1	Spatial mapping of key plant functional traits in terrestrial
2	ecosystems across China
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#### 15 Abstract

16 Trait-based approaches are of increasing concern in predicting vegetation changes and 17 linking ecosystem structures to functions at large scales. However, a critical challenge for such 18 approaches is acquiring spatially continuous plant functional trait maps. Here, six key plant 19 functional traits were selected as they can reflect plant resource acquisition strategies and 20 ecosystem functions, including specific leaf area (SLA), leaf dry matter content (LDMC), leaf N 21 concentration (LNC), leaf P concentration (LPC), leaf area (LA) and wood density (WD). A total 22 of 34589 in-situ trait measurements of 3447 seed plant species were collected from 1430 sampling 23 sites in China and were used to generate spatial plant functional trait maps (~1 km), together with 24 environmental variables and vegetation indices based on two machine learning models (random 25 forest and boosted regression trees). To obtain the optimal estimates, a weighted average algorithm 26 was further applied to merge the predictions of the two models to derive the final spatial plant 27 functional trait maps. The models showed a good accuracy in estimating WD, LPC and SLA, with average  $\mathbb{R}^2$  values ranging from 0.48 to 0.68. In contrast, both the models had weak performance 28 29 in estimating LDMC, with average  $R^2$  values less than 0.30. Meanwhile, LA showed considerable 30 differences between the two models in some regions. Climatic effects were more important than 31 those of edaphic factors in predicting the spatial distributions of plant functional traits. Estimates 32 of plant functional traits in the northeast China and the Qinghai-Tibet Plateau had relatively high 33 uncertainties due to sparse samplings, implying a need of more observations in these regions in the 34 future. Our spatial trait maps could provide critical supports for trait-based vegetation models and 35 allow exploration into the relationships between vegetation characteristics and ecosystem 36 functions at large scales. The six plant functional traits maps for China with 1 km spatial 37 resolution are now available at https://figshare.com/s/c527c12d310cb8156ed2 (An et al., 2023).

#### 38 **1 Introduction**

39 Climate change has been affecting vegetation distributions and biogeochemical cycling globally 40 and altering their feedbacks to the climate system (Kirilenko et al., 2000; Finzi et al., 2011; 41 Jónsdóttir et al., 2022). Dynamic global vegetation models (DGVMs) are powerful tools for 42 predicting changes in vegetation and ecosystem-atmosphere exchanges (e.g., water, carbon and 43 nutrient cycling) in a changing climate (Foley et al., 1996; Peng, 2000). However, conventional 44 DGVMs are still insufficient realistic, largely due to their dependence on the plant functional types 45 (PFTs) assumption (Sitch et al., 2008; Yurova and Volodin, 2011; Scheiter et al., 2013). PFTs in 46 conventional DGVMs commonly have fixed attributes (mostly trait values) (Van-van Bodegom et 47 al., 2012; Wullschleger et al., 2014) that do not reflect plant adaptation to environments, limiting 48 the quantification of carbon-water-nutrient feedbacks between terrestrial ecosystems and the 49 atmosphere (Zaehle and Friend, 2010; Liu and Yin, 2013). Trait-based approaches can provide a 50 robust theoretical basis for developing the next generation of DGVMs (Van-van Bodegom et al., 51 2012; Sakschewski et al., 2015; Matheny et al., 2017). Plant functional traits, which are closely 52 associated with ecosystem functions (Diaz et al., 2004; Yan et al., 2023), can effectively reflect 53 response and adaptation of plants to environmental conditions (Myers-Smith et al., 2019; Qiao et 54 al., 2023).

55 Attempts to predict spatially continuous trait maps have been conducted at regional to global 56 scales (e.g., Madani et al., 2018; Moreno-Mart nez et al., 2018; Boonman et al., 2020; Loozen et 57 al., 2020; Dong et al., 2023). Webb et al. (2010) proposed that the environment creates a filtered 58 trait distribution along an environmental gradient, and such trait-environment relationships offer 59 fundamental supports to predict the spatial distributions of plant functional traits through 60 extrapolating local trait measurements. Boonman et al. (2020) mapped the global patterns of 61 specific leaf area (SLA), leaf N concentration (LNC) and wood density (WD) based on a set of 62 climate and soil variables. As the number of available regional and global trait databases increases 63 (Wang et al., 2018; Kattge et al., 2020), trait-environment relationships are becoming increasingly 64 quantitative and accurate (Bruelheide et al., 2018; Myers-Smith et al., 2019). Alternatively, remote 65 sensing approaches, such as empirical methods and physical radiative transfer models (e.g., partial least squares regression and PROSPECT model), have been developed to estimate plant 66 67 physiological, morphological and chemical traits (e.g., leaf chlorophyll content, SLA, LNC and 68 leaf dry matter content (LDMC)) (Darvishzadeh et al., 2008; Romero et al., 2012; Ali et al., 2016). 69 Vegetation indices, such as normalized difference vegetation index (NDVI) and enhanced 70 vegetation index (EVI), have been successful in estimating plant functional traits of croplands, 71 grasslands and forests (Clevers and Gitelson, 2013; Li et al., 2018; Loozen et al., 2018). Loozen et 72 al. (2020) demonstrated that EVI was the most important predictor for mapping the spatial pattern 73 of canopy nitrogen in European forests. Admittedly, a recent study has suggested that combining environmental variables and vegetation indices can improve the predictive accuracy of canopy 74

nitrogen compared to those based on vegetation indices alone (Loozen et al., 2020).

76 Although there have been reports on plant functional trait distributions in China in some 77 global or regional researches (e.g., Yang et al., 2016; Butler et al., 2017; Madani et al., 2018; 78 Moreno-Mart nez et al., 2018; Boonman et al., 2020), they-there are still large uncertainties in 79 characterizing the spatial distributions of plant functional traits in China. First, global studies 80 generally have relatively few- and unevenly distributed sampling sites across China (Butler et al., 81 2017; Madani et al., 2018; Boonman et al., 2020), impeding our understanding of the true spatial 82 characteristics of trait variability. Second, the spatial patterns of traits among these studies are 83 usually inconsistent. For example, Moreno-Mart nez et al. (2018) and Madani et al. (2018) 84 demonstrated that SLA values were low in the southeast areas but high in the southwest areas of 85 China, whereas Boonman et al. (2020) found the opposite. Third, most studies focused on leaf 86 traits (Yang et al., 2016; Loozen et al., 2018; Moreno-Mart nez et al., 2018), whereas traits 87 associated with the whole-plant strategies, such as WD, were ignored. Therefore, mapping and 88 verifying the spatial patterns of key functional traits that reflect the whole plant economics 89 spectrum in China is a top priority.

90 In this study, our main objective was to generate spatial maps for several key plant functional 91 traits, through combining field measurements, environmental variables and vegetation indices. We 92 selected six plant functional traits including SLA, LDMC, LNC, LPC, LA and WD. As key leaf 93 economics traits, SLA, LDMC, LNC and LPC were selected because they are closely linked to 94 plant growth rate, resource acquisition and ecosystem functions (Wright et al., 2004; Diaz et al., 95 2016). LA is indicative of the trade-off between carbon assimilation and water-use efficiency (Wright et al., 2017), and WD reflects the trade-off between plant growth rate and support cost, 96 97 with a higher WD linked to a lower growth rate, a higher survival rate and a higher biomass 98 support cost (King et al., 2006). For each plant functional trait, we predicted spatial patterns at a 1 99 km resolution using an ensemble modelling algorithm based on two machine learning methods 100 (i.e., random forest and boosted regression trees).

101 **2 Materials and Methods** 

#### 102 **2.1 Overview**

103 The spatial maps of plant functional traits in China were generated based on machine learning 104 methods trained by a large dataset of in-situ field measurements, environmental variables and 105 vegetation indices in three steps (Fig. 1). First, in-situ field measurements of six plant functional 106 traits were collected from TRY and China databases as well as published literature, and the PFTs 107 of plant species were classified based on plant growth form, leaf type and leaf phenology. Multiple 108 gridded predictors of climate, soil, topography and vegetation indices were used after avoiding the 109 collinearity among them. Second, random forest and boosted regression trees were used to train 110 the relationships between plant functional traits and predictors for each PFT individually. Third, the spatial abundance of each PFT within 1 km grid cell was calculated using land cover map (100 m). Community-weighted trait values within 1 km grid cell wereas calculated based on these abundances of each PFT and their predicted trait values in Step 2. To reduce the variability of different single-models, we derived the final spatial maps of plant functional traits using an ensemble model algorithm to merge the predictions of random forest and boosted regression trees according to their cross-validated  $R^2$  values.



#### 117

**Figure 1.** Methodological workflow for spatial mapping of plant functional traits. Trait mapping is performed in three steps. Step 1: in-situ field measurements of plant functional traits, PFT classification of plant species and gridded predictors were collected. Step 2: two machine learning methods were used to predict trait values by training the field measurements and predictors for each PFT. Step 3: spatialization of trait maps by calculating the abundance of each PFT using 100 m land cover map and predicted trait values within 1 km grid cells. PFT, plant functional type; RF, random forest; BRT, boosted regression trees.

#### 125 **2.2 Plant functional trait collection and data processing**

126 The information on the six plant functional traits and their ecological meanings are described in 127 Table 1. Plant trait data was obtained and collected via two main sources. The first source was 128 public trait databases, including the TRY database (Kattge et al., 2020) and the China Plant Trait 129 Database (Wang et al., 2018). The second source was from literature (listed in Appendix A). To 130 ensure data quality and comparability, we only included trait observations that met the following 131 five criteria: 1) Measurements must be obtained from natural terrestrial fields in order to minimize 132 the influences of management disturbance, and observations from croplands, aquatic habitats, 133 control experiments and gardens were excluded; 2) According to the mass ratio hypothesis, the effect of plant species on ecosystem functioning is determined to an overwhelming extent by the 134 135 traits and functional diversity of the dominant species and is relatively insensitive to the richness

of subordinate species (Grime, 1998). Thus, we only included studies that measured plant trait 136 137 observations from all species or dominant species within a community; 3) In order to consider the intraspecific trait variation, when the same species occurred in at the same sampling site from 138 139 different studies, we included all original observed data from different studies rather than 140 averaging the values at the species level (Jung et al., 2010; Siefert et al., 2015); 4) Plant trait 141 observations must be made on mature and healthy plant individuals, so some specific growth 142 stages (e.g., seedling) and size classes (e.g., sapling) were excluded to reduce the confounding 143 effect of ontogeny and seasonality (Thomas, 2010); 5) We only included studies with clear 144 geographical coordinates to match predictor variables. The sampling location and sampling time 145 information were also included in the dataset. The sampling time mostly focused on the growing 146 season of a year (i.e., May-October), which <u>can</u> ensures the relative consistency of sampling time 147 to minimize the effects of seasonality. Plant functional traits must be sampled and measured 148 according to standardized measurement procedures (Perez-Harguindeguy et al., 2013) to reduce 149 the variation and uncertainty among different data sources. In this study, we included SLA 150 measurements on sun-leaves, and WD measurements on main stem of woody species.

151 **Table 1** Description of plant functional traits selected in this study and their relevant

152 ecosystem functions.

Trait	Abbreviation	Description	Relevant ecosystem functions
Specific leaf	SLA	As a core leaf economics trait (Wright et al.,	Productivity, litter decomposition,
area		2004), it is related to trade-off between leaf	competitive ability (Bakker et al., 2011;
		lifespan and C-carbon_acquisition as well as	Smart et al., 2017)
		light competition (Reich et al., 1991)	
Leaf dry matter	LDMC	Strongly related to resource availability and	Productivity, litter decomposition, herbivore
content		potential growth rate (Hodgson et al., 2011)	resistance, and drought tolerance (Bakker et
			al., 2011; Smart et al., 2017; Blumenthal et
			al., 2020)
Leaf N	LNC	As a core leaf economics trait, it is strongly	Productivity, nutrient cycling, litter
concentration		related to photosynthetic capacity (Wright et	decomposition (LeBauer and Treseder, 2008;
		al., 2004)	Bakker et al., 2011)
Leaf P	LPC	As a core leaf economics trait, it is strongly	Productivity, nutrient cycling, litter
concentration		related to photosynthetic capacity (Wright et	decomposition (LeBauer and Treseder, 2008;
		al., 2004)	Bakker et al., 2011)
Leaf area	LA	Trade-off between C-carbon assimilation and	Productivity (Li et al., 2020)
		water use efficiency, it is related to energy	
		balance (Wright et al., 2017)	
Wood density	WD	A measure of carbon investment, representing	Drought tolerance, productivity (Hoeber et
		the trade-off between growth and mechanical	al., 2014; Liang et al., 2021)
		support (Mart ńez-Vilalta et al., 2010)	

153 The plant trait data was checked for possible errors and corrected in three steps as follows.

154 First, species name and taxonomic nomenclature were corrected and standardized according to the

155 Plant List (http://www.theplantlist.org/) using the "plantlist" package. Second, illogical values,

repeated values and outliers were removed, which were defined by observations exceeding 1.5 156 157 standard deviations of from the mean trait value for a given species (Kattge et al., 2011). Third, we appended information on plant growth form, leaf type and leaf phenology from the TRY 158 159 categorical traits database (https://www.try-db.org/TryWeb/Data.php#3) and Flora Reipublicae Popularis Sinicae (http://www.iplant.cn/frps), which were used to match species names to PFTs. 160 161 We associated each species with a corresponding PFT based on plant growth form (tree, shrub and grass), leaf type (broadleaf and needleleaf) and leaf phenology (evergreen and deciduous). For 162 example, the information on Salix matsudana is: tree, deciduous and broadleaf, thus, we were able 163 164 to associate the PFT of deciduous broadleaf forest (DBF) to this species. The species that did not 165 correspond to any PFT were discarded. After these treatments, we collected a total of 34589 trait measurements from 1430 sampling sites for our database, representing 3447 species from 195 166 167 families and 1066 genera (Fig. 2). Information on the statistics for the six plant functional traits 168 collected in this study is shown in Table B1 in Appendix B.

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170 171

172 different ecosystems in China. The white areas represent artificial land cover types.

#### 173 **2.3 Preparing predictor variables**

#### 174 **2.3.1 Climate data**

Twenty-one climate variables were used in this study, including 19 bioclimate variables, solar
radiation (RAD) and aridity index (AI) (Table B2 in Appendix B). The 19 bioclimate variables and
RAD were obtained from the WorldClim version 2.1 for the period from 1970 to 2000
(https://www.worldclim.org/data/worldclim21.html). The AI data was extracted from the CGIAR
Consortium of Spatial Information (CGIAR-CSI) website for the period from 1970 to 2000
(http://www.csi.cgiar.org) (Trabucco and Zomer, 2018). The spatial resolution of climate data is 1
km.

182 **2.3.2 Soil data** 

Twelve soil variables were included in this study, representing the different aspects of soil properties, i.e., soil texture, bulk density (BD), pH and soil nutrients (Table B2 in Appendix B). All soil variables were extracted from the Soil Database of China for Land Surface Modeling (http://globalchange.bnu.edu.cn/research/soil2) (Shangguan et al., 2013). Given the importance of topsoil properties on community composition (Bohner, 2005), we averaged the first four layers to represent the topsoil properties (~ 30 cm) in our study. The spatial resolution is 1 km.

#### 189 **2.3.3 Topography**

The topographic variable was elevation. Elevation data was extracted from the STRM 90m dataset
in China, based on the SRTM V4.1 database (<u>https://www.resdc.cn/data.aspx?DATAID=123</u>). The
spatial resolution is 1 km.

193 Given the collinearity among climate and soil variables, we reduced the dimensionality of 194 these predictors based on Pearson's correlation coefficient (r) (Figs. B1 and B2 in Appendix B). 195 Among a set of highly correlated variables (r > 0.75), only one variable was retained in subsequent 196 analysis to ensure a combination of different environmental variables. The final selection of 197 environment predictors included nineteen-twenty variables: mean annual temperature (MAT), 198 mean diurnal range (MDR), min temperature of the coldest quarter (Tmin), max temperature of 199 the warmest quarter (Tmax), temperature seasonality (TS), mean annual precipitation (MAP), 200 precipitation seasonality (PS), precipitation of the wettest quarter (PEQ), precipitation of the driest 201 quarter (PDQ), AI, RAD, elevation, soil sand content (SAND), pH, BD, soil total N (STN), soil 202 total P (STP), soil available P (SAP), soil alkali-hydrolysable N (SAN) and cation exchange 203 capacity (CEC).

#### 204 2.3.4 Vegetation indices

205 Three categories of vegetation indices were included in this study (Table B2 in Appendix B). First, 206 EVI was extracted from the MOD13A3 V006 product 207 (https://lpdaac.usgs.gov/products/mod13a3v006/). This product is available as a monthly average 208 with the spatial resolution of 1 km, ranging from January 2000 to December 2018. Second, 209 MODIS reflectance data was also extracted from the MOD13A3 V006 product, including MIR 210 reflectance, NIR reflectance, red reflectance and blue reflectance. Third, the MERIS terrestrial 211 chlorophyll index (MTCI) was extracted from the Natural Environment Research Council Earth 212 Observation Data Centre (NERC-NEODC, 2005) (https://data.ceda.ac.uk/). MTCI data is 213 available globally as a monthly average at 4.63 km spatial resolution, and ranges from June 2002 214 to December 2011. It is noted that valid MTCI values should be greater than 1, so our study 215 deleted any values less than 1.

To avoid collinearity, we also reduced the dimensionality of vegetation indices based on Pearson's correlation coefficient (r values) (Fig. B3 in Appendix B). Most selected variables were related to growing seasons due that plant functional traits were measured during the growing season. Furthermore, based on the results of Pearson's correlation coefficient (r)analysis, MTCI, MIR, NIR, red and blue in January showed low correlations with those in growing season, thus they were included in subsequent analysis. The final selection included 36 variables: annual EVI,
 <u>monthly</u> EVI (May, June, July, August and September), <u>monthly</u> MTCI, MIR, NIR, red and blue
 (all for January, June, July, August and September).

Both environmental variables and vegetation indices variables were resampled to a consistent spatial resolution of 1 km using the nearest neighborhood method.

226 PFT is also an important factor in influencing the variation of plant functional traits 227 (Verheijen et al., 2016; Loozen et al., 2020), thus the trait predictions were performed for each 228 PFT individually. We used the 2015 land cover map at a 100 m spatial resolution to calculate the 229 relative abundance of each PFT within 1 km grid cells, which was extracted from the Copernicus 230 Global Land Service (CGLS-LC100, Version 3) (https://land.copernicus.eu/global/products/lc) 231 (Buchhorn et al., 2020). We focused on natural terrestrial vegetation, so all artificial land cover 232 types or (e.g., croplands) areas were thus eliminated in our dataset. Seven categories were 233 included: evergreen needleleaf forest (ENF), evergreen broadleaf forest (EBF), deciduous 234 needleleaf forest (DNF), deciduous broadleaf forest (DBF), shrubland (SHL), grassland (GRL) and bare/sparse vegetation. 235

#### 236 **2.4 Model fitting and validation**

To predict spatial patterns of plant functional traits, we used two machine learning models, i.e.,random forest and boosted regression trees.

239 Random forest is an ensemble machine learning method based on classification and 240 regression trees using collections of regression trees to classify observations according to a set of 241 predictive variables (Breiman, 2001). This method repeatedly constructs a set of trees from 242 random samples of training data, and the final prediction is produced by integrating the results of all individual trees, which makes it a robust method. The model is controlled by two main 243 244 parameters: the number of sampled variables (mtry) and the number of trees (ntree). The mtry was 245 set to range from 1 to 57 (at an interval of 1), and the ntree was set as 500, 1000, 2000, 5000 and 246 10000 in subsequent runs. This analysis was performed using the 'randomForest' function in the 247 'randomForest' package (Liaw and Wiener, 2002).

248 Boosted regression trees are machine learning methods based on generalized boosted 249 regression models and using a boosting algorithm to combine many sample tree models to 250 optimize predictive performance (Elith et al., 2006). There is no need for prior data transformation 251 or the elimination of outliers, and this method can fit complex non-linear relationships while 252 automatically handling interaction effects between predictors (Elith et al., 2008). The four 253 parameters to optimize in these models are the number of trees, interaction depth, learning rate 254 and bag fractions. We varied the parameter settings to find the optimal parameter combination that 255 achieves minimum predictive error. The number of trees was set to 3000, the interaction depth 256 varied from 1 to 7 (at an interval of 1), the learning rate was set to 0.001, 0.01, 0.05 and 0.1, and 257 the bag fraction was set to 0.5, 0.6, 0.7 and 0.75. PFT was used as a dummy variable in the boosted regression trees models. This analysis was conducted using the 'gbm' function in the
'gbm' package (Ridgeway, 2006).

260 We built separate predictive model for each plant functional trait. To select the optimal 261 parameter combination and to evaluate the final model performance for each trait, we calibrated 262 the models 10 times using randomly selected 80% of the data for training the-models and 263 validating against the remaining 20% based on cross-validation (Table B3 in Appendix B). The 264 predictive performance was evaluated by regressing the predicted and observed trait values from 265 all repetitions of the cross-validation. The fitting performances of the random forest and boosted 266 regression trees were was evaluated using determinate coefficient (R<sup>2</sup>), normalized root-mean-267 square error (NRMSE) and mean absolute error (MAE). These scores are calculated following Eq. 268 (1), Eq. (2) and Eq. (3):

269 
$$R^{2} = 1 - \frac{\sum_{i=1}^{n} (p_{i} - o_{i})^{2}}{\sum_{i=1}^{n} (p_{i} - \hat{o}_{i})^{2}}$$
(1)

270 NRMSE = 
$$\frac{\sqrt{\frac{1}{n}\sum_{i=1}^{n}(p_i - o_i)^2}}{p_{max} - p_{min}}$$
 (2)

271 
$$MAE = \frac{1}{n} \sum_{i=1}^{n} |o_i - p_i|$$
 (3)

where  $p_i$  and  $o_i$  are the predictive values and observed values, respectively;  $\hat{o}_i$  is the mean of the observed values.

274 To quantify the relative importance of each predictor across the two models consistently, we 275 used the method proposed by Thuiller et al. (2009). This method applies correlation between the 276 standard predictions fitted with the original data and predictions where the variable under 277 investigation has been randomly permutated. If the correlation is high, which indicates little 278 difference between the two predictions, the variable permutated is considered not important for the 279 model. This step was repeated multiple times for each predictor, and the mean correlation 280 coefficient over runs was recorded. Then the relative importance of each predictor was quantified 281 as one minus the Spearman rank correlation coefficient (see Boonman et al., 2020). In addition, 282 we used generalized additive models to fit the relationships between plant functional traits and the 283 most important variables using the 'gam' function in the 'mgcv' package.

#### 284 **2.5** Generation of plant functional trait maps and model performance

The generation of spatial maps of plant functional <u>traits</u> was performed in three steps. First, we predicted trait values for each natural PFT (e.gi.e., EBF, ENF, DBF, DNF, SHL and GRL) within 1 km grid cell separately. Second, the abundance of individual natural PFT within 1 km grid cell was estimated using a land cover map with a spatial resolution of 100 m. Third, refer to the Eq. (4) that has been widely applied in a community (Garnier et al., 2004), the final trait value in a given 1 km grid cell was calculated as the sum of the predicted trait values multiplying by corresponding abundance of each natural PFT.

292 
$$CWM = \sum_{i=1}^{n} W_i X_i$$

(4)

where *n* is the total number of PFT in a given grid;  $W_i$  is the relative abundance of the *i*th natural PFT; and  $X_i$  is the predicted trait value of the *i*th natural PFT.

To reduce the variability of different single-models and to construct a more stable and accurate model, the ensemble model was further applied to merge the predictions of random forest and boosted regression trees according to their cross-validated R<sup>2</sup> values. The <u>predictive-predicted</u> value of ensemble model was calculated in a given grid cell as described by Eq. (5) (Marmion et al., 2009). The model accuracy was calculated by regressing the <u>predictive-predicted</u> values of ensemble model against the observed trait values.

301 
$$Pred_EM_t = \frac{\sum_{m=1}^{2} (pred_{m,t} \times r_{m,t}^2)}{\sum_{m=1}^{2} r_{m,t}^2}$$
 (5)

where  $Pred\_EM_t$  is the <u>predictive-predicted</u> values of *t* trait in the ensemble model;  $pred_{m,t}$  is the predictive-predicted values of *t* trait in *m* model;  $r_{m,t}^2$  is the cross-validated R<sup>2</sup> of *t* trait in *m* model. To evaluate the model performance (i.e., the variability in the prediction across models), the coefficient of variation (CV) was calculated as the difference between the predictions of random forest and boosted regression trees methods and the ensemble <u>predictionmodel</u>. CV is calculated as following Eq. (6):

$$CV_{t} = \frac{\frac{\sqrt{\sum_{m=1}^{2} (pred_{m,t} - obs_{t})^{2} * r_{m,t}^{2}}}{\sum_{m=1}^{2} r_{m,t}^{2}}}{obs_{t}}$$
(6)

309 where  $pred_{m,t}$  is the <u>predictive-predicted</u> values of t trait in m model;  $obs_t$  is the values of t trait 310 in the ensemble model;  $r_{m,t}^2$  is the cross-validated R<sup>2</sup> of t trait in m model.

#### 311 **2.6 Uncertainty assessments**

Multivariate environmental similarity surface analysis (MESS) was used to identify the range of the extrapolated predictor values across the locations in the plant trait dataset (Elith et al., 2010). This method is often used to evaluate the extent of extrapolation and the applicability domain. If the values are is negative, this indicates that at a given grid cell, at least one predictor variable is outside the extent of the referenced predictor layer. This analysis was conducted using the 'mess' function in the 'dismo' package.

318 All analyses were performed in R 4.0.2 (R Core Team, 2020).

#### 319 **3 Results**

308

#### 320 **3.1 Performances of prediction models**

321 Cross-validation showed that the performance of the predictive models differed greatly among the

322 plant functional traits (Table 2, Tables C1 and C2 in Appendix C). WD had the best performance

323 in all three models, with  $R^2$  values of 0.64, 0.68 and 0.67 for random forest, boosted regression

- 324 trees and ensemble model, respectively. SLA and LPC had R<sup>2</sup> values greater than 0.45, while
- 325 LDMC performed the worst, with  $R^2$  values below 0.30.

	Random forest			Boosted regression trees				Ensemble model		
Traits	$\mathbb{R}^2$	NRMSE	MAE	$\mathbb{R}^2$	NRMSE	MAE	$\mathbb{R}^2$	NRMSE	MAE	
SLA	0.48	0.22	5.10	0.48	0.20	5.08	0.49	0.21	5.07	
LDMC	0.23	0.21	0.07	0.28	0.18	0.07	0.24	0.20	0.07	
LNC	0.33	0.19	4.92	0.34	0.18	4.85	0.34	0.19	4.85	
LPC	0.51	0.24	0.53	0.51	0.22	0.53	0.51	0.27	0.53	
LA	0.37	0.45	26.76	0.39	0.51	27.47	0.40	0.58	26.59	
WD	0.64	0.20	0.10	0.68	0.13	0.10	0.67	0.17	0.10	

Table 2 Results of plant functional traits for cross-validated R<sup>2</sup>, NRMSE and MAE for random forest, boosted regression trees and ensemble model.

328 SLA, specific leaf area (m<sup>2</sup> kg<sup>-1</sup>); LDMC, leaf dry matter content (g g<sup>-1</sup>); LNC, leaf N concentration

(mg g<sup>-1</sup>); LPC, leaf P concentration (mg g<sup>-1</sup>); LA, leaf area (cm<sup>2</sup>); WD, wood density (g cm<sup>-3</sup>); R<sup>2</sup>,
 determinate coefficient; NRMSE, normalized root-mean-square error; MAE, mean absolute error.

#### 331 **3.2 Spatial patterns of predicted plant functional traits**

332 There were relatively consistent spatial patterns for SLA, LNC and LPC, with high values in the northeastern and northwestern regions-China and the southeastern Qinghai-Tibet Plateau, and low 333 334 values in the southwestern China (Figs. 3a, 3c and 3d, Figs. D1, D2, D3, D5 and D6 in Appendix 335 D). SLA and LPC increased with latitude, while LNC did not vary significantly along the 336 latitudinal gradient. For SLA, LNC and LPC, the variability was low among random forest, 337 boosted regression trees and ensemble model, with an overall CV less than 0.30 (Figs. 4a, 4c and 338 4d). LDMC values were relatively high in most regions of China, and the low values were mainly 339 located in the eastern Yunnan Province and the Loess Plateau (Fig. 3b, Figs. D1, D2 and D4 in 340 Appendix D). LA showed high values in the northeastern and southern regions (except for the 341 Sichuan Basin), and the southeastern Qinghai-Tibet Plateau (Fig. 3e, Figs. D1, D2 and D7 in 342 Appendix D). The strong latitudinal gradient was observed in LA, where the values decreased 343 with latitude.

The CV values of LPC decreased with latitude, but other traits did not show latitudinal patterns (Fig. 4). The CV values of LA were relatively high, especially in the northwestern region <u>China</u> and the Inner Mongolia-Loess Plateau region (Fig. 4e). WD had high values in the northeastern and southern regions (Fig. 2f, Figs. D1, D2 and D8 in Appendix D), while CV values for WD in China were low throughout China (Fig. 4f).





350 351

**Figure 3.** Spatial patterns of predicted plant functional traits in China based on the ensemble model. The grey curves to the right of the maps display trait distribution along with latitude. The white areas represent artificial land cover types and bare vegetation. The lines in grey, blue and purple represent the boundaries of province, the Qinghai-Tibet Plateau and the Loess Plateau, respectively. RF, random forest; BRT, boosted regression trees; ensemble, ensemble model; SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA, leaf area; WD, wood density.







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**Figure 4.** The variability in plant functional trait predictions among random forest, boosted regression trees and ensemble model. The grey curves to the right of the maps display coefficient of variation along with latitude. The white areas represent artificial land cover types <u>and bare</u> <u>vegetation</u>. The lines in grey, blue and purple represent the boundaries of province, the Qinghai-<u>Tibet Plateau and the Loess Plateau, respectively.</u> SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA, leaf area; WD, wood density.

#### 366 **3.3 Relative importance of predictive variables**

The dominant factors explaining spatial variation differed greatly among plant functional traits (Table 3). Overall, climate variables were more important for predicting plant functional traits than were soil variables. Temperature variables (i.e., MAT, MDR and TS) showed close relationships with SLA, LDMC, LPC and WD, while precipitation variables (i.e., PS, PEQ, MAP

371 and PDQ) were more important for predicting the spatial patterns of LNC, LPC and LA. RAD was 372 the fourth most dominant factor in predicting the spatial patterns of SLA and WD. Elevation also 373 played an important role in the LDMC and LPC predictions. Within soil variables, soil nutrients 374 (i.e., pH and SAP) showed close associations with SLA and LNC. In addition to the environmental 375 variables, MTCI emerged as an important predictor for explaining SLA, LDMC and LA. Finally, 376 EVI was the most important predictor for LA, and MIR in January and May were the primary 377 predictors of WD. The relationships between plant functional traits and the most important 378 variables were shown in Figs. E1 and E2 in Appendix E.

Rank	SLA	LDMC	LNC	LPC	LA	WD
1	SAP	MAT	PS	MDR	EVI5	MIR1
2	TS	Elevation	SAP	PDQ	PEQ	TS
3	blue9	MTCI5	pH	Elevation	MTCI9	MIR5
4	RAD	blue8	MDR	MIR8	NIR9	RAD
5	MTCI4	MTCI4	MAP	Tmax	AI	MIR6
6	MTCI6	MTCI6	PEQ	MTCI6	MTCI6	pН
7	Elevation	NIR1	MIR1	MIR7	MAP	red5
8	MTCI7	CEC	Tmax	MIR9	red5	PS

**Table 3** List of the eight most important variables for plant functional trait predictions.

SLA, specific leaf area (m<sup>2</sup> kg<sup>-1</sup>); LDMC, leaf dry matter content (g g<sup>-1</sup>); LNC, leaf N concentration 380 (mg g<sup>-1</sup>); LPC, leaf P concentration (mg g<sup>-1</sup>); LA, leaf area (cm<sup>2</sup>); WD, wood density (g cm<sup>-3</sup>); SAP, soil 381 382 available P; TS, temperature seasonality; blue, blue reflectance; RAD, solar radiation; MTCI, MERIS 383 terrestrial chlorophyll index; MAT, mean annual temperature; NIR, near-infrared reflectance; CEC, 384 cation exchange capacity; PS, precipitation seasonality; MDR, mean diurnal range; MAP, mean annual 385 precipitation; PEQ, precipitation of the wettest quarter of a year; MIR, middle infrared reflectance; 386 Tmax, max temperature of the warmest month of a year; PDQ, precipitation of the driest quarter of a 387 year; EVI, enhanced vegetation index; AI, aridity index; red, red reflectance.

#### 388 **3.4 Model performance**

379

389 The distributions of the <u>predictive predicted trait</u> values based on random forest, boosted 390 regression trees, and ensemble model were consistent with the original trait\_observations,

391 especially the peak values (Fig. 5). The mean values of trait observations were relatively higher

392 than those of the predictive predicted values.





**Figure 5.** Comparison of trait distribution between observations and <u>predictions predictive</u> values in each of the different three models. Each panel depicts the distribution of observations in solid red, of the random forest (RF) model in yellow, of the boosted regression trees (BRT) model in blue, and of the ensemble model in green. The dashed vertical lines indicate mean values. SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA, leaf area; WD, wood density.

400 **3.5 Uncertainty assessments** 

401 The MESS values of all plant functional traits were positive in most regions, indicating a wide 402 applicability domain of our models (Fig. 6). Nevertheless, trait predictions should be interpreted 403 carefully for <u>the</u> northeastern China and the Qinghai-Tibet Plateau due to <u>the</u> sparse samplings in 404 these regions.





Figure 6. Multivariate environmental similarity surface (MESS) assessments for the six plant functional traits. The blue line represents the boundary of the Qinghai-Tibet Plateau. The black dots represented the locations of trait observations. More intense shades indicate greater similarity (blue) or difference (red) in environmental conditions of the location compared to the predictive factors covered by the training dataset. The white areas represent artificial land cover types and bare vegetation. SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA, leaf area; WD, wood density.

## 413 **4 Discussion**

#### 414 **4.1 Comparison with previous work**

415 Our study predicted the spatial patterns of six key plant functional traits across China using 416 machine learning methods and identified the applicability domain of the models. WD had the 417 highest precision with an average of  $R^2$  of 0.66, which was higher than the global WD prediction 418 (Boonman et al., 2020). This improvement in precision may be attributed to the large number and 419 dense occurrence of sample sites as well as the inclusion of vegetation indices in our study. In 420 addition, SLA and LPC also showed good accuracy with  $R^2$  values of 0.50, which was higher than that of Boonman et al. (2020) and consistent with that of Moreno-Mart frez et al. (2018). However,
LNC and LA showed relatively poor performance, which may be related to the reason that these
two traits were more influenced by phylogeny than environmental variables (Yang et al., 2017; An
et al., 2021). In addition, we found that mean values of trait predictions were lower than those of
observations, which may be attributable to the reason that the mean values of trait observations
were from the individual level, while the mean values of predicted values were based on the
relative abundance of PFTs and corresponding predicted values within 1 km grid cell.

428 The frequency distributions of plant functional traits in China differed between our study and 429 previous studies (Fig. 7, Fig. F1, Table F1 in Appendix F). Given that the spatial resolution of trait maps in most previous studies is-was 0.5° (except for Moreno-Mart nez et al. (2018) and 430 431 Vallicrosa et al. (2022)), we resampled the data products of previous studies and our study to  $0.5^{\circ}$ 432 spatial resolution. The distributions in our study contained more predictions at lower values of 433 SLA, LNC and LPC and was were broader than those for SLA and LNC in previous global studies. 434 However, the distribution of LNC in our study was consistent with that in the study of Vallicrosa 435 et al. (2022) at-with a 1 km spatial resolution the 1 km spatial resolution (Fig. F1 in Appendix F). 436 LA in our study contained more predictions at higher values and was also broader than those in 437 previous global studies. WD did not show the lower and higher predictive predicted values in this 438 study, however, the WD values in the studies of Boonman et al. (2020) and Schiller et al. (2021) 439 had more predictions at higher values and no lower values (< 0.30 g cm<sup>-3</sup>). Our predicted values of 440 SLA showed the highest spatial correlation with those of Dong et al. (2023), and LNC showed the 441 strongest spatial correlation with those of Butler et al. (2017) (Table 54). LA and WD showed the 442 best spatial correlation with those of Schiller et al. (2021), but LPC showed relatively weak spatial 443 correlation with those of published studies.

In addition, we compared our results to-with the other studies focused on China. Yang et al. (2016) predicted the spatial distributions of leaf mass per area (i.e., 1/SLA) and LNC based on trait-environment relationships in China and had an-R<sup>2</sup> values of 0.13-0.16. The lower predictive precision may be because Yang et al. (2016) only used MAT, MAP and RAD as predictors in estimating the spatial patterns of leaf mass per area and LNC, which likely led to poor performance and low heterogeneity. These results also demonstrated the advantage of our methods in mapping the spatial patterns of plant functional traits at a regional scale.

451 **Table 5**<u>4</u> Spatial correlations for SLA, LNC, LPC, LA and WD between this study and other

452 previous trait maps, labelled by the first author of the corresponding publication (see Table F1 in

453 Appendix F for citations)

Spatial	Dong	Vallicrosa	Schiller	Boonman	Moreno	Madani	Butler	Bodegom
correlation								
SLA	<del>0.398</del>		-0.08 <mark>2</mark>	0. <del>327<u>33</u></del>	0.24 <mark>2</mark>	0. <del>136<u>14</u></del>	-0.04 <del>2</del>	0. <del>319<u>32</u></del>
	<u>0.40</u>							
LNC	0.1 <del>5</del> 6	0. <del>359<u>36</u></del>	0. <del>229</del> 23	0.25 <mark>2</mark>			0.394	
LPC		0. <del>136<u>14</u></del>					0. <del>057<u>06</u></del>	
LA			0.514					
WD			0. <del>647<u>65</u></del>	0. <del>107<u>11</u></del>				

The spatial correlation of leaf dry matter content (LDMC) between our study and previous studies was not included, as the LDMC maps were not available. SLA, specific leaf area ( $m^2 kg^{-1}$ ); LNC, leaf N concentration (mg g<sup>-1</sup>); LPC, leaf P concentration (mg g<sup>-1</sup>); LA, leaf area ( $cm^2$ ); WD, wood density (g cm<sup>-3</sup>).





459 Figure 7. Frequency distributions of plant functional traits in our study ("This study", dashed
460 black lines) and other trait maps, identified by the first author of the corresponding publication

461 (see Table F1 for citations). SLA, specific leaf area ( $m^2 kg^{-1}$ ); LNC, leaf N concentration (mg g<sup>-1</sup>);

462 LPC, leaf P concentration (mg  $g^{-1}$ ); LA, leaf area (cm<sup>2</sup>); WD, wood density (g cm<sup>-3</sup>).

#### 463 **4.2 Spatial patterns of plant functional traits in China**

464 Our study revealed the spatial patterns of different plant functional traits across China, and the 465 variability among the two machine learning methods was relatively low. We compared the spatial differences of trait maps between our study and previous studies at the global scale (Figs. F2-F6 in 466 467 Appendix F). For example, our study showed high SLA values in the southeastern Oinghai-Tibet 468 Plateau, which concurred with the global study of Boonman et al. (2020). The spatial difference of 469 SLA between our study and van Bodegom et al. (2014) was relatively low, and the predictive 470 predicted values in most regions were slightly lower in our study than those in van Bodegom et al. 471 (2014). The spatial pattern of difference in SLA between our study and Moreno et al. (2018), 472 Bulter et al. (2017) and van Bodegom et al. (20202014) was consistent, and the values were higher 473 in the northeastern China and the southwestern Qinghai-Tibet Plateau in our study than those 474 studies. Our study showed higher LNC values in the northern Inner Mongolia-the Loess Plateau-475 the eastern Qinghai-Tibet Plateau and the northwestern China than those global studies at the 476 global studies (Butler et al., 2017; Moreno-Mart nez et al., 2018; Boonman et al., 2020; Vallicrosa 477 et al., 2022; Dong et al., 2023), reflecting the consistent spatial pattern among these studies. 478 However, Yang et al. (2016) predicted high LNC values in the northeastern and the northwestern 479 China, the northern Inner Mongolia and the entire Qinghai-Tibet Plateau, and SLA and LNC had 480 low heterogeneity overall. The discrepancy with Yang et al. (2016) may be attributed to spatial 481 extrapolation based on trait-climate relationships with a low predictive precision. There was no 482 consistent spatial pattern in LPC between our study and previous studies. Consistent with the 483 global pattern (Wright et al., 2017), LA was larger in the southern regions than in the northern 484 regions and showed a decreasing trend with latitude. In addition, LA and WD values in our study 485 were lower in most regions than those ones at the global scale. These discrepancies between our 486 study and previous studies at the global scale may be related to three reasons. First, there is bias in 487 the available in-situ field measurement data from China in these-global studies, with a large gaps 488 in the western China for SLA and no data in China for WD (Boonman et al., 2020). Second, some 489 trait-environment relationships may be scale-dependent (Bruelheide et al., 2018), and these studies 490 we compared are from the global scale because the trait maps in China are not available. Third, the 491 methods used for trait mapping were different among studies, including eco-evolutionary 492 optimality models (Dong et al., 2023), Convolutional Neural Networks based on RGB 493 photographs (Schiller et al., 2021), machine learning algorithms (Vallicrosa et al., 2022; Boonman 494 et al., 2020) and multiple regression analysis (van Bodegom et al., 2014).

Moreover, our study also identified the applicability domain of our models for predicting the
 spatial patterns of plant functional traits across China. Five leaf traits and WD appeared to have
 poor applicability in <u>the</u> northeastern China and the Qinghai-Tibet Plateau, primarily due to sparse

498 samplings. Future studies predicting plant functional traits across a large scale through remote 499 sensing observations or other supplementary data will be needed to re-evaluate our results.

### 500

#### 4.3 The role of predictive variables

501 Our study indicates indicated that environmental variables are-were important for predicting the 502 spatial patterns of plant functional traits, especially climate variables. Temperature variables were 503 primary predictors for SLA, LDMC, LPC and WD. The relationships between leaf traits and 504 temperature have been widely discussed in global and regional studies (Reich and Oleksyn, 2004; 505 Bruelheide et al., 2018). The positive linkage between WD and temperature may be driven by 506 changes in water viscosity. Plants can adapt to the low water viscosity at high temperatures by 507 reducing the diameter and density of their vessels and by thickening cell walls (Roderick and 508 Berry, 2002; Thomas et al., 2004). Precipitation variables were important predictors for leaf 509 nutrient traits and LA. For example, precipitation of the wettest quarter of a year was the factor 510 that most influenced LA variation, which has been confirmed by a previous study (An et al., 2021). 511 A smaller LA could be an adaptive strategy to decrease water loss via reducing the surface area for 512 transpiration under dry environmental conditions (Du et al., 2019). Although the effects of soil on 513 trait predictions were relatively weak, we found that SAP and pH played key roles in SLA and 514 LNC predictions. These results were similar with the previous studies that reported reporting that 515 soil pH was an important driver of trait variation at the global scale and in tundra regions (Maire et 516 al., 2015; Kemppinen et al., 2021). Additionally, from the perspective of cost-efficient theory, the 517 strong effects of SAP reflected that high SLA may be an adaptation for facilitating soil exploration 518 more efficiently in fertile soils (Freschet et al., 2010).

519 Vegetation indices have recently been proposed as important predictors of spatial patterns of 520 plant functional traits (Loozen et al., 2018). Our results corroborated these findings and further 521 suggested that EVI, MTCI and MIR reflectance are were important predictors in models. Here, the 522 underlying mechanisms between vegetation indices and plant functional traits are-were not further 523 discussed due to their complexity-and uncertainty. However, our results indicated that vegetation 524 indices and NIR reflectance are-were not key predictors of LNC estimation, which contrasts 525 contrasted the findings from global and regional studies (Wang et al., 2016; Loozen et al., 2018; 526 Moreno-Mart nez et al., 2018). This may be related to the multitude of factors that influence the 527 relationships between LNC and vegetation indices and NIR reflectance, such as forest type and 528 canopy structure (Dahlin et al., 2013).

#### 529 **4.4 Uncertainties**

530 Although our study mapped the spatial patterns of key functional traits in terrestrial ecosystems of 531 seed plants in across China through large-scale field investigations and compared the predictions 532 with previous studies performed at global and regional scales, there persists persisted some 533 uncertainties in the interpretation of these results. First, the predictive ability of models was

534 relatively worse for certain traits, especially LDMC. Beyond the environmental effects, the 535 variation in plant functional traits is also regulated by phylogenetic structure among plant species 536 (e.g., family, order and phylogenetic clade) (Li et al., 2017). Consequently, incorporating the 537 phylogenetic information will be a promising avenue for further improving the accuracy of spatial 538 predictions of plant functional traits (Butler et al., 2017). A second potential issue is sampling bias; 539 there were are major spatial gaps in field investigations in both the northeastern China and the 540 Qinghai-Tibet Plateau. Due to the few measurements for shrubs and the lack of herbs, WD data is 541 mainly confined to eastern forests, and the overall quantity of WD data was-is much lower than 542 that of leaf traits, even in the TRY database. The environmental information of sampling sites was 543 not always obtained from original literature, thus using the public environmental products is a 544 common resolution in large-scale plant trait studies (Boonman et al., 2020; Vallicrosa et al., 2022). 545 Such mismatch between in-situ trait measurements and predictors should be resolved in further 546 work. Finally, an additional key challenges in data availability must be resolved to scale up from 547 the species to the community levels, in particular with data surrounding species co-occurrence and 548 their relative cover or abundance in ecological communities (He et al., 2023). For example, Global 549 biodiversity data (e.g., sPlot and Global Biodiversity Information Agency databases) that contains 550 information on species occurrence or the proportion of species in a community has the potential 551 for enabling the calculation of community-weighted trait values and the re-evaluation of our 552 results in future work (Telenius, 2011; Bruelheide et al., 2019). The lack of consistent time period 553 and spatial resolution of predictors due to limitation of data availability is another a key challenge 554 in the spatial mapping of plant functional traits. In addition, although WorldClim version 2.1 555 product has high spatial resolution and includes various aspects of climatic parameters, there 556 exists certain limitation and uncertainty in predicting trait maps. Therefore, integrating satellite 557 remote sensing monitoring methods with in-situ trait data collection can also provide an effective 558 way to estimate and assess the species diversity at large scales (Cavender-Bares et al., 2022).

#### 559 **4.5 Potential applications**

560 Maps of these key functional traits in terrestrial ecosystemsof seed plants highlighted large-scale 561 variability in space, which will significantly advance ecological analyses and future 562 interdisciplinary research. First, using the spatially continuous trait maps, one can optimize and 563 develop trait-flexible vegetation models to reduce uncertainties uncertainty of conventional 564 vegetation models based on PFTs, which allows for the exploration of the community assembly 565 rules based on how plants with different trait combinations perform under a given set of 566 environmental conditions (Berzaghi et al., 2020). When trait-flexible vegetation models are 567 available, incorporating trait maps into models will bridge the gap for vegetation classifications 568 and predictions of vegetation distribution under global change (Van-van Bodegom et al., 2012; 569 Yang et al., 2019). Second, most studies focused on the effects of plant functional traits on 570 ecosystem carbon processes at individual, species and community scales, while how such effects

- 571 scale up to regional or larger scales remains challenging. In addition, the assessments of China's
- 572 terrestrial ecosystem carbon sink have had-large uncertainties-so far (Piao et al., 2022). The spatial
- 573 continuous trait maps will provide an effective way to link ecosystem characteristics to ecosystem
- 574 carbon sink estimates in China (Madani et al., 2018; Šímová et al., 2019). These analyses will help
- 575 shed light on the mechanisms underlying plant functional traits and terrestrial ecosystem carbon
- 576 storage at a large scale.

#### 577 **5 Data availability**

The original plant functional trait data collected in this study that <u>were was</u> used for machine learning models (named by Data file used for machine learning models.csv) and final maps of plant functional traits <u>in a GeoTIFF format in terrestrial ecosystems in a GeoTIFF format across</u> <u>China</u> (named by plant functional trait category) are now available for the private link <u>https://figshare.com/s/c527c12d310cb8156ed2 (An et al., 2023)</u>. Once the article is accepted, we will publicly publish the<u>se maps data</u> at the figshare website.

#### 584 6 Conclusions

585 We generated a set of spatial continuous trait maps at a 1-km spatial resolution using machine 586 learning methods in combination with field measurements, environmental variables and vegetation 587 indices. Models for leaf traits (except for LDMC) and WD showed good accuracy and robustness, 588 whereas models of LDMC had relatively poor precision and robustness. Temperature variables 589 were the most important predictors for leaf traits (except for LA) and WD, and precipitation 590 variables were the most important predictors for leaf nutrient traits and LA. We caution that plant 591 functional trait predictions should be interpreted carefully for the northeastern China and the 592 Qinghai-Tibet Plateau. The spatial continuous trait maps generated in our study are 593 complementary to current terrestrial in-situ observations and offer new avenues for predicting 594 large-scale changes in vegetation and ecosystem functions under climate scenarios in China.

595

#### 596 Appendix A Data collection from literature

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# 885 Appendix B

_	Trait	Unit	Range	Mean	CV (%)	No. of species	Entries	Sites
	SLA	m <sup>2</sup> kg <sup>-1</sup>	0.06-81.68	17.88	54.96	2463	9195	1032
	LDMC	g g <sup>-1</sup>	0.06-0.95	0.34	100.00	1582	3957	193
	LNC	mg g⁻¹	3.41-66.02	21.52	37.44	2335	7407	567
	LPC	mg g <sup>-1</sup>	0.09–9.70	1.83	62.19	2074	6266	515
	LA	cm <sup>2</sup>	0.0033-2553.33	36.16	259.64	1838	5976	691
	WD	g cm <sup>-3</sup>	0.25–1.37	0.68	33.16	768	1788	639
	Altitude	m	-144–5454					1430
	MAT	${}^{\mathbb{C}}$	-12.07-24.32					1430
	MAP	mm	15–2982					1430
	Soil total N	g kg <sup>-1</sup>	0.11-10.25					1430
	Bulk density	g cm <sup>-3</sup>	0.83-1.45					1430

886 Table B1 Summary of statistics in plant functional traits, environmental variables and887 geographical distribution in China.

888 SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA,

leaf area; WD, wood density; MAT, mean annual temperature; MAP, mean annual precipitation.

## 891 in this study.

Type of variables	Variable name	Abbreviations	Units	Time periods	Spatial resolution	Source
Climate	Mean annual temperature	MAT	C	1970-2000	1 km	WorldClim version 2.1
	Mean diurnal range	MDR	C	1970-2000	1 km	WorldClim version 2.1
	Temperature seasonality	TS	C	1970-2000	1 km	WorldClim version 2.1
	Max temperature of the	Tmin	C	1970-2000	1 km	WorldClim version 2.1
	Min temperature of <u>the</u>	Tmax	C	1970-2000	1 km	WorldClim version 2.1
	coldest month Temperature annual range	TAR	C	1970-2000	1 km	WorldClim version 2.1
	Isothermality	IS	%	1970-2000	1 km	WorldClim version 2.1
	Mean temperature of the	MTEQ	C	1970-2000	1 km	WorldClim version 2.1
	wettest quarter Mean temperature of <u>the</u>	MTDQ	C	1970-2000	1 km	WorldClim version 2.1
	driest quarter Mean temperature of <u>the</u>	MTWQ	C	1970-2000	1 km	WorldClim version 2.1
	warmest quarter Mean temperature of the	MTCO	C	1970-2000	1 km	WorldClim version 2.1
	coldest quarter Mean annual precipitation	МАР	mm	1970-2000	1 km	WorldClim version 2.1
	Precipitation of the wettest	PEM	mm	1970-2000	1 km	WorldClim version 2.1
	month Precipitation of the driest	PDM		1970 2000	1 km	WorldClim version 2.1
	month	T DIVI		1970-2000	1 KIII	We difference in 2.1
	Precipitation seasonality	PS	%	1970-2000	I KM	WorldClim version 2.1
	quarter	PEQ	mm	1970-2000	I KM	WorldClim version 2.1
	Precipitation of <u>the</u> driest quarter	PDQ	mm	1970-2000	1 km	WorldClim version 2.1
	Precipitation of <u>the</u> warmest guarter	PWQ	mm	1970-2000	1 km	WorldClim version 2.1
	Precipitation of <u>the</u> coldest	PCQ	mm	1970-2000	1 km	WorldClim version 2.1
	Aridity index	AI	/	1970-2000	1 km	Global CGIAR-CSI
	Solar radiation	RAD	kJ m <sup>-2</sup>	1970-2000	1 km	WorldClim version 2.1
Topography	Elevation	/	m		1 km	SRTM 90m V4.1
Soil	Soil sand content	SAND	%	/	1 km	Shangguan et al. (2013)
	Soil silt content	SILT	%	/	1 km	Shangguan et al. (2013)
	Soil clay content	CLAY	%	/	1 km	Shangguan et al. (2013)
	Bulk density	BD	g cm <sup>-3</sup>	/	1 km	Shangguan et al. (2013)
	Soil pH	pН	/	/	1 km	Shangguan et al. (2013)
	Soil organic matter	SOC	g kg <sup>-1</sup>	/	1 km	Shangguan et al. (2013)
	Soil total N	STN	g kg-1	/	1 km	Shangguan et al. (2013)
	Soil total P	STP	g kg <sup>-1</sup>	/	1 km	Shangguan et al. (2013)
	Soil alkali-hydrolysable N	SAN	mg kg <sup>-1</sup>	/	1 km	Shangguan et al. (2013)
	Soil available P	SAP	mg kg <sup>-1</sup>	/	1 km	Shangguan et al. (2013)
	Soil available K	SAK	mg kg <sup>-1</sup>	/	1 km	Shangguan et al. (2013)
	Cation exchange capacity	CEC	me kg <sup>-1</sup>	/	1 km	Shangguan et al. (2013)

Table B2 List of all the predictors including environment and remote sensing variables used

Continued

Type of variables	Variable name	Abbreviations	Units	Time periods	Spatial resolution	Source			
EVI	MODIS EVI long-term		/	2001-2018	1 km	MOD13A3 V006			
NIR	MODIS NIR long-term		/	2001-2018	1 km	MOD13A3 V006			
MIR	MODIS MIR long-term		/	2001-2018	1 km	MOD13A3 V006			
Red	MODIS red long-term monthly averages		/	2001-2018	1 km	MOD13A3 V006			
Blue	MODIS blue long-term monthly averages		/	2001-2018	1 km	MOD13A3 V006			
MTCI	MTCI long-term monthly		/	2003-2011	4.63 km	MTCI level 3 product			
Land cover	Land cover map		/	2015	100 m	Copernicus Global Land Service Collection 3			
892 The veg	<sup>192</sup> The vegetation indices <del>remote sensing variables</del> are calculated as long-term monthly averages from 2001 to 2018.								

The <u>vegetation indices</u>remote sensing variables are calculated as long-term monthly averages from 2001 to 2018.
 Thus thus 12 variables of each remote sensingvegetation index category are obtained.

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Table B3 The number of samples of eight six plant functional traits used for model training
(80%) and validation (20%).

Traits	No. of samples	No. of samples used for model	No. of samples used for model
		training	validation
SLA	9195	7356	1839
LDMC	3957	3166	791
LNC	7407	5926	1481
LPC	6266	5013	1253
LA	5976	4781	1195
WD	1787	1430	357

SLA, specific leaf area (m<sup>2</sup> kg<sup>-1</sup>); LDMC, leaf dry matter content (g g<sup>-1</sup>); LNC, leaf N concentration (mg g<sup>-1</sup>); LPC,
leaf P concentration (mg g<sup>-1</sup>); LA, leaf area (cm<sup>2</sup>); WD, wood density (g cm<sup>-3</sup>).





903 **Figure B1.** Correlations among climate variables. The blank indicates that the correlations are not 904 significant (P > 0.05). The size of the circles is proportional to the correlation coefficient. The 905 abbreviations of climate variables is are seen in Table B2.

STN												- 1
0.54	STP								•	•		- 0.8
0.95	0.49	SOC						•	•		ŏ	- 0.6
0.83	0.48	0.77	SAN		•				•		ŏ	- 04
0.50	0.48	0.52	0.55	SAP		•	•	•		•	$\overline{\bullet}$	0.4
0.26	0.44	0.23	0.31	0.36	SAK	•		•	•	•		- 0.2
-0.42	-0.19	-0.42	-0.38	-0.19	-0.21	BD				•	•	- 0
-0.33	0.18	-0.37	-0.37	-0.06	0.27	0.31	PH		•		•	0.2
-0.14		-0.12	-0.19	-0.14	-0.11	0.30	0.40	SAND			•	0.4
0.24	0.20	0.19	0.29	0.33	0.25	-0.24	-0.09	-0.81	SILT		•	0.6
	-0.28			-0.12	-0.09	-0.24	-0.56	-0.78	0.26	CLAY		0.8
0.78	0.54	0.74	0.65	0.51	0.37	-0.37	-0.07	-0.21	0.30		CEC	1

907 **Figure B2.** Correlations among soil variables. The blank indicates that the correlations are not 908 significant (P > 0.05). The size of the circles is proportional to the correlation coefficient. The 909 abbreviations of soil variables is are seen in Table B2.





911Figure B3. Correlations among monthly vegetation indexremote sensingvariables. The blank912indicates that the correlations are not significant (P > 0.05). The size of the circles is proportional913to the correlation coefficient. (a) enhanced vegetation index (EVI); (b) MERIS terrestrial914chlorophyll index (MTCI); (c) MIR reflectance; (d) NIR reflectance; (e) red reflectance; (f) blue915reflectance.

#### 916 Appendix C

917 Table C1 Optimal parameter combination and model performance of random forest for plant 918 functional traits.

Traits	ntree	mtry	<b>R</b> <sup>2</sup>	NRMSE	MAE			
SLA	1000	24	0.4 <del>76<u>48</u></del>	0.22	5.134			
LDMC	1000	11	0.234	0.20	0.07 <del>2</del>			
LNC	1000	57	0.39 <mark>2</mark>	0.00	0. <del>098<u>10</u></del>			
LPC	1000	20	0. <del>587<u>59</u></del>	0.05	0. <del>129<u>13</u></del>			
LA	1000	18	0.278	0.48	26.62 <mark>2</mark>			
WD	1000	9	0.534	0.02	0.07 <mark>2</mark>			

919 SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA,

920 leaf area; WD, wood density-; R<sup>2</sup>, determinate coefficient; NRMSE, normalized root-mean-square error; MAE, mean absolute error.

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Table C2 Optimal parameter combination and model performance of boosted regression trees

924	for plant functional traits.	
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Traits	n.tree	interaction <del>.</del> depth	shrinkage	learning rate	bag fractions	$\mathbb{R}^2$	NRMSE	MAE
SLA	3000	6	0.01	10	0.75	0.4 <u>864</u> <u>9</u>	0.20	5.08 <del>2</del>
LDMC	3000	2	0.01	10	0.75	0. <del>247<u>2</u> <u>8</u></del>	0.19	0.07 <del>1</del>
LNC	3000	6	0.01	10	0.70	0.414	0.00	0. <del>096<u>10</u></del>
LPC	3000	7	0.01	10	0.75	0.59 <mark>1</mark>	0.05	0. <del>129<u>13</u></del>
LA	3000	3	0.001	10	0.75	0.28 <mark>2</mark>	0.55	27.5 <del>5</del> 6
WD	3000	4	0.01	10	0.70	0. <u>6276</u> <u>3</u>	0.01	0. <del>066<u>07</u></del>

925 SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA, 926 leaf area; WD, wood density; R<sup>2</sup>, determinate coefficient; NRMSE, normalized root-mean-square error; MAE,

927 mean absolute error.



929

Figure D1. Spatial distributions of plant functional traits based on random forest. The grey curves
on the right of maps were are trait distribution along with latitude. The white areas represent
artificial land cover types and bare vegetation. SLA, specific leaf area; LDMC, leaf dry matter
content; LNC, leaf N concentration; LPC, leaf P concentration; LA, leaf area; WD, wood density.





Figure D2. Spatial distributions of plant functional traits based on boosted regression trees. The
grey curves on the right of maps were are trait distribution along with latitude. The white areas
represent artificial land cover types and bare vegetation. SLA, specific leaf area; LDMC, leaf dry
matter content; LNC, leaf N concentration; LPC, leaf P concentration; LA, leaf area; WD, wood
density.





Figure D3. Spatial distribution of specific leaf area (SLA) for each plant functional type. The left
penal was-is\_obtained from RF method (random forest) method, the right penal was-is\_obtained
from BRT method (boosted regression trees) method. The white areas represent other natural
vegetation types and artificial land cover types. EBF, evergreen broadleaf forest; ENF, evergreen
needleleaf forest; DBF, deciduous broadleaf forest; DNF, deciduous needleleaf forest; SHRUB,
shrubland; GRASS, grassland.





Figure D4. Spatial distribution of leaf dry matter content (LDMC) for each plant functional type.
The left penal was-is obtained from RF method (random forest) method, the right penal was-is
obtained from BRT method (boosted regression trees) method. The white areas represent other
natural vegetation types and artificial land cover types.\_EBF, evergreen broadleaf forest; ENF,
evergreen needleleaf forest; DBF, deciduous broadleaf forest; DNF, deciduous needleleaf forest;
SHRUB, shrubland; GRASS, grassland.



Figure D5. Spatial distribution of leaf N concentration (LNC) for each plant functional type. The
left penal was is obtained from RF method (random forest) method, the right penal was is obtained
from BRT method (boosted regression trees) method. The white areas represent other natural
vegetation types and artificial land cover types. EBF, evergreen broadleaf forest; ENF, evergreen
needleleaf forest; DBF, deciduous broadleaf forest; DNF, deciduous needleleaf forest; SHRUB,
shrubland; GRASS, grassland.



961

**Figure D6.** Spatial distribution of leaf P concentration (LPC) for each plant functional type. The left penal was is obtained from RF method (random forest) method, the right penal was is obtained from BRT method (boosted regression trees) method. The white areas represent other natural vegetation types and artificial land cover types. EBF, evergreen broadleaf forest; ENF, evergreen needleleaf forest; DBF, deciduous broadleaf forest; DNF, deciduous needleleaf forest; SHRUB, shrubland; GRASS, grassland.





Figure D7. Spatial distribution of leaf area (LA) for each plant functional type. The left penal was
is obtained from RF method (random forest) method, the right penal was is obtained from BRT
method (boosted regression trees) method. The white areas represent other natural vegetation
types and artificial land cover types. EBF, evergreen broadleaf forest; ENF, evergreen needleleaf
forest; DBF, deciduous broadleaf forest; DNF, deciduous needleleaf forest; SHRUB, shrubland;
GRASS, grassland.



Figure D8. Spatial distribution of wood density (WD) for each plant functional type. The left
penal was is obtained from RF method (random forest) method, the right penal was is obtained
from BRT method (boosted regression trees) method. The white areas represent other natural
vegetation types and artificial land cover types. EBF, evergreen broadleaf forest; ENF, evergreen
needleleaf forest; DBF, deciduous broadleaf forest; DNF, deciduous needleleaf forest; SHRUB,
shrubland.





Figure E1. The relationships between SLA (specific leaf area), LDMC (leaf dry matter content),
LNC (leaf N concentration), LPC (leaf P concentration) and their eight most important predictors.



Figure E2. The relationships between LA (leaf area), WD (wood density) and their eight most
important predictors.

# Appendix F Comparisons between our study with trait maps from previous studies

991 Given that the trait maps predicted for China were not available from the literature and their 992 authors, we compared our study with those studies performed at the global scale (see Table F1). 993 Thus, we extracted the data in China from global trait maps. Before the quantitative comparisons 994 with previous studies, we performed two steps to make the data products as comparable as 995 possible and improve the consistency between different studies. First, due to different spatial 996 resolution of global trait maps (mainly 0.5 °) and our study, we resampled the data products of 997 previous studies and our maps to 0.5 ° spatial resolution. In addition, Vallicrosa et al. (2022) 998 generated the global maps of LNC and LPC with a 1 km spatial resolution, we also compared the 999 frequency distribution of Vallicrosa et al. (2022) with that of our study at a 1 km spatial resolution. Second, our study focused on natural vegetation, so the global trait maps were used to filter out 1000 1001 non-natural vegetation (e.g., croplands). For example, Madani et al. (2018) predicted the spatial 1002 distributions of SLA that included croplands. We quantitatively compared our maps with previous 1003 studies from two perspectives. The comparisons among trait maps were made using frequency plots and spatial correlations (Figure Fig. 7, and Table 54 and Fig. F1 in Appendix F). And the 1004 005 maps of spatial differences between our study and previous studies were displayed as Figs. F4F2-006 F5-F6 in Appendix F.

1007

Table F1 Summary table of related trait maps of previous studies used in this study.

					2
References	Related	Methods	Predictors	Consideration	<u>Spatial</u>
	traits			of PFT	Resolution resolution
Dong et al.	SLA	Optimality models	Climate	Yes	0.5 °
(2023)	LNC				
Vallicrosa et	LNC	Neural networks	Climate	Yes	0.0083 °
al. (2022)	LPC		Soil		
			N and P		
			deposition		
Schiller et al.	SLA	Convolutional	Climate	No	0.5 °
(2021)	LNC	Neural Networks	In-situ RGB		
	LA		images		
	WD				
Boonman et	SLA	Generalized linear	Climate	No	0.5 °
al. (2020)	LNC	model, Generalized	Soil		
	WD	additive model,			
		Random forest,			
		Boosted regression			
		trees, Ensemble			
		model			
Moreno et al.	SLA	Regularized linear	Climate	Yes	$0.0045~^{\circ}$
(2018)	LNC	regression, Random	Elevation		

	LPC	forest, Neural	Reflectance		
	LDMC	networks, Kernel			
		networks			
Madani et al.	SLA	Generalized	Climate	No	$0.5~^{\circ}$
(2018)		additive model			
Butler et al.	SLA	Bayesian model	Climate	Yes	$0.5~^{\circ}$
(2017)	LNC		Soil		
	LPC				
Bodegom et	SLA	Multiple regression	Climate	No	$0.5^{\circ}$
al. (2014)	WD	analysis	Soil		

1009 The resolutions 0.5 °, 0.0083 ° and 0.0045 ° correspond to square grid cell sizes of about 50 km, 1 km and 500 m at

the equator. PFT, plant functional type; SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf Nconcentration; LPC, leaf P concentration; LA, leaf area; WD, wood density.



1012

1013 Figure F1. Frequency distributions of plant functional traits in our study ("This study", dashed

black lines) and Vallicrosa et al. (2022) at 1 km spatial resolution. (a) LNC, leaf N concentration
(mg g<sup>-1</sup>); (b) LPC, leaf P concentration (mg g<sup>-1</sup>).



1016 **Figure F2.** Spatial differences in SLA (specific leaf area,  $m^2 \text{ kg}^{-1}$ ) between our study and trait 1018 maps from previous studies (see Table F1 for citations).



Figure F3. Spatial differences in LNC (leaf N concentration, mg g<sup>-1</sup>) between our study and trait
 maps from previous studies (see Table F1 for citations).



Figure F4. Spatial differences in LPC (leaf P concentration, mg g<sup>-1</sup>) between our study and trait
 maps from previous studies (see Table F1 for citations).



Figure F5. Spatial differences in LA (leaf area, cm<sup>2</sup>) between our study and trait maps from
previous studies (see Table F1 for citations).



Figure F6. Spatial differences in WD (wood density, g cm<sup>-3</sup>) between our study and trait maps
from previous studies (see Table F1 for citations).

1031 Author contributions. NA and NL designed the research. NA did the analysis, processed the data 1032 and wrote the draft of the paper. All co-authors commented on the manuscript and agreed upon the 1033 final version of the paper.

1034

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