



1 **Multispecies and high spatiotemporal resolution database of vehicular**
2 **emissions in Brazil**

3
4 Leonardo Hoinaski^{a,*}, Thiago Vieira Vasques^b, Camilo Bastos Ribeiro^b and Bianca Meotti^b

5
6 ^aDepartment of Sanitary and Environmental Engineering, Federal University of Santa Catarina, Flo-
7 rianópolis, Santa Catarina, Brazil, leonardo.hoinaski@ufsc.br

8 ^bGraduate Program in Environmental Engineering, Federal University of Santa Catarina,
9 Florianópolis, Santa Catarina, Brazil, vvthiago@hotmail.com, cb_ambiental@hotmail.com,
10 meottibianca@gmail.com

11

12

13

14 *Corresponding author

15 *E-mail address:* leonardo.hoinaski@ufsc.br (L. Hoinaski).



16

17 **Abstract** – In this article, we present the BRAZilian Vehicular Emissions inventory Software
18 (BRAVES) database, a multispecies and high spatiotemporal resolution database of vehicular emis-
19 sions in Brazil. We provide this database using spatial disaggregation based on road density, temporal
20 disaggregation using vehicular flow profiles, and chemical speciation based on SPECIATE database
21 from the United States Environmental Protection Agency. Our BRAVES database provides hourly
22 and annual emissions of 41 gaseous and particle pollutants. Users can define the spatial resolution,
23 which ranges from a coarse to very refined scale Spatial correlation analysis reveals that the BRAVES
24 database reaches similar performance to the vehicular emissions inventory from Emissions Database
25 for Global Atmospheric Research (EDGAR). A comparison with Modern-Era Retrospective analysis
26 for Research and Applications, Version 2 (MERRA-2) surface concentration confirms the con-
27 sistency and reliability of the BRAVES database on representing the spatial pattern of vehicular emis-
28 sions. Compared to EDGAR, the BRAVES database brings more spatial, temporal, and chemical
29 details. These additional features are crucial to understanding important atmospheric chemistry pro-
30 cesses in Brazil. All codes and inputs are freely available, and the outputs are compatible with the
31 input requirements of sophisticated chemical transport models. We envision that our database will
32 enable the scientific and environmental community to gain new insights into vehicular emissions and
33 their effects in Brazil, where emissions inventories are scarce and urgently needed.

34

35 **Keywords:** Vehicular emissions, temporal disaggregation, chemical speciation, spatial disaggrega-
36 tion, Brazil.

37

38

39



40 **1. Introduction**

41 Vehicular emissions threaten urban air quality (Brito et al. 2018; Sawyer, 2010) and cause
42 several environmental damages from local to global scales. These emissions deleteriously affect hu-
43 man health (Anenberg et al. 2017; Krzyzanowski et al. 2005) and contribute to the increase in the
44 concentration of greenhouse gases in the atmosphere (Shindell et al. 2011; Unger et al. 2010).

45 It is challenging to control vehicular emissions in developing countries where city growth is
46 disorganized manner and vehicle population increases dramatically (Lyu et al. 2020; Sun et al. 2020).
47 Furthermore, vehicular emissions inventories, an essential tool to control air pollution, are scarce and
48 limited to wealthy cities (Huneus et al. 2020) and developed countries. When available, an emissions
49 inventory often does not provide the required data to design air quality management systems.

50 Brazil has experienced a rapid rise in its vehicular fleet (Carvalho et al. 2015) and transport
51 volume. Even though the program to control vehicular emissions has reduced the emissions from the
52 transport sector in Brazil (Andrade et al. 2017), vehicles are still potentially the dominant source of
53 air pollution in several municipalities.

54 The impact of vehicular emissions in many Brazilian municipalities is still unknown (Ribeiro
55 et al. 2021). Current inventories provide only annual emissions from national to municipality scales,
56 not reaching the spatial and temporal resolution necessary for air quality modeling (Álamos et al.
57 2022), nor the concentration of chemical species that participate in chemical reactions in the atmos-
58 phere. For this reason, most regional air quality assessments in Brazil rely on global emissions inven-
59 tories, which have been proved to be biased against local inventories. Also, global inventories do not
60 present enough spatial and temporal resolution for regional and local studies (Ibarra-Espinosa et al.
61 2018). Even in the megacity of São Paulo, where the air quality network is well developed and mul-
62 tiple inventories have been developed, there is still room for improvement in emissions inventories
63 (Andrade et al. 2017), especially regarding chemical species involved in the photochemical process
64 in the atmosphere.



65 In this article, we present the first comprehensive multispecies high spatiotemporal resolution
66 database of vehicular emissions for the entire Brazilian territory. The BRAZilian Vehicular Emissions
67 inventory Software (BRAVES) database has spatial disaggregation based on road density, temporal
68 disaggregation using vehicular flow profiles, and chemical speciation from the US EPA SPECIATE
69 database. The BRAVES database provides hourly and annual emissions of 41 gaseous and particle
70 pollutants. Users can define the spatial resolution from coarse to very refined scales. The emissions
71 are derived from the BRAVES model (Vasques and Hoinaski, 2021), which uses a probabilistic ap-
72 proach that accounts for the fleet characteristics, fuel consumption, vehicle deterioration, and inten-
73 sity of use, to calculate the vehicular emissions from the exhaust, tires, roads, brake wear, soil resus-
74 pension, refueling, and evaporative emissions. Here, we present methods and a comparison between
75 the BRAVES database and independent databases. We also make all codes and inputs freely availa-
76 ble.

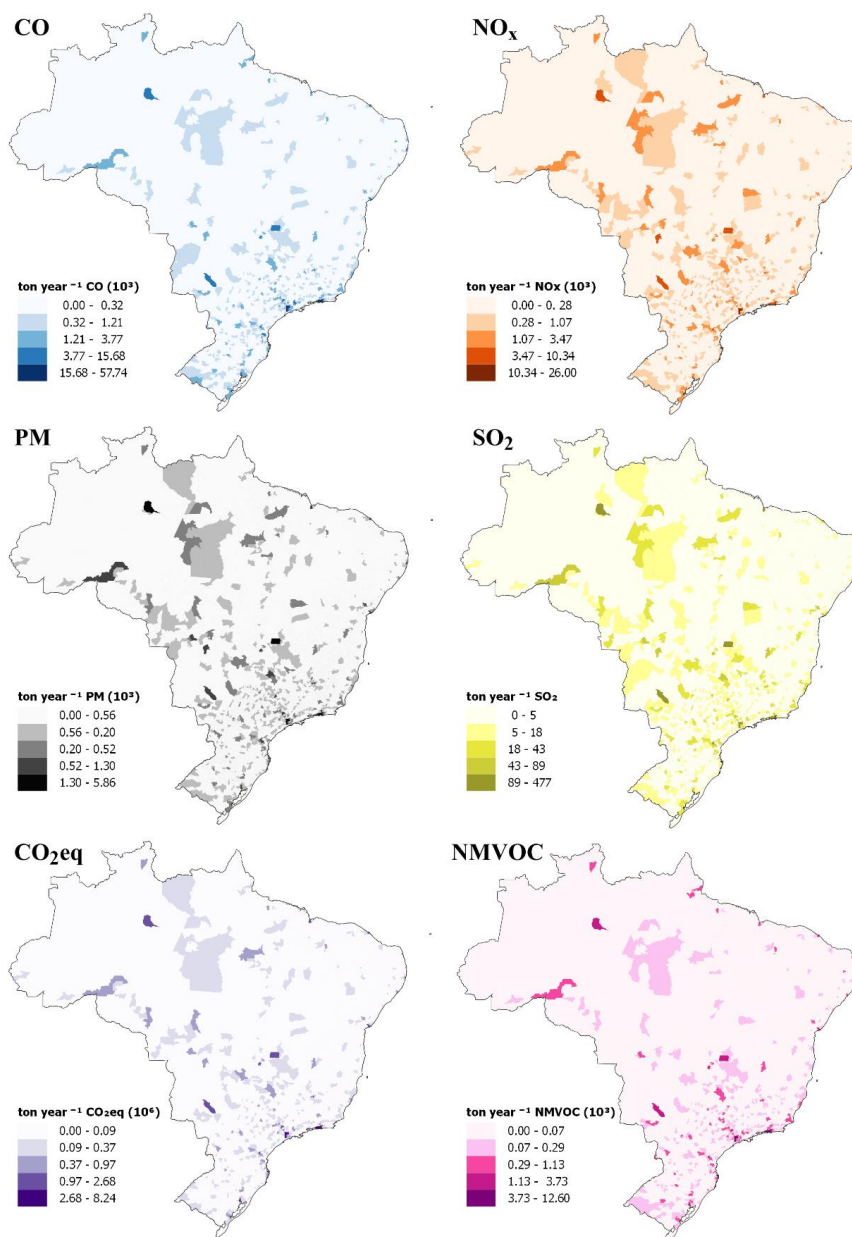


77 **2. Vehicular emissions data**

78

79 Our database uses output data from the BRAZilian Vehicular Emissions inventory Software –
80 BRAVES (Vasques and Hoinaski, 2021), which employs a probabilistic bottom-up method to esti-
81 mate vehicular emissions aggregated by municipality (Figure 1). BRAVES estimates the vehicular
82 emissions from the exhaust, tires, roads, brake wear, soil resuspension, refueling, and evaporative
83 emissions. The software provides, by fleet category (i.e., commercial-light vehicles, motorcycles,
84 light-duty and heavy-duty vehicles), annual emissions of carbon monoxide (CO), carbon dioxide
85 (CO₂), methane (CH₄), hydrocarbons (HC), aldehydes (RCHO), non-methane volatile organic com-
86 pounds (NMVOC), nitrogen oxides (NO_x), particulate material (PM), nitrous oxide (N₂O), and sulfur
87 oxide (SO₂). Throughout this article we call current database the BRAVES database. Codes and out-
88 puts from BRAVES are available by registering at <https://hoinaski.prof.ufsc.br/BRAVES/> and
89 <https://github.com/leohoinaski/BRAVES>, where users can access instructions to run the database and
90 download the input files. The outputs are generated in netCDF format and with annual or hourly
91 resolutions.

92



93

94 **Figure 1.** Vehicular emissions inventory in 2019 of a) CO, b) NO_x, c) PM, d) SO₂, e) CO₂, and f)

95

NMVOC provided by BRAVES (Vasques and Hoinaski, 2021).

96



97 **3. Spatial disaggregation**

98

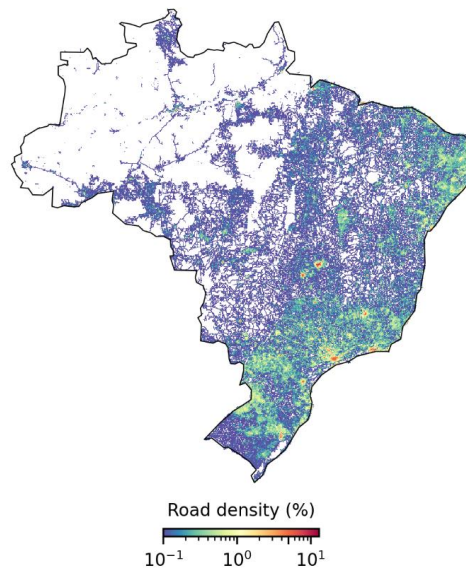
99 Since vehicular emissions from BRAVES are aggregated by municipality, we use a road den-
100 sity approach to distribute the emissions of each municipality in pixels with user-defined resolution.
101 Previous work of Tuia et al. (2007) and Gomez et al. (2018) shows that road density is one of the
102 most reliable approaches to disaggregate vehicular emissions. In this article, road density factor
103 ($RD_{p,m}$) is calculated by the sum of road lengths (L_p) on each pixel (p) divided by the total road
104 length (L_m) inside a municipality (m) (Equation 1). Road shapefile data derived from OpenStreetMap
105 (www.openstreetmap.org) can be downloaded at (<https://download.geofabrik.de/south-america/brazil.html>)
106 for Brazilian territory. Figure 2 shows the spatial distribution of $RD_{p,m}$ in Brazil. Multiplying
107 $RD_{p,m}$ by the vehicular emissions in each municipality derived from BRAVES provides the spatial-
108 ized emission ($E_{p,m,c}$) of compound c in pixel p within a municipality m (Equation 2). We provide a
109 parallelized method to estimate the road density in Brazil at <https://github.com/leohoinaski/BRAVES>.
110

111 $RD_{p,m} = \frac{L_p}{L_m}$ (Eq.1)

112

113 $E_{c,p,m} = RD_{p,m} \times E_{c,m}$ (Eq.2)

114



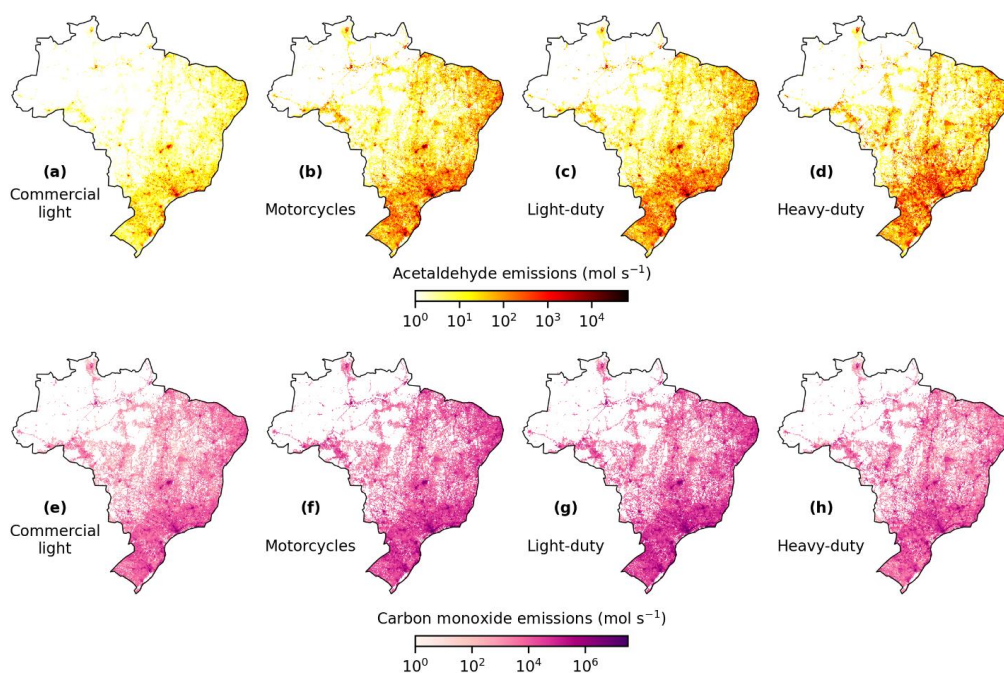
115

116

Figure 2. Road density factor in Brazil.

117

118 Figure 3 shows the annual emissions of aldehydes and CO from 2019 in Brazil by fleet cate-
119 gory using spatial distribution based on the road density approach. Hotspots of vehicular emissions
120 are concentrated in urbanized areas in Brazil (Figure 3). Among fleet categories, heavy-duty is the
121 major emitter of aldehydes, while light-duty emits the most CO. Vasques and Hoinaski (2021) pre-
122 sents a full comparison of vehicular emissions from each fleet category in Brazil using BRAVES.
123 Figures SM1 to SM5 demonstrate the BRAVES database by Brazilian state.



124

125 **Figure 3.** Vehicular emissions of aldehydes a-d) and carbon monoxide e-h) from commercial light,
126 motorcyle, light, and heavy-duty fleets in 2019 in Brazil.

127

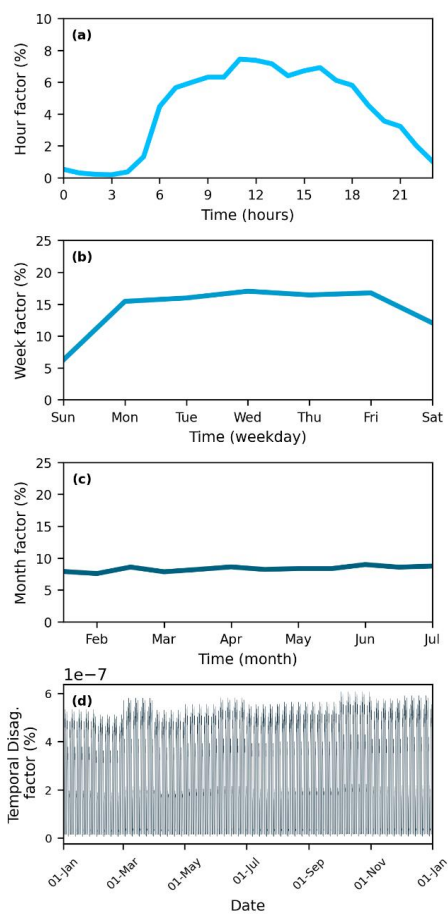
128 4. Temporal disaggregation

129

130 Temporal disaggregation based on traffic flow observations from Environment and Water Re-
131 sources Institute from Espírito Santo state (IEMA ES, 2019) splits original annual emissions from
132 BRAVES into hourly basis emissions. In this article, the temporal disaggregation factor (Figure 4) is
133 composed of hourly, weekly, and monthly traffic factors. Hourly emissions ($E_{c,p,m,h}$) of each air pol-
134 lutant are obtained by the multiplication of $E_{p,m,c}$ and the temporal disaggregation factor (T_f) (Equa-
135 tion 3).

$$136 E_{c,p,m,h} = T_f \times E_{c,p,m} \quad (\text{Eq.3})$$

137



138

139 **Figure 4.** Temporal disaggregation factor and its components a) Hourly factor, b) Weekly factor, c)

140 Monthly factor, d) Annual to hourly basis temporal disaggregation factor.



141 **5. Chemical speciation**

142

143 We use data from SPECIATE 5.1 (US EPA, 2020; Eyth et al. 2020) from the United States
144 Environmental Protection Agency - US EPA (<https://www.epa.gov/air-emissions-modeling/speciate>)
145 to speciate emissions of chemical constituents of Volatile Organic Compounds (VOC) and PM, which
146 has not been previously estimated by BRAVES. The chemical speciation method also converts NO_x
147 to NO and NO₂.

148 Regarding the speciation procedures, in this article we group light duty and commercial light
149 vehicles, and motorcycles together as light vehicles. We select profiles to speciate PM emissions from
150 the exhaust of heavy and light vehicles, soil resuspension (road dust), tire wear, and brake wear. VOC
151 emissions from the exhaust and evaporative process are also speciated. Table SM6 in the supplemen-
152 tary material summarizes the profiles from SPECIATE 5.1 used in the chemical speciation. We target
153 the species required for Carbon Bond Chemical Mechanism (Yarwood et al. 2010ab) version 6
154 (CB06), which describe tropospheric oxidant chemistry in a concise manner suitable for use in com-
155 plex 3-dimensional atmospheric models (Yarwood et al. 2010ab). Table 1 presents the VOC and PM
156 compounds considered in the chemical speciation in this database.

157

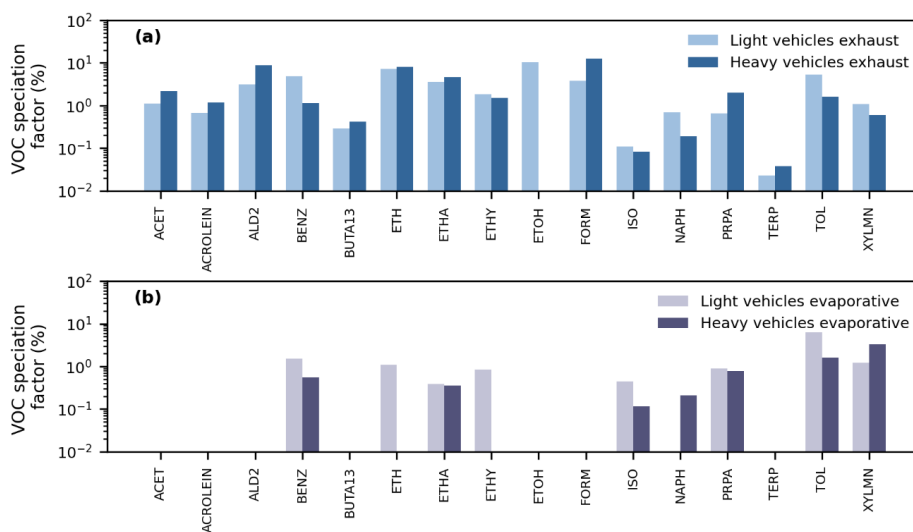


158 Table 1. List of chemical species included in the BRAVES database (Yarwood et al. 2010b).

Specie	ID	Specie	ID
Acetone	ACET	Aluminum	PAL
total Acrolein	ACROLEIN	Calcium ion	PCA
total Acetaldehyde	ALD2	Chloride ion	PCL
Benzene	BENZ	Elemental carbon	PEC
1,3-Butadiene	BUTA13	Iron	PFE
Ethanol	ETH	Potassium ion	PK
Ethane	ETHA	Magnesium ion	PMG
Ethyne	ETHY	Manganese	PMN
Formaldehyde	FORM	Sodium ion	PNA
Isoprene	ISO	Ammonium	PNH4
Naphthalene	NAPH	Nitrate	PNO3
Propane	PRPA	Silicon	PSI
Monoterpenes	TERP	Sulfate	PSO4
Toluene	TOL	Titanium	PTI
xylene	XYLMN		

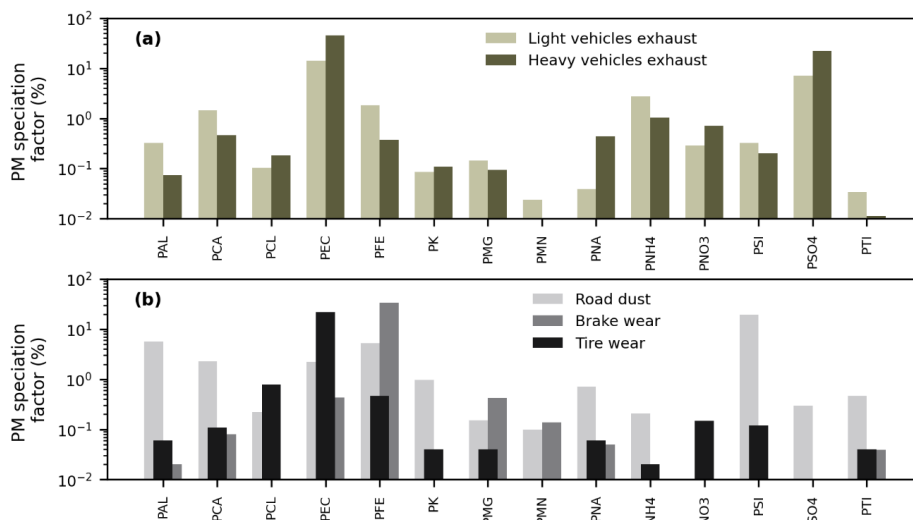
159

160 The chemical speciation factors employed to split VOC and PM emissions are calculated by
161 the average of the weighting percentage of the corresponding species from SPECIATE 5.1. We con-
162 sider exhaust, evaporative, and particulate emissions of light and heavy vehicles. Figures 5 and 6
163 show the speciation factor used to generate the database. Multiplication factors of 0.495 and 0.505
164 derived from SPECIATE 5.1 convert NO_x emissions to NO and NO₂, respectively. Table SM7 sum-
165 marizes the speciation factors used to build this database.



166

167 **Figure 5.** VOC chemical speciation factor for exhaust and evaporative emissions from light and
 168 heavy vehicles used in this article.



169

170 **Figure 6.** PM chemical speciation factors a) exhaust emissions from light and heavy vehicles b)
 171 road dust resuspension, brake wear, and tire wear.

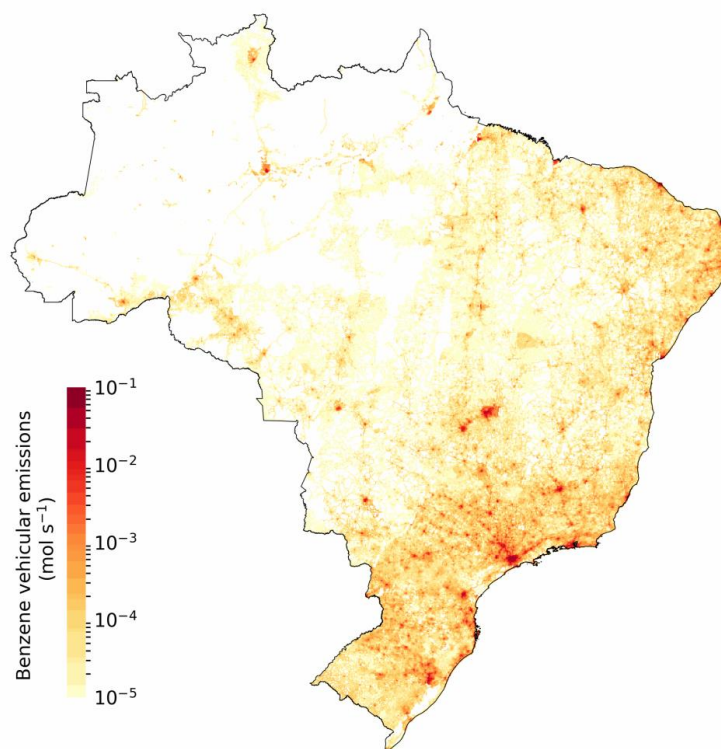
172



173 **6. The database and codes**

174

175 The database contains hourly emissions of 41 chemical species, such as ACET, ACROLEIN,
176 ALD2, BENZ, BUTADIENE13, CH₄, CO, CO₂, ETH, ETHA, ETHY, ETOH, FORM, ISO, N₂O,
177 NAPH, NO, NO₂, PAL, PCA, PCL, PEC, PFE, PK, coarse mode primary PM (PMC), PMG, PMN,
178 unspciated PM_{2.5} (PMOTHR), PNA, PNH₄, PNO₃, POC, PRPA, PSI, PSO₄, PTI, SO₂, TERP, TOL,
179 VOC, and XYLMN. We provide a code to generate hourly resolved files with a user-defined grid for
180 a single or whole group of species (<https://github.com/leohoinaski/BRAVES>). These files are com-
181 patible with the input requirements of sophisticated chemical transport models, such as the Commu-
182 nity Multiscale Air Quality Model (CMAQ), the Weather Research and Forecasting (WRF) model
183 coupled with Chemistry (WRF-Chem), the Comprehensive Air Quality Model with Extensions
184 (CMAx), and others. Smaller domains and finer resolution can be easily created by modifying in the
185 python codes. Figure 7 shows the vehicular emissions of Benzene in Brazil on January 1st, 2019 using
186 the BRAVES database.



January 01, 2019 00:00:00 UTC

187

188 **Figure 7.** Vehicular emissions of Benzene in Brazil on January 1st, 2019 using the BRAVES data-
189 base.

190 We provide the BRAVES database annual speciated emissions with 0.05°x0.05° of resolution
191 covering the entire Brazilian territory at: <https://doi.org/10.5281/zenodo.6141109>

192

193

194

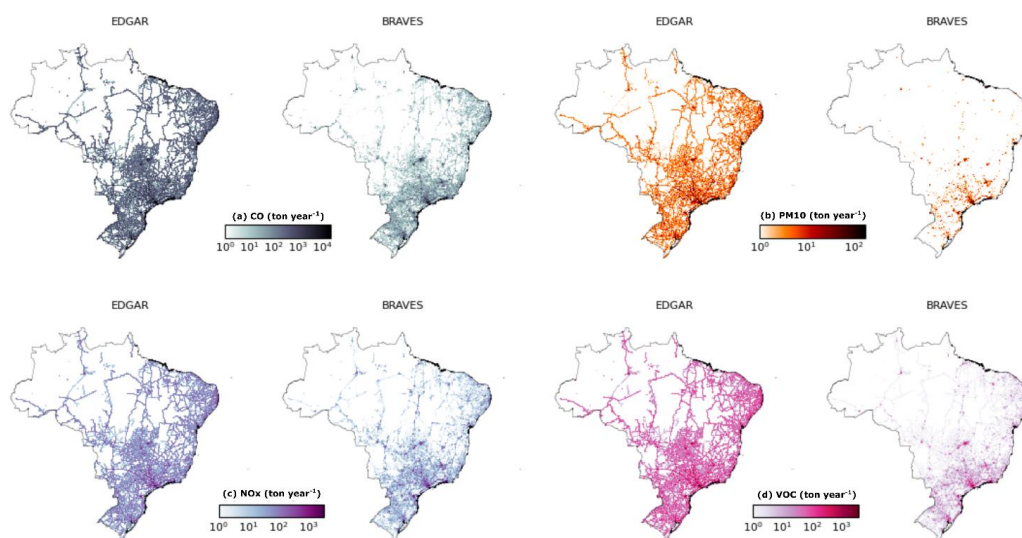
195



196 7. Comparison with independent databases

197

198 We analyze the spatial correlation and bias between the BRAVES database and the Emissions
199 Database for Global Atmospheric Research (EDGAR – version 5.0 https://edgar.jrc.ec.europa.eu/dataset_ap50) annual gridmaps (Crippa et al. 2018; European Commission 2022). We performed the com-
200 parison using the “Road Transportation” emissions from EDGAR for the Brazilian territory, includ-
201 ing soil resuspension emission rates of PM₁₀ from EDGAR. The BRAVES database emission rates
202 in tons per year from 2015 were regridded to the same spatial resolution of EDGAR. The Spearman
203 coefficient estimates the spatial correlation, while the difference in absolute emissions calculates the
204 bias between the datasets. We compare the disaggregated emissions of CO, PM₁₀, NO_x, and COV
205 from BRAVES and EDGAR.
206

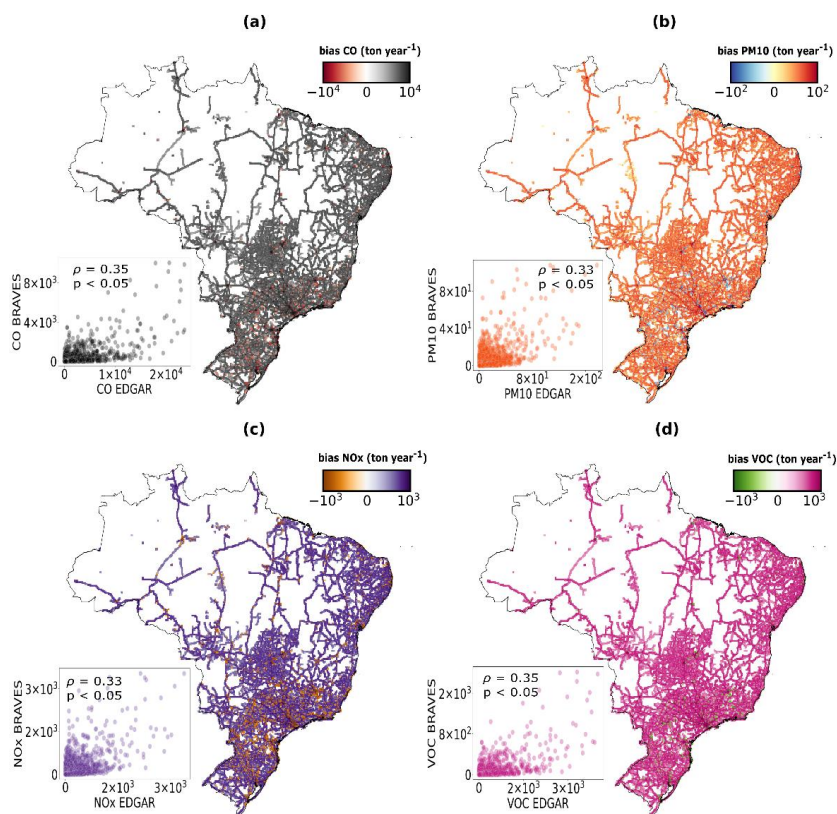


207

208 **Figure 8.** Comparison of a) CO, b) PM₁₀, c) NO_x, and d) VOC spatial distribution (log scale) pro-
209 vided by the EDGAR and BRAVES databases.



210



211

212 **Figure 9.** Bias (log scale) and scatter plots of a) CO, b) PM₁₀, c) NO_x, and d) VOC emission rates
213 provided by the BRAVES database and EDGAR in 2015.

214

215 Emission from BRAVES and EDGAR presents overall spatial correlation ($p < 0.05$) of $\rho = 0.35$
216 for CO, $\rho = 0.33$ for PM₁₀, $\rho = 0.33$ for NO_x, and $\rho = 0.35$ for VOC (Figure 9). Emissions from
217 EDGAR are consistently higher (Figure 9) than from BRAVES, as also reported by Vasques and
218 Hoinaski (2021). The largest differences are observed in CO emissions, followed by VOC, NO_x, and
219 PM₁₀. Madrazo et al. (2018) explains that most of the road transport Emission Factors are
220 overestimated in EDGAR, while Huneus et al. (2020) found discrepancies between EDGAR and



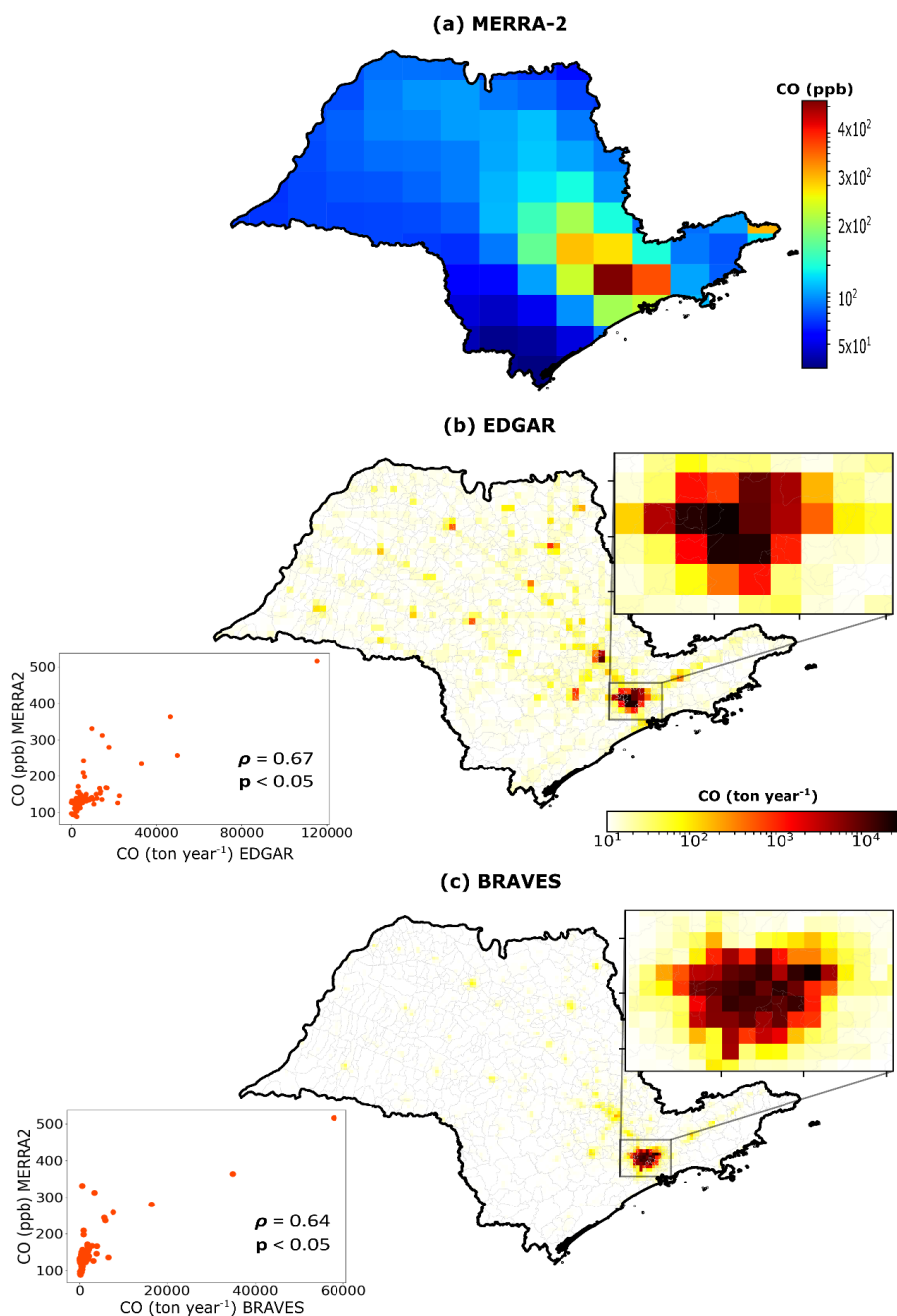
221 local/national city emissions. Álamos et al. (2022) also reported an overestimation of EDGAR
222 emissions.

223 We analyze the spatial correlation between vehicular emissions of CO from the BRAVES
224 database, EDGAR, and surface concentration of CO from the Modern-Era Retrospective Analysis for
225 Research and Applications - MERRA-2. The Global Modeling and Assimilation Office (GMAO),
226 managed by NASA, provides MERRA-2 reanalysis products in a spatial resolution of $0.625^{\circ} \times 0.5^{\circ}$,
227 covering from 1980 to the present (Gelaro et al. 2017; Randles et al., 2017). We calculate the annual
228 average concentration in 2015 from monthly data in netCDF files available at GES-DISC platform
229 (https://disc.gsfc.nasa.gov/datasets/M2TMNXCHM_5.12.4/summary). All grids are realigned to
230 match the MERRA-2 spatial resolution. We analyze the spatial correlation by Brazilian state since
231 the vehicular emission has more influence in urbanized ones. We assume that those cells which have
232 the vehicular emissions as the major source of air pollutant also have higher surface concentration.
233 However, this assumption has several limitations and should be carefully evaluated since it does not
234 account for the dispersion process and other source types (i.e., industrial, biomass burning, biogenic
235 sources).

236 Figure 10 shows the (a) CO concentrations from MERRA-2 in São Paulo (SP) state, (b) ve-
237 hicular emission of CO from EDGAR, and (c) vehicular emission of CO from the BRAVES database.
238 We highlight the state of SP since it has approximately 7 million vehicles, being considered the state
239 with the highest vehicular emission in Brazil (Vasques and Hoinaski, 2021). The BRAVES database
240 and EDGAR reach similar spatial correlation with MERRA-2 (Figure 10). However, the zoom-in
241 quadrant in the São Paulo metropolitan region reveals the greater level of details from the BRAVES
242 database compared to EDGAR. In addition, BRAVES has higher temporal resolution and chemical
243 speciated emissions. In other Brazilian states, such as Minas Gerais (MG) and Rio Grande do Sul
244 (RS), there is also a positive correlation between vehicle emissions and surface concentrations of CO



245 (Table SM8). It shows that both databases consistently capture the spatial variability of vehicular
246 emissions and the BRAVES database brings additional features for air quality studies in Brazil.



247

248 **Figure 10.** CO spatial distribution provided by (a) MERRA2, (b) EDGAR, and (c) BRAVES. Scatter
249 plots of CO vehicle emission and CO surface concentrations in SP.



250 **8. Data availability**

251 The BRAVES database is freely available at <https://doi.org/10.5281/zenodo.6141109>
252 (Hoinaski et al., 2022). We provide annual speciated emissions with 0.05°x0.05° of resolution cov-
253 ering the entire Brazilian territory. Codes to generate the database are available at:
254 <https://github.com/leohoinaski/BRAVES>. Using the annual files, users can derive hourly basis emis-
255 sions through the available codes.

256



257 **9. Conclusions**

258 Here, we introduce the BRAVES database, the first high-resolution and chemical speciated
259 database of vehicular emissions covering the entire Brazilian territory. The BRAVES database con-
260 tains emissions of 41 air pollutants, from annual to hourly basis temporal resolution and user-defined
261 spatial resolution. The attributes of this emission database are fully compatible with sophisticated air
262 quality models. Moreover, the emissions of multiple chemical species presented here provide essen-
263 tial information to understand important atmospheric chemistry processes in Brazil. We also provide
264 python scripts for users who want to create their custom gridded inventory.

265 Even though detailed emission inventories are required to control air pollution, vehicular
266 emissions are scarce in most developing countries. So far, Brazil has lacked a comprehensive and
267 easily accessible database of vehicular emissions, and, creating gridded inventories in South America
268 is urgently needed. This work contributes to overcoming this gap.

269 The spatial correlation analysis reveals that the BRAVES database agrees with the vehicular
270 emissions from EDGAR, even though EDGAR emissions are consistently higher than those
271 BRAVES ones. We conclude that this database can be a better alternative to represent the spatial
272 variability of vehicular emissions in Brazil. The BRAVES database has a similar performance repre-
273 senting the spatial pattern of vehicular emissions, with more spatial, temporal, and chemical details
274 when compared with EDGAR. Moreover, the BRAVES database is in closer agreement to local and
275 very detailed emissions inventories. A comparison with MERRA-2 surface concentration confirms
276 the consistency of the BRAVES database.

277 Even though the present database is a step forward for air pollution research in Brazil, there
278 are several opportunities for expanding and improving this work. Most heavy-duty emissions occur
279 in high flow and high-speed limit roads, such as expressways. Future versions could improve the
280 spatial disaggregation in pixels containing roads with high traffic flow and/or high-speed traffic
281 through the optimization of the disaggregation factors. Different criteria for light and heavy vehicles



282 would also be needed. Moreover, the chemical speciation could include profiles to consider the Bra-
283 zilian reality as biofuels, fleet motorization, and regionalized soil resuspension properties. Temporal
284 variability would also be improved by regionalizing the profiles to account for the traffic flow in each
285 location.



286 **Acknowledgments**

287

288 Authors would like to thank the Fundação de Amparo à Pesquisa e Inovação de Santa Catarina

289 - FAPESC, for financial support of project number 2018TR499.

290



291 **References**

292

293 Andrade MF, Kumar P, Freitas ED, Ynoue RY, Martins J, Martins LD, Nogueira T, Perez-Martinez
294 P, de Miranda RM, Albuquerque T, Gonçalves FLT, Oyama B, Zhang Y (2017). Air quality in the
295 megacity of São Paulo: evolution over the last 30 years and future perspectives. Atmospheric Envi-
296 ronment 159:66–82. <https://doi.org/10.1016/j.atmosenv.2017.03.051>

297

298 Álamos N, Huneeus N, Opazo M, Osses M, Puja S, Pantoja N, Denier van der Gon H, Schueftan A,
299 Reyes R, Calvo R (2022). High-resolution inventory of atmospheric emissions from transport, indus-
300 trial, energy, mining and residential activities in Chile. Earth System Science Data 14:361–379.
301 <https://doi.org/10.5194/essd-14-361-2022>

302

303 Anenberg S, Miller J, Minjares R, Du L, Henze DK, Lacey F, Malley CS, Emberson L, Franco V,
304 Klimont Z, Heyes C (2017). Impacts and mitigation of excess diesel-related NO_x emissions in 11
305 major vehicle markets. Nature 545:467–471. <https://doi.org/10.1038/nature22086>

306

307 Brito J, Carbone S, Santos DAM, Dominutti P, Alves NO, Rizzo LV, Artaxo P (2018). Disentangling
308 vehicular emission impact on urban air pollution using ethanol as a tracer. Scientific Reports 8:10679.
309 <https://doi.org/10.1038/s41598-018-29138-7>

310

311 Carvalho VSB, Freitas ED, Martins LD, Martins JA, Mazzoli CR, Andrade MF (2015). Air quality
312 status and trends over the Metropolitan Area of São Paulo, Brazil as a result of emission control
313 policies. Environmental Science & Policy 47:68–79. <https://doi.org/10.1016/j.envsci.2014.11.001>

314

315



316 Crippa M, Guizzardi D, Muntean M, Schaaf E, Dentener F, Van Aardenne JA, Monni S, Doering U,
317 Olivier JGJ, Pagliari V, Janssens-Maenhout G (2018). Gridded emissions of air pollutants for the
318 period 1970–2012 within EDGAR v4.3.2. *Earth System Science Data* 10:1987–2013.
319 <https://doi.org/10.5194/essd-10-1987-2018>
320
321 European Commission, Joint Research Centre (JRC). Emission Database for Global Atmospheric
322 Research (EDGAR), v5.0. Accessed January 2022. Available online: [https://edgar.jrc.ec.eu-](https://edgar.jrc.ec.eu-ropa.eu/index.php/dataset_ap50)
323 [ropa.eu/index.php/dataset_ap50](https://doi.org/10.2904/JRC_DATASET_EDGAR), https://doi.org/10.2904/JRC_DATASET_EDGAR
324
325 Eyth A, Strum M, Murphy B, Shah T, Shi Y, Beardsley R, Yarwood G (2020). Speciation Tool User’s
326 Guide Version 5.0. Available online: [https://www.cmascenter.org/speciation_tool/documenta-](https://www.cmascenter.org/speciation_tool/documentation/5.0/Ramboll_sptool_users_guideV5.pdf)
327 [tion/5.0/Ramboll_sptool_users_guideV5.pdf](https://www.cmascenter.org/speciation_tool/documentation/5.0/Ramboll_sptool_users_guideV5.pdf).
328
329 Gelaro R, McCarty W, Suárez MJ, Todling R, Molod A, Takacs L, Randles CA, Darmenov A, Bos-
330 ilovich MG, Reichle R, Wargan K, Coy L, Cullather R, Draper C, Akella S, Buchard V, Conaty A,
331 Silva AM, Gu W, Kim G, Koster R, Lucchesi R, Merkova D, Nielsen JE, Partyka G, Pawson S,
332 Putman W, Rienecker M, Schubert SD, Sienkiewicz M, Zhao B (2017). The Modern-Era Retrospec-
333 tive Analysis for Research and Applications, Version 2 (MERRA-2). *Journal of Climate* 30(14):5419-
334 5454. <https://doi.org/10.1175/JCLI-D-16-0758.1>
335
336 Gómez CD, González CM, Osses M, Aristizábal BH (2018). Spatial and temporal disaggregation of
337 the on-road vehicle emission inventory in a medium-sized Andean city. Comparison of GIS-based
338 top-down methodologies. *Atmospheric Environment* 179:142–155. [https://doi.org/10.1016/j.at-](https://doi.org/10.1016/j.atmosenv.2018.01.049)
339 [mosenv.2018.01.049](https://doi.org/10.1016/j.atmosenv.2018.01.049)
340



- 341 Hoinaski L, Vasques T V, Ribeiro C B, Meotti B. (2022). BRAVES database Version 1: vehicular
342 emissions in Brazil. <https://doi.org/10.5281/zenodo.6141109>
343
- 344 Huneus N, Van der Gon HD, Castesana P, Menares C, Granier C, Granier L, Alonso M, Andrade
345 MF, Dawidowski L, Gallardo L, Gomez D, Klimont Z, Janssens-Maenhout G, Osses M, Puliafito SE,
346 Rojas N, Sánchez-Ccoyllo O, Tolvett S, Ynoue RY (2020). Evaluation of anthropogenic air pollutant
347 emission inventories for South America at national and city scale. Atmospheric Environment
348 235:117606. <https://doi.org/10.1016/j.atmosenv.2020.117606>
349
- 350 Ibarra-Espinosa S, Ynoue R, O'Sullivan S, Pebesma E, Andrade MDF, Osses M (2018). VEIN v0.2.2:
351 an R package for bottom-up vehicular emissions inventories. Geoscientific Model Development
352 11:2209–2229. <https://doi.org/10.5194/gmd-11-2209-2018>
353
- 354 Instituto de Meio Ambiente e Recursos Hídricos do Espírito Santo (IEMA-ES) (2019). Inventário de
355 Emissões Atmosféricas Região da Grande Vitória – Ano Base 2015 (in Portuguese). Available online:
356 <https://iema.es.gov.br/qualidadedoar/inventariodefuentes/2015>
357
- 358 Krzyzanowski M, Kuna-Dibbert Birgit, Schneider J (2005). Health effects of transport-related air
359 pollution. World Health Organization. Regional Office for Eu-
360 rope. [https://www.euro.who.int/en/publications/abstracts/health-effects-of-transport-related-air-pol-
361 lution](https://www.euro.who.int/en/publications/abstracts/health-effects-of-transport-related-air-pollution)
362
- 363 Lyu M, Bao X, Zhu R, Matthews R (2020). State-of-the-art outlook for light-duty vehicle emission
364 control standards and technologies in China. Clean Technologies and Environmental Policy
365 22(4):757–771. <https://doi.org/10.1007/s10098-020-01834-x>



366

367 Madrazo J, Clappier A, Belalcazar LC, Cuesta O, Contreras H, Golay F (2018). Screening differences
368 between a local inventory and the Emissions Database for Global Atmospheric Research (EDGAR).
369 Science of The Total Environment 631–632:934-941. <https://doi.org/10.1016/j.scitotenv.2018.03.094>
370

371 Randles CA, Silva AM, Buchard V, Colarco PR, Darmenov A, Govindaraju R, Smirnov A, Holben
372 B, Ferrare R, Hair J, Shinozuka Y, Flynn CJ (2017). The MERRA-2 Aerosol Reanalysis, 1980 On-
373 ward. Part I: System Description and Data Assimilation Evaluation. Journal of Climate 30(17):6823-
374 6850. <https://doi.org/10.1175/JCLI-D-16-0609.1>

375

376 Ribeiro CB, Rodella FHC, Hoinaski L (2021). Regulating light-duty vehicle emissions: an overview
377 of US, EU, China and Brazil programs and its effect on air quality. Clean Technologies and Environ-
378 mental Policy. <https://doi.org/10.1007/s10098-021-02238-1>

379

380 Sawyer R (2010). Vehicle emissions: progress and challenges. Journal of Exposure Science & Envi-
381 ronmental Epidemiology 20:487–488. <https://doi.org/10.1038/jes.2010.44>

382

383 Shindell D, Faluvegi G, Walsh M, Anenberg SC, Van Dingenen R, Muller NZ, Austin J, Koch D,
384 Milly G (2011). Climate, health, agricultural and economic impacts of tighter vehicle-emission stand-
385 ards. Nature Climate Change 1:59–66. <https://doi.org/10.1038/nclimate1066>

386

387 Sun S, Jin J, Xia M, Liu Y, Gao M, Zou C, Wang T, Lin Y, Wu L, Mao H, Wang P (2020). Vehicle
388 emissions in a middle-sized city of China: current status and future trends. Environmental Internati-
389 onal 137:105514. <https://doi.org/10.1016/j.envint.2020.105514>

390



391 Tuia D, Ossés de Eicker M, Zah R, Osses M, Zarate E, Clappier A (2007). Evaluation of a simplified
392 top-down model for the spatial assessment of hot traffic emissions in mid-sized cities. Atmospheric
393 Environment. 41(17):3658–3671. <https://doi.org/10.1016/j.atmosenv.2006.12.045>
394
395 Unger N, Bond TC, Wang JS, Koch DM, Menon S, Shindell DT, Bauer S (2010) Attribution of cli-
396 mate forcing to economic sectors. Proceedings of the National Academy of Sciences 107(8):3382-
397 3387. <https://doi.org/10.1073/pnas.0906548107>
398
399 United States Environmental Protection Agency (US EPA) (2020). Addendum Speciate Version 5.1
400 Database Development Documentation. Available online: [https://www.epa.gov/air-emissions-mod-](https://www.epa.gov/air-emissions-modeling/speciate-51-and-50-addendum-and-final-report)
401 [eling/speciate-51-and-50-addendum-and-final-report](https://www.epa.gov/air-emissions-modeling/speciate-51-and-50-addendum-and-final-report)
402
403 Vasques TV, Hoinaski L (2021). Brazilian vehicular emission inventory software – BRAVES. Trans-
404 portation Research Part D: Transport and Environment 100:103041.
405 <https://doi.org/10.1016/j.trd.2021.103041>
406
407 Yarwood G, Jung J, Whitten GZ, Heo G, Mellberg J, Estes M (2010a). Updates to the Carbon Bond
408 Mechanism for version 6 (CB6). 9th CMAS Conference, Chapel hill, NC. Available online:
409 https://www.cmascenter.org/conference/2010/abstracts/emery_updates_carbon_2010.pdf
410
411 Yarwood G, Whitten GZ, Jung J (2010b). Development, Evaluation and Testing of Version 6 of the
412 Carbon Bond Chemical Mechanism (CB6). Texas Commission on Environmental Quality. Available
413 online:[https://www.tceq.texas.gov/assets/public/implementation/air/am/contracts/re-](https://www.tceq.texas.gov/assets/public/implementation/air/am/contracts/reports/pm/5820784005FY1026-20100922-environ-cb6.pdf)
414 [ports/pm/5820784005FY1026-20100922-environ-cb6.pdf](https://www.tceq.texas.gov/assets/public/implementation/air/am/contracts/reports/pm/5820784005FY1026-20100922-environ-cb6.pdf)