1	Spatial mapping of key plant functional traits in terrestrial
2	ecosystems across China
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17 Abstract

18 Trait-based approaches are of increasing concern in predicting vegetation changes and linking 19 ecosystem structures to functions at large scales. However, a critical challenge for such approaches 20 is acquiring spatially continuous plant functional trait distributionmaps. Here, eight six key plant 21 functional traits were selected to as they can reflect plant resource acquisition strategies and ecosystem functionsrepresent two-dimensional spectrum of plant form and function, including 22 23 specific leaf area (SLA)leaf area (LA), leaf dry matter content (LDMC), leaf N concentration (LNC), leaf P concentration (LPC), plant height, seed mass (SM), leaf area (LA) specific leaf area (SLA) 24 25 and wood density (WD). A total of 52477-34589 in-situ trait measurements of 4291-3447 seed plant 26 species were collected from 1541-1430 sampling sites in China and were used to generate a spatial 27 plant functional trait plant functional trait dataset maps (~1 km), together with environmental 28 variables and vegetation indices based on two machine learning models (random forest and boosted 29 regression trees). The two models showed a good accuracy in estimating WD, LPC and SLA, with average R^2 values ranging from 0.45 to 0.66. In contrast, both the two models had-a weak 30 performance in estimating SM and LDMC, with average R² values below 0.25. Meanwhile, LA₅ 31 32 SM and plant height showed considerable differences between two models in some regions. To 33 obtain the optimal estimates, a weighted average algorithm was further applied to merge the 34 predictions of the two models to derive the final spatial plant functional trait datasetmaps. The 35 optimal estimates showed that cClimatic effects were more important than those of edaphic factors in predicting the spatial distribution of plant functional traits. Estimates of plant functional traits in 36 37 northeast China and the Qinghai-Tibet Plateau had relatively high uncertainties due to sparse 38 samplings, implying a need of more observations in these regions in future. Our spatial trait dataset 39 maps could provide critical support for trait-based vegetation models and allows exploration into 40 the relationships between vegetation characteristics and ecosystem functions at large scales. The 41 eight six plant functional traits datasets maps for China with 1 km spatial resolution are now 42 available at https://figshare.com/s/c527c12d310cb8156ed2 (An et al., 2023).

43 **1 Introduction**

44 Climate change has been affecting vegetation distributions and biogeochemical cycling globally and 45 altering their feedbacks to the climate system (Kirilenko et al., 2000; Finzi et al., 2011; Jónsdóttir et al., 2022). Dynamic global vegetation models (DGVMs) are powerful tools for predicting changes 46 47 in vegetation and ecosystem-atmosphere exchanges (e.g., water, carbon, and nutrient cycling) in a 48 changing climate (Foley et al., 1996; Peng, 2000). However, conventional DGVMs are still 49 insufficient realistic, largely due to their dependence on the plant functional types (PFTs) assumption 50 (Sitch et al., 2008; Yurova and Volodin, 2011; Scheiter et al., 2013). PFTs in conventional DGVMs 51 commonly have fixed attributes (mostly trait values) (Van Bodegom et al., 2012; Wullschleger et 52 al., 2014) that do not reflect plant adaptation to environments, limiting the quantification of carbon-53 water-nutrient feedback between terrestrial ecosystems and the atmosphere (Zaehle and Friend, 54 2010; Liu and Yin, 2013). Trait-based approaches can provide robust theoretical basis for developing 55 the next generation of DGVMs (Van Bodegom et al., 2012; Sakschewski et al., 2015; Matheny et 56 al., 2017). Plant functional traits, which are closely associated with ecosystem functions (Diaz et al., 57 2004; Yan et al., 2023), can effectively reflect response and adaptation of plants to environmental 58 conditions (Myers-Smith et al., 2019; Qiao et al., 2023).

59 Attempts to predict spatially continuous trait data-maps have been conducted at regional to global scales (Madani et al., 2018; Moreno-Martínez et al., 2018; Boonman et al., 2020; Loozen et 60 al., 2020; Dong et al., 2023). Webb et al. (2010) proposed that the environment creates a filtered 61 62 trait distribution along an environmental gradient, and such trait-environment relationships offer 63 fundamental supports to predict the spatial distribution of plant functional traits through 64 extrapolating local trait measurements. Boonman et al. (2020) mapped the global patterns of specific 65 leaf area (SLA), leaf N concentration (LNC) and, wood density (WD) and plant height based on a 66 set of climate and soil variables. As the number of available global trait databases increases (Kattge 67 et al., 2011; Wang et al., 2018; Kattge et al., 2020), trait-environment relationships are becoming increasingly quantitative and accurate (Bruelheide et al., 2018; Myers-Smith et al., 2019). 68 Alternatively, remote sensing approaches, such as empirical methods and physical radiative transfer 69 70 models (e.g., partial least squares regression, PROSPECT model), have been developed to estimate 71 plant physiological, morphological, and chemical traits (e.g., leaf chlorophyll content, SLA, LNC 72 and leaf dry matter content (LDMC)) (Darvishzadeh et al., 2008; Romero et al., 2012; Ali et al., 73 2016). Vegetation indices, such as the normalized difference vegetation index (NDVI) and the 74 enhanced vegetation index (EVI), have been successful in estimating plant functional traits of crops, grasslands and forests (Clevers and Gitelson, 2013; Li et al., 2018; Loozen et al., 2018). Loozen et 75 76 al. (2020) demonstrated that EVI was the most important predictor for mapping the spatial pattern 77 of canopy nitrogen in European forests. Admittedly, recent studies have suggested that combining 78 environmental variables and vegetation indices can improve the predictive accuracy of canopy 79 nitrogenN compared to those based on vegetation indices alone (Loozen et al., 2020).

80 Although there have been reports on plant functional trait distribution in China in some global 81 or regional research (Yang et al., 2016; Butler et al., 2017; Madani et al., 2018; Moreno-Martínez et 82 al., 2018; Boonman et al., 2020), they are still of large uncertainties in characterizing the spatial 83 distribution of plant functional traits in China. First, global studies generally have relatively few, 84 unevenly distributed sampling sites in China (Butler et al., 2017; Madani et al., 2018; Boonman et 85 al., 2020), impeding our understanding of the true spatial characteristics of trait variability. Second, the spatial pattern of traits among these studies are usually inconsistent. For example, Moreno-86 Martínez et al. (2018) and Madani et al. (2018) demonstrated that SLA values were low in the 87 88 southeast areas but high in the southwest areas of China, whereas Boonman et al. (2020) found the 89 opposite. Third, most studies mainly focused on leaf traits (Yang et al., 2016; Loozen et al., 2018; 90 Moreno-Martínez et al., 2018), whereas traits associated with the whole-plant and reproductive 91 strategies, such as WD-and seed mass (SM), were ignored. Therefore, mapping and verifying the 92 spatial patterns of key functional traits that reflect the whole plant economics spectrum in China is 93 a top priority.

94 In this study, our main objective was to generate a spatial datasetmaps- for several key plant 95 functional traits, through combining field measurements, environmental variables and vegetation 96 indices. To achieve this goal, wWe used a processing routine to predict the spatial distribution of 97 plant functional traits. selected six plant functional traits including SLA, LDMC, LNC, LPC, LA 98 and WD. As key leaf economics traits, SLA, LDMC, LNC and LPC were selected because they are 99 closely linked to plant growth rate, resource acquisition and ecosystem function (Wright et al., 2004; 100 Diaz et al., 2016). LA is indicative of the trade-off between carbon assimilation and water-use 101 efficiency (Wright et al., 2017), and WD reflects the trade-off between plant growth rate and support 102 cost, with a higher WD linked to a lower growth rate, a higher survival rate and a higher biomass 103 support cost (King et al., 2006). For each plant functional trait, we predicted spatial patterns at a 1 104 km resolution using an ensemble modelling algorithm based on two machine learning methods (i.e., 105 random forest and boosted regression trees).

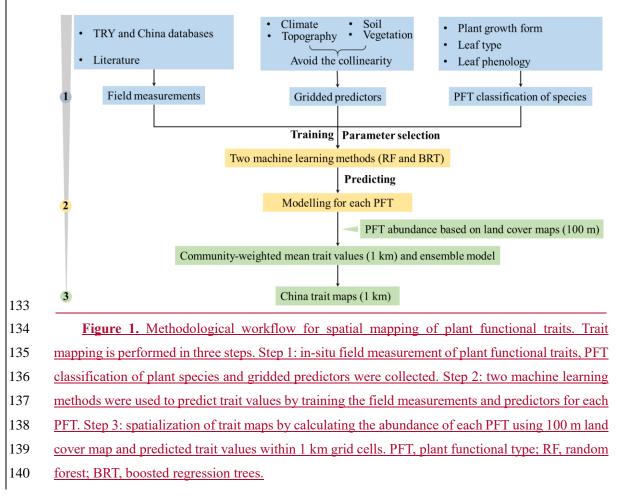
106 First, eight plant functional traits (i.e., SLA, LDMC, LNC, LPC, LA, plant height, WD and 107 SM) were selected because they reflect plant adaptation to environment constraints and trade offs 108 between plant form and function (Reich and Cornelissen, 2014; Diaz et al., 2016). Second, we used 109 random forest and boosted regression trees to predict the spatial patterns of plant functional traits 110 by training the relationships between plant functional traits and environmental variables and vegetation indices. To obtain the optimal estimates, an ensemble model (i.e., weighted average 111 112 algorithm) was further applied to merge the predictions of the two models. Finally, plant species 113 were aggregated to PFTs, and the spatial abundance of PFTs at 1 km resolution was calculated using 114 land cover map (100 m). We derived the spatial trait datasets via calculating community-weighted 115 trait values within grid cells (1 km) based on these abundances of each PFT and predicted trait 116 values.

117 2 Materials and Methods

118 **2.1 Overview**

119 The spatial maps of plant functional traits in China were generated based on machine learning 120 algorithms trained by a large dataset of in-situ field measurements, environmental variables and 121 vegetation indices in three steps (Fig. 1). First, in-situ field measurements of six plant functional 122 traits were collected from TRY and China databases as well as published literature, and the PFTs of 123 plant species were classified based on plant growth form, leaf type and leaf phenology. Multiple 124 gridded predictors of climate, soil, topography and vegetation indices were used after avoiding the 125 collinearity among them. Second, random forest and boosted regression trees were used to train the 126 relationships between plant functional traits and predictors for each PFT individually. Third, the 127 spatial abundance of each PFT within 1 km grid cell was calculated using land cover map (100 m). 128 Community-weighted trait values within 1 km grid cell were calculated based on these abundances 129 of each PFT and their predicted trait values in Step 2. To reduce the variability of different single-130 models, we derived the final spatial maps of plant functional traits using ensemble model to merge 131 the predictions of random forest and boosted regression trees according to their cross-validated R²

132 values.



141 2.1-2 Plant functional trait collection and data processing

142 The information on the eight six plant functional traits and their ecological meanings are described 143 in Table 1._Plant trait data was obtained and collected via two main sources. The first source was 144 public trait databases, including the TRY database (Kattge et al., 2020) and the China Plant Trait 145 Database (Wang et al., 2018). The second source was from literature (listed in Appendix A). To ensure data quality and comparability, we only included trait observations that met the following 146 147 five criteria: 1) Measurements must be obtained from natural terrestrial fields in order to minimize 148 the influences of management disturbance, and observations from cropland, aquatic habitat, control 149 experiments or gardens were excluded; 2) According to the mass ratio hypothesis, the effect of plant 150 species on ecosystem functioning is determined to an overwhelming extent by the traits and 151 functional diversity of the dominant species and is relatively insensitive to the richness of subordinate species (Grime, 1998). Thus, we only included studies that measured plant trait 152 153 observations from all species or dominant species within a community; 3) In order to consider the 154 intraspecific trait variation, when the same species occurred in the same sampling site from different 155 studies, we included all original observed data from different studies rather than averaging the values 156 at the species level In order to account for the trait variation within and between communities, we 157 only considered real measurements of traits from individual plants, and not species-level averages 158 (Jung et al., 2010; Siefert et al., 2015); 4) Plant trait observations must be made on mature and 159 healthy plant individuals, so some specific growth stages (e.g., seedling) and size classes (e.g., 160 sapling) were excluded to reduce the confounding effect of ontogeny and seasonality (Thomas, 161 2010); 5) We only included studies with clear geographical coordinates to ensure alignment with 162 predictor variables. The sampling location and sampling sample-timetime information from the 163 original studies were also included in the dataset. The sampling time mostly focused on the growing 164 season of a year (i.e., May-October), which ensures the relative consistency of sampling time to 165 minimize the effects of seasonality.

Plant functional traits must be sampled and measured according to standardized measurement procedures (Perez-Harguindeguy et al., 2013) to reduce the variation and uncertainty among different data sources. In this study, we included SLA measurements on both-sun-leaves and shadeleaves, and WD measurements on both heartwood and sapwood of tree species. SM measurements on both seeds and fruits, and plant height measurements on both vegetative and generative organs.

171 Table 1 Description of plant functional traits selected in this study and their relevant ecosystem

172 functions.

Trait	Abbreviation	Description	Relevant ecosystem functions
Specific leaf	SLA	As a core leaf economics trait (Wright et al.,	Productivity, litter decomposition,
area		2004), it is related to trade-off between leaf	competitive ability (Bakker et al., 2011;
		lifespan and C acquisition as well as light	Smart et al., 2017)
		competition (Reich et al., 1991)	
Leaf dry matter	LDMC	Strongly related to resource availability and	Productivity, litter decomposition, herbivore
content		potential growth rate (Hodgson et al., 2011)	resistance, and drought tolerance (Bakker et
			al., 2011; Smart et al., 2017; Blumenthal et
			al., 2020)
Leaf N	LNC	As a core leaf economics trait, it is strongly	Productivity, nutrient cycling, litter
concentration		related to photosynthetic capacity (Wright et	decomposition (LeBauer and Treseder,
		al., 2004)	2008; Bakker et al., 2011)
Leaf P	LPC	As a core leaf economics trait, it is strongly	Productivity, nutrient cycling, litter
concentration		related to photosynthetic capacity (Wright et	decomposition (LeBauer and Treseder,
		al., 2004)	2008; Bakker et al., 2011)
Leaf area	LA	Trade-off between C assimilation and water	Productivity (Li et al., 2020)
		use efficiency, it is related to energy balance	
		(Wright et al., 2017)	
Plant height	4	A major determinant of a plant's ability to -	C storage, animal diversity (Conti et al., -
		compete for light (Moles et al., 2009)	2013; Roll et al., 2015)
Wood density	WD	A measure of carbon investment, representing	Drought tolerance, productivity (Hoeber et
		the trade-off between growth and mechanical	al., 2014; Liang et al., 2021)
		support (Martínez-Vilalta et al., 2010)	
Seed mass	SM	Closely related to seed dispersal and seed	Competitive ability (Zepeda and Martorell,
		predation, as well as seedling establishment, -	2019)
		growth, and survival (Leishman et al., 2000)	

173

The plant trait data was checked for possible errors and corrected as perin three steps as follows. 174 First, species name and taxonomic nomenclature were corrected and standardized according to the 175 Plant List (http://www.theplantlist.org/) using the "plantlist" package. Second, illogical values, repeated values and outliers were removed, which were defined by observations exceeding 1.5 176 177 standard deviations of the mean trait value for a given species (Kattge et al., 2011). Third, we 178 appended information on plant growth form (tree, shrub, and herb), leaf type (broadleaved and 179 needle-leaved) and leaf phenology (evergreen and deciduous) from the TRY categorical traits 180 database (https://www.try-db.org/TryWeb/Data.php#3) and Flora Reipublicae Popularis Sinicae (http://www.iplant.cn/frps), which were used . Furthermore, in order to match species names to 181 182 PFTs,-. we We associated each species (i.e., plant growth form, leaf type and leaf phenology) with 183 a corresponding PFT based on plant growth form (tree, shrub and grass), leaf type (broadleaf and 184 needleleaf) and leaf phenology (evergreen and deciduous). For example, the information on Salix 185 matsudana is: tree, deciduous and broadleaf, thus, we were able to associate the PFT of deciduous

broadleaf forest (DBF) to this species. The species that did not correspond to any PFT were 186 187 discarded. After these treatments, we collected a total of 3458952477 trait measurements from 1541 188 1430 sampling sites for our database, representing 4291-3447 species from 212-195 families and 189 1230-1066 genera (Fig. 12 and Fig. B1 in Appendix B). Information on the statistics for the eight 190 six plant functional traits collected in this study is shown in Table B1 in Appendix B.

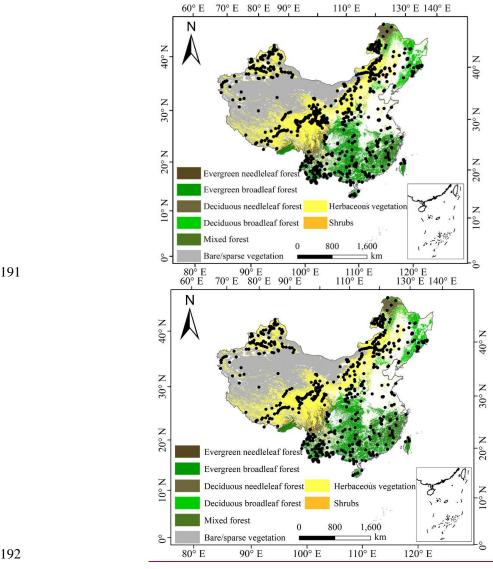




Figure 12. Location distribution and land cover map in China.

2.2-3 Preparing predictor variables Environmental data 194

195 2.23.1 Climate data

196 Twenty-one climate variables were used in this study, including 19 bioclimate variables, solar 197 radiation (RAD), and aridity index (AI) (Table B2 in Appendix B). The 19 bioclimate variables and 198 RAD were obtained from the WorldClim dataset (version 2.11) for the period from 1970 to 2000 199 (https://www.worldclim.org/data/worldclim21.html). The AI data was extracted from the CGIAR 200 Consortium of Spatial Information (CGIAR-CSI) website for the period from 1950-1970 to 2000 201 (http://www.csi.cgiar.org) (Trabucco and Zomer, 2018). The spatial resolution of climate data was 202 <u>is</u> 1 km.

203 2.23.2 Soil data

Twelve soil variables were included in this study, representing the different aspects of soil properties, i.e. soil texture, bulk density (BD), $pH_{\overline{7}}$ and soil nutrients (Table B2 in Appendix B). All soil variables were extracted from the Soil Database of China for Land Surface Modeling (<u>http://globalchange.bnu.edu.cn/research/soil2</u>) (Shangguan et al., 2013). Given the importance of topsoil properties on community composition (Bohner, 2005), <u>we averaged the first four layers to</u> the soil data was averaged to represent the top<u>soil-30 cm of soil-properties (~ 30 cm)</u> in our study. The spatial resolution was_is 1 km.

211 2.23.3 Topography

The topographic variable was elevation. Elevation data was extracted from the STRM 90m dataset in China, based on the SRTM V4.1 database (<u>https://www.resdc.cn/data.aspx?DATAID=123</u>). The spatial resolution <u>was is</u> 1 km.

215 Given the collinearity among climate and soil variables, we reduced the number of 216 environmental predictors based on Pearson's correlation coefficient (r) (Figs. B1-B2 and B2-B3 in 217 Appendix B). Among a set of highly correlated variables (r > 0.75), only one variable was retained 218 in subsequent analysis to ensure a combination of different environmental variables. The final 219 selection of environment predictors included nineteen variables: mean annual temperature (MAT), 220 mean diurnal range (MDR), min temperature of coldest quarter (Tmin), max temperature of coldest 221 warmest quarter (Tmax), temperature seasonality (TS), mean annual precipitation (MAP), 222 precipitation seasonality (PS), precipitation of wettest quarter (PEQ), precipitation of driest quarter 223 (PDQ), AI, RAD, elevation, soil sand content (SAND), pH, BD, soil total N (STN), soil total P 224 (STP), soil available P (SAP), soil alkali-hydrolysable N (SAN), and cation exchange capacity 225 (CEC).

226 2.3-3.4 Vegetation indices

227 Three categories of vegetation indices were included in this study (Table B2 in Appendix B). First, EVI The first selected was EVI, which was extracted from the MOD13A3 V006 product 228 (https://lpdaac.usgs.gov/products/mod13a3v006/). This product is available as a monthly average 229 230 with spatial resolution of 1 km, ranging from January 2000 to December 2018. Second, MODIS 231 reflectance data was also extracted from the MOD13A3 V006 product. This, included including 232 MIR reflectance, NIR reflectance, red reflectance, and blue reflectance. Third, the MERIS terrestrial 233 chlorophyll index (MTCI) was extracted from the Natural Environment Research Council Earth Observation Data Centre (NERC-NEODC, 2005) (https://data.ceda.ac.uk/). MTCI data is available 234 235 globally as a monthly average at 4.63 km spatial resolution, and ranges from June 2002 to December 236 2011. It is noted that valid MTCI values should be greater than 1, so our study deleted any values 237 less than 1.

To avoid collinearity, we also reduced the number of vegetation indices based on Pearson's correlation coefficient (r) (Figs. <u>B3-B4</u> in Appendix B). <u>Most selected variables were related to</u> growing seasons due that plant functional traits were measured during the growing season.
Furthermore, based on the results of Pearson's correlation coefficient (r), MTCI, MIR, NIR, red and

- blue in January showed low correlations with those in growing season, thus they were included in
- 242 <u>Side in Jundary showed low correlations with those in growing season, thus they were included</u>

<u>subsequent analysis.</u> Furthermore, given that most plant functional traits were measured during the
 growing season, the variables related to the growing season were determined to be important
 predictors. The final selection included 36 variables: annual EVI, EVI (May, June, July, August and
 September), MTCI, MIR, NIR, red and blue (all for January, June, July, August and September).

247 Both environmental variables and vegetation indices variables were resampled to a consistent 248 spatial resolution of 1 km using the nearest neighborhood method.

249 PFT is also an important factor in influencing the variation of plant functional traits (Verheijen 250 et al., 2016; Loozen et al., 2020), thus the trait predictions were performed for each PFT 251 individually, thus PFT was included as a predictor in this analysis. We used the 2015 land cover map 252 at a 100 m spatial resolution to calculate the relative abundance of each PFT within 1 km grid cells, 253 which was extracted from the Copernicus Global Land Service (CGLS-LC100, Version 3) 254 (https://land.copernicus.eu/global/products/lc) (Buchhorn et al., 2020). We focused on natural 255 terrestrial vegetation, so all artificial or crop areas were thus eliminated in our dataset. Seven 256 categories were included: evergreen needleleaf forest (ENF), evergreen broadleaf forest (EBF), 257 deciduous needleleaf forest (DNF), deciduous broadleaf forest (DBF), shrubland (SHL), grassland (GRL) and bare/sparse vegetation. Furthermore, in order to match species names to PFTs, we 258 259 associated each species (i.e., plant growth form, leaf type and leaf phenology) with a corresponding PFT. For example, the information on Salix matsudana is: tree, deciduous and broadleaf, thus, 260 261 were able to associate the PFT of deciduous broadleaf forest (DBF) to this species. The species that 262 did not correspond to any PFT were disearded.

263 **2.4-4 Model fitting and validation**

To predict spatial patterns of plant functional traits, we used two machine learning models, i.e., random forest and boosted regression trees.

266 Random forest is an ensemble machine learning method based on classification and regression trees using collections of regression trees to classify observations according to a set of predictive 267 268 variables (Breiman, 2001). This method repeatedly constructs a set of trees from random samples 269 of training data, and the final prediction is produced by integrating the results of all individual trees, 270 which makes it a robust method. The models are-is controlled by two main parameters: the number 271 of sampled variables (mtry) and the number of trees (ntree). The parameter mtry was set to range 272 from 1 to 57 (at an interval of 1), and the ntree was set as 500, 1000, 2000, 5000 and 10000 in 273 subsequent runs. This analysis was performed using the 'randomForest' function in the 274 'randomForest' package (Liaw and Wiener, 2002).

Boosted regression trees <u>is_are a</u>-machine learning methods based on generalized boosted regression models, and using a boosting algorithm to combine many sample tree models to optimize 277 predictive performance (Elith et al., 2006). There is no need for prior data transformation or the elimination of outliers, and this method can fit complex non-linear relationships while automatically 278 handling interaction effects between predictors (Elith et al., 2008). The four parameters to optimize 279 280 in these models are the number of trees, interaction depth, learning rate and bag fractions. We varied 281 the parameter settings to find the optimal parameter combination that achieves minimum predictive 282 error. The number of trees was set to 3000, the interaction depth varied from 1 to 7 (at an interval of 1), the learning rate was set to 0.001, 0.01, 0.05 and 0.1, and the bag fraction was set to 0.5, 0.6, 283 284 0.7 and 0.75. PFT was used as a dummy variable in the boosted regression trees models. This 285 analysis was conducted using the 'gbm' function in the 'gbm' package (Ridgeway, 2006).

We built separate predictive model for each plant functional trait. We used a 10-fold cross 286 287 validation To select the optimal parameter combination and to evaluate the final model performance 288 for each trait, we calibrated the models 10 times using randomly selected 80% of the data for training 289 the models and validating against the remaining 20% based on cross-validation (Table B3 in 290 Appendix B).- We spilt the data into two parts: 80% of the trait data was used to train the models, and the remaining 20% was used to assess model's performance. The predictive performance was 291 292 evaluated by regressing the predicted and observed trait values from all repetitions of the cross-293 validation. The fitting performances of the random forest and boosted regression trees methods were 294 evaluated using determinate coefficient (R²), normalized root-mean-square error (NRMSE), and 295 mean absolution error (MAE). These scores are calculated following Eq. (1), Eq. (2) and Eq. (3):

296
$$R^{2} = 1 - \frac{\sum_{i=1}^{n} (p_{i} - o_{i})^{2}}{\sum_{i=1}^{n} (p_{i} - \delta_{i})^{2}} - (1)$$
297
$$RMSE = \sqrt{-NRMSE} = \frac{\sqrt{\frac{1}{n} \sum_{i=1}^{n} (p_{i} - o_{i})^{2}}}{p_{max} - p_{min}} - (2)$$
298
$$MAE = \frac{1}{n} \sum_{i=1}^{n} |o_{i} - p_{i}| - (3)$$

where p_i and o_i are the predictive values and observed values, respectively, \hat{o}_i is the mean of the observed values.

301 To quantify the relative importance of each predictor consistently across the two models 302 consistently, we used the method proposed by Thuiller et al. (2009). This method applies correlation 303 between the standard predictions fitted with the original data and predictions where the variable 304 under investigation has been randomly permutated. If the correlation is high, which indicates little 305 difference between the two predictions, the variable permutated is considered not important for the 306 model. The correlations between fitted values and predictions were calculated using permuted 307 values for the predictor of concern. This step was repeated a user defined number of multiple times 308 for each predictor, and the mean correlation coefficient over runs was recorded. Then the relative 309 importance of each predictor was quantified as one minus the Spearman rank correlation coefficient 310 (see Boonman et al., 2020). In addition, we used generalized additive models to fit the relationships 311 between plant functional traits and the most important variables using the 'gam' function in the 312 'mgcv' package.

313 **2.5-5** Generation of plant functional trait maps and model performance

314 The generation of spatial maps of plant functional was performed in three steps. First, we predicted trait values for each PFT within 1 km grid cell separately. Second, the abundance of individual PFT 315 316 within 1 km grid cell was estimated using a land cover map with a spatial resolution of 100 m. Third, 317 refer to the Eq. (4) that has been widely applied in a community (Garnier et al., 2004), the The-final community weighted mean trait values in a given 1 km grid cell wereas calculated according to as 318 319 the sum of the predicted trait values and multiplying by corresponding abundance of each PFT. $CWM = \sum_{i=1}^{n} W_i X_i$ 320 (4)where *n* is the total number of PFT in a given grid; W_i is the relative abundance of the *i*th PFT; and 321 322 X_i is the predicted trait value of the *i*th PFT. To calculate community weighted mean trait values, 323 the abundance of individual PFT within 1 km grid cell was estimated using a land cover map with 324 a spatial resolution of 100 m. The final community weighted mean trait values were calculated

325 according to the predicted trait values and corresponding abundance of each PFT.

326 To reduce the variability of different single-models and to construct a more stable and accurate 327 model, To obtain the optimal estimates, the ensemble model was further applied to merge the 328 predictions of random forest and boosted regression trees according to their cross-validated R² 329 values. The predictive value of ensemble model was calculated in a given grid cell as described by 330 Eq. (5) (Marmion et al., 2009). The model accuracy was calculated by regressing the predictive 331 values of ensemble model against the observed trait values. The accuracy of the ensemble model 332 was calculated by regressing the 20% of cross-validation data used for testing against the observed 333 trait values.

341

$$EM_t = \frac{\sum_{m=1}^{2} (pred_{m,t} \times r_{m,t}^2)}{\sum_{m=1}^{2} r_{m,t}^2}$$
(5)

335 where $Pred_EM_t$ is the predictive values of t trait in the ensemble model; $pred_{m,t}$ is the 336 predictive values of t trait in m model; $r_{m,t}^2$ is the cross-validated R² of t trait in m model.

To evaluate the model performance (i.e. the variability in the prediction across models), the coefficient of variation (CV) was calculated as the difference between the predictions of random forest and boosted regression trees methods and the ensemble prediction weighted by the predictive performance of each of the models. CV is calculated <u>as</u> following Eq. (4<u>6</u>):

$$CV_t = \frac{\frac{\sqrt{\sum_{m=1}^{2} (pred_{m,t} - obs_t)^2 * r_{m,t}^2}}{\sum_{m=1}^{2} r_{m,t}^2}}{obs_t}$$
(46)

where $pred_{m,t}$ is the predictive values of t trait in m model, jobs is the values of t trait in the ensemble model, $r_{t,t}^2 r_{m,t}^2$ is the cross-validated R² of t trait in m model.

344 2.6-6 Uncertainty assessments

Multivariate environmental similarity surface analysis (MESS) was used to identify the range of the extrapolated predictor values across the locations in the plant trait dataset (Elith et al., 2010). This 347 method is often used to evaluate the extent of extrapolation and the applicability domain. If the

- 348 values are negative, this indicates that at a given grid cell, at least one predictor variable is outside
- 349 the extent of referenced predictor layer. This analysis was conducted using the 'mess' function in
- 350 the 'dismo' package.
- 351 All analyses were performed in R 4.0.2 (R Core Team, 2020).

352 **3 Results**

353 **3.1 Performances of prediction models**

Cross-validation showed that the performance of the predictive models differed greatly among the plant functional traits (Table 2, <u>Tables C1 and C2 in Appendix C</u>). WD had the best performance in all three models, with R² values of 0.64, 0.68 and 0.67 for random forest, boosted regression trees and ensemble model, respectively. <u>SLALPC, height</u> and <u>LPC SLA-had R² values greater than 0.45</u>, while <u>SM and-LDMC</u> performed the worst, with R² values below 0.25. <u>In addition, the ensemble</u> model performed better than the random forest and boosted regression trees alone (Tables C1 and <u>C2 in Appendix C)</u>.

Table 2 Results of plant functional traits for cross-validated R², <u>N</u>RMSE and MAE for random
 forest, boosted regression trees, and ensemble model.

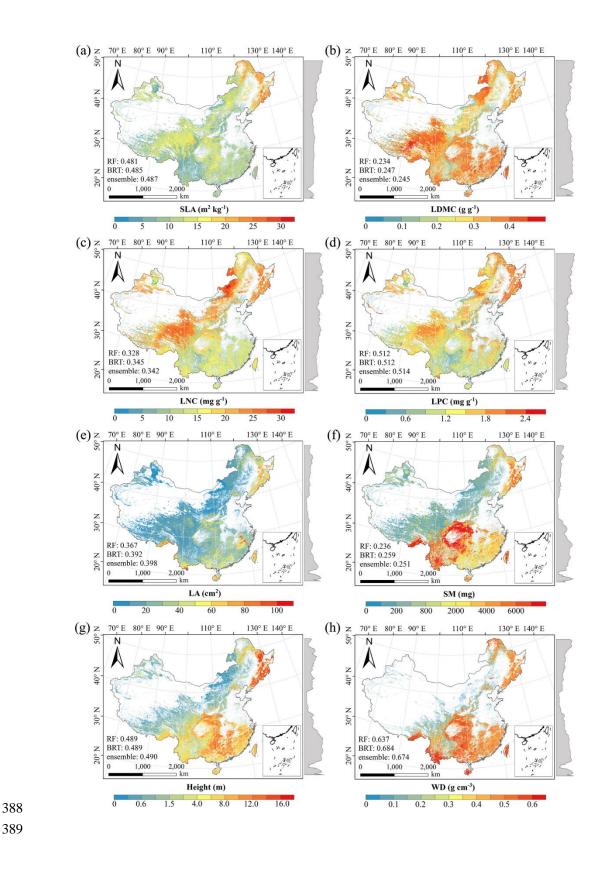
		Random forest		Boosted regression trees			Ensemble model		
Traits	\mathbb{R}^2	<u>N</u> RMSE	MAE	\mathbb{R}^2	<u>N</u> RMSE	MAE	\mathbb{R}^2	<u>N</u> RMSE	MAE
SLA	0.48	7.02 0.22	5.10	0.48	<u>6.990.20</u>	5.08	0.49	<u>6.980.21</u>	5.07
LDMC	0.23	<u>0.100.21</u>	0.07	0.28	<u>0.090.18</u>	0.07	0.24	<u>0.090.20</u>	0.07
LNC	0.33	<u>6.640.19</u>	4.92	0.34	<u>6.520.18</u>	4.85	0.34	<u>6.540.19</u>	4.85
LPC	0.51	0.80 <u>0.24</u>	0.53	0.51	0.80 <u>0.22</u>	0.53	0.51	0.80 <u>0.27</u>	0.53
LA	0.37	<u>68.940.45</u>	26.76	0.39	<u>67.69</u> 0.51	27.47	0.40	<u>67.280.58</u>	26.59
SM	0.24	4547.22	1228.07	0.26	4478.88	1183.61	0.25	4499.67	1201.83
Height	0.49	2.89	2.09	0.49	<u>2.89</u>	2.10	0.49	2.89	2.10
WD	0.64	<u>0.130.20</u>	0.10	0.68	<u>0.12</u> 0.13	0.10	0.67	0.13<u>0.17</u>	0.10

363 SLA, specific leaf area (m² kg⁻¹); LDMC, leaf dry matter content (g g⁻¹); LNC, leaf N
364 concentration concertation (mg g⁻¹); LPC, leaf P concentration concertation (mg g⁻¹); LA, leaf area (cm²);
365 WD, wood density (g cm⁻³); R², determinate coefficient; NRMSE, normalized root-mean-square error;
366 MAE, mean absolution error. SM, seed mass.

367 **3.2 Spatial patterns of predicted plant functional traits**

There were relatively consistent spatial patterns for SLA, LNC and LPC, with high values in the northeastern and northwestern regions and the southeastern Qinghai-Tibet Plateau, and low values in southwestern China (Figs. 2a3a, 2e3c and 2d3d, Figs. D1 and, D2 in Appendix D). SLA and LPC increased with latitude, while LNC did not vary significantly along the latitudinal gradient. For SLA, LNC and LPC, the variability was low among the random forest, boosted regression trees and ensemble model, with an overall CV less than 0.3 (Figs. 3a4a, 3e4c, and 3d4d). LDMC values were

- relatively high in most regions of China, and the low values were mainly located in eastern Yunnan
 and the Loess Plateau (Fig. 2b3b, Figs. D1 and, D2 in Appendix D). LA and SM showed consistent
 spatial patterns, withshowed high values in the northeastern and southern regions (except for the
 Sichuan Basin), and the southeastern Qinghai-Tibet Plateau (Figs. 23e and 2f, Figs. D1, and D2
 in Appendix D). The strong latitudinal gradients were was observed in LA and SM, where values
 decreased with latitude.
- The CV values of LPC and SM decreased with latitude, but other traits did not show latitudinal patterns (Figs. <u>34</u>). The CV values <u>of LA</u> were relatively high, especially in the northwestern region, <u>and the Inner Mongolia-Plateau-Loess Plateau region (only for LA) and Yunnan province (only for</u> <u>SM) (Figs. 34e and 3f)</u>. <u>Plant height and WD had consistent spatial patterns, with high values in the</u> northeastern and southern regions (Figs. <u>2g-2f and 2h</u>, Figs. D1, <u>and D2</u> in Appendix D). <u>The CV</u> <u>values across models for plant height were higher in northwestern China and Inner Mongolia</u> <u>Plateau-Loess Plateau region, while CV values for WD in China were low throughout China (Figs.</u>
- 387 3g <u>4f</u>and 3h).



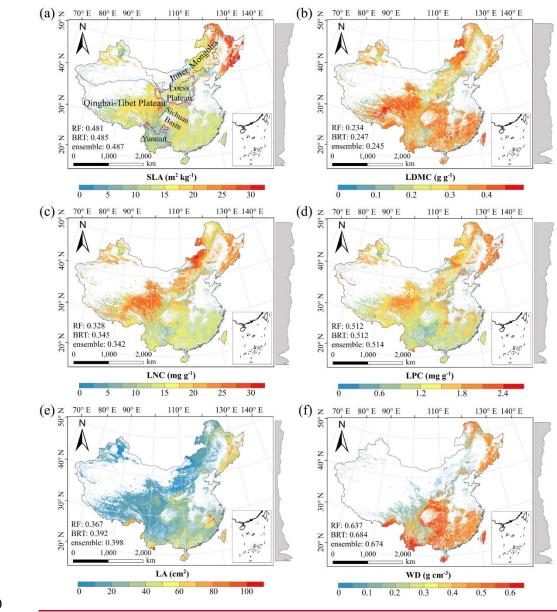
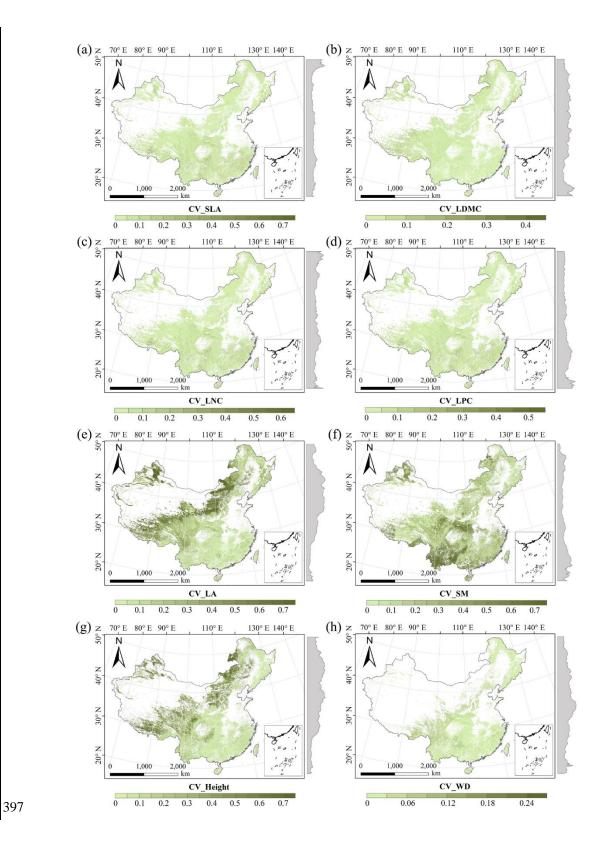
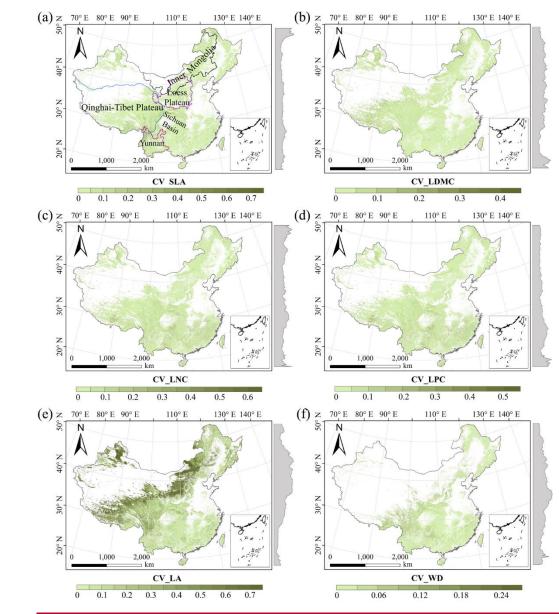


Figure 23. Spatial patterns of predicted plant functional traits in China based on the ensemble
 model. The grey curves to the right of the maps display trait distribution along with latitude. RF,
 random forest; BRT, boosted regression trees; ensemble, ensemble model; SLA, specific leaf area;
 LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>concertation; LPC, leaf P
 <u>concentration</u>concertation; LA, leaf area; WD, wood density; <u>SM, seed mass</u>.





398

Figure <u>34</u>. The variability in plant functional trait predictions among random forest, boosted
 regression trees and ensemble model. The grey curves to the right of the maps display coefficient of
 variation along with latitude. SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N
 <u>concentrationconcertation</u>; LPC, leaf P <u>concentrationconcertation</u>; LA, leaf area; WD, wood density;
 <u>SM, seed mass</u>.

404 **3.3 Relative importance of predictive variables**

The dominant factors explaining spatial variation differed greatly among the eight plant functional traits (Table 3). Overall, climate variables were more important for predicting plant functional traits than were soil variables. Temperature variables (i.e., MAT, MDR and TS) showed close relationships with SLA, LDMC, LPC and WD, while precipitation variables (i.e., PS, PEQ, MAP and PDQ) were more important for predicting the spatial patterns of LNC, LPC, <u>and LA and plant</u> 410 height. RAD was the fourth most dominant factor in predicting the spatial patterns of SLA, SM, 411 and WD-and plant height. Elevation also played an important role in the LDMC and LPC predictions. 412 Within soil variables, soil nutrients {(i.e., pH and and soil available P (SAP)}) showed close associations with SLA, and LNC and SM. In addition to the environmental variables, MTCI 413 414 emerged as an important predictor for explaining SLA, LDMC, and LA and plant height. Finally, 415 EVI was the most important predictor for LA-and SM, and MIR in January and May were the 416 primary predictors of WD. The relationships between plant functional traits and the most important 417 variables were shown in Figs. E1 and E2 in Appendix E.

418

Table 3 List of the eight most important variables for plant functional trait predictions.

-								
Ra	ink SLA	LDMC	LNC	LPC	LA	SM	Height	WD
1	I SAP	MAT	PS	MDR	EVI5	EVI8	PEQ	MIR1
2	2 TS	Elevation	SAP	PDQ	PEQ	SAP	EVI6	TS
3	3 blue9	MTCI5	pH	Elevation	MTCI9	MAT	MTCI6	MIR5
4	4 RAD	blue8	MDR	MIR8	NIR9	RAD	RAD	RAD
5	5 MTCI4	MTCI4	MAP	Tmax	AI	NIR1	pH	MIR6
e	6 MTCI6	MTCI6	PEQ	MTCI6	MTCI6	NIR8	MAP	pН
7	7 Elevation	NIR1	MIR1	MIR7	MAP	SAND	red9	red5
	8 MTCI7	CEC	Tmax	MIR9	red5	BD	red5	PS

419 SLA, specific leaf area (m² kg⁻¹); LDMC, leaf dry matter content (g g⁻¹); LNC, leaf N concentration (mg g⁻¹)concertation; LPC, leaf P concentration (mg g⁻¹)concertation; LA, leaf area (cm²); WD, wood density 420 (g cm⁻³); SM, seed mass; SAP, soil available P; TS, temperature seasonality; blue, blue reflectance; RAD, 421 422 solar radiation; MTCI, MERIS terrestrial chlorophyll index; MAT, mean annual temperature; NIR, near-423 infrared reflectance; CEC, cation exchange capacity; PS, precipitation seasonality; MDR, mean diurnal 424 range; MAP, mean annual precipitation; PEQ, precipitation of wettest quarter of a year; MIR, middle 425 infrared reflectance; Tmax, max temperature of warmest month of a year; PDQ, precipitation of driest 426 quarter of a year; EVI, enhanced vegetation index; AI, aridity index; red, red reflectance; SAND, soil 427 sand content; BD, bulk density.

3.4 Model performance 428

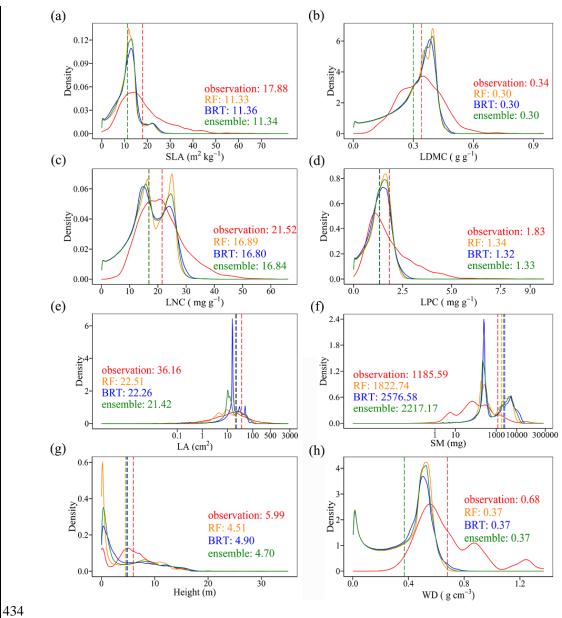
429 The distributions of the predictive trait values based on random forest, boosted regression trees, and

430 ensemble model were consistent with the original trait observations, especially the peak values (Fig.

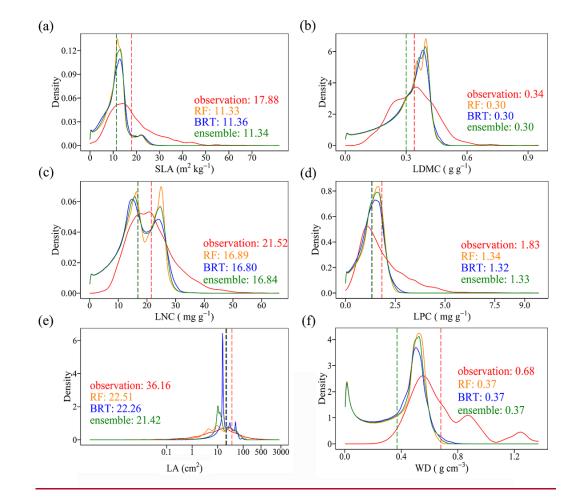
431 432

45). Except for SM, tThe mean values of trait observations were relatively higher than those of the predictive values.

433



1.5 1



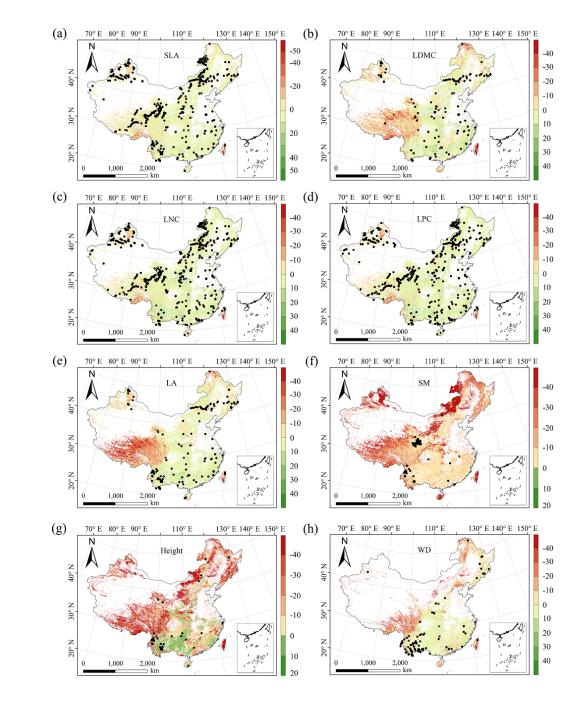
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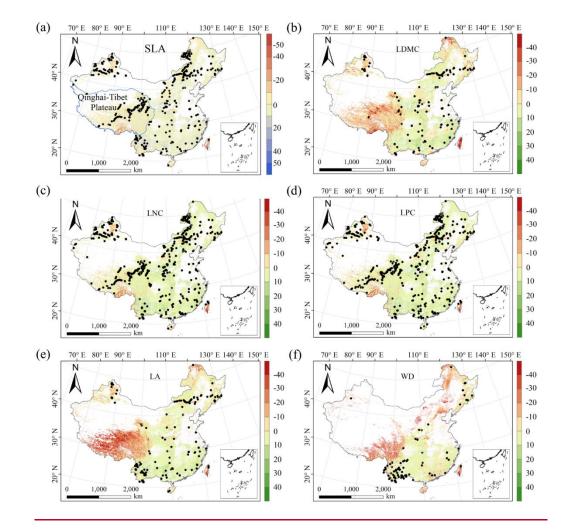
Figure 45. Comparison of trait distribution between observations and predictive values in each of the different models. Each panel depicts the distribution of observations in solid red, of the random forest (RF) model in yellow, of the boosted regression trees (BRT) model in blue, and of the ensemble model in green. The dashed vertical lines indicate mean values. SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>eoncertation; LPC, leaf P <u>concentrationconcertation</u>; LA, leaf area; WD, wood density; <u>SM, seed mass</u>.

442 **3.5 Uncertainty assessments**

The MESS values of five leaf traits and WDall plant functional traits were positive in most regions, indicating a wide applicability domain of our models (Fig. 56). Nevertheless, trait predictions should be interpreted carefully for northeastern China and the Qinghai-Tibet Plateau due to the sparse samplings in these regions. In addition, spatial predictions for SM and plant height were extrapolated to a larger extent than were the other plant functional traits.

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450

Figure 56. Multivariate environmental similarity surface (MESS) assessments for the eight six plant functional traits. The black dots represented the locations of trait observations. More intense shades indicate greater similarity (blue) or difference (red) in environmental conditions of the location compared to the predictive factors covered by the training dataset. SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>concertation; LPC, leaf P <u>concentration</u>concertation; LA, leaf area; WD, wood density; SM, seed mass.

457 **4 Discussion**

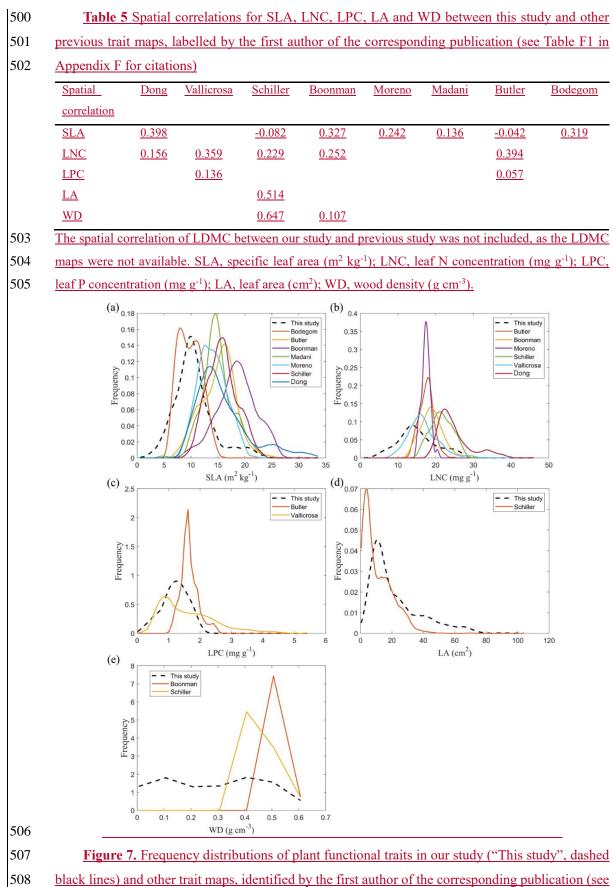
458 **4.1 Comparison with previous work**

Our study predicted the spatial patterns of <u>eight six</u> key plant functional traits across China using machine learning methods and identified the applicability domain of the models. WD had the highest precision with an R^2 of 0.66, which was higher than the global WD prediction (Boonman et al., 2020). This improvement in precision may be attributed to the large number and dense occurrence of sample sites as well as the inclusion of vegetation indices in our study. In addition, SLA and LPC <u>also</u> showed good accuracy with R^2 values of 0.50, which was higher than that of Boonman et al. (2020) and consistent with that of <u>Moreno-MartínezMartínez</u> et al. (2018). <u>However</u>, 466 <u>LNC and LA showed relatively poor performance, which may be related to the reason that these</u>
467 two traits were more influenced by phylogeny than environmental variables (Yang et al., 2017; An
468 et al., 2021).

469 The frequency distribution of plant functional traits in China differed between our study and 470 previous studies (Fig. 7, Fig. F1, Table F1 in Appendix F). Given that the spatial resolution of trait 471 maps in most previous studies is 0.5° (except for Moreno-Martínez et al. (2018) and Vallicrosa et 472 al. (2022)), we resampled the data products of previous studies and our study to 0.5° spatial 473 resolution. The distribution in our study contained more predictions at lower values of SLA, LNC 474 and LPC and was broader than those for SLA and LNC in previous global studies. However, the 475 distribution of LNC in our study was consistent with that in Vallicrosa et al. (2022) at the 1 km 476 spatial resolution (Fig. F1 in Appendix F). LA in our study contained more predictions at higher 477 values and was also broader than those in previous global studies. WD did not show the lower and 478 higher predictive values, however, the WD values in the studies of Boonman et al. (2020) and 479 Schiller et al. (2021) had more predictions at higher values and no lower values ($< 0.3 \text{ g cm}^{-3}$). Our 480 predicted values of SLA showed the highest spatial correlation with those of Dong et al. (2023), and 481 LNC showed the strongest spatial correlation with those of Butler et al. (2017) (Table 5). LA and 482 WD showed the best spatial correlation with those of Schiller et al. (2021), but LPC showed 483 relatively weak spatial correlation with those of published studies. 484 Our study also predicted the spatial pattern of SM with an R² of 0.24, which was lower than in the

485 global study of Madani et al. (2018) in which environmental variables explained 45.5% of SM 486 variation. The low precision of SM may be explained in two ways. First, with few sampling sites 487 included in our study, the environmental gradients were narrow. Second, previous studies and our 488 unpublished study have suggested that SM variation is primarily controlled by a phylogenetic effect, 489 whereas the environmental effect is weak (Moles et al., 2006). Therefore, phylogenetic relatedness 490 among species should be considered to improve the predictive precision of the spatial pattern of SM 491 in further work.

492 There was no consensus in the spatial patterns of plant functional traits among the global 493 studies. ThusIn addition, we compared our results to the other studies focused on China. Yang et al. 494 (2016) predicted the spatial distribution of leaf mass per area (1/SLA) and LNC based on trait-495 environment relationships in China and had an R^2 of 0.13-0.16. The lower predictive precision may 496 be because Yang et al. (2016) only used MAT, MAP₃ and RAD as predictors in estimating the spatial 497 patterns of leaf mass per area and LNC, which likely led to poor performance and low heterogeneity. 498 These results also demonstrated the advantage of our methods in mapping the spatial patterns of 499 plant functional traits at a regional scale.



509 <u>Table F1 for citations). SLA, specific leaf area (m² kg⁻¹); LNC, leaf N concentration (mg g⁻¹); LPC,
 510 leaf P concentration (mg g⁻¹); LA, leaf area (cm²); WD, wood density (g cm⁻³).
</u>

511 4.2 Spatial patterns of plant functional traits in China

512 Our study revealed the spatial patterns of different plant functional traits across China, and the 513 variability among the two machine learning methods was relatively low. We compared the spatial 514 differences of trait maps between our study and previous studies at the global scale In some regions, 515 there were consistent patterns in plant functional traits between our study and the previous 516 studies(Figs. F2-F6 in Appendix F). For example, our study showed high SLA values in the 517 southeastern Qinghai-Tibet Plateau, which concurred with the global study of Boonman et al. (2020). 518 The spatial difference of SLA between our study and Bodegom et al. (2014) was relatively low, and 519 the predictive values in most regions were slightly lower in our study than those in Bodegom et al. 520 (2014). The spatial pattern of difference in SLA between our study and Moreno et al. (2018), Bulter 521 et al. (2017) and Bodegom et al. (2020) was consistent, and the values were higher in northeastern 522 China and southwestern Qinghai-Tibet Plateau in our study than those studies. Our study showed 523 higher LNC values in the northern Inner Mongolia-the Loess Plateau-the eastern Qinghai-Tibet 524 Plateau and northwestern China than those studies at the global studies (Butler et al., 2017; Moreno-Martínez et al., 2018; Boonman et al., 2020; Vallicrosa et al., 2022; Dong et al., 2023), reflecting 525 526 the consistent spatial pattern among these studies. However, Yang et al. (2016) predicted high LNC 527 values in northeastern and northwestern China, northern Inner Mongolia and the entire Qinghai-528 Tibet Plateau, and SLA and LNC had low heterogeneity overall. The discrepancy with Yang et al. 529 (2016) may be attributed to spatial extrapolation based on trait-climate relationships with a low 530 predictive precision. There was no consistent spatial pattern in LPC between our study and previous studies. Consistent with the global pattern (Wright et al., 2017), LA was larger in southern regions 531 532 than in northern regions and showed a decreasing trend with latitude. In addition, LA and WD values 533 in our study were lower in most regions than those ones at the global scale. These discrepancies 534 between our study and previous studies at the global scale may be related to three reasons. First, 535 there is bias in the available in-situ field measurement data from China in these global studies, with 536 large gaps in western China for SLA and no data in China for WD (Boonman et al., 2020). Second, 537 some trait-environment relationships may be scale dependent (Bruelheide et al., 2018), and these 538 studies we compared are from the global scale due to the trait maps in China are not available. Third, 539 the methods used for trait mapping were different among studies, including eco-evolutionary 540 optimality models (Dong et al., 2023), Convolutional Neural Networks based on RGB photographs 541 (Schiller et al., 2021), machine learning algorithms (Vallicrosa et al., 2022; Boonman et al., 2020) 542 and multiple regression analysis (Bodegom et al., 2014). 543 Consistent with the global pattern (Wright et al., 2017), LA was larger in southern regions than in 544 northern regions and showed a decreasing trend along a latitudinal gradient. Potapov et al. (2021)

544 northern regions and showed a decreasing trend along a latitudinal gradient. Potapov et al. (2021) 545 mapped global forest canopy height at a spatial resolution of 30 m by integrating GEDI and Landsat 546 data, and their resulting spatial pattern of plant height in China was consistent with our study. This 547 confirms the accuracy and reliability of predicting plant height through spatial extrapolation as in 548 our study. However, in some regions there were contradictory patterns in the plant functional traits 549 between our study and previous ones. Our study showed high LNC values in the northern Inner 550 Mongolia Plateau Loess Plateau eastern Qinghai-Tibet Plateau and high heterogeneity in southern 551 China, whereas Yang et al. (2016) predicted high LNC values in northeastern and northwestern 552 China, northern Inner Mongolia Plateau, and the entire Qinghai Tibet Plateau, but low heterogeneity 553 overall. In addition, Yang et al. (2016) predicted low SLA variability in China, especially in the 554 Qinghai-Tibet Plateau. These discrepancies may be attributed to spatial extrapolation based on trait-555 climate relationships with a low predictive precision.

Moreover, our study also identified the applicability domain of our models for predicting the spatial patterns of plant functional traits across China. Five leaf traits and WD appeared to have poor applicability in northeastern China and the Qinghai-Tibet Plateau, primarily due to sparse samplings. Although the predictions of plant height and SM were obtained in most regions through spatial extrapolation, the results were consistent with previous studies. This demonstrates the reliability of our methods. Future studies predicting plant functional traits across a large scale through remote sensing observations or other supplementary data will be needed to re-evaluate our results.

563 **4.3 The role of predictive variables**

564 Our study indicates that environmental variables are important for predicting the spatial patterns of 565 plant functional traits, especially climate variables. Temperature variables were primary predictors for SLA, LDMC, LPC and WD. The relationships between leaf traits and temperature have been 566 widely discussed in global and regional studies (Reich and Oleksyn, 2004; Bruelheide et al., 2018). 567 The positive linkage between WD and temperature may be driven by changes in the viscosity of 568 569 water. Plants can adapt to the low water viscosity at high temperatures by reducing the diameter and 570 density of their vessels and by thickening cell walls (Roderick and Berry, 2002; Thomas et al., 2004). Precipitation variables were important predictors for leaf nutrient traits, __and LA, and plant height. 571 572 For example, precipitation of wettest quarter of a year was the factor that most influenced LA 573 variation, which has been confirmed by a previous study (An et al., 2021). A smaller LA could be 574 an adaptive strategy to decrease water loss via reducing the surface area for transpiration under dry 575 environmental conditions (Du et al., 2019). Water availability emerged as the primary predictor of 576 plant height, which may be explained by the hydraulic limitation hypothesis. Relative to shorter 577 plants, taller plants tend to be at a higher risk of embolisms due to their increased xylem length and 578 conduit width. Thus, tall plants may be forced to close their stomata and reduce the amount of 579 photosynthesis (Renninger et al., 2009; Wang et al., 2019). Although the effects of soil on trait 580 predictions were relatively weak, we found that SAP and pH played key roles in SLA, and LNC 581 and SM predictions. These results were similar with the previous studies that reported that soil pH 582 was an important driver of trait variation at the global scale and in tundra regions (Kemppinen et al., 2021; Maire et al., 2015; Kemppinen et al., 2021). Additionally, from the perspective of cost-583 584 efficient theory, the strong effects of SAP reflected that high SLA and small seeds-may be an 585 adaptation for facilitating soil exploration more efficiently in fertile soils (Freschet et al., 2010).

Vegetation indices have recently been proposed as important predictors of spatial patterns of 586 plant functional traits (Loozen et al., 2018). Our results corroborate these findings and further 587 suggest that EVI, MTCI and MIR reflectance are important predictors in models. Here, the 588 589 underlying mechanisms between vegetation indices and plant functional traits are not further 590 discussed due to their complexity and uncertainty. However, our results indicated that vegetation 591 indices and NIR reflectance are not key predictors of LNC estimation, which contrasts the findings 592 from global and regional studies (Wang et al., 2016; Loozen et al., 2018; Moreno-Martínez et al., 593 2018; Wang et al., 2016). This may be related to the multitude of factors that influence the relationship between LNC and vegetation indices and NIR reflectance, such as forest type and 594 595 canopy structure (Dahlin et al., 2013).

596 **4.4 Uncertainties**

Although our study mapped the spatial patterns of key functional traits of seed plants in China 597 598 through large-scale field investigations and compared the predictions with previous studies 599 performed at global and regional scales, there persists some uncertainties in the interpretation of 600 these results. First, the predictive ability of machine learning methodmodels was relatively worse 601 for certain traits, especially for SM and LDMC. Beyond the environmental effects, the variation in 602 plant functional traits is also regulated by phylogenetic structure among plant species (e.g., family, 603 order and phylogenetic clade) (Li et al., 2017). Consequently, incorporating the phylogenetic 604 information will be a promising avenue in future studies for further improving the accuracy of 605 spatial predictions of plant functional traits (Butler et al., 2017). A second potential issue is sampling 606 bias; there were major spatial gaps in field investigation in both the northeastern China and the 607 Oinghai-Tibet Plateau., especially for SM and plant height. There are an increasing number of 608 available datasets and studies on SM and plant height, but many did not provide the corresponding 609 geographic coordinates, thus rendering the data unusable. In addition, dDue to the challenges-lack 610 of measurements for small shrubs and low vegetation, WD data is mainly confined to eastern forests 611 (Perez-Harguindeguy et al., 2013), and the overall quantity of WD data was much lower than that 612 of leaf and reproductive traits, even in the TRY database, which is the largest trait database in the 613 world (Kattge et al., 2020). The environmental information of sampling sites was not always 614 obtained from original literature, thus using the public environmental products is a common 615 resolution in large-scale plant trait studies (Boonman et al., 2020; Vallicrosa et al., 2022). Such mismatch between in-situ trait measurements and predictors should be resolved in further work. 616 617 Finally, additional key challenges in data availability must be resolved to scale up from the species 618 to the community levels, in particular with data surrounding species co-occurrence and their relative 619 cover or abundance in ecological communities (He et al., 2023). Global biodiversity data (e.g., sPlot 620 and Global Biodiversity Information Agency databases) that contains information on species 621 occurrence or the proportion of species in a community has the potential for enabling the calculation of community-weighted trait values and the re-evaluation of our results in future work (Telenius, 622

- 623 2011; Bruelheide et al., 2019). The lack of consistent time period and spatial resolution of predictors
- 624 due to limitation of data availability is another key challenges in the spatial mapping of plant
- 625 functional traits. In addition, although Worldclim version 2.1 product has high spatial resolution and
- 626 includes various aspects of climatic parameters, there exists certain limitation and uncertainty in
- 627 predicting trait maps. In addition, Therefore, integrating satellite remote sensing monitoring
- 628 methods with in-in-situ trait data collection can also provide an effective way to estimate and assess
- 629 the species diversity at a large scale (Cavender-Bares et al., 2022).

630 **4.5 Potential applications**

631 Maps of these key functional traits of seed plants highlighted large-scale variability in space, which 632 will significantly advance ecological analyses and future interdisciplinary research. First, using the 633 spatially continuous trait datasetsmaps, one can optimize and develop trait-flexible vegetation models, which allows for the exploration of the community assembly rules based on how plants 634 635 with different trait combinations perform under a given set of environmental conditions (Berzaghi et al., 2020). When trait-flexible vegetation models are available, incorporating trait maps into 636 637 models will bridge the gap for vegetation classifications and predictions of vegetation distribution 638 under global change, which is not feasible in conventional vegetation models (Van Bodegom et al., 639 2012; Yang et al., 2019). Second, the assessments of China's terrestrial ecosystem carbon sink have 640 had large uncertainties so far (Piao et al., 2022), but the spatial continuous trait datasets-maps will 641 provide an effective way to link ecosystem characteristics to ecosystem carbon sink estimates in 642 China (Madani et al., 2018; Šímová et al., 2019). These analyses will help shed light on the 643 mechanisms underlying plant functional traits and terrestrial ecosystem carbon storage at a large 644 scale.

645 **5 Data availability**

646 The original-eight plant functional traits dataset data collected in this study that were used for 647 machine learning models (named by Data file used for machine learning models.csv) and final maps 648 of eight-plant functional traits in terrestrial ecosystems in a GeoTIFF format across China (named 649 by plant functional available the trait category) are now for private link 650 https://figshare.com/s/c527c12d310cb8156ed2 (An et al., 2023). Once the article is accepted, we 651 will publicly publish these datasets maps at the figshare website.

652 6 Conclusions

We <u>ereated-generated</u> a set of spatial continuous trait <u>datasets-maps</u> at a 1-km spatial resolution using machine learning methods in combination with field measurements, environmental variables and vegetation indices. Models for leaf traits (except for LDMC), and WD and plant height showed good accuracy and robustness, whereas models of SM and LDMC had relatively poor precision and robustness. Temperature variables were the most important predictors for leaf traits (except for LA) and WD, and precipitation variables were the most important predictors for leaf nutrient traits, <u>and</u>

- 659 LA-and plant height. We caution that plant functional trait predictions should be interpreted carefully
- 660 for northeastern China and the Qinghai-Tibet Plateau. The spatial continuous trait datasets-maps
- 661 generated in our study are complementary to current terrestrial in-situ observations and offer new
- 662 avenues for predicting large-scale changes in vegetation and ecosystem function under climate 663 scenarios in China.
- 664

665 Appendix A Data collection from literature

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Appendix B 981

982 Table B1 Summary of statistics in plant functional traits, environmental variables and geographical 983 distribution in China.

Trait	Unit	Range	Mean	CV (%)	No. of species	Entries	Sites
SLA	m ² kg ⁻¹	0.06-81.68	17.88	54.96	2463	9195	1032
LDMC	g g ⁻¹	0.06-0.95	0.34	100.00	1582	3957	193
LNC	mg g ⁻¹	3.41-66.02	21.52	37.44	2335	7407	567
LPC	mg g ⁻¹	0.09-9.70	1.83	62.19	2074	6266	515
LA	cm ²	0.0033-2553.33	36.16	259.64	1838	5976	691
Height	m	0.01 35.00	5.99	67.58	1171	16324	636
WD	g cm ⁻³	0.25-1.37	0.68	33.16	768	1788	639
SM	mg	0.10 201300.00	1185.59	562.32	1163	3080	134
Altitude	m	-144–5454					1541<u>1430</u>
MAT	°C	-12.07-24.32					<u>15411430</u>
MAP	mm	15–2982					<u>15411430</u>
Soil total N	g kg ⁻¹	0.11-10.25					1541<u>1430</u>
Bulk density	g cm ⁻³	0.83-1.45					<u>15411430</u>

984 SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N concentrationeoncertation; LPC, leaf P

concentrationconcertation; LA, leaf area; WD, wood density; SM, seed mass; MAT, mean annual temperature; MAP, 986 mean annual precipitation.

Type of variables	Variable name	Abbreviations	Units	Time periods	Spatial	Source
					resolution	
Climate	Mean annual temperature	MAT	°C	1970-2000	1 km	WorldClim version 2.41
	Mean diurnal range	MDR	°C	1970-2000	1 km	WorldClim version 2.41
	Temperature seasonality	TS	°C	1970-2000	1 km	WorldClim version 2.41
	Max temperature of	Tmin	°C	1970-2000	1 km	WorldClim version 2.41
	warmest month					
	Min temperature of coldest	Tmax	°C	1970-2000	1 km	WorldClim version 2.11
	month					
	Temperature annual range	TAR	°C	1970-2000	1 km	WorldClim version 2.41
	Isothermality	IS	%	1970-2000	1 km	WorldClim version 2.41
	Mean temperature of	MTEQ	°C	1970-2000	1 km	WorldClim version 2.41
	wettest quarter					
	Mean temperature of driest	MTDQ	°C	1970-2000	1 km	WorldClim version 2.1
	quarter					
	Mean temperature of	MTWQ	°C	1970-2000	1 km	WorldClim version 2.1
	warmest quarter					
	Mean temperature of	MTCQ	°C	1970-2000	1 km	WorldClim version 2.1
	coldest quarter					
	Mean annual precipitation	MAP	mm	1970-2000	1 km	WorldClim version 2.1
	Precipitation of wettest	PEM	mm	1970-2000	1 km	WorldClim version 2.1
	month					
	Precipitation of driest	PDM	mm	1970-2000	1 km	WorldClim version 2.1
	month					
	Precipitation seasonality	PS	%	1970-2000	1 km	WorldClim version 2.1
	Precipitation of wettest	PEQ	mm	1970-2000	1 km	WorldClim version 2.1
	quarter					
	Precipitation of driest	PDQ	mm	1970-2000	1 km	WorldClim version 2.1
	quarter					
	Precipitation of warmest	PWQ	mm	1970-2000	1 km	WorldClim version 2.1
	quarter					
	Precipitation of coldest	PCQ	mm	1970-2000	1 km	WorldClim version 2.1
	quarter					
	Aridity index	AI	/	1970-2000	1 km	Global CGIAR-CSI
	Solar radiation	RAD	kJ m ⁻²	1970-2000	1 km	WorldClim version 2.1
			day-1			
Fopography	Elevation	/	m		1 km	SRTM 90m V4.1
Soil	Soil sand content	SAND	%	/	1 km	Shangguan et al. (2013)
	Soil silt content	SILT	%	/	1 km	Shangguan et al. (2013)
	Soil clay content	CLAY	%	/	1 km	Shangguan et al. (2013)
	Bulk density	BD	g cm ⁻³	/	1 km	Shangguan et al. (2013)
	Soil pH	pН	/	/	1 km	Shangguan et al. (2013)

Table B2 List of all the predictors including environment and remote sensing variables used in this study.

Type of variables	Variable name	Abbreviations	Units	Time periods	Spatial	Source
					resolution	
	Soil organic matter	SOC	g kg-1	/	1 km	Shangguan et al. (2013)
	Soil total N	STN	g kg-1	/	1 km	Shangguan et al. (2013)
	Soil total P	STP	g kg-1	/	1 km	Shangguan et al. (2013)
	Soil alkali-hydrolysable N	SAN	mg	/	1 km	Shangguan et al. (2013)
			kg ⁻¹			
	Soil available P	SAP	mg	/	1 km	Shangguan et al. (2013)
			kg ⁻¹			
	Soil available K	SAK	mg	/	1 km	Shangguan et al. (2013)
			kg ⁻¹			
	Cation exchange capacity	CEC	me	/	1 km	Shangguan et al. (2013)
			kg ⁻¹			
EVI	MODIS EVI long-term		/	2001-2018	1 km	MOD13A3 V006
	monthly averages					
NIR	MODIS NIR long-term		/	2001-2018	1 km	MOD13A3 V006
	monthly averages					
MIR	MODIS MIR long-term		/	2001-2018	1 km	MOD13A3 V006
	monthly averages					
Red	MODIS red long-term		/	2001-2018	1 km	MOD13A3 V006
	monthly averages					
Blue	MODIS blue long-term		/	2001-2018	1 km	MOD13A3 V006
	monthly averages					
MTCI	MTCI long-term monthly		/	2003-2011	4.63 km	MTCI level 3 product
	averages					
Land cover	Land cover map		/	2015	100 m	Copernicus Global Land
						Service Collection 3

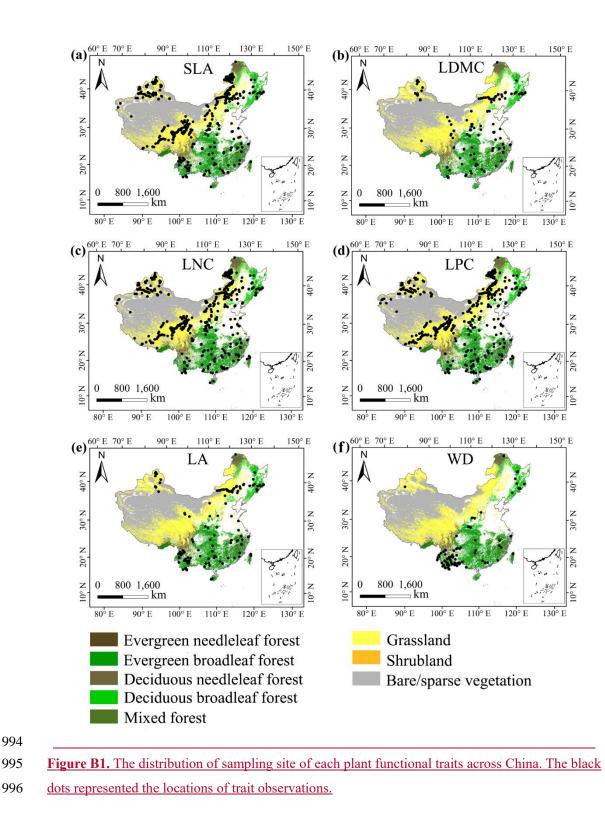
988 The remote sensing variables are calculated as long-term monthly averages from 2001 to 2018. Thus 12 variables of

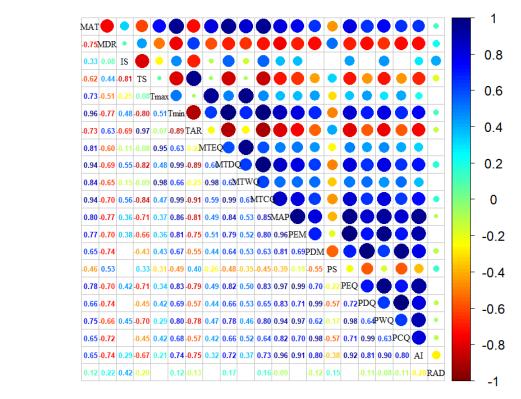
989 each remote sensing category are obtained.

Continued

<u>Traits</u>	<u>No. of samples</u>	<u>No. of samples used</u>	for No. of samples use
		model training	model validation
<u>SLA</u>	<u>9195</u>	<u>7356</u>	<u>1839</u>
<u>LDMC</u>	<u>3957</u>	<u>3166</u>	<u>791</u>
LNC	<u>7407</u>	<u>5926</u>	<u>1481</u>
<u>LPC</u>	<u>6266</u>	<u>5013</u>	<u>1253</u>
LA	<u>5976</u>	<u>4781</u>	<u>1195</u>
WD	<u>1787</u>	<u>1430</u>	<u>357</u>
SLA, specific	c leaf area (m ² kg ⁻¹); LD	MC, leaf dry matter content (g g ⁻¹); LNC, leaf N concentra

Table B3 The number of samples of eight plant functional trait used for model training (80%) and







998 Figure **B1B2**. Correlations among climate variables. The blank indicates that the correlations are 999 not significant (P > 0.05). The size of the circles is proportional to the correlation coefficient. The 1000 abbreviation of climate variables is seen in Table B2.

STN												- 1
									-			- 0.8
0.54	SIP					-			-			
0.95	0.49	SOC	lacksquare		•			•	•			- 0.6
0.83	0.48	0.77	SAN	•	•	•	•	•	•		\bullet	- 0.4
0.50	0.48	0.52	0.55	SAP			•	•	•	•		- 0.2
0.26	0.44	0.23	0.31	0.36	SAK			•	•	•	\bullet	- 0
-0.42	-0.19	-0.42	-0.38	-0.19	-0.21	BD						0
-0.33	0.18	-0.37	-0.37	-0.06	0.27	0.31	PH		•		•	- 0.2
-0.14		-0.12	-0.19	-0.14	-0.11	0.30	0.40	SAND			•	- 0.4
0.24	0.20	0.19	0.29	0.33	0.25	-0.24	-0.09	-0.81	SILT			0.6
	-0.28			-0.12	-0.09	-0.24	-0.56	-0.78	0.26	CLAY		0.8
0.78	0.54	0.74	0.65	0.51	0.37	-0.37	-0.07	-0.21	0.30		CEC	1

Figure B2B3. Correlations among soil variables. The blank indicates that the correlations are not significant (P > 0.05). The size of the circles is proportional to the correlation coefficient. The abbreviation of soil variables is seen in Table B2.

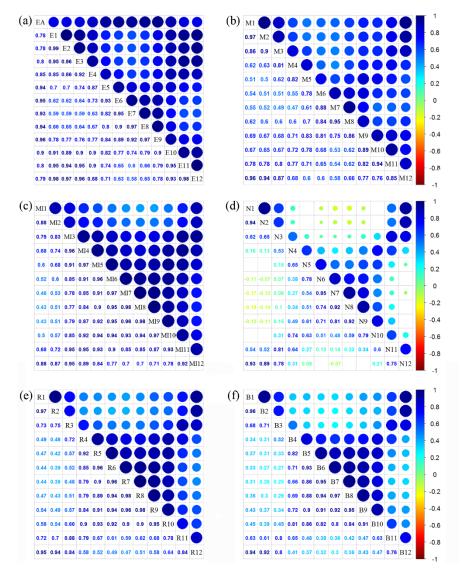


Figure B3B4. Correlations among monthly remote sensing variables. The blank indicates that the correlations are not significant (P > 0.05). The size of the circles is proportional to the correlation coefficient. (a) enhanced vegetation index (EVI); (b) MERIS terrestrial chlorophyll index (MTCI); (c) MIR reflectance; (d) NIR reflectance; (e) red reflectance; (f) blue reflectance.

1010 Appendix C

 1011
 Table C1 Optimal parameter combination and model performance of random forest-(RF) for plant

 1012
 functional traits

Traits	ntree	mtry	R ²	<u>N</u> RMSE	MAE
SLA	1000	24	0.476	7.049<u>0.22</u>	5.134
LDMC	1000	11	0.234	<u>0.0950.20</u>	0.072
LNC	1000	57	0.392	<u>0.1290.00</u>	0.098
LPC	1000	20	0.587	<u>0.176</u> 0.05	0.129
LA	1000	18	0.278	72.996<u>0.48</u>	26.622
Height	1000	38	0.871	0.234	0.178
WD	1000	9	0.531	0.0 <mark>9</mark> 2	0.072
SM	1000	22	0.197	6043.95	1290.866

013 SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>concertation; LPC,

014 leaf P <u>concentration</u>concertation; LA, leaf area; WD, wood density; SM, seed mass.

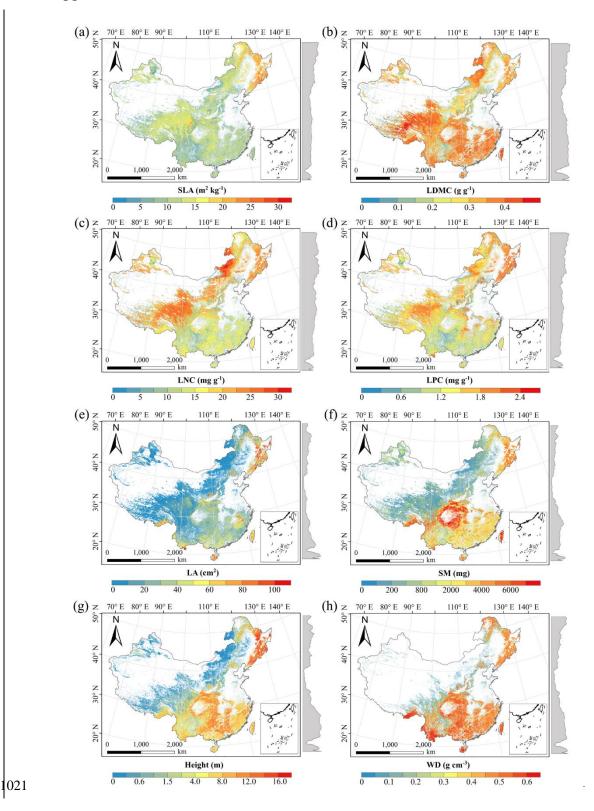
1015

Table C2 Optimal parameter combination and model performance of boosted regression trees (BRT) for plant functional traits

Traits	n.tree	interaction. depth	shrinkage	learning rate	bag fractions	R ²	<u>N</u> RMSE	MAE
SLA	3000	6	0.01	10	0.75	0.486	<u>6.9860.20</u>	5.082
LDMC	3000	2	0.01	10	0.75	0.247	0.094<u>0.19</u>	0.071
LNC	3000	6	0.01	10	0.70	0.414	<u>0.1260.00</u>	0.096
LPC	3000	7	0.01	10	0.75	0.591	0.175<u>0.05</u>	0.129
LA	3000	3	0.001	10	0.75	0.282	72.308<u>0.5</u> <u>5</u>	27.556
Height	3000	3	0.05	10	0.6	0.871	0.234	0.178
WD	3000	4	0.01	10	0.70	0.627	<u>0.0820.01</u>	0.066
SM	3000	7	0.001	10	0.50	0.192	6070.703	1268.386

018 SLA, specific leaf area; LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>concertation; LPC,

019 leaf P <u>concentration</u>concertation; LA, leaf area; WD, wood density; SM, seed mass.



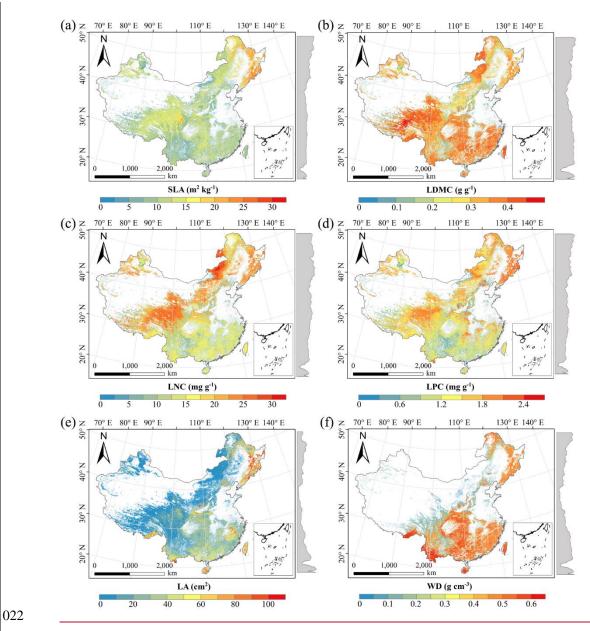
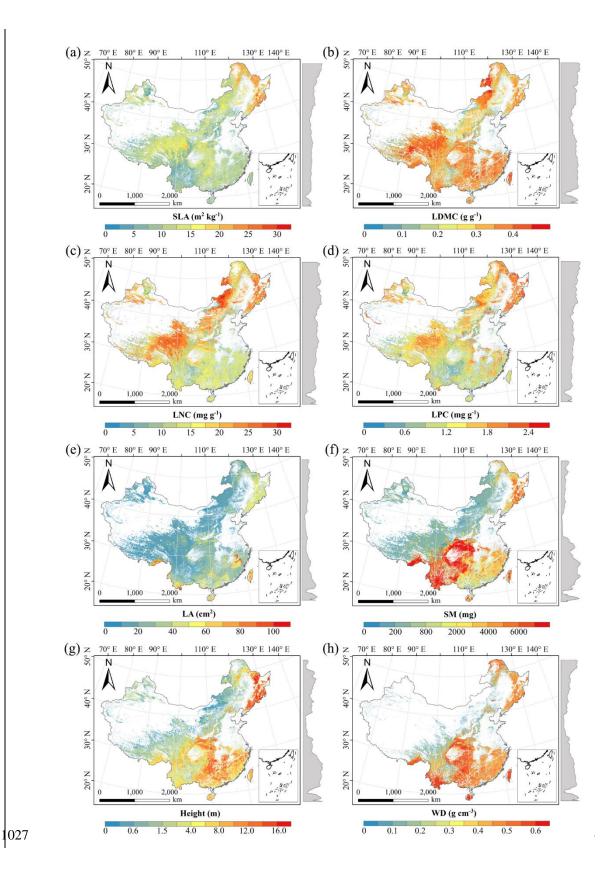


Figure D1. Spatial distribution of plant functional traits based on random forest-(RF). The grey
 curves on the right of maps were trait distribution along with latitude. SLA, specific leaf area;
 LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>; LPC, leaf P
 <u>concentration</u>; LA, leaf area; WD, wood density; SM, seed mass.



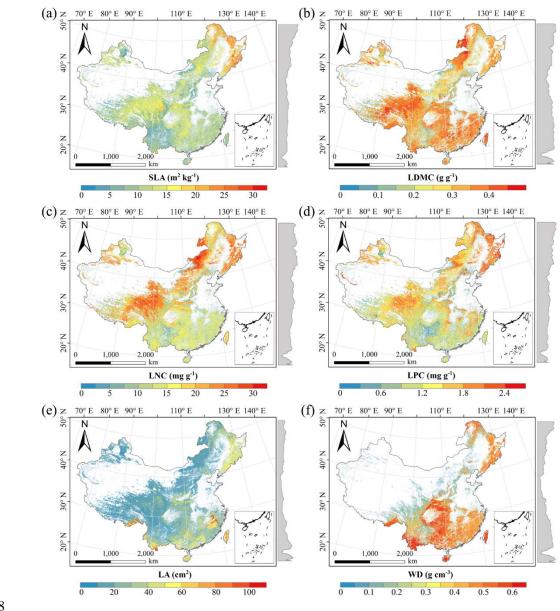
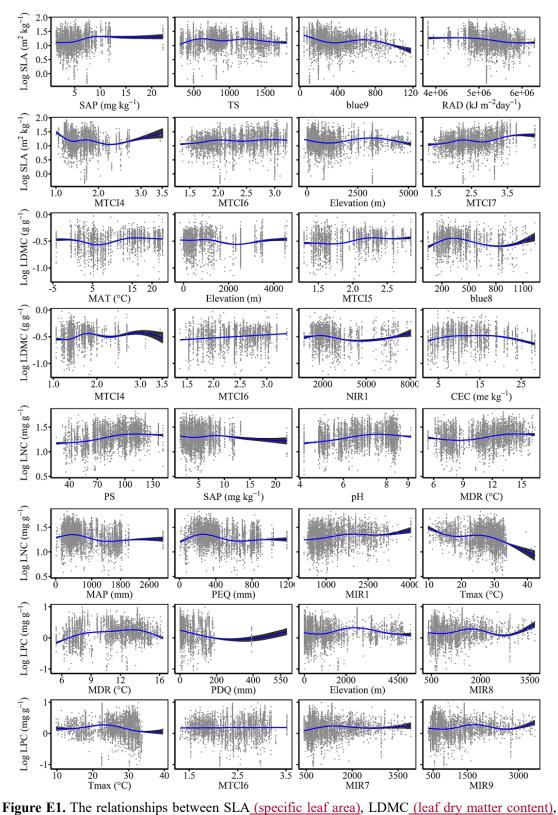
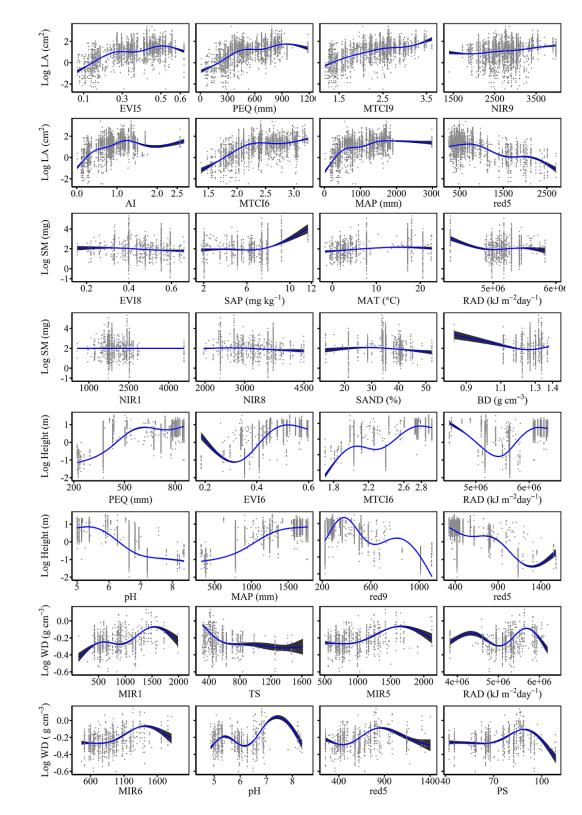


Figure D2. Spatial distribution of plant functional traits based on boosted regression trees (BRT).
 The grey curves on the right of maps were trait distribution along with latitude. SLA, specific leaf
 area; LDMC, leaf dry matter content; LNC, leaf N <u>concentration</u>eoncertation; LPC, leaf P
 <u>concentration</u>concertation; LA, leaf area; WD, wood density; SM, seed mass.





LNC (leaf N concentration), LPC (leaf P concentration) and their eight most important predictors.



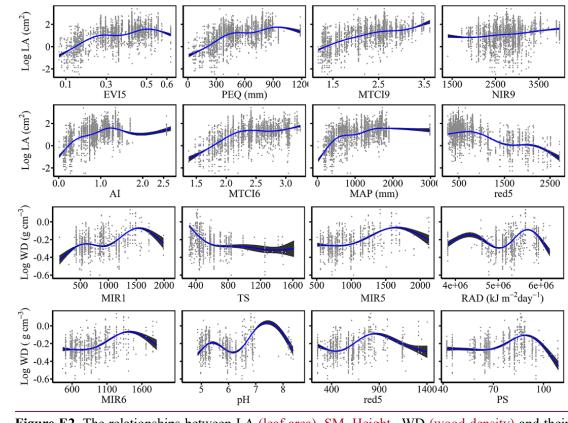


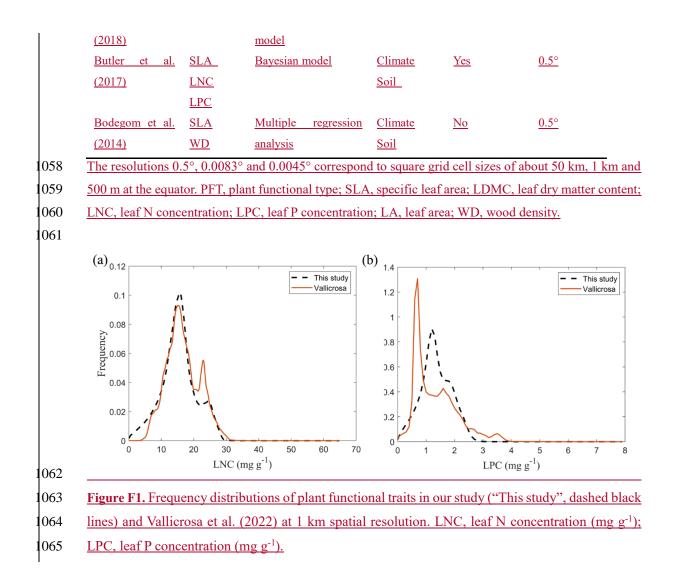
Figure E2. The relationships between LA <u>(leaf area)</u>, <u>SM</u>, <u>Height</u>, WD <u>(wood density)</u> and their eight most important predictors.

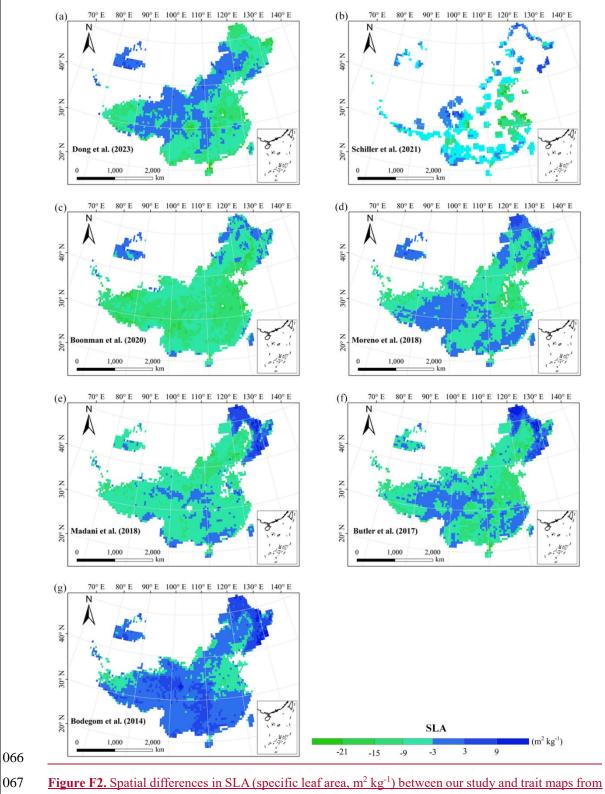
1041 Appendix F Comparisons between our study with trait maps from previous studies

042 Given that the trait maps predicted for China were not available from the literature and authors, 043 we compared our study with those studies performed at the global scale (see Table F1). Thus, we 044 extracted the data in China from global trait maps. Before the quantitative comparisons with 045 previous studies, we performed two steps to make the data products as comparable as possible and 046 improve the consistency between different studies. First, due to different spatial resolution of global 047 trait maps (mainly 0.5°) and our study, we resampled the data products of previous studies and our maps to 0.5° spatial resolution. In addition, Vallicrosa et al. (2022) generated the global maps of 048 049 LNC and LPC with a 1 km spatial resolution, we also compared the frequency distribution of 050 Vallicrosa et al. (2022) with that of our study at a 1 km spatial resolution. Second, our study focused 051 on natural vegetation, so the global trait maps were used to filter out non-natural vegetation (e.g., 052 croplands). For example, Madani et al. (2018) predicted the spatial distributions of SLA that 053 included croplands. We quantitatively compared our maps with previous studies from two 054 perspectives. The comparisons among trait maps were made using frequency plots and spatial 055 correlations (Figure 7 and Table 5). And the maps of spatial differences between our study and 056 previous studies were displayed as Figs F1-F5 in Appendix F.

057 <u>Table F1. Summary table of related trait maps of previous studies used in this study.</u>

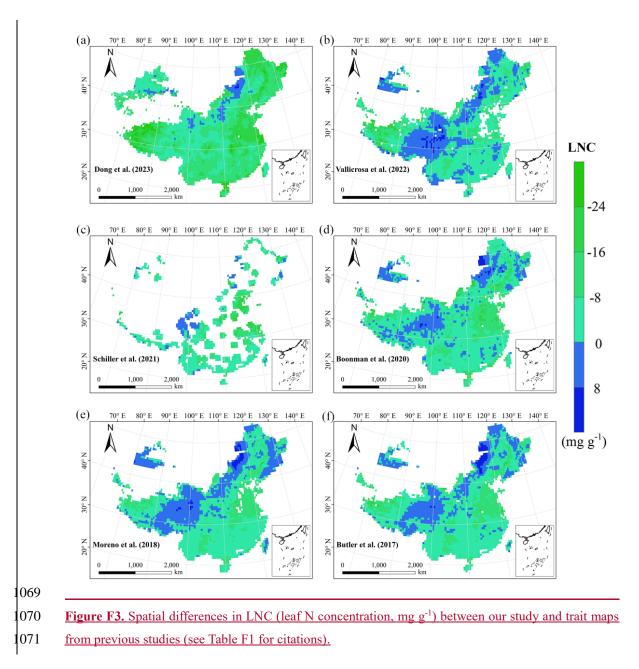
<u>References</u>	<u>Related</u>	Methods	Predictors	Consideration of PFT	<u>Resolution</u>
Dong et al.	SLA	Optimality models	Climate	Yes	<u>0.5°</u>
(2023)	LNC				
Vallicrosa et al.	LNC	Neural networks	<u>Climate</u>	Yes	<u>0.0083°</u>
<u>(2022)</u>	<u>LPC</u>		<u>Soil</u>		
			<u>N and P</u>		
			deposition		
<u>Schiller et al.</u>	<u>SLA</u>	Convolutional Neural	<u>Climate</u>	<u>No</u>	<u>0.5°</u>
<u>(2021)</u>	LNC	<u>Networks</u>	In-situ RGB		
	LA		images		
	WD				
<u>Boonman et al.</u>	<u>SLA</u>	Generalized linear	Climate	No	<u>0.5°</u>
<u>(2020)</u>	<u>LNC</u>	model, Generalized	<u>Soil</u>		
	<u>WD</u>	additive model,			
		Random forest,			
		Boosted regression			
		trees, Ensemble model			
<u>Moreno et al.</u>	<u>SLA</u>	Regularized linear	<u>Climate</u>	Yes	<u>0.0045°</u>
<u>(2018)</u>	<u>LNC</u>	regression, Random	<u>Elevation</u>		
	<u>LPC</u>	forest, Neural	<u>Reflectance</u>		
	LDMC	networks, Kernel			
		networks			
<u>Madani et al.</u>	<u>SLA</u>	Generalized additive	<u>Climate</u>	<u>No</u>	<u>0.5°</u>







previous studies (see Table F1 for citations).



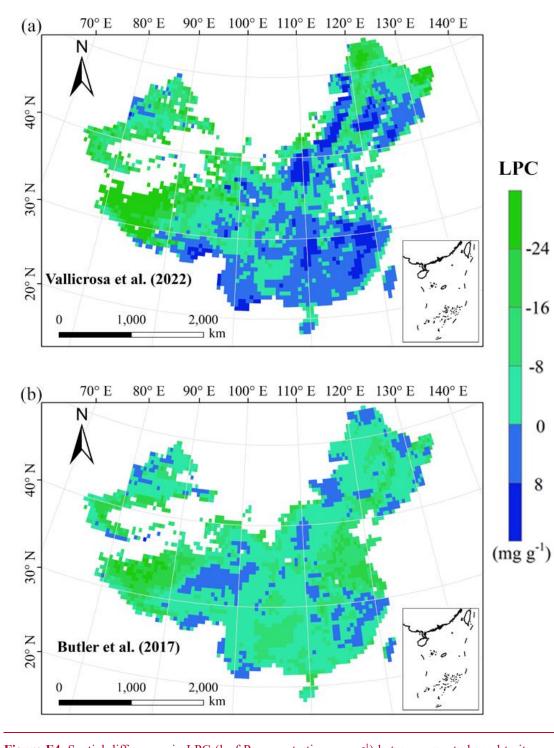
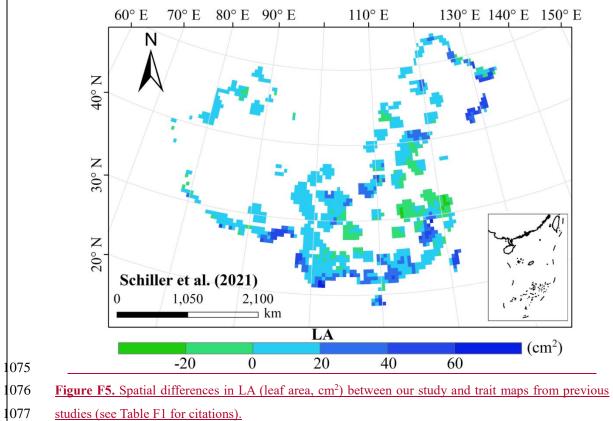
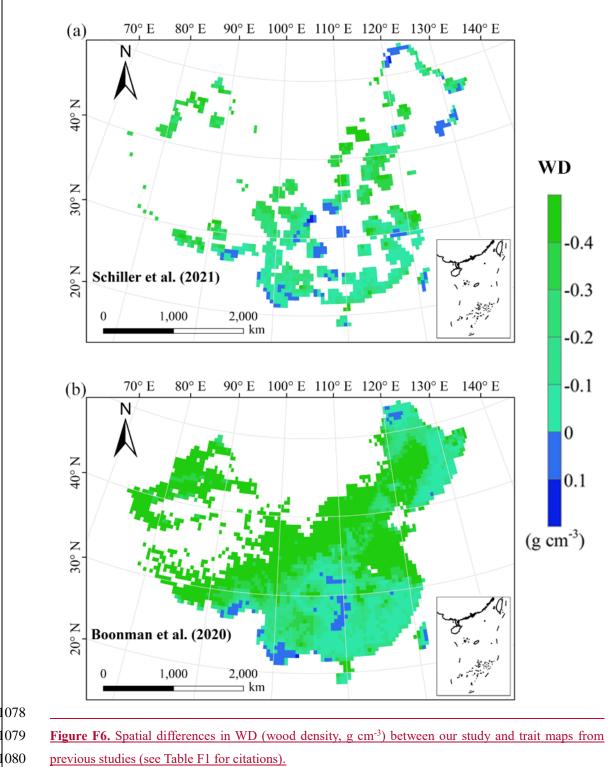




Figure F4. Spatial differences in LPC (leaf P concentration, mg g⁻¹) between our study and trait maps
 from previous studies (see Table F1 for citations).







1081 Author contributions. NA and NL designed the research. NA did the analysis, processed the data 1082 and wrote the draft of the paper. All co-authors commented on the manuscript and agreed upon the 1083 final version of the paper.

1084

1085 Competing interests. The contact author has declared that none of the authors has any competing1086 interests.

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