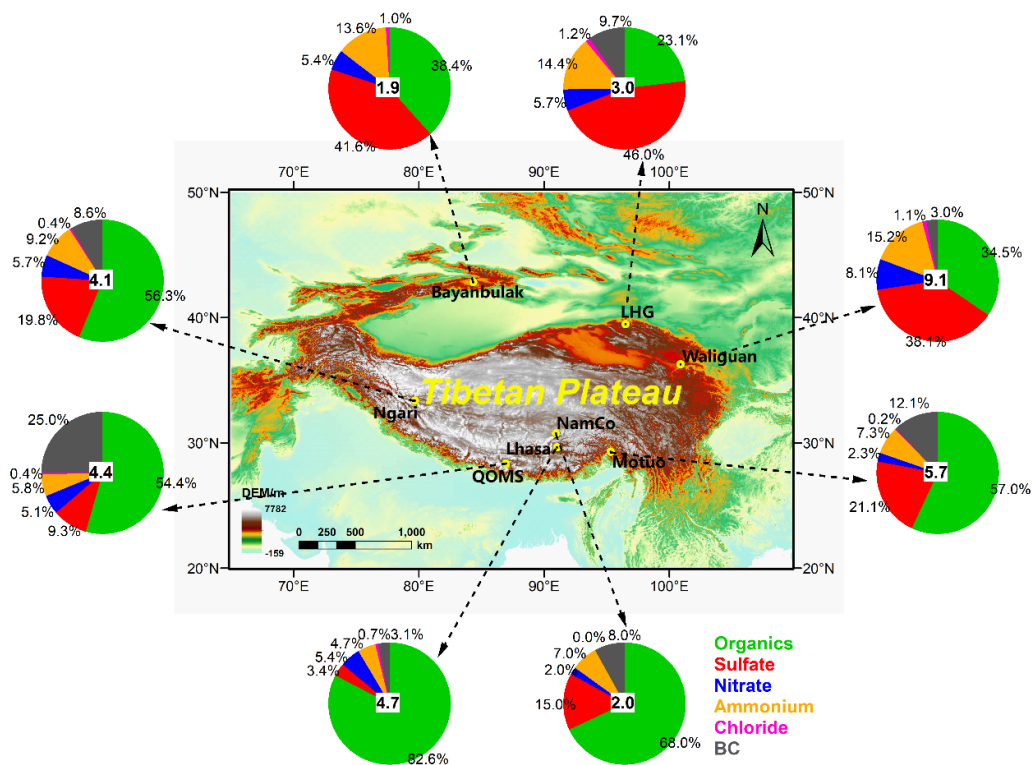
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1164 **Figures**

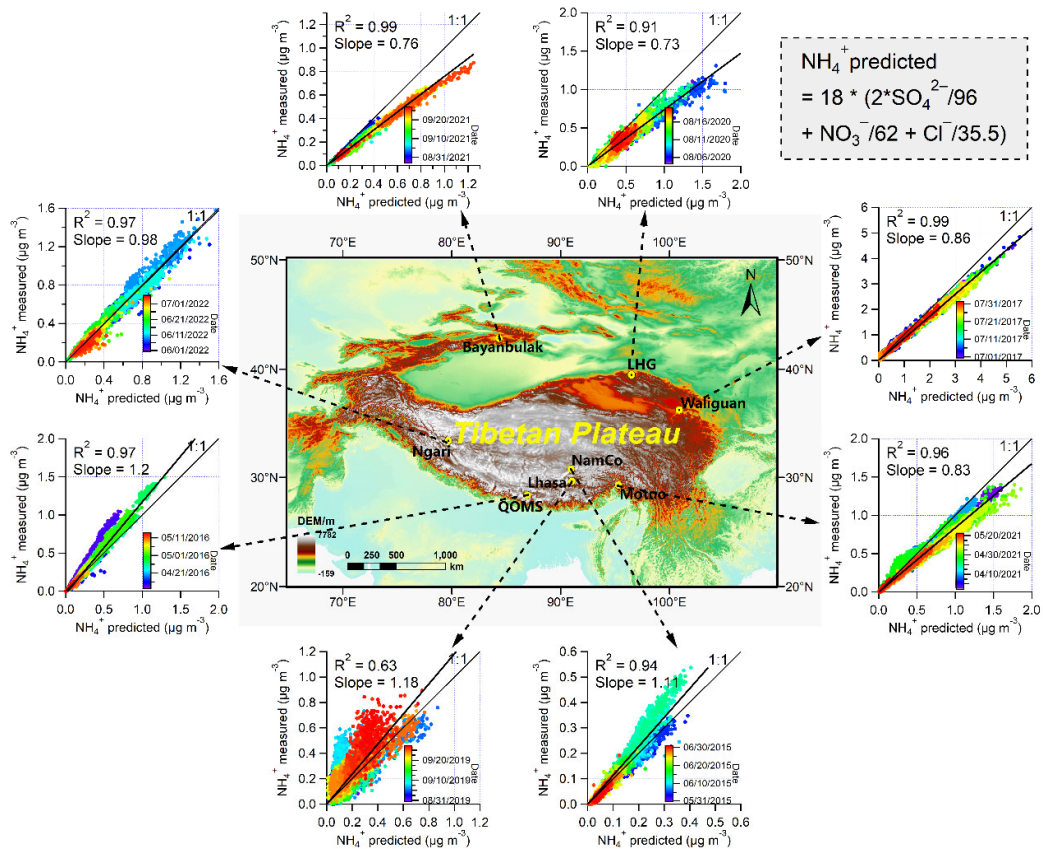


1165
 1166 **Figure 1.** (a) Geographical locations of the observation sites (see Table 1 for full name and
 1167 characteristics of each site) in the Tibetan Plateau and its surroundings in this study (The
 1168 geographical base map is created with ArcGIS). Fieldwork photographs illustrate the real
 1169 observation conditions and surroundings at each site. (b) The normal sampling setups of instruments
 1170 during the online aerosol observations.



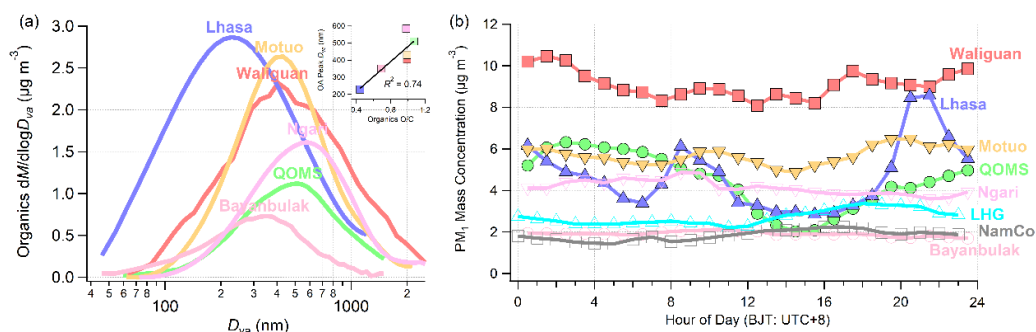
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1172 **Figure 2.** Regional distribution of average mass concentrations (values marked in the central of
 1173 each pie chart with unit of $\mu\text{g m}^{-3}$) and chemical compositions (percentage values around each pie
 1174 chart) of submicron aerosols (PM_{10}) during the eight online aerosol field measurements in the Tibetan
 1175 Plateau and its surroundings (The geographical base map is created with ArcGIS). The concentration
 1176 at each site is presented in ambient condition.



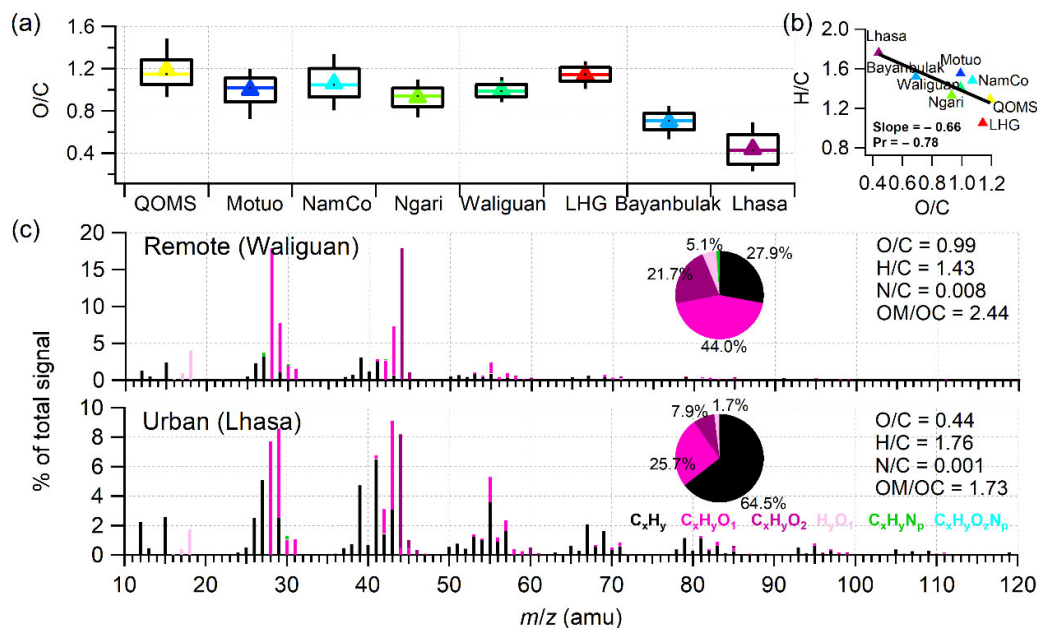
1177

1178 **Figure 3.** Regional difference of bulk acidity of submicron aerosols based on the scatterplot analysis
 1179 and linear regression of measured NH_4^+ versus predicted NH_4^+ during the eight aerosol field
 1180 measurement campaigns in the Tibetan Plateau and its surroundings (The geographical base map is
 1181 created with ArcGIS).



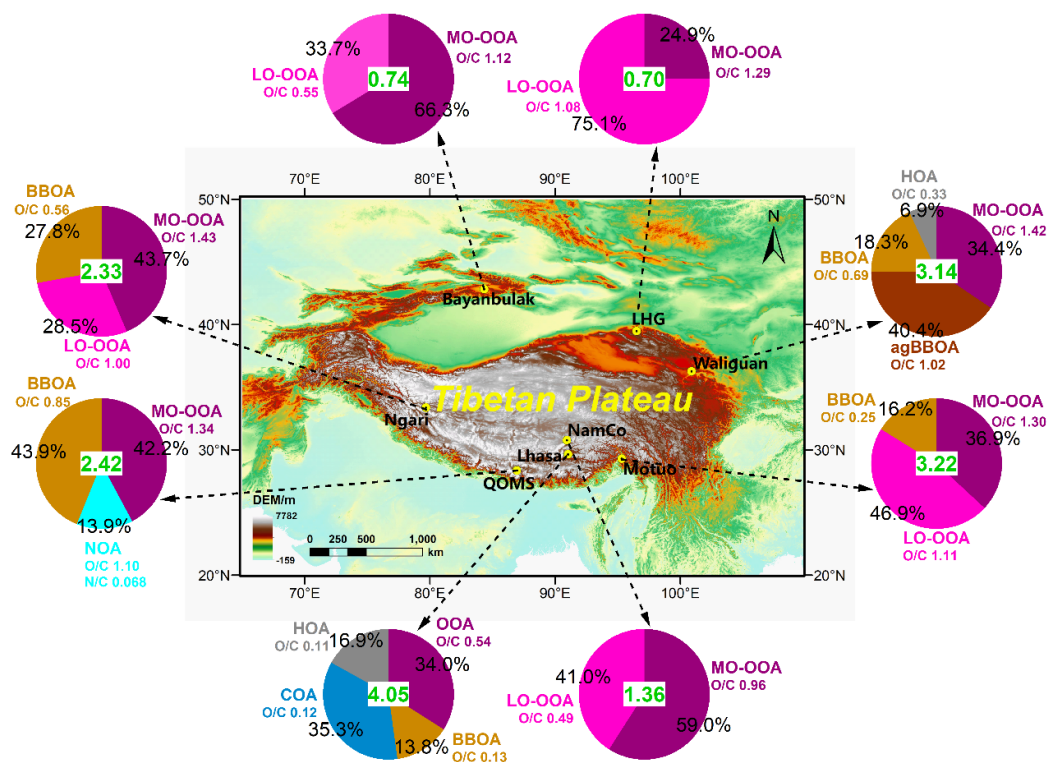
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1183 **Figure 4.** (a) Average size distributions of organic mass concentrations during six field
 1184 measurement campaigns in the Tibetan Plateau and its surroundings. (b) Diurnal variations of total
 1185 PM₁ mass concentrations during the eight field measurement campaigns in the Tibetan Plateau and
 1186 its surroundings. Insert graph in (a) is the scatter plot of peak diameters in their size distributions
 1187 versus the average O/C ratio of organics.



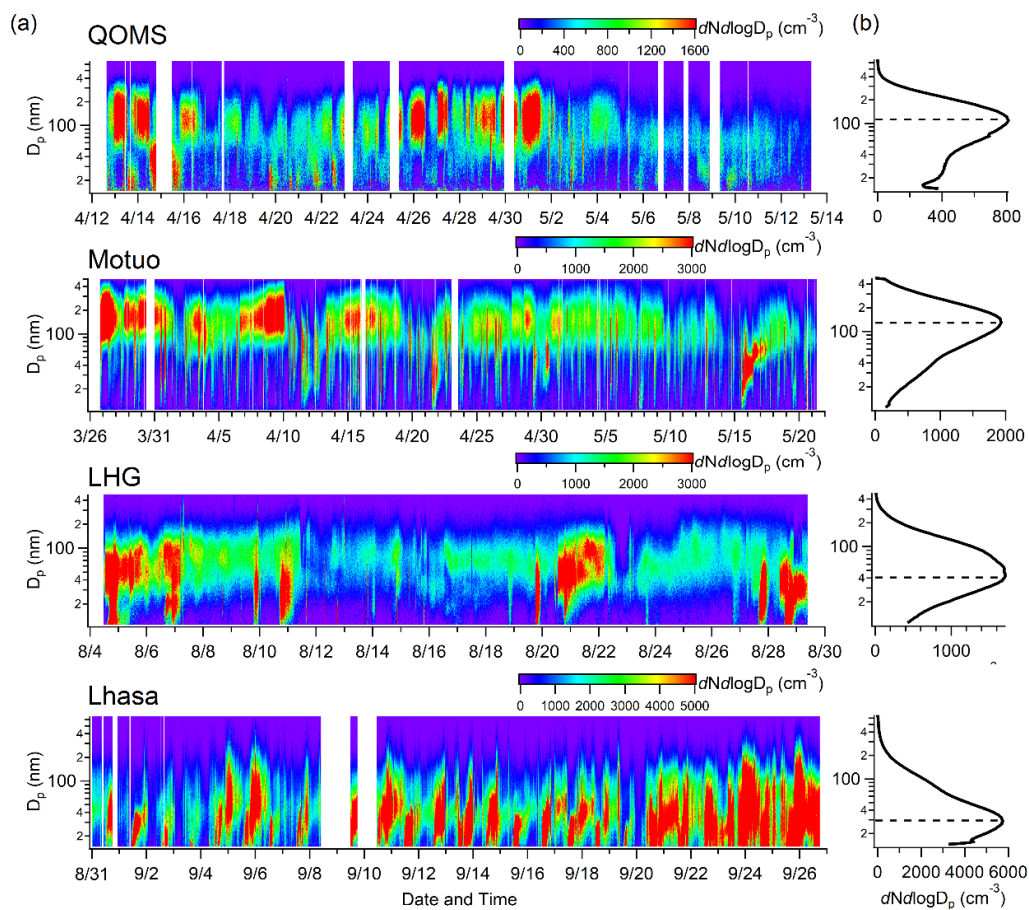
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1189 **Figure 5.** (a) Box plots of the average O/C ratios and (b) Van Krevelen diagram of H/C versus O/C
 1190 among the eight field measurement campaigns in this study. (c) The average HRMSs of OA colored
 1191 with different ion categories during the Waliguan and Lhasa measurement campaigns. The whiskers
 1192 of boxes indicate the 90th and 10th percentiles, the upper and lower boundaries of boxes indicate the
 1193 75th and 25th percentiles, the lines in the boxes indicate the median values, the markers indicate the
 1194 mean values, and similarly hereinafter.



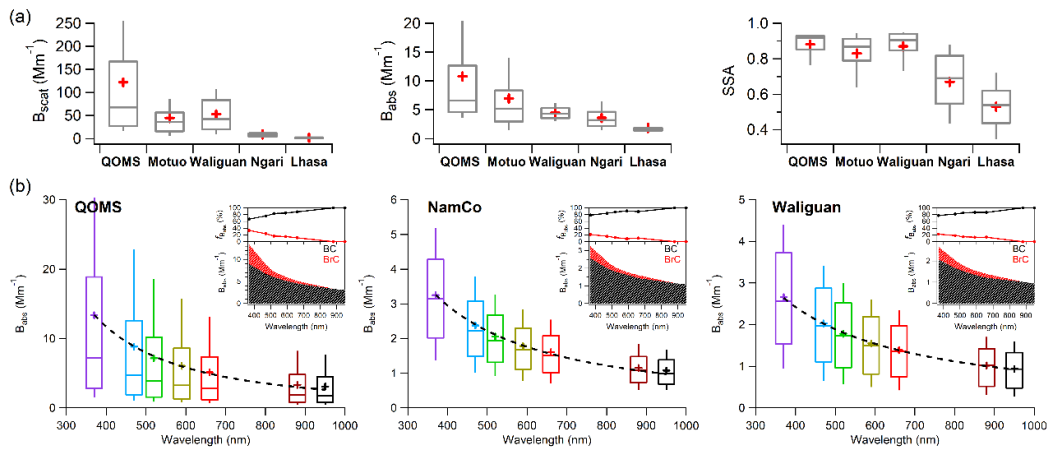
1195

1196 **Figure 6.** Regional distribution of OA components from PMF source apportionment during the eight
 1197 online aerosol field measurements in the Tibetan Plateau and its surroundings (The geographical
 1198 base map is created with ArcGIS). Values marked in the central of each pie chart are average OA
 1199 mass with unit of $\mu\text{g m}^{-3}$, while the percentage values around the pie chart are the mass contributions
 1200 of each OA component. The O/C ratio of each OA component is also marked around each pie chart.



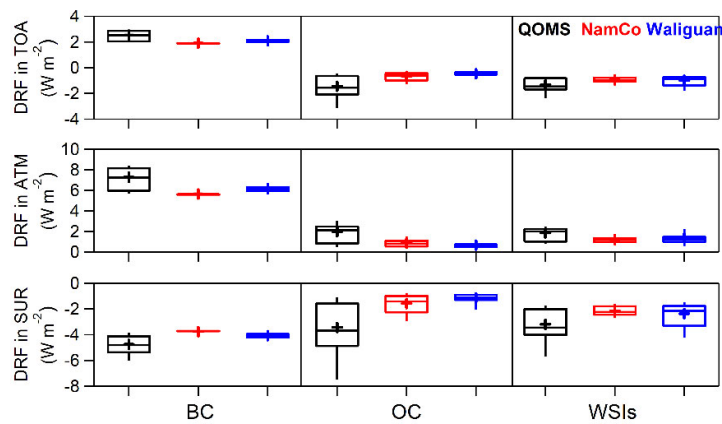
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1202 **Figure 7. (a)** Temporal variations of the size distributions of particle number concentrations during
 1203 the aerosol field measurement campaigns at QOMS, Motuo, LHG, and Lhasa sites. **(b)** The average
 1204 size distribution of particle number concentration during entire measurement period at each site.



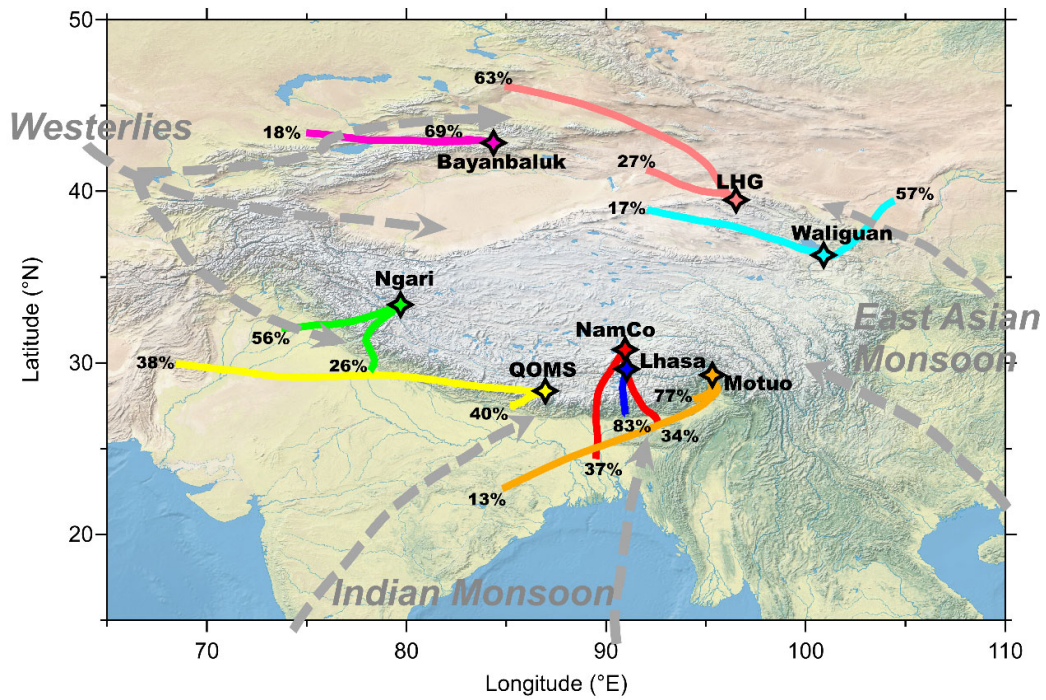
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1206 **Figure 8.** Box plots of (a) the average particle light scattering coefficient (B_{scat}), light absorption
 1207 coefficient (B_{abs}), and single scattering albedo (SSA) during the five aerosol field measurement
 1208 campaigns at QOMS, Motuo, Waliguan, Ngari, and Lhasa sites, and (b) the particle B_{abs} at seven
 1209 wavelengths measured by aethalometers at QOMS, NamCo, and Waliguan sites. The dashed lines
 1210 in the boxes in (b) show the power-law fit of the average B_{abs} as a function of wavelength. The
 1211 inserted plots in (b) are the apportioned contributions of BC and BrC to total B_{abs} at different
 1212 wavelengths.



1213

1214 **Figure 9.** Box-plots of the modelled direct radiative forcing (DRF) at the top of the atmosphere
 1215 (TOA), the atmosphere (ATM), and the earth's surface (SUR) caused by black carbon (BC), organic
 1216 carbon (OC), and water-soluble ions (WSIs) during the QOMS, NamCo, and Waliguan campaigns.



1217

1218 **Figure 10.** Air mass backward trajectory clusters during the eight field campaigns in the Tibetan
 1219 Plateau and its surroundings in our study (The geographical base map is created with Igor Pro). Only
 1220 major trajectory clusters during each field campaign are displayed with the contributions marked
 1221 correspondingly.

1222 **Tables**

1223 **Table 1.** Detailed information about the full name and geographic characteristic of observation
 1224 station, sample period, online instruments, and corresponding references during each aerosol field
 1225 measurement campaigns over the Tibetan Plateau and its surroundings in this study.

Station	Full Station Name	Lat. (°N)	Long. (°E)	Alt. (m)	Sample Period	Online Instruments					References	
						HR-ToF-AMS		SMPS	PAX	Aethal ometer		CCN -100
						MS	PToF					
QOMS	Qomolangma Station for Atmospheric and Environmental Observation and Research, Chinese Academy of Sciences	28.36	86.95	4276	12 April to 12 May 2016	✓	✓	✓	✓	✓	Zhang et al. (2018) An et al. (2019) Xu et al. (2020) Zhang et al. (2021) Xu et al. (2022)	
Motuo	Motuo County, Linzhi City, Tibet Autonomous Region, China	29.30	95.32	1305	26 Mar to 22 May 2021	✓	✓	✓	✓	✓	This study	
NamCo	Nam Co Station for Multisphere Observation and Research, Chinese Academy of Sciences	30.77	90.95	4730	31 May to 1 July 2015	✓				✓	Xu et al. (2018) Zhang et al. (2021)	
Ngari	Ngari Station for Desert Environment Observation and Research, Chinese Academy of Sciences	33.39	79.70	4270	1 Jun to 5 Jul 2022	✓	✓		✓		This study	
Waliguan	China Global Atmospheric Watch Baseline Observatory, Mount Waliguan Base	36.28	100.90	3816	1 July to 31 July 2017	✓	✓		✓	✓	Zhang et al. (2019) Zhang et al. (2020) Xu et al. (2020) Zhang et al. (2021) Xu et al. (2022)	
LHG	Qilian Observation and Research Station of Cryosphere and Ecologic Environment, Chinese Academy of Sciences	39.50	96.51	4180	4 August to 29 August 2020	✓		✓		✓	This study	
Bayanbulak	Bayanbulak Town, Hejing County, Bayingolin Mongolian Autonomous Prefecture, Xinjiang Uygur Autonomous Region, China	42.83	84.35	2454	29 August to 26 September 2021	✓	✓				This study	
Lhasa	Lhasa City, Tibet Autonomous Region, China	29.65	91.03	3650	31 August to 26 September 2019	✓	✓	✓	✓		Zhao et al. (2022)	

1226

1227 **Table 2.** Summary of the average values measured with various instruments during the eight aerosol
 1228 field measurement campaigns in the TP and its surroundings in this study.

Measurement items	QOMS	Motuo	NamCo	Ngari	Waliguan	LHG	Bayanbulak	Lhasa
HR-ToF-AMS measurements								
PM ₁ mass conc. ($\mu\text{g m}^{-3}$)	4.4	5.7	2.0	4.1	9.1	3.0	1.9 ^a	4.7
PM ₁ chemical compositions (%)								
OA	54.4	57.0	68.0	56.3	34.5	23.1	38.4	82.6
Sulfate	9.3	21.1	15.0	19.8	38.1	46.0	41.6	3.4
Nitrate	5.1	2.3	2.0	5.7	8.1	5.7	5.4	5.4
Ammonium	5.8	7.3	7.0	9.2	15.2	14.4	13.6	4.7
Chloride	0.4	0.2	0	0.4	1.1	1.2	1.0	0.7
BC	25.0	12.1	8.0	8.6	3.0	9.7	N/A	3.1
Peak diameter in mass size distribution (nm)								
OA	510.2	430.5		584.4	405.5		350.8	228.1
SNA	510.2	471.9		634.5	504.7		379.6	250.0
OA components (%)								
MO-OOA	42.2	36.9	59.0	43.7	34.4	24.9	66.3	
LO-OOA		46.9	41.0	28.5		75.1	33.7	
OOA								34.0
BBOA	3.9	16.2		27.8	18.3			13.8
agBBOA					40.4			
NOA	13.9							
HOA					6.9			16.9
COA								35.3
OA elemental ratios								
O/C	1.19	0.99	1.07	0.98	0.99	1.14	0.69	0.44
H/C	1.29	1.55	1.48	1.33	1.41	1.05	1.52	1.76
OM/OC	2.70	2.48	2.57	2.44	2.45	2.62	2.09	1.74
N/C	0.030	0.020	0.016	0.019	0.008	0.011	0.026	0.001
SMPS measurements								
Number conc. (cm^{-3})	709.3	1639.2				1462.0		3994.4
Peak diameter in PNSD (nm)	109.4	131.0				42.9		28.9
PAX measurements								
B_{scat} (Mm^{-1})	121.9	44.9		8.9	36.3			2.1
B_{abs} (Mm^{-1})	10.8	7.0		3.6	4.1			1.9
B_{ext} (Mm^{-1})	132.7	51.9		12.6	40.4			4.0
SSA	0.89	0.83		0.67	0.86			0.52
Aethalometer measurements								
$B_{\text{abs},370}$ (Mm^{-1})	13.40		3.25		2.66			
Absorption Ångström exponent	1.73		1.28		1.12			
$B_{\text{abs},\text{BrC},370}$ (Mm^{-1})	4.42		0.69		0.60			
$B_{\text{abs},\text{BC},370}$ (Mm^{-1})	8.94		2.56		2.06			
$fB_{\text{abs},\text{BrC},370}$ (%)	33.1		21.3		22.4			
$fB_{\text{abs},\text{BC},370}$ (%)	66.9		78.7		77.6			
CCN-100 measurements (cm^{-3})								
CCN number conc. (SS 0.2%)		974.0			233.7	120.5		
CCN number conc. (SS 0.4%)		1142.6			857.8	340.1		
CCN number conc. (SS 0.6%)		1240.1			1138.7	417.8		
CCN number conc. (SS 0.8%)		1296.5			1313.1	468.0		
CCN number conc. (SS 1.0%)		1337.9			1407.0	504.5		

1229 ^aonly non-refractory PM₁ is reported at Bayanbulak due to the absence of BC observation.

1230 **Table 3.** Summary of the average PM₁ mass concentrations ($\mu\text{g m}^{-3}$) measured by the Aerodyne
 1231 AMSs at various high-altitude and remote sites worldwide.

Observation Sites	Latitude (°N)	Longitude (°E)	Altitude (m a.s.l.)	PM ₁ mass ($\mu\text{g m}^{-3}$)	References
QOMS, China	28.36	86.95	4276	4.4	This study & Zhang et al. (2018)
Motuo, China	29.30	95.32	1305	5.7	This study
NamCo, China	30.77	90.95	4730	2.0	This study & Xu et al. (2018)
Ngari, China	33.39	79.70	4270	4.1	This study
Waliguan, China	36.28	100.90	3816	9.1	This study & Zhang et al. (2019)
LHG, China	39.50	96.51	4180	3.0	This study
Bayanbulak, China	42.83	84.35	2454	1.9 ^a	This study
Lhasa, China	29.65	91.03	3650	4.7	This study & Zhao et al. (2022)
NamCo, China	30.77	90.95	4730	1.06	Wang et al. (2017)
Mt. Yulong, China	27.20	100.20	3410	5.7	Zheng et al. (2017)
Menyuan, China	37.61	101.26	3295	11.4	Du et al. (2015)
Mt. Wuzhi, China	18.84	109.49	958	10.9	Zhu et al. (2016)
Mt. Jungfrauoch, Switzerland	46.55	7.98	3580	0.55	Fröhlich et al. (2015)
Mt. Jungfrauoch, Switzerland	46.55	7.98	3580	2.24	Zhang et al. (2007a)
Mt. Bachelor, USA	43.98	-121.69	2800	15.10	Zhou et al. (2017)
Mt. Whistler, Canada	50.01	-122.95	2182	1.91	Sun et al. (2009)
Mt. Cimone, Italy	44.18	10.70	2165	4.5	Rinaldi et al. (2015)
Puy de Dôme, France	45.77	2.95	1465	5.58	Frenay et al. (2011)
Sub-Antarctic Bird Island	-54.00	-38.04		0.46	Schmale et al. (2013)
Mace Head, Ireland	53.30	-9.80		1.53	Zhang et al. (2007a)
Hyytiala, Finland	61.90	24.30		2.04	Zhang et al. (2007a)
Storm Peak, USA	40.50	-106.70		2.11	Zhang et al. (2007a)
Duke Forest, USA	36.00	-79.10		2.82	Zhang et al. (2007a)
Chebogue, Canada	43.80	-66.10		2.91	Zhang et al. (2007a)
Okinawa Island, Japan	26.87	33.51		7.89	Jimenez et al. (2009)
Fukue Island, Japan	32.69	128.84		12.03	Takami et al. (2005)
Cheju Island, Korea	33.51	126.50		10.66	Jimenez et al. (2009)

1232 ^aonly non-refractory PM₁ is reported at Bayanbulak due to the absence of BC observation.