



1 **HTAP_v3 emission mosaic: a global effort to tackle air quality**
2 **issues by quantifying global anthropogenic air pollutant**
3 **sources**

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31 **Abstract.** This study, performed under the umbrella of the Task Force on Hemispheric Transport of Air Pollution
32 (TF-HTAP), responds to the need of the global and regional atmospheric modelling community of having a mosaic
33 emission inventory of air pollutants that conforms to specific requirements: global coverage, long time series,
34 spatially distributed emissions with high time resolution, and a high sectoral resolution. The mosaic approach of
35 integrating official regional emission inventories based on locally reported data, with a global inventory based on
36 a globally consistent methodology, allows modellers to perform simulations of a high scientific quality while also
37 ensuring that the results remain relevant to policymakers.

38 HTAP_v3, an ad-hoc global mosaic of anthropogenic inventories, has been developed by integrating official
39 inventories over specific areas (North America, Europe, Asia including Japan and Korea) with the independent
40 Emissions Database for Global Atmospheric Research (EDGAR) inventory for the remaining world regions. The
41 results are spatially and temporally distributed emissions of SO₂, NO_x, CO, NMVOC, NH₃, PM₁₀, PM_{2.5}, Black
42 Carbon (BC), and Organic Carbon (OC), with a spatial resolution of 0.1 x 0.1 degree and time intervals of months



1 and years covering the period 2000-2018 (DOI 10.5281/zenodo.7516361,
2 https://edgar.jrc.ec.europa.eu/dataset_htap_v3). The emissions are further disaggregated to 16 anthropogenic
3 emitting sectors. This paper describes the methodology applied to develop such an emission mosaic, reports on
4 source allocation, differences among existing inventories, and best practices for the mosaic compilation. One of
5 the key strengths of the HTAP_v3 emission mosaic is its temporal coverage, enabling the analysis of emission
6 trends over the past two decades. The development of a global emission mosaic over such long time series
7 represents a unique product for global air quality modelling and for better-informed policy making, reflecting the
8 community effort expended by the TF-HTAP to disentangle the complexity of transboundary transport of air
9 pollution.

10 1 Introduction

11 Common international efforts have procured an agreement to reduce global air pollutant emissions. For this
12 purpose, the United Nations Economic Commission for Europe (UNECE) Convention on Long Range
13 Transboundary Air Pollution (CLRTAP) and the Task Force on Hemispheric Transport of Air Pollution (TF-
14 HTAP) have been instrumental in developing the understanding of intercontinental transport of air pollution and
15 thus contributing to the reduction of key pollutants in Europe and North America.

16 The success of CLRTAP is based on meeting strict reduction targets for pollutant releases. Therefore, evaluating
17 the resulting implications of these reductions requires an ongoing improvement of global emission inventories in
18 terms of emission updating and of methodological refinements. These aspects are instrumental to gain
19 understanding of transboundary air pollution processes and drivers and to measure the effectiveness of emissions
20 reduction and air quality mitigation policies. New guidance is available to achieve further emission reductions
21 across all emitting sectors. For example, the 2019 establishment of the Task Force for International Cooperation
22 on Air Pollution, which is intended to promote international collaboration for preventing and reducing air pollution
23 and improving air quality globally (UNECE, 2021). As part of the ongoing effort by CLRTAP to reduce emissions
24 and to set out more effective and accountable mitigation measures, the 2005 Gothenburg Protocol (UNECE, 2012)
25 has been revised, including the review of the obligations in relation to emission reductions and mitigation
26 measures (e.g., black carbon and ammonia) and the review of the progress towards achieving the environmental
27 and health objectives of the Protocol.

28
29 The Task Force on Hemispheric Transport of Air Pollution (TF-HTAP) of the Convention has a mandate to
30 promote the scientific understanding of the intercontinental transport of air pollution to and from the UNECE area
31 (<https://unece.org/geographical-scope>), to quantify its impacts on human health, vegetation and climate, and to
32 identify emission mitigation options that will shape future global policies.

33 This paper describes and discusses a consistent global emission inventory of air pollutants emitted by
34 anthropogenic activities. This important tool has been developed to assess the contribution of anthropogenic air
35 pollution emission sources within and outside the UNECE-area through atmospheric modelling. This inventory
36 has been compiled based on officially reported emissions, and an independent global inventory where officially
37 reported emissions are not used. This harmonised emissions “mosaic” dataset, hereafter referred to as the
38 HTAP_v3, contains annual and monthly:

- 39 - emission time series (from 2000 to 2018) of SO₂, NO_x, CO, NMVOC, NH₃, PM₁₀, PM_{2.5}, BC, OC by
40 emitting sector and country, and
- 41 - spatially distributed emissions on a global grid with spatial spacing of 0.1x0.1 degree.

42 A global emission mosaic that gathers officially reported air pollutant emissions is essential since it aggregates
43 the officially available information on air pollution as input for atmospheric modellers and policy evaluations.
44 The creation of a global emission mosaic requires the harmonisation of several data sources, detailed analysis of
45 contributing sectors for the different input inventories, development of data quality control procedures, and a
46 robust and consistent gap-filling methodology when lacking information. The development of HTAP_v3 builds
47 upon the previous experience of the HTAPv1 (Janssens-Maenhout et al., 2012) and HTAPv2.2 (Janssens-
48 Maenhout et al., 2015) global inventories. HTAP_v3, as requested by the TF-HTAP modelling community,
49 provides a more refined sectoral disaggregation compared to the previous HTAP emission mosaics. It also
50 includes tools (https://edgar.jrc.ec.europa.eu/htap_tool/) that allows the extraction of emission data over selected
51 domains (detailed later in section 4).

52 This paper describes the development the HTAP_v3 database as a global anthropogenic air pollutant emissions
53 inventory mosaic for the period 2000-2018. The HTAP_v3 mosaic has been composed by integrating official,
54 spatially distributed emissions data from CAMS-REF-v5.1 (Kuenen et al., 2022), US EPA (U.S. Environmental
55 Protection Agency, 2021a, b), Environment and Climate Change Canada (ECCC) (NPRI, 2017), REAS, CAPSS-
56 KU, and JAPAN (<https://www.env.go.jp/air/osen/pm/info.html>) (Kurokawa and Ohara, 2020; Chatani et al.,



1 2018; Chatani et al., 2020) inventories. As the information gathered from the official reporting covers only part
2 of the globe, HTAP_v3 has been completed using emissions from the Emissions Database for Global Atmospheric
3 Research (EDGAR) version 6.1 (https://edgar.jrc.ec.europa.eu/dataset_ap61).

4 One of the key strengths of the HTAP_v3 emission mosaic is the wide temporal coverage of the emissions,
5 spanning the 2000-2018 period, enabling the analysis of emission trends over the past two decades. The
6 development of a global emission mosaic over such long time series represents a unique product for air quality
7 modelling and for better-informed policy making, reflecting the effort of the TF-HTAP community to improve
8 understanding of the transboundary transport of air pollution. The year 2000 was chosen as the baseline year since
9 it often represents the year from which complete datasets of annual air pollutant emissions can be generated. It
10 also represents a turning point for several emerging economies (e.g., China) and the strengthening of mitigation
11 measures in historically developed regions (e.g., EU, USA, etc.).

12 The two previous HTAP emission mosaics had limited temporal coverage. HTAPv1 covered the period 2000-
13 20005 with annual resolution (https://edgar.jrc.ec.europa.eu/dataset_htap_v1, (Janssens-Maenhout et al., 2012)),
14 while HTAPv2.2 covered two recent years (2008 and 2010), but with monthly resolution (Janssens-Maenhout et
15 al., 2015) (https://edgar.jrc.ec.europa.eu/dataset_htap_v2). However, the needs of the TF-HTAP modelling
16 community are continuously evolving to both foster forward-looking air quality science and produce more fit-for-
17 purpose analyses in support of efficient policy making. HTAP_v3 therefore not only covers the time period of the
18 previous HTAP phases, but also extends it forward by almost a decade, to provide the most up-to-date picture of
19 global air pollutant emission trends. Another distinguishing feature of the HTAPv3 mosaic is a considerably
20 higher sectoral resolution than previous iterations of the HTAP mosaic inventories (section 2.2), enabling more
21 policy-relevant use of the inventory.

22 The methodology and data sources for the HTAP_v3 emission mosaic are described in section 2. The long-time
23 coverage of two decades, allows comprehensive trend analysis (see section 3), the HTAP_v3 data format and
24 data-set access are presented in section 4 and conclusions are provided in section 5.

25

26 **2 HTAP_v3 emission mosaic overview: data sources, coverage, and methodology**

27 **2.1 Data input**

28 The HTAP_v3 mosaic is a database of monthly- and sector-specific global air pollutant emission gridmaps
29 developed by integrating spatially explicit regional information from recent officially-reported national or regional
30 emission inventories. Data from six main regional inventories were integrated into HTAP_v3, which covered only
31 North America, Europe, and a portion of Asia (including Japan, China, India, and South Korea) (Fig.1). The
32 geographical domain covered by each of these inventories is depicted in Fig. 1, while further details on each
33 contributing inventory are presented in section 2.3. The emissions for all other countries, international shipping
34 and aviation (international and domestic) have been retrieved from the Emissions Database for Global
35 Atmospheric Research (EDGARv6.1, https://edgar.jrc.ec.europa.eu/dataset_ap61) as represented by the grey
36 areas in Fig.1. Depending on the pollutant, more than half of global emissions are provided by region-specific
37 inventories, while the remaining contribution is derived from the EDGAR global inventory as reported in the bar
38 graph of Fig.1, where the share of each individual inventory to global emissions is represented. For all pollutants,
39 the Asian domain is contributing most to global emissions, hence the importance of having accurate emission
40 inventories for this region.

41 Table 1 provides an overview of all data providers, in terms of geographical and temporal coverage, data format,
42 and sectoral and pollutant data availability. Table 2 defines the HTAP_v3 sectors and corresponding IPCC codes.
43 Table 3 further details the sector-pollutant data availability for each inventory and the gap-filling approach
44 required for some sectors and pollutants.

45 **2.2 Pollutant, spatial, temporal and sectoral coverage**

46 The HTAP_v3 emission mosaic helps to address the transboundary role of air pollutants by providing a key input
47 for atmospheric modellers and supporting the evaluation of environmental impact analyses for poor air quality.
48 For this reason, HTAP_v3 provides global 0.1 x 0.1 degree emission gridmaps for all air pollutants and specifically
49 for acidifying and eutrophying gases (such as SO₂, NH₃, NO_x), ozone precursors (NMVOC, CO, NO_x), and
50 primary particulate matter (PM₁₀, PM_{2.5}, BC, OC).



1 Emissions from each officially-reported inventory were submitted to HTAP on 0.1 x 0.1 degree regional gridmaps.
2 Spatial allocation was performed to these gridmaps for each sector by each inventory group using the best
3 available set of subsector spatial surrogate fields used by each group (e.g., <https://www.cmascenter.org/sa-tools>).
4 EDGARv6.1 global gridmaps are also on a 0.1 x 0.1 degree grid.

5 Compared to the two previous HTAP emission mosaics, HTAP_v3 input emission gridmaps were provided with
6 monthly time distributions to better reflect the regional seasonality of sector specific emissions (e.g., household,
7 power generation, and agricultural activities). Information on emission peaks over certain months of the year is
8 also a useful information for the development of territorial policies to mitigate localised emission sources in space
9 and time (e.g., emissions from residential heating over winter months, agricultural residue burning, etc.).

10 The HTAP_v3 mosaic provides emissions for gaseous and particulate matter air pollutants arising from all
11 anthropogenic emitting sectors except for wildfires and savannah burning, which represent major sources of
12 particulate matter and CO emissions. Wildfires and savannah burning are not included in the current mosaic since
13 community efforts are ongoing to tackle these sources specifically. Modellers can find these additional sources
14 on several publicly available global wildfire emission datasets compiled based on the best available scientific
15 knowledge, such as the Global Fire Emission Database (GFED, <https://www.globalfiredata.org/>) or the Global
16 Wildfire Information System (GWIS, <https://gwis.jrc.ec.europa.eu/>). When using satellite retrieved emissions
17 from fires, they should be treated with caution to avoid double counting the emissions released by e.g. agricultural
18 crop residue burning activities.

19 HTAP_v3 provides emissions at higher sectoral disaggregation than previous HTAP experiments¹ to better
20 understand drivers of emission trends and the effectiveness of sector-specific policy implementation. Emissions
21 from 16 sectors are provided by the HTAP_v3 mosaic, namely: International Shipping; Domestic Shipping;
22 Domestic Aviation; International Aviation; Energy; Industry; Fugitives; Solvent Use; Road Transport; Brake and
23 Tyre Wear; Other Ground Transport; Residential; Waste; Agricultural Waste Burning; Livestock; and Agricultural
24 Crops. Further details on the sector definitions as well as their correspondence with the IPCC codes (IPCC, 1996,
25 2006) are provided in Table 2. The selection of the number of sectors was constrained by the sectoral
26 disaggregation of the input inventories (see Table S1). Table 3 provides the complete overview of the emission
27 data provided by each inventory group indicating the pollutants covered for each sector and eventual gap-filling
28 information included using the EDGARv6.1 data.

29 2.3 Inventory overviews

30 In the following sub-sections, details are provided on each officially-reported inventory used to construct the
31 HTAP_v3 emission mosaic.

32 2.3.1 CAMS-REG-v5.1 inventory

33 The CAMS-REG-v5.1 emission inventory was developed to support air pollutant and greenhouse gas modelling
34 activities at the European scale. The inventory builds largely on the official reported data to the UN Framework
35 Convention on Climate Change (UNFCCC) for greenhouse gases (for CO₂ and CH₄), and the Convention on
36 Long-Range Transboundary Air Pollution (CLRTAP) for air pollutants. For the latter, data are collected for NO_x,
37 SO₂, CO, NMVOC, NH₃, PM₁₀ and PM_{2.5}, including all major air pollutants. For each of these pollutants, the
38 emission data are collected at the sector level at which these are reported for the time series 2000-2018 for each
39 year and country. The CAMS-REG inventory covers UNECE-Europe, extending eastward until 60°E, therefore
40 including the European part of Russia. For some non-EU countries, the reported data are found to be partially
41 available or not available at all. In other cases, the quality of the reported data is found to be insufficient, i.e. with
42 important data gaps or following different formats or methods. In this case, emission data from the IIASA GAINS
43 model instead (IIASA, 2018) are used. This model is the main tool used to underpin pan-European and EU level
44 air quality policies such as the UNECE Convention on Long Range Transboundary Air Pollution (UNECE, 2012)
45 and the EU National Emission reduction Commitments Directive (European Commission, 2016).

¹HTAPv1 covered 10 broad emission sectors (Aircraft, Ships, Energy, Industry Processes, Ground Transport, Residential, Solvents, Agriculture, Agriculture Waste Burning, and Waste), while even broader sectoral emissions were provided in HTAPv2.2 (Air, Ships, Energy, Industry, Transport, Residential (including waste), and Agriculture (only for NH₃)).



1 After collecting all the emission data from reporting and GAINS, the source sectors are harmonised, distinguishing
2 around 250 different subsectors. For each detailed sector, a speciation is applied to the PM_{2.5} and PM₁₀ emissions,
3 distinguishing elemental carbon (representing BC in the HTAP_v3 inventory), organic carbon and other non-
4 carbonaceous emissions for both the coarse (2.5-10 µm) and fine (<2.5 µm) mode.

5 A consistent spatial resolution is applied across the entire domain, where a specific proxy is selected for each
6 subsector to spatially distribute emissions, including for instance the use of point source emissions, e.g., from the
7 European Pollutant Release and Transfer Register (E-PRTR), complemented with additional data from the
8 reporting of EU Large Combustion Plants (European Commission, 2001) and the Platts/WEPP commercial
9 database for power plants (Platts, 2017). Road transport emissions are spatially disaggregated using information
10 from OSM (Open Street Map, 2017), combined with information on traffic intensity in specific road segments
11 from OTM (OpenTransportMap, 2017). Agricultural livestock emissions are spatially distributed using global
12 gridded livestock numbers (FAO, 2010). Furthermore, CORINE land cover (Copernicus Land Monitoring
13 Service, 2016) and population density are other key spatial distribution proxies.

14 After having spatially distributed the data, the ~ 250 different source categories are aggregated to fit with the
15 HTAP_v3 sector classification (Table S1). CAMS-REG-v5.1 is an update of an earlier version, CAMS-REG-v4.2
16 and based on the 2020 submissions to cover the years 2000-2018. A detailed description of the CAMS-REG-v4.2
17 inventory is provided in Kuenen et al. (2022).

18 The data are provided as gridded annual totals at a resolution of 0.05°x0.1° (lat-lon). Along with the grids,
19 additional information is available including height profiles as well as temporal profiles to break down the annual
20 emissions into hourly data (monthly profiles, day-of-the-week profiles and hourly profiles for each day).
21 Furthermore, the CAMS-REG inventory provides dedicated speciation profiles for NMVOC per year, country
22 and sector.

23 2.3.2 US EPA inventory

24 Emissions estimates for the United States were based primarily on estimates produced for the EPA's Air Quality
25 Time Series Project (EQUATES), which generated a consistent set of modelled emissions, meteorology, air
26 quality, and pollutant deposition for the United States spanning the years 2002 through 2017
27 (<https://www.epa.gov/cmaq/equates>). For each sector, a consistent methodology was used to estimate emissions
28 for each year in the 16-year period, in contrast to the evolving methodologies applied in the triennial U.S. National
29 Emissions Inventories (NEIs) produced over that span. The HTAPv3 time series was extended back one year to
30 2001 and forward one year to 2018 using country, sector, and pollutant specific trends from EDGARv6.1.

31

32 Emissions estimates were calculated for more than 8000 Source Classification Codes grouped into 101 sectors
33 and then aggregated to the 16 HTAP_v3 emission sectors. The 2017 NEI (U.S. Environmental Protection Agency,
34 2021a) served as the base year for the time series. For each sector, emissions estimates were generated for
35 previous years using one of four methods: 1) applying new methods to create consistent emissions for all years,
36 2) scaling the 2017 NEI estimates using annual sector-specific activity data and technology information at the
37 county level, 3) using annual emissions calculated consistently in previous NEIs and interpolating to fill missing
38 years, and 4) assuming emissions were constant at 2017 levels constant. The assumption of constant emissions
39 was applied to a very limited number of sources. Foley et al. (submitted, 2022) provides a detailed explanation of
40 the assumptions used for each sector.

41 Emissions from electric generating units were estimated for individual facilities, combining available hourly
42 emissions data for units with continuous emissions monitors (CEMs) and applying regional fuel-specific profiles
43 to units without CEMS. On-road transport and non-road mobile emissions were estimated using emission factors
44 from the MOVES v3 model (U.S. Environmental Protection Agency, 2021b). A complete MOVES simulation
45 was completed only for the NEI years with national adjustment factors applied for years plus or minus one from
46 the NEI year. For California, emission factors for all on-road sources for all years were based on the California
47 Air Resources Board Emission Factor Model (EMFAC) ([https://ww2.arb.ca.gov/our-work/programs/mobile-
48 source-emissions-inventory/](https://ww2.arb.ca.gov/our-work/programs/mobile-source-emissions-inventory/)). New non-road emissions estimates for Texas were provided by the Texas
49 Commission on Environmental Quality. Emissions from oil and gas exploration and production were calculated
50 using point source specific data and the EPA Oil and Gas Tool (U.S. Environmental Protection Agency, 2021a),
51 incorporating year-specific spatial, temporal, and speciation profiles. Residential wood combustion estimates
52 were developed with an updated methodology incorporated into the 2017 NEI and scaled backward to previous



1 years using a national activity as a scaling factor. Solvent emissions were estimated using the Volatile Chemical
2 Product (VCPy) framework of Seltzer et al. (2021). Emissions from livestock waste were calculated with revised
3 annual animal counts to address missing data and methodological changes over the period. Emissions for
4 agricultural burning were developed using a new suite of activity data with the same methodology and input data
5 sets for all years except 2002. County-level estimates were only available for 2002 because activity data based
6 on satellite information was not yet available. Emissions for forest wildfires, prescribed burns, grass and
7 rangeland fires were also calculated in EQUATES but not included in the HTAP_v3 data. For EQUATES,
8 fugitive dust emissions (e.g., unpaved road dust, coal pile dust, dust from agricultural tilling) were reduced to
9 account for precipitation and snow cover by grid cell. For use in HTAP_v3, however, no meteorological
10 adjustments (which decrease annual PM₁₀ emissions by about 75% on average) were applied to fugitive dust
11 emissions. Wind-blown fugitive dust emissions are not included in the estimates for other regions in the HTAP_v3
12 mosaic

13 Non-point source emissions were allocated spatially based on a suite of activity surrogates (e.g. population, total
14 road miles, housing, etc.), many of which are sector specific. The spatial allocation factors were calculated for
15 the 0.1 degree grid used by EDGARv6.1 with no intermediate re-gridding. The spatial allocation factors for all
16 sectors were held constant for the entire time series except for oil and gas sectors which were year-specific.
17 Depending on the sector, either 2017-based or 2014-based surrogates were developed for the same sectors as in
18 the EQUATES.

19 Emissions from the US EPA inventory were provided from 2002-2017 (Table 1). Emissions for the year 2018
20 were estimated applying country, sector and pollutant specific trends from EDGAR and for the 2000 and 2001 to
21 complete the entire time series. Table S1 provides an overview about the US EPA inventory sector mapping to
22 the HTAP_v3 sectors.

23 2.3.3 Environment and Climate Change Canada (ECCC) inventory

24 The Canadian emissions inventory data were obtained from 2018-released edition of Canada's Air Pollutant
25 Emissions Inventory (APEI) originally compiled by the Pollutant Inventories and Reporting Division (PIRD) of
26 Environment and Climate Change Canada (ECCC) (APEI, 2018). This inventory contains a comprehensive and
27 detailed estimate of annual emissions of seven criteria air pollutants (SO₂, NO_x, CO, NMVOC, NH₃, PM₁₀, PM_{2.5})
28 at the national and provincial/territorial level for each year for the period from 1990 to 2016. The APEI inventory
29 was developed based on a bottom-up approach for facility-level data reported to the National Pollutant Release
30 Inventory (NPRI) (NPRI, 2017), as well as an in-house top-down emission estimates based on source-specific
31 activity data and emissions factors. In general, methodologies used to estimate Canadian emissions are consistent
32 with those developed by the U.S. EPA (EPA, 2009) or those recommended in the European emission inventory
33 guidebook (EMEP/EEA, 2013). These methods are often further adjusted by PIRD to reflect the Canadian climate,
34 fuels, technologies and practices.

35 To prepare emissions in the desired HTAP classification, the APEI sector emissions were first mapped to the
36 United Nations Economic Commission for Europe (UNECE) Nomenclature for Reporting (NFR) categories,
37 which involved dividing the sector emissions into their combustion and process components. The NFR categories
38 were then mapped to the HTAP 16 sector categories provided in the sector disaggregation scheme guide. Table
39 S1 provides an overview of ECCC sector mapping to the HTAP_v3 sectors.

40 The HTAP-grouped APEI inventory emissions files were further processed by the Air Quality Policy-Issue
41 Response (REQA) Section of ECCC to prepare the air-quality-modelling version of inventory files in the standard
42 format (i.e., FF10 format) supported by the U.S EPA emissions processing framework. To process emissions into
43 gridded, speciated and total monthly values, a widely-used emissions processing system called the Sparse Matrix
44 Operator Kernel Emissions (SMOKE) model, version 4.7 (UNC, 2019) was used. As part of the preparation for
45 SMOKE processing, a gridded latitude-longitude North American domain at 0.1 x 0.1 degree resolution was
46 defined with 920 columns and 450 rows covering an area of -142W to -50W and 40N to 85N. The point-source
47 emissions in the APEI include latitude and longitude information so those sources were accurately situated in the
48 appropriate grid cell in the Canadian HTAP gridded domain. However, to allocate provincial-level non-point
49 source emissions into this domain, a set of gridded spatial surrogate fields was generated for each province from
50 statistical proxies, such as population, road network, dwellings, crop distributions, etc. Over 80 different surrogate
51 ratio files were created using the 2011 Canadian census data obtained from Statistics Canada website
52 (<https://www12.statcan.gc.ca/census-recensement/2011/index-eng.cfm>) and other datasets, such as the Canadian
53 National Road Network (<https://open.canada.ca/data/en/dataset/3d282116-e556-400c-9306-ca1a3cada77f>).



1 To map the original APEI inventory species to the HTAP's desired list of species, PM speciation profiles from
2 the SPECIATE version 4.5 database (EPA, 2016) were used to calculate source-type-specific EC and OC
3 emissions. As a final step in SMOKE processing, the monthly emissions values were estimated using a set of
4 sector-specific temporal profiles developed and recommended by the U.S. EPA (Sassi, 2021). For the point
5 sources the NPRI annually reported monthly emissions proportions were applied. Emissions for the years 2017
6 and 2018 were calculated by applying sector- and pollutant-specific trends from EDGAR.

7 **2.3.4 REASv3.2.1 inventory**

8 The Regional Emission inventory in ASia (REAS) series have been developed for providing historical trends of
9 emissions in the Asian region including East, Southeast, and South Asia. REASv3.2.1, the version used in
10 HTAP_v3, runs from 1950 to 2015. REASv3.2.1 includes emissions of SO₂, NO_x, CO, NMVOCs, NH₃, CO₂,
11 PM₁₀, PM_{2.5}, BC, and OC from major anthropogenic sources: fuel combustion in power plant, industry, transport,
12 and domestic sectors; industrial processes; agricultural activities; evaporation; and others.

13 Emissions from stationary fuel combustion and non-combustion sources are traditionally calculated using activity
14 data and emission factors, including the effects of control technologies. For fuel consumption, the amount of
15 energy consumption for each fuel type and sector was obtained from the International Energy Agency World
16 Energy Balances for most countries and province-level tables in the China Energy Statistical Yearbook were used
17 for China. Other activity data such as the amount of emissions produced from industrial processes were obtained
18 from related international and national statistics. For emission factors, those without effects of abatement measures
19 were set and then, effects of control measures were considered based on temporal variations of their introduction
20 rates. Default emission factors and settings of country- and region-specific emission factors and removal
21 efficiencies were obtained from scientific literature studies as described in Kurokawa and Ohara (2020) and
22 references therein.

23 Emissions from road transport were calculated using vehicle numbers, annual distance travelled, and emission
24 factors for each vehicle type. The number of registered vehicles were obtained from national statistics in each
25 country and the World Road Statistics. For emission factors, year-to-year variation were considered by following
26 procedures: (1) Emission factors of each vehicle type in a base year were estimated; (2) Trends of the emission
27 factors for each vehicle type were estimated considering the timing of road vehicle regulations in each country
28 and the ratios of vehicle production years; (3) Emission factors of each vehicle type during the target period were
29 calculated using those of base years and the corresponding trends.

30 In REASv3.2.1, only large power plants were treated as point sources. For emissions from cement, iron, and steel
31 plants, grid allocation factors were developed based on positions, production capacities, and start and retire years
32 for large plants. Gridded emission data of EDGARv4.3.2 were used for grid allocation factors for the road
33 transport sector. Rural, urban, and total population data were used to allocation emissions from the residential
34 sector. For other sources, total population were used for proxy data.

35 For temporal distribution, if data for monthly generated power and production amounts of industrial products were
36 available, monthly emissions were estimated by allocating annual emissions to each month using the monthly data
37 as proxy. For the residential sector, monthly variation of emissions was estimated using surface temperature in
38 each grid cell. If there is no appropriate proxy data, annual emissions were distributed to each month based on
39 number of dates in each month.

40 Monthly gridded emission data sets at 0.25° × 0.25° resolution for major sectors and emission table data for major
41 sectors and fuel types in each country and region during 1950-2015 are available in text format from a data
42 download site of REAS (<https://www.nies.go.jp/REAS/>). Table S1 provides an overview about the REASv3.2.1
43 sector mapping to the HTAP_v3 sectors.

44 More details of the methodology of REASv3.2.1 are available in Kurokawa and Ohara (2020) and its supplement.
45 (Note that REASv3.2.1 is the version after error corrections of REASv3.2 of Kurokawa and Ohara (2020)). Details
46 of the error corrections are described in the data download site of REAS.) Table S1 provides an overview about
47 the REASv3.2.1 sector mapping to the HTAP_v3 sectors.

48 The MEIC inventory (<http://meicmodel.org>, 2021) is not currently included in the HTAP_v3 mosaic. Since the
49 REAS inventory only includes emissions until 2015, the REAS-based HTAP_v3 mosaic is only complete until
50 this year. Emissions beyond 2015 were extrapolated using trends derived from a combination of MEIC and
51 EDGAR. To extend the Chinese emission estimates to most recent years, MEIC data were used to adjust sector
52 and pollutant specific trend for China for the years 2016 and 2017 (refer to Table S2 for the mapping sectors of
53 MEIC and HTAP_v3). Then, the 2018 data were calculated based on the 2015-2017 trend. For all the other



1 countries belonging to the REAS domain, the emissions were extended beyond 2015 applying the sector-, country-
2 , and pollutant-specific trends from EDGAR.

3 **2.3.5 CAPSS-KU inventory**

4 In the Republic of Korea, the National Air Emission Inventory and Research Center (NAIR) estimates annual
5 emissions of the air pollutants CO, NO_x, SO_x, TSP, PM₁₀, PM_{2.5}, BC, VOCs, and NH₃ via the Clean Air Policy
6 Support System (CAPSS). The CAPSS inventory is divided into four source-sector levels (high, medium, low and
7 detailed) based on the European Environment Agency's (EEA) CORE Inventory of AIR emissions
8 (EMEP/CORINAIR). For activity data, various national- and regional-level statistical data collected from 150
9 domestic institutions are used. For large point sources, emissions are estimated directly using real-time stack
10 measurements. For small point, area and mobile sources, indirect calculation methods using activity data, emission
11 factors, and control efficiency are used.

12 Even though CAPSS (Clean Air Policy Support System) has been estimating annual emissions since 1999, some
13 inconsistencies exist in the time series because of the data and methodological changes over the period. For
14 example, emissions of PM_{2.5} were initiated from the year 2011 and not from 1999. Therefore, in the CAPSS
15 emission inventory, PM_{2.5} emissions were calculated from 2011, and post-2011 the PM₁₀ to PM_{2.5} emission ratio
16 was used to calculate the emissions from 2000 to 2010. These limitations make it difficult to compare and analyse
17 emissions inter-annually. To overcome these limitations, re-analysis of the annual emissions of pollutants was
18 conducted using upgrades of the CAPSS inventory, such as missing source addition and emission factor updates.

19 The biomass combustion and fugitive dust sector emissions from 2000 to 2014 were estimated and added in the
20 inventory, which are newly calculated emission sources from 2015. As for the on-road mobile sector, new
21 emission factors using 2016 driving conditions were applied from the year 2000 to 2015. Since the emissions from
22 the combustion of imported anthracite coal were calculated only from 2007, the coal use statistics of imported
23 anthracite from 2000 to 2006 were collected to estimate emissions for those years.

24 After all the adjustments, a historically re-constructed emissions inventory using the latest emission estimation
25 method and data was developed. Table S1 provides an overview about the CAPSS sector mapping to the
26 HTAP_v3 sectors.

27 **2.3.6 JAPAN inventory (PM2.5EI and J-STREAM)**

28 The Japanese emission inventory contributing to the HTAP_v3 mosaic is jointly developed by the Ministry of the
29 Environment, Japan (MOEJ) for emissions arising from mobile sources and by the National Institute of
30 Environmental Studies (NIES) for estimating emissions from fixed sources.

31 The mobile source emissions data for the HTAP_5.1, 5.2, and 5.4 sectors are based on the air pollutant emission
32 inventory named "PM2.5 Emission Inventory (PM2.5EI), <https://www.env.go.jp/air/osen/pm/info.html>).
33 PM2.5EI has been developed for the years 2012 and 2015, while for 2018 is currently under development. Almost
34 all anthropogenic sources are covered, but emissions from vehicles are estimated in particular detail based on
35 JATOP (Shibata, 2021). The emission factor of automobiles is constructed by MOEJ as a function of the average
36 vehicle speed over several kilometres in a driving cycle that simulates driving on a real road. Emission factors are
37 organized by 7 types of vehicles, 2 fuel types, 5 air pollutants, and regulation years, and have been implemented
38 since 1997 as a project of MOEJ. By using these emission factors and giving the average vehicle speed on the
39 road to be estimated, it is possible to estimate the air pollutant emissions per kilometre per vehicle. The hourly
40 average vehicle speed of trunk roads, which account for 70% of Japan's traffic volume, is obtained at intervals of
41 several kilometres nationwide every five years, so the latest data for the target year is used. For narrow roads, the
42 average vehicle speed by prefecture measured by probe information is applied. It is 20 km/h in Tokyo, but slightly
43 faster in other prefectures. Starting emission is defined as the difference between the exhaust amount in the
44 completely cold state and the warm state in the same driving cycle and is estimated by the times the engine started
45 in a day. Chassis dynamometer tests are performed in a well-prepared environment, so for more realistic emissions
46 estimates, temperature correction factor, humidity correction factor, deterioration factor, DPF regeneration factor,
47 and soak time correction factor are used. In addition to running and starting emissions, evaporative emissions
48 from gasoline vehicles and non-exhaust particles such as road dust (including brake wear particles) and tire wear
49 particles are combined to provide a vehicle emissions database with a spatial resolution of approximately 1 km ×
50 1 km (30° latitude, 45° longitude), and a temporal resolution of an hour by month, including weekdays and
51 holidays.



1 Off-road vehicle emissions are estimated separately for 17 types of construction machinery, industrial machinery
2 (forklifts), and 5 types of agricultural machinery. In all cases, emission factors by type and regulatory year per
3 workload are used, as researched by the MOEJ. Although not as precise as automobiles, the off-road database is
4 provided with the same temporal and spatial resolution as the automobile database.

5 Emissions from stationary sources in Japan are derived from the emission inventory developed in the Japan's
6 Study for Reference Air Quality Modelling (J-STREAM) model intercomparison project (Chatani et al., 2018;
7 Chatani et al., 2020). In this emission inventory, emissions from stationary combustion sources are estimated by
8 multiplying emission factors and activities including energy consumption, which is available in the comprehensive
9 energy statistics. Large stationary sources specified by the air pollution control law need to report emissions to
10 the government every three years. The emission factors and their annual variations were derived from the
11 emissions reported by over 100,000 sources (Chatani et al., 2020). For fugitive VOC emissions, MOEJ maintains
12 a special emission inventory to check progress on regulations and voluntary actions targeting 30% reduction of
13 fugitive VOC emissions starting from 2000. VOC emissions estimated in this emission inventory are used.
14 Emissions from agricultural sources are consistent with the emissions estimated in the national greenhouse gas
15 emission inventory (Center for Global Environmental Research et al., 2022). Emissions of all the stationary
16 sources are divided into prefecture, city, and grid (approximately 1 x 1 km) levels based on spatial proxies specific
17 to each source. Emissions for the year 2018 were estimated applying sector- and pollutant-specific trends from
18 EDGAR. Table S1 provides an overview about the Japanese inventory sector mapping to the HTAP_v3 sectors.

19 2.4 Gap-filling methodology with EDGARv6.1

20 EDGAR is a globally consistent emission inventory of air pollutant and greenhouse gases developed and
21 maintained by the Joint Research Centre of the European Commission (<https://edgar.jrc.ec.europa.eu/>). The
22 EDGAR methodology used to compute GHG and air pollutant emissions has been described in detail in several
23 publications (Janssens-Maenhout et al., 2019; Crippa et al., 2018) and summarised here after. In EDGAR, air
24 pollutant emissions are computed making use of international statistics as activity data (e.g., International Energy
25 Balance data, Food and Agriculture Organisation statistics, US Commodity statistics, etc.), region- and/or country-
26 specific emission factors by pollutant/sector and technology and abatement measures, following Eq. 1:

$$27 \quad EM_{i(C,t,x)} = \sum_{j,k} AD_{i(C,t)} * TECH_{i,j(C,t)} * EOP_{i,j,k(C,t)} * EF_{i(C,t,x)} * (1 - RED)_{i,j,k(C,t,x)}$$

28 (Eq. 1)

29 where EM are the emissions from a given sector *i* in a country *C* accumulated during a year *t* for a chemical
30 compound *x*, AD the country-specific activity data quantifying the human activity for sector *i*, TECH the mix of
31 *j* technologies (varying between 0 and 1), EOP the mix of *k* (end-of-pipe) abatement measures (varying between
32 0 and 1) installed with a share *k* for each technology *j*, and EF the uncontrolled emission factor for each sector *i*
33 and technology *j* with relative reduction (RED) by abatement measure *k*. Emission factors are typically derived
34 from the EMEP/EEA Guidebooks (EMEP/EEA, 2013, 2019, 2016), the AP-42(EPA, 2009) inventory and
35 scientific literature.

36 Annual country and sector specific air pollutant emissions are then disaggregated into monthly values (Crippa et
37 al., 2020) and subsequently spatially distributed making use of detailed proxy data (Janssens-Maenhout et al.,
38 2019; Crippa et al., 2021).

39 As the most comprehensive and globally consistent emission database, the latest update of the EDGAR air
40 pollutant emissions inventory, EDGARv6.1 (https://edgar.jrc.ec.europa.eu/dataset_ap61), is used in the HTAP_v3
41 mosaic to complete missing information from the officially reported inventories, as reported in Table 3.
42 EDGARv6.1 includes important updates to estimate air pollutant emissions such as the improvement of road
43 transport emission estimates for many world regions, the inclusion of agricultural NMVOC emissions, revised
44 monthly emission profiles (Crippa et al., 2020) and updated spatial proxies to distribute national emissions by
45 sector over the globe (Crippa et al., 2021). EDGARv6.1 also includes new international shipping proxies and their
46 monthly distribution based on the STEAM model (Jalkanen et al., 2012; Johansson et al., 2017). In the
47 Supplementary Material (section S2), the assessment of EDGAR emission data is reported in comparison with
48 global and regional inventories.

49 3 Results

50 3.1 Annual time series analysis: trends and regional and sectoral contributions



1 Having a consistent set of global annual emission inventories for a two-decade period allows the investigation of
2 global emissions trends for the inventory pollutants and regional and sectoral contributions. Figure 2 presents
3 annual time series (2000-2018) of the global emissions of the nine air pollutants included in the HTAP_v3 mosaic
4 separated into the actual contributions of 12 regions. Figure 3 shows the corresponding relative contributions of
5 (a) 16 sectors and (b) 12 regions to the 2018 global emissions of these same pollutants. We can then consider each
6 pollutant in turn.

7 Global SO₂ emissions declined from 99.4 to 72.9 Mt over the past two decades. This decreasing pattern is found
8 for several world regions with the fastest decline in Eastern Asia, where after the year 2005 SO₂ emissions began
9 to decrease steadily. This is consistent with the use of cleaner fuels with lower sulphur content and the
10 implementation of desulphurisation techniques in power plants and industrial facilities in China in accordance
11 with the 11th Five-Year Plan (FYP, 2006–2010 (Planning Commission, 2008)) and the 12th Five-Year Plan (FYP,
12 2011–2015 (Hu, 2016)) (Sun et al., 2018). Similarly, industrialised regions, such as North America and Europe,
13 are characterised by a continuous decreasing trend in SO₂ emission, which had started well before the year 2000
14 due to the implementation of environmental and air quality legislation (EEA, 2022). Increasing SO₂ emissions,
15 on the other hand, are found for Southern Asia (+112% compared to 2000), South-East Asia and developing
16 Pacific (+62%), and Africa (+40%). These increases mostly arise from the energy, industry, and (partly)
17 residential sectors, and reflect the need for emerging and developing economies to mitigate these emissions.
18 Emissions estimated using satellite retrievals and model inversions confirm the trends provided by the HTAP_v3
19 mosaic (Liu et al., 2018). SO₂ is mostly emitted by power generation and industrial activities, which in 2018
20 represent 42% and 26%, respectively, of the global total. Despite measures in some specific sea areas to mitigate
21 sulphur emissions, globally they have been rising steadily with increasing activity. International shipping
22 represents 13.8% of global SO₂ emissions in 2018, and it is 41% higher compared to the 2000 levels (Fig. 3).

23 Global NO_x emissions increased from 110.4 Mt in 2000 to 117.4 Mt in 2018 as a result of the increase in energy-
24 and industry-related activities for most of the world regions (in particular over the Asian domain). The strongest
25 decreases are found for North America (-63%), Europe (-42%), Asia-Pacific Developed (-32%) and to a lower
26 extent for Eurasia (-6%). Comparable spatio-temporal patterns are found by satellite OMI data and ground based
27 measurements of NO₂ concentrations (Jamali et al., 2020). NO_x is mainly produced at high combustion
28 temperatures (e.g., power and industrial activities, 38% of the global total), but also by transportation (27% of the
29 global total) and international shipping (14% of the global total).

30 CO is mostly emitted by incomplete combustion processes from residential combustion, transportation and the
31 burning of agricultural residues. Globally, CO emissions showed little change over the past two decades (502.7
32 Mt in 2000 vs. 499.8 in 2018), but different regional trends are present. Historically industrialised regions have
33 reduced their emissions over the years (-42% in Europe and -62% in North America), while CO emissions
34 increased in Africa by 45% and in Southern Asia by 49%. Road transport CO emissions halved over the past two
35 decades (-55%), while the emissions from all other sectors increased. These results are consistent with MOPITT
36 satellite retrievals, which mostly show the same trends over the different regional domains over the past decades
37 (Yin et al., 2015).

38 NMVOC emissions increased from 115.2 Mt in 2000 to 146 Mt in 2018. These emissions are mostly associated
39 with the use of solvents (25% of the 2018 global total), fugitive emissions (23%), road transportation (including
40 both combustion and evaporative emissions, 15%) and small-scale combustion activities (19%). The most
41 prominent increases in the emissions at the global level are found for the energy (+98%) and solvents (+81%). In
42 2018, NMVOC emissions from solvents were 3.7 and 3.5 times higher than in 2000 in China and India,
43 respectively, while a rather stable trend in found for US and Europe.

44 Global NH₃ emissions increased from 47.4 Mt in 2000 to 58.9 Mt in 2018 due to enhanced emissions from
45 agricultural activities. In particular, NH₃ emissions strongly increased in Africa (+60.5%), South-East Asia and
46 developing Pacific (48.3%), Southern Asia (+38.7%), and Latin America and Caribbean (+41.1%).

47 Particulate matter emissions showed little change over the past two decades at the global level, whereas regional
48 emission increases are found for Africa (e.g., +47.0% for PM₁₀), Latin America and Caribbean (+39.1%), Middle
49 East (48.3%), and Southern Asia (+56%), mostly associated with increases in agricultural waste burning and the
50 livestock, energy, and waste sectors. By contrast, Eastern Asia (-39.5%), Europe (24.3%), and Asia-Pacific
51 Developed (-36.8%) significantly decreased their PM₁₀ emissions over the past two decades due to the continuous
52 implementation of reduction and abatement measures for the energy, industry, road transport and residential
53 sectors ((Crippa et al., 2016). As shown in Fig. 3, the relative contribution of North America to global PM₁₀ is
54 quite high compared to other substances due to fugitive dust emissions (e.g., unpaved road dust, coal pile dust,
55 dust from agricultural tilling) which have not been adjusted for meteorological conditions (e.g., rain, snow) and
56 near-source settling and mitigation (e.g., tree wind breaks) because these removal mechanisms are better



1 addressed by the chemical transport models. Additional uncertainty may be therefore introduced for these
2 emissions, depending on the modelling assumptions of each official inventory. Similarly, particulate matter
3 speciation into its carbonaceous components it is often challenging and subjected to higher level of uncertainty,
4 for instance because different definitions are used for PM in inventories, including condensable emissions or not
5 (Denier van der Gon et al., 2015). Attempts to improve the accuracy of such emissions (e.g. BC and OC emissions
6 over the European domain) are ongoing.

7 3.2 Emission maps

8 Spatially distributed emission data describe where emissions take place, as input for local, regional and global air
9 quality modelling. As noted in section 2.2, nationally aggregated air pollutant emissions are spatially distributed
10 over the corresponding national territory using spatial proxy data which are believed to provide a relatively good
11 representation of where emissions takes place. Depending on the emitting sector, air pollutants can be associated
12 with the spatial distributions of point sources (e.g., in the case of power plant or industrial activities), road
13 networks (e.g., for transportation related emissions), settlement areas (e.g., for small-scale combustion emissions),
14 crop and livestock distribution maps, ship tracks etc. Using reliable and up-to-date spatial information to distribute
15 national emissions is therefore relevant, although challenging. Multiple assumptions are often made by inventory
16 compilers when developing their inventories, which may result in differences when analysing spatially distributed
17 emissions provided by different inventory compilers over the same geographical domain.

18 One key goal of the HTAP_v3 mosaic is to collate in one inventory the most accurate spatially-distributed
19 emissions for all air pollutants at the global level, based on the best available local information. Point sources
20 related with emissions from power plant and industrial facilities represent one the most critical spatial information
21 to be retrieved, and their misallocation can significantly affect the characterisation of local air quality. This
22 challenge is also present in the HTAP_v3 mosaic. For example, the REASv3.2.1 inventory is still using limited
23 information to distribute emissions from these two sectors especially for industrial plants. Depending on the
24 region, point source information could be limited compared to datasets used in inventories of North America and
25 Europe. To overcome this issue, the participation of national emission inventory developers not only from China,
26 but also India and other Asian countries is recommended. The impact can be seen in Fig. 4, which shows the
27 global map of SO₂ emissions in 2018 based on the HTAP_v3 mosaic compilation, where information about the
28 magnitude and the type of emission sources for the different regions can be retrieved. The energy and industry
29 sectors contribute a large fraction of SO₂ emissions (Fig. 3a), but the spatial distribution of these emissions is
30 qualitatively different in North America and Europe than in Asia (i.e., more “spotty”, less smooth and widely
31 distributed). Ship tracks cover the entire geographical marine domain, consistent with emissions from the STEAM
32 model (Jalkanen et al., 2012; Johansson et al., 2017) included in the EDGARv6.1 database, although showing
33 marked emissions over the Mediterranean Sea, Asian domain, Middle East and North American coasts.
34 Furthermore, emissions from power plant and industrial activities, as well as small-scale combustion are
35 prominent over the Asian domain, Eastern Europe, and some African regions.

36 Sector-specific case studies are presented in the maps of Figs. 5-8. Figure 5 shows the comparison of annual NO_x
37 emissions for the year 2000 and 2018. The road transport sector is a key source of NO_x emissions (cf. Fig. 3a),
38 and this contribution is reflected in the visible presence of road networks in the maps. Decreasing emissions are
39 found for industrialised regions (USA, Europe, Japan) thanks to the introduction of increasingly restrictive
40 legislation on vehicle emissions since the 1990s, whereas a steep increase is found for emerging economies and
41 in particular India, China, and the Asian domain. Figure 6 shows the different spatial allocation of PM₁₀ emissions
42 from the residential sector during the month of January 2018, with higher emission intensities evident in the
43 Northern Hemisphere (cold season) and the lower values in the Southern Hemisphere (warm season). Figures 7
44 and 8 show the spatio-temporal allocation of agriculture-related emissions, and specifically, PM₁₀ emissions from
45 agricultural waste burning and NH₃ emissions from agricultural soil activities.

46 3.3 Monthly temporal distribution

47 3.3.1 Monthly variability by region

48 The magnitude of air pollutant emissions varies by month because of the seasonality of different anthropogenic
49 activities and their geographical location (e.g., Northern vs. Southern Hemisphere regions). Figures 9 and 10 (and
50 S3.1, S3.2 and S3.3) show the monthly distribution of regional emissions for those pollutants and sectors for
51 which higher variability is expected. The year 2015 was chosen since it is the last year for which all of the official
52 data providers have data. Figure 9 shows monthly NH₃ emissions by region from three agricultural activities
53 (agricultural waste burning, livestock, and crops). These sectors display the largest variability by month, reflecting
54 the seasonal cycle and the region-specific agricultural practices, such as fertilisation, crop residue burning, manure
55 and pasture management, animal population changes, etc. In Figure 10, NO_x emissions from residential activities



1 show a particular monthly distribution, with the highest emissions occurring during the cold months shifted for
2 the Northern and Southern Hemispheres. By contrast, regions in the equatorial zone do not show a marked monthly
3 profile even for residential activities. The energy sector also follows monthly-seasonal cycles related to the
4 demand for power generation, which is also correlated with ambient temperature and local day length. Transport-
5 related emissions do not show a large variation by month, whereas daily and weekly cycles for transport-related
6 emissions, which are typically more relevant, are beyond the temporal resolution of this work.

7 Although a spatio-temporal variability of the HTAP_v3 emissions is found in these figures, a more in-depth
8 analysis reveals that with the exception of few regions and sectors (e.g., Canada, USA and regions gap-filled with
9 EDGAR), no inter-annual variability of the monthly profiles is present, meaning that the majority of official
10 inventories assume the same monthly distribution of the emissions for the past two decades (refer to Figs. S3.4-
11 S3.9). This is different from the approach used for example by EDGAR (Crippa et al., 2020), ECCO for Canada,
12 and U.S. EPA for the USA, where year-dependent monthly profiles are used for specific sectors, in particular for
13 residential, power generation, and agricultural activities. Further analysis has shown that for the European domain
14 regional rather than country-specific monthly profiles are applied. Improvements of the monthly distribution could
15 be achieved through the implementation of the state-of-the-art information on this topic (Crippa et al., 2020;
16 Guevara et al., 2021).

17

18 3.3.2 Spatially-distributed monthly emissions

19 An important added value of HTAP_v3 comes from the availability of monthly gridmaps that reflect the
20 seasonality of the emissions for different world regions. Access to spatially distributed monthly emissions is
21 essential to design effective mitigation actions, providing information on hot spots of emissions and critical
22 periods of the year when emissions are highest.

23 Figure 11 shows mid-season PM_{2.5} monthly emissions arising from the residential sector in 2018. The global map
24 shows higher emissions in the Northern Hemisphere during January, while the opposite pattern is found for the
25 Southern Hemisphere in July. Agriculture is an important activity characterised by strong seasonal patterns, as
26 shown in Figs. 12 and 13. Figure 12 shows PM₁₀ monthly emission maps from agricultural residue burning in
27 2018 from HTAP_v3, highlighting higher emissions over certain months of the year related with specific burning
28 practices of agricultural residues for different world regions. For example, during the month of April, intense
29 burning of crop residues is found in Africa (Nigeria, Ethiopia, Sudan, South Africa, etc.), South America (Brazil,
30 Argentina, Colombia, etc.), Northern India, and South-Eastern Asia (e.g., Vietnam, Thailand, Indonesia,
31 Philippines, etc.). Figure 13 represents the yearly variability of NH₃ emissions from agricultural soils activities,
32 mostly related with fertilisation. During the month of March and April, intense agricultural soils activities are
33 found over Europe and North America compared to other months, while during the month of October the highest
34 emissions are for this sector are found in China, India, several countries of the Asian domain, but also in USA,
35 Australia, and Latin America. These results are consistent with satellite based observations performed using
36 Cross-track Infrared Sounder (Shephard et al., 2020).

37 3.4 Vertical distribution of the emissions

38 3.4.1 Aircraft emissions

39 In EDGAR6.1 the emissions are provided at three effective altitude levels (landing/take-off, ascent/descent, and
40 cruising). The spatial proxy for the aviation sector is derived from International Civil Aviation Organization
41 (ICAO, 2015) which specifies a typical flight pattern with landing/take-off cycle within few km of the airport,
42 followed by climb-out/descending phase during the first 100 km and the last 100km of a flight and finally the
43 remaining part from 101 km until the last 101 km as the cruise phase. Routes and airport locations are taken from
44 the Airline Route Mapper of ICAO (2015). In HTAP_v3, aircraft emissions are provided as domestic and
45 international, but with no information about altitude ranges. We recommend modellers to use the corresponding
46 EDGARv6.1 data (https://edgar.jrc.ec.europa.eu/dataset_ap61) including the vertical distribution of the
47 emissions.

48

49 3.4 Speciation of NMVOC emissions

50 For emission data to be useful for modellers, total NMVOC emissions must be decomposed into emissions of
51 individual NMVOC species. As the chemical mechanisms used by models can differ with respect to the NMVOC
52 species they include, it is not practical to provide an NMVOC speciation which is usable by all models. Instead,
53 a speciation is provided here for the set of 25 NMVOCs defined by Huang et al. (2017) and the corresponding
54 data are made available on the HTAP_v3 website. The absolute values of 25-category speciated NMVOC



1 emissions were obtained for all countries for the 28 EDGAR sectors from here:
2 https://edgar.jrc.ec.europa.eu/dataset_ap432_VOC_spec. The absolute NMVOC emissions of each species from
3 each sector in this dataset were remapped to the HTAP_v3 sectors following the mapping from Table 2, then
4 converted to a speciation by dividing by the total emissions of each individual species for the four world regions
5 defined by Huang et al. (2017): Asia; Europe; North America; and Other. The resulting NMVOC speciation is
6 provided in the supplementary material to this paper for the 25 NMVOC species, 4 world regions, and 15 emitting
7 NMVOC sectors² following the HTAP_v3 sector classification (including 13 sectors defined over the 4 world
8 regions, and the two international sectors: international shipping and international aviation). The list of countries
9 comprising each region is also provided in the supplement.

10

11 4 Data availability

12 The HTAP_v3 emission mosaic data can be freely accessed and cited using DOI 10.5281/zenodo.7516361. All
13 data can be also accessed through the EDGAR website at the following link:
14 https://edgar.jrc.ec.europa.eu/dataset_htap_v3.

15 Data are made available in the following formats:

- 16 • Monthly gridmaps of emissions (in Mg/month) at 0.1x0.1degree resolution: there is one .NetCDF file
17 per year and substance that includes the emissions for each sector for the 12 months.
- 18 • Monthly gridmaps of emission fluxes (in kg/m²/s) at 0.1x0.1degree resolution: there is one .NetCDF file
19 per year and substance that includes the emission fluxes for each sector the emission fluxes for the 12
20 months.
- 21 • Annual gridmaps of emissions (in Mg/year at 0.1x0.1degree resolution: there is one .NetCDF file per
22 year and substance that includes the emissions for each sector.
- 23 • Annual gridmaps of emission fluxes (in kg/m²/s) at 0.1x0.1degree resolution: there is one .NetCDF file
24 per year and substance that includes the emission fluxes for each sector.

25 The full set of HTAP_v3 data is quite large, requiring substantial network bandwidth and time for download, and
26 substantial storage space. To make it easier for users to query and use the data, additional products are available.
27 For global modellers who may not require such high spatial resolution, gridmaps at 0.5x0.5 degree resolution are
28 made available following the abovementioned specifications of the higher spatial resolution data. Furthermore, to
29 allow regional modellers to download only the data for the regions they need, the JRC EDGAR group has also
30 developed an interface to allow the users of the HTAP_v3 mosaic to extract emission data over arbitrarily specified
31 geographical domains. The HTAP tool is accessible after creation of an ECAS account
32 (<https://webgate.ec.europa.eu/cas/login>) and it is available at: https://edgar.jrc.ec.europa.eu/htap_tool/.

33

34 5 Conclusions

35 The global air pollution mosaic inventory HTAP_v3 presented and discussed in this paper is a state-of-the-art tool
36 for addressing the present status and the recent evolution of a set of policy-relevant air pollutants. The inventory
37 derives from the harmonization and blending of six regional inventories, gapfilled using the most recent release
38 of EDGAR (EDGARv6.1). By directly incorporating the best available local information, including the spatial
39 distribution of emissions, the HTAP_v3 mosaic inventory can be used for policy-relevant studies at both regional
40 and global levels. As such, the HTAP_v3 mosaic inventory provides a complement to globally consistent emission
41 inventories such as EDGAR. The global and regional trends of air pollutant emissions in the HTAP_v3 mosaic
42 are comparable with other commonly available global emission datasets.

43 By providing consistent times series for almost two decades, HTAP_v3 allows an evaluation of the impact and
44 success of the pollution control measures deployed across various regions of the world since 2000. Similarly, its
45 finer sectoral resolution is suitable for understanding how and where technological changes have resulted in
46 emissions reductions, suggesting possible pathways for strengthening appropriate policy actions.

² No speciation profile is provided for the ‘tyre and brake wear sector’ not being a source of NMVOC emissions.



1 All these features make HTAP_v3 a tool of certain interest for policy makers active in the air quality regulatory
2 efforts. Indeed, although HTAP_v3 provides a picture of a world where most pollutants are following a steady or
3 decreasing path, several areas of the world show an opposite emission increasing trend, while wide portions of
4 the world remain subject to unsatisfactory levels of ambient air pollution.

5 Similar to its predecessor (e.g. HTAP_v2.2 mosaic inventory), we expect that this new HTAP_v3 mosaic
6 inventory will be used as a basis for global assessments of long-range, transboundary transport of air pollution
7 under the Task Force on Hemispheric Transport of Air Pollution, while also providing a convenient and useful
8 tool for regional modellers seeking the best available regional emissions with a consistent gap-filling
9 methodology. Future versions of the HTAP mosaic inventory will incorporate new regional emission inventories
10 and updated datasets as time and resources permit.

11

12 **Author contributions.**

13 MC and DG developed the mosaic gathering input from all data providers. The co-chairs of the TF-HTAP (TK,
14 TB, RW and JaKa) fostered the dialogue with international institutions contributing to this work with their data.
15 PM, RM, JR, JZ, DN, MS, MDM, RW provided data for Canada, JuKu, SC, TM provided data for Japan, JeKu
16 provided data for Europe, J-HW, JK provided data for Korea, TK, GP provided data for USA, JiKi provided data
17 for Asia. The JRC EDGAR group (MC, ES, DG, EP, MM, FM, ES, MB, FP) lead the drafting of the publication
18 with input from colleagues contributing to the HTAP_v3 mosaic. SJS and HS performed detailed data comparison
19 among available emission inventories.

20 **Competing interests.** The authors declare that they have no conflicts of interest nor competing interests.

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26 Fund (JPMEERF20165001 and JPMEERF20215005) of the Environmental Restoration and Conservation Agency
27 Provided by the Ministry of Environment of Japan, and FRIEND (Fine Particle Research Initiative in East Asia
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29 the Ministry of Science and ICT (2020M3G1A1114622).

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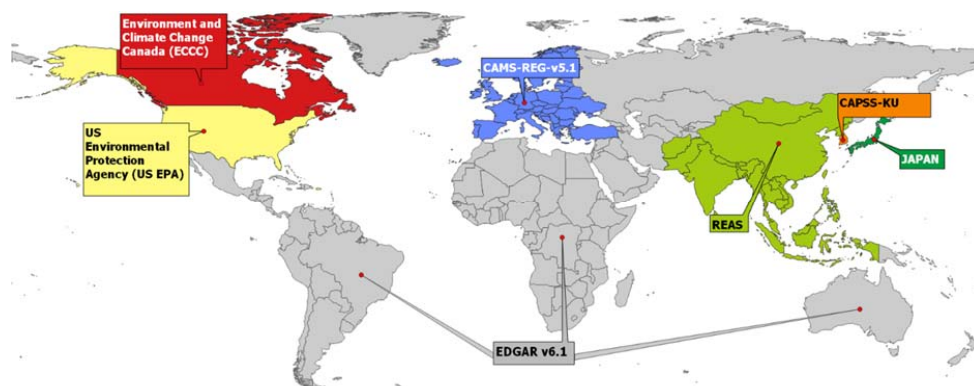
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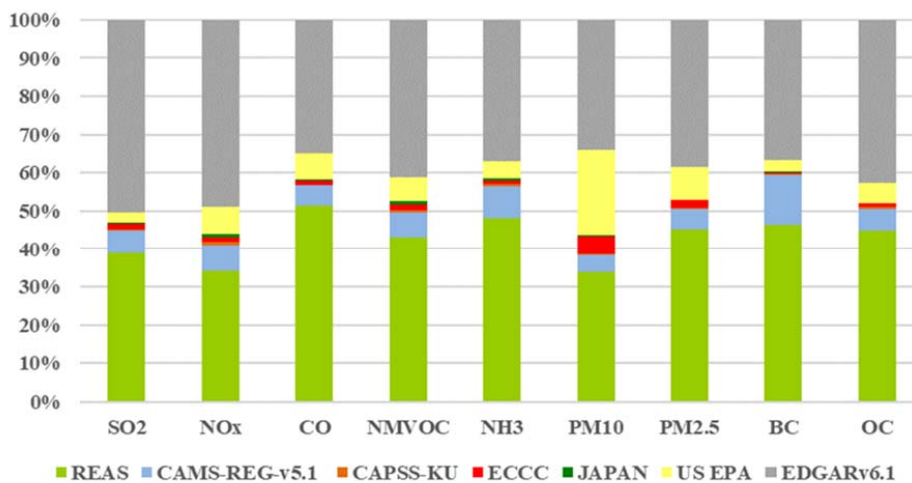
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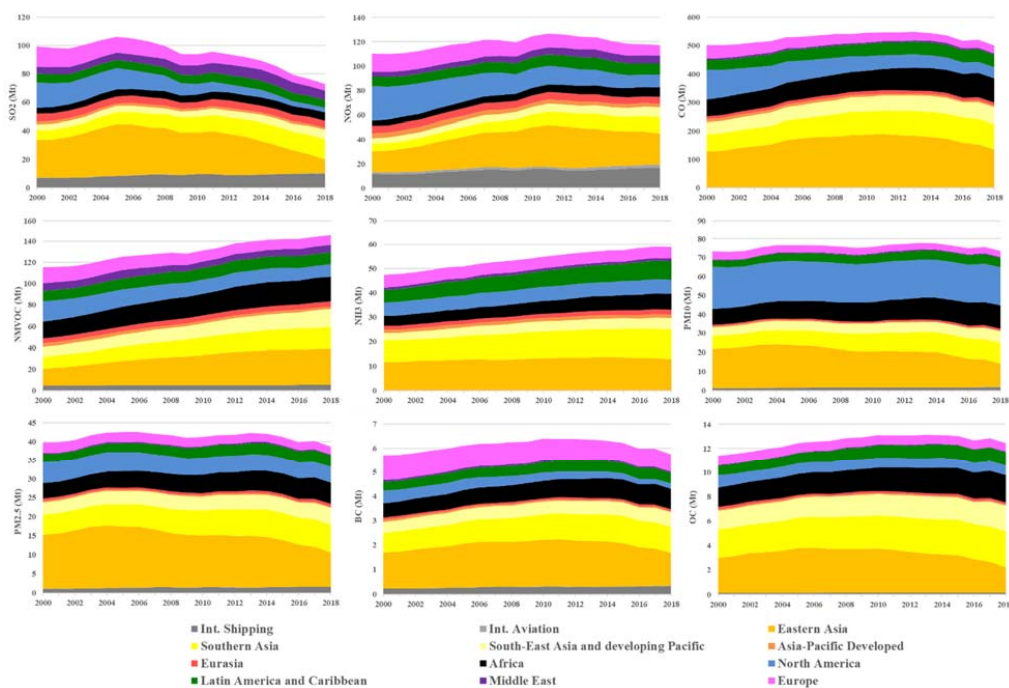


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2 **Figure 1 – Overview of the HTAP_v3 mosaic data providers. Data from officially reported emission**
 3 **gridmaps were collected from the US Environmental Protection Agency, Environment and Climate**
 4 **Canada, CAMS-REG-v5.1 for Europe, REASv3.2.1 for most of the Asian domain, CAPSS-KU for Korea**
 5 **and JAPAN (PM2.5EI and J-STREAM) for Japan. The share of the total emissions covered by each data**
 6 **provider is reported in the bar chart at the bottom.**

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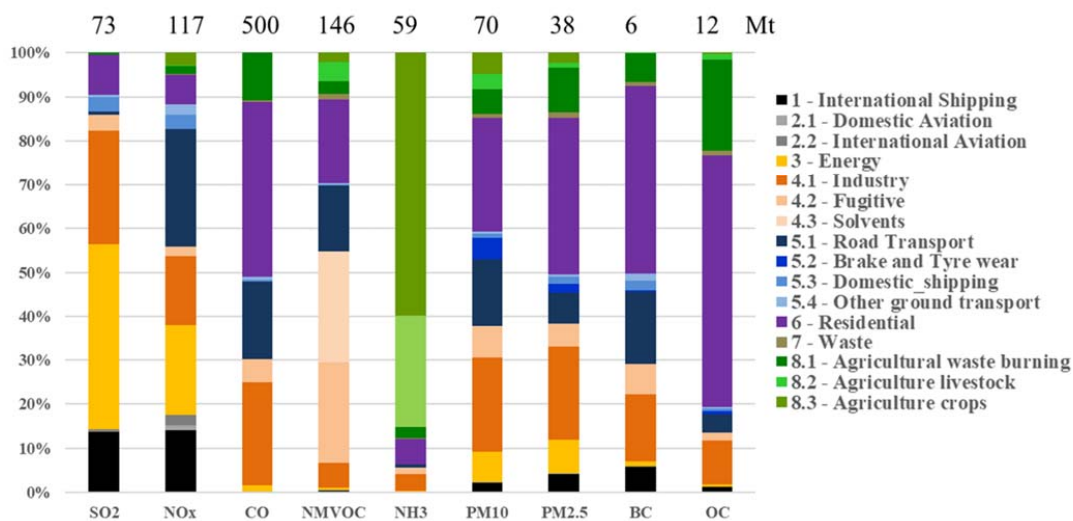


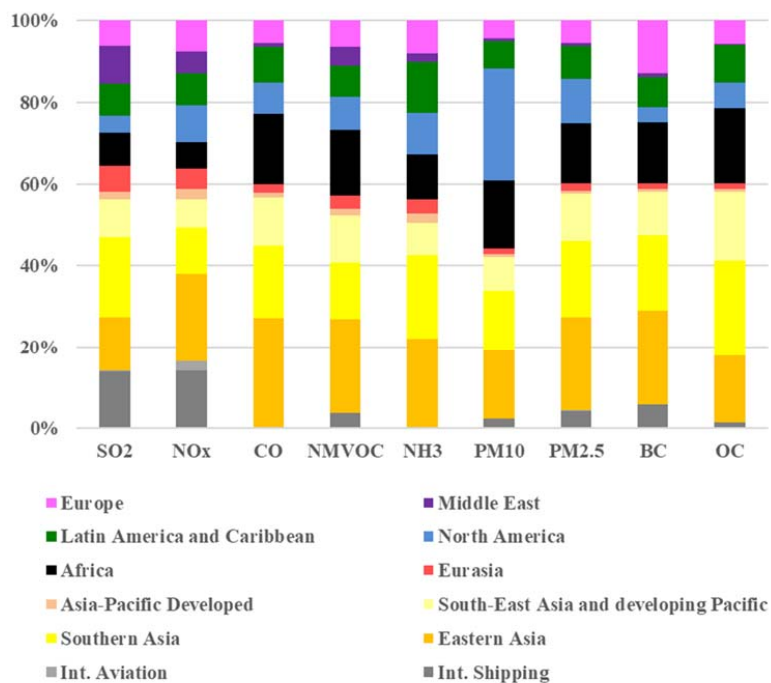
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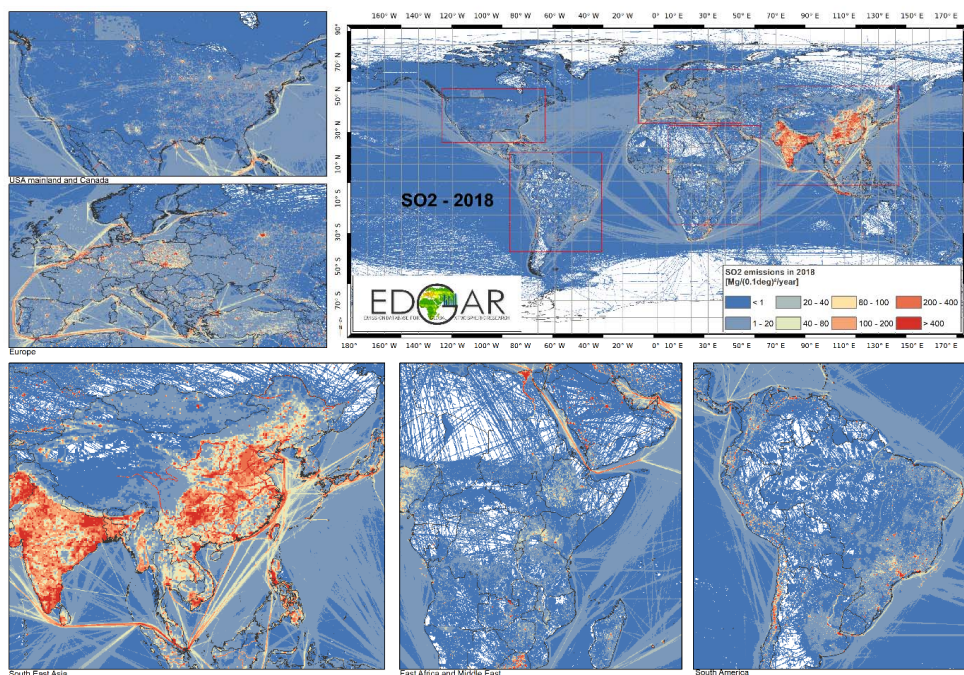
1 **Figure 2 – Time series of gaseous and particulate matter pollutants from HTAP_v3 by aggregated regions.**
 2 **Regional grouping follows the Intergovernmental Panel on Climate Change Sixth Assessment Report**
 3 **(IPCC AR6) definitions. Table S3 provides information on the regional belonging of each country to the**
 4 **IPCC AR6 regions.**

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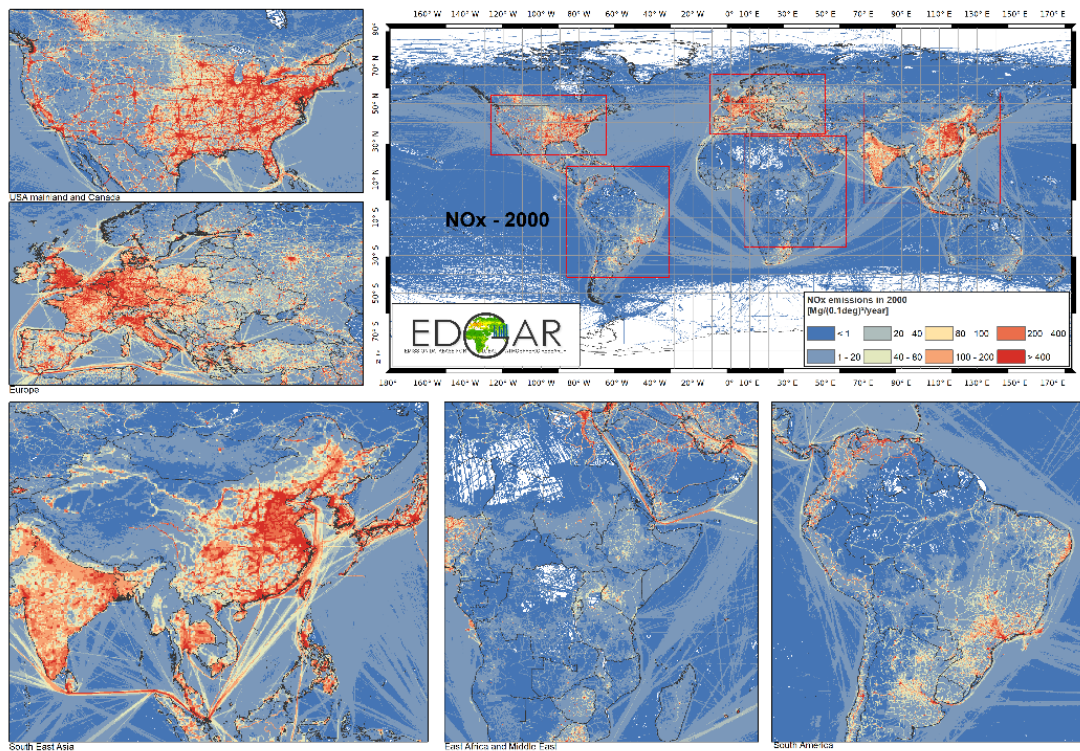


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2 **Figure 3 - 2018 sectoral (panel a) and regional (panel b) breakdown of air pollutant emissions from**
3 **HTAP_v3. At the top of each bar in panel a, total emissions for each pollutant are reported (in Mt).**

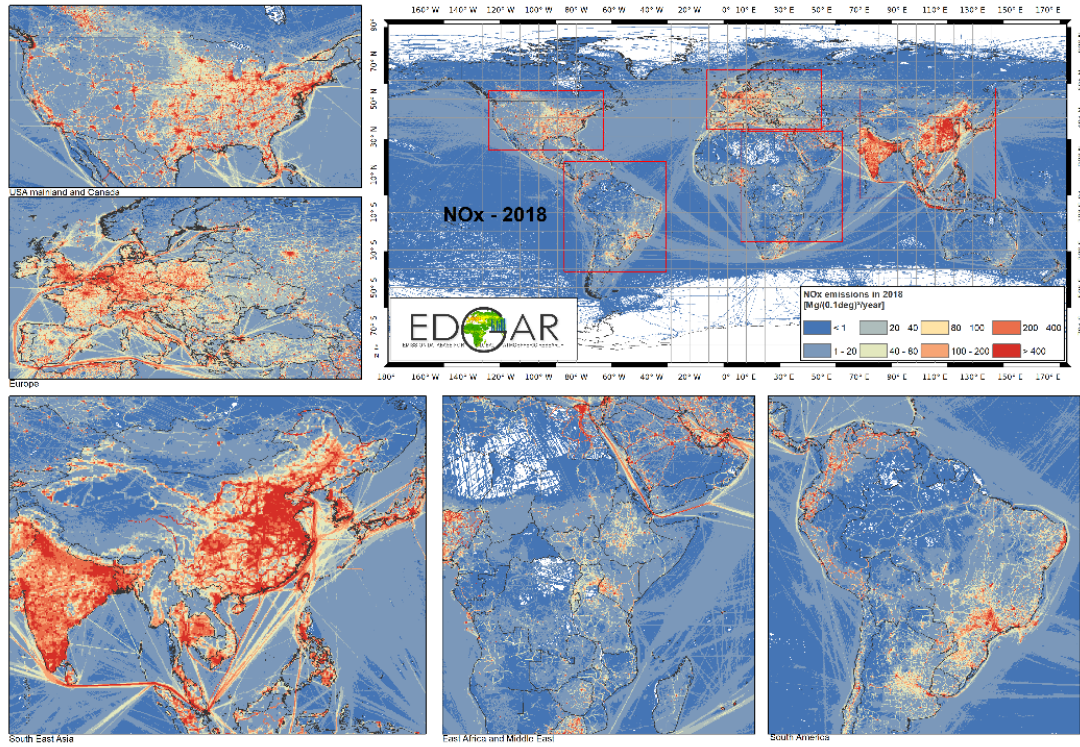


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Figure 4 – HTAP_v3 mosaic: SO₂ emission gridmaps for the year 2018.



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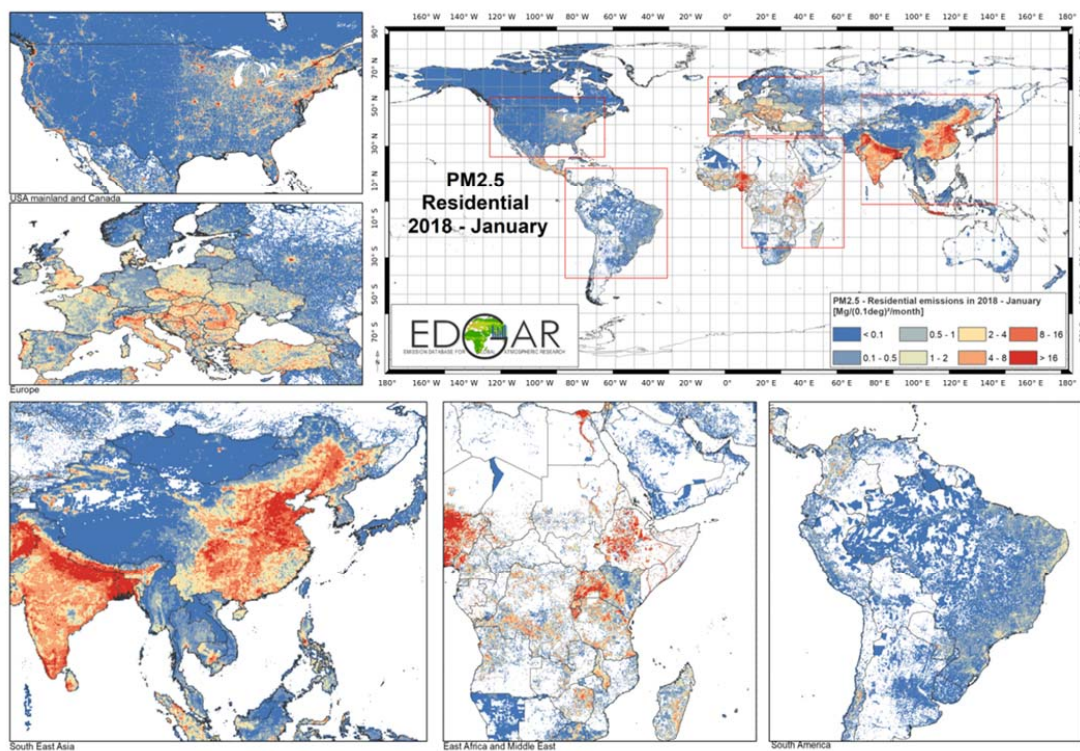


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2 **Figure 5 – HTAP_v3 mosaic: NO_x emission gridmaps in 2000 (top panel) and 2018 (bottom panel).**

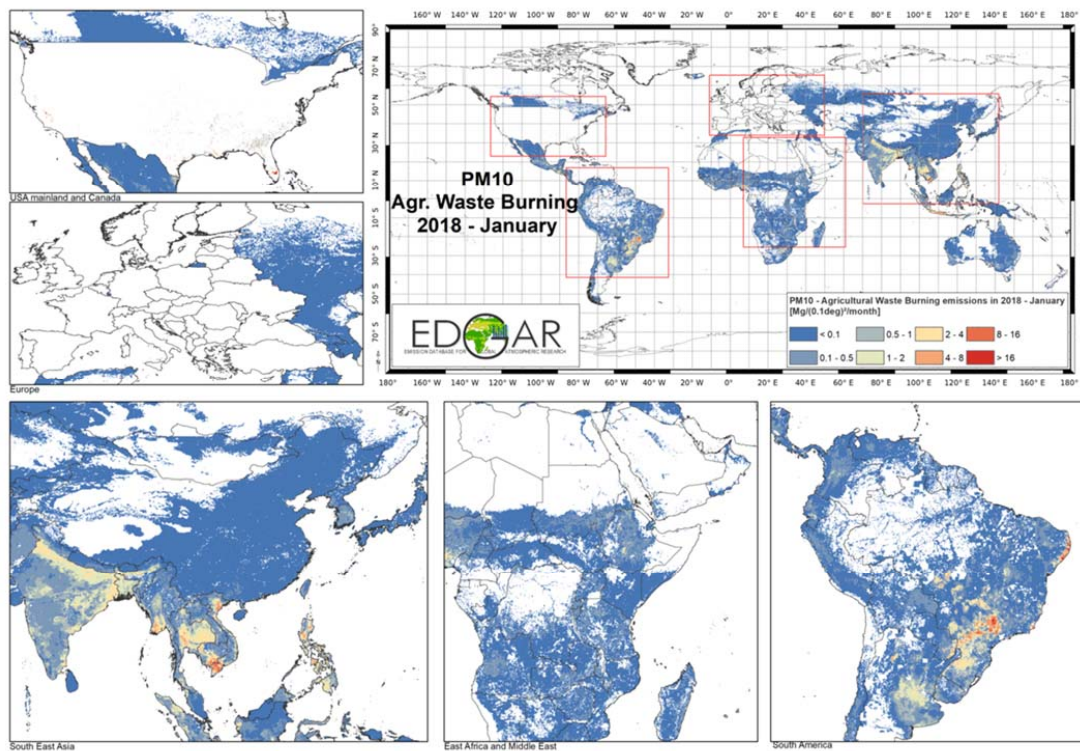


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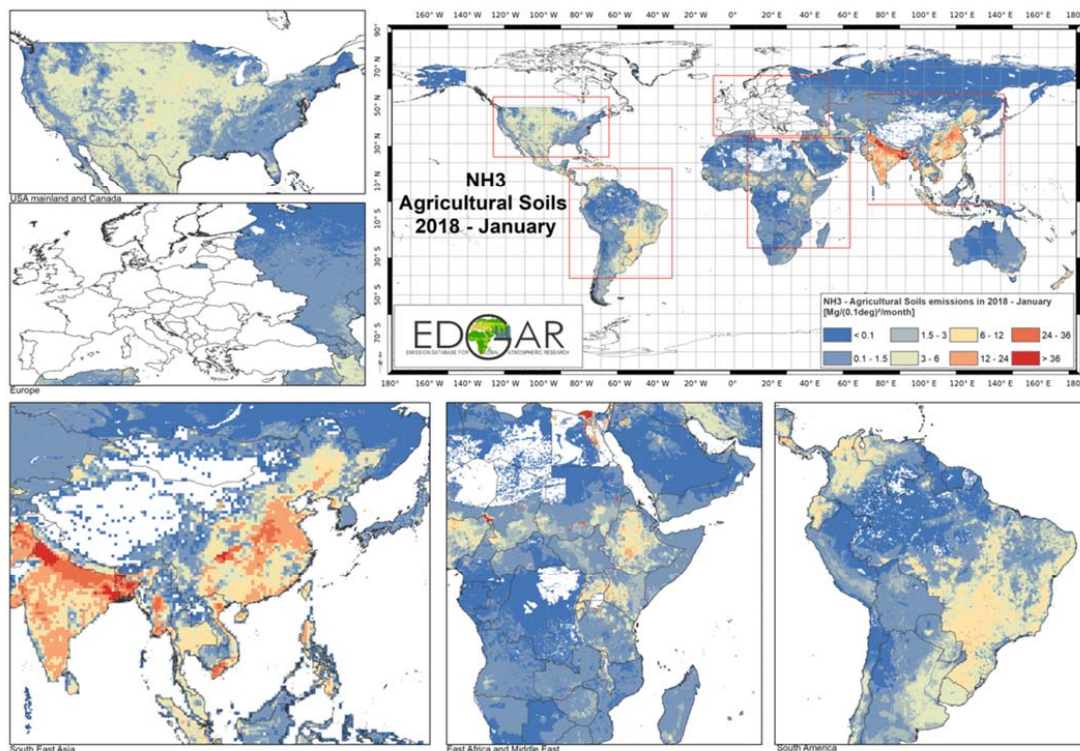
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3 **Figure 6 – HTAP_v3 mosaic: PM_{2.5} emissions from residential activities in January 2018.**



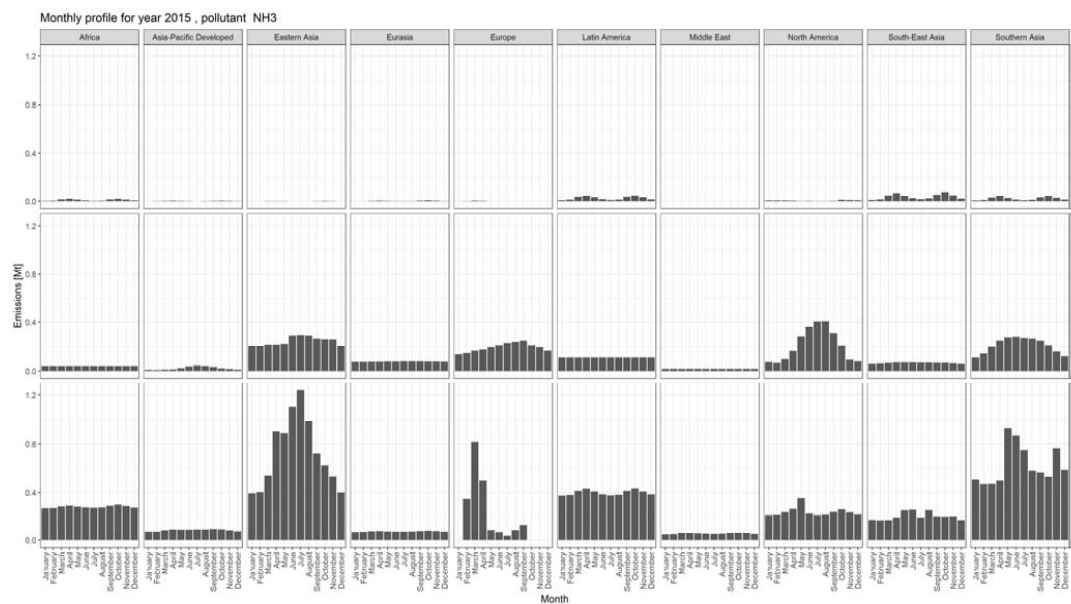
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2 **Figure 7 – HTAP_v3 mosaic: PM₁₀ emissions from agricultural waste burning in January 2018.**



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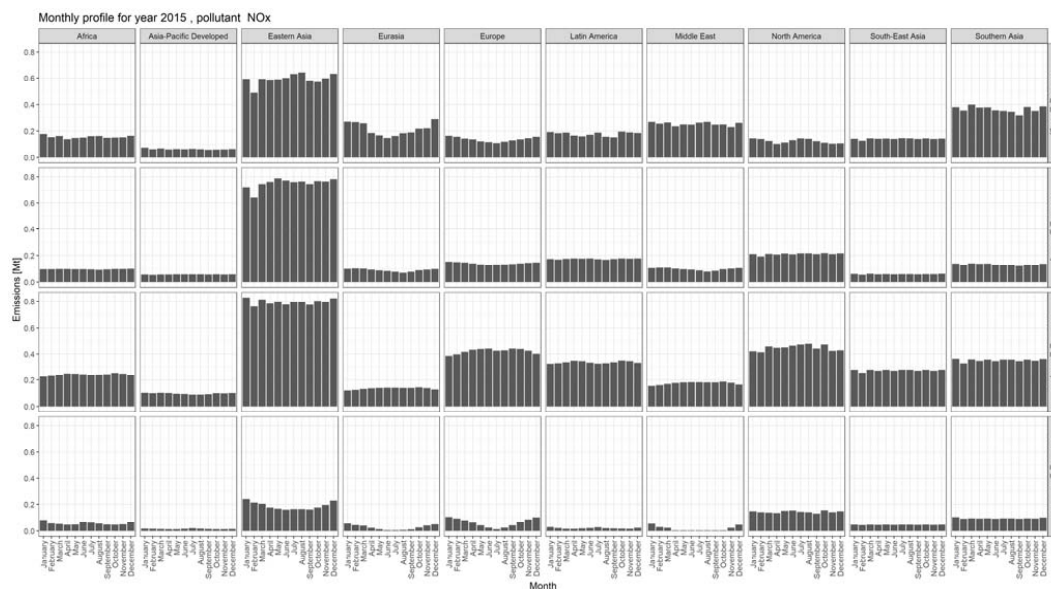
Figure 8 – HTAP_v3 mosaic: NH₃ emissions from agricultural soils activities in January 2018.



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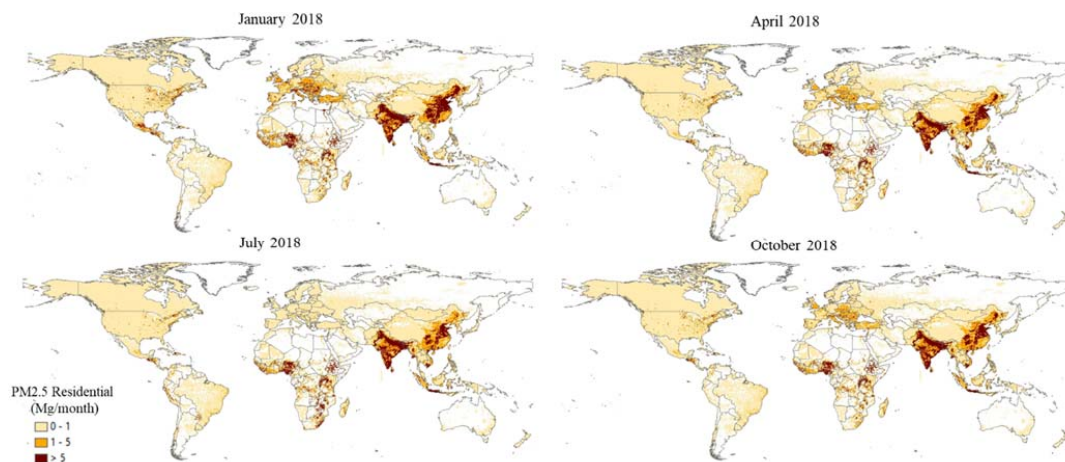
1 **Figure 9 – Monthly variability of NH₃ emissions for agriculture related activities for the different world**
2 **regions in 2015.**



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4 **Figure 10 – Monthly variability of NO_x emissions for relevant emission sectors for the different world**
5 **regions in 2015.**
6



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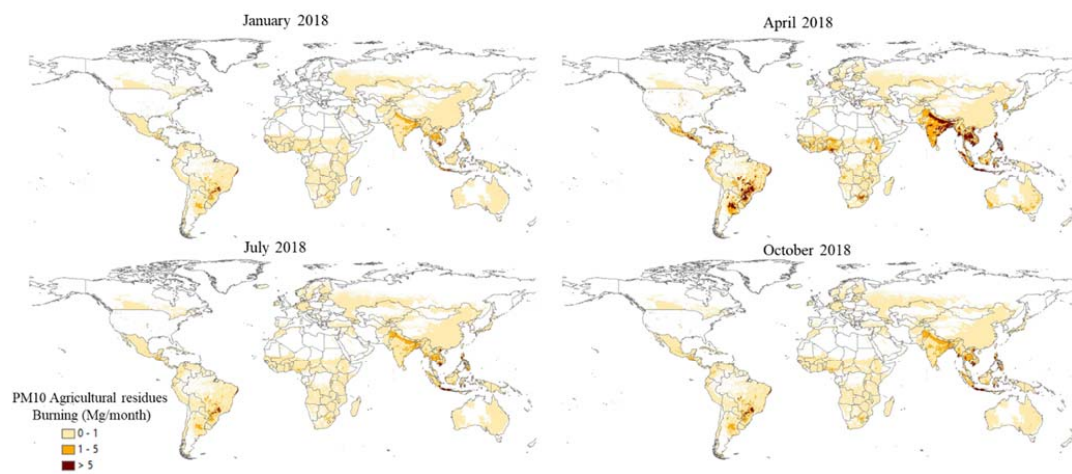


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Figure 11 – PM_{2.5} monthly emission maps from the residential sector in 2018 from HTAP_v3.

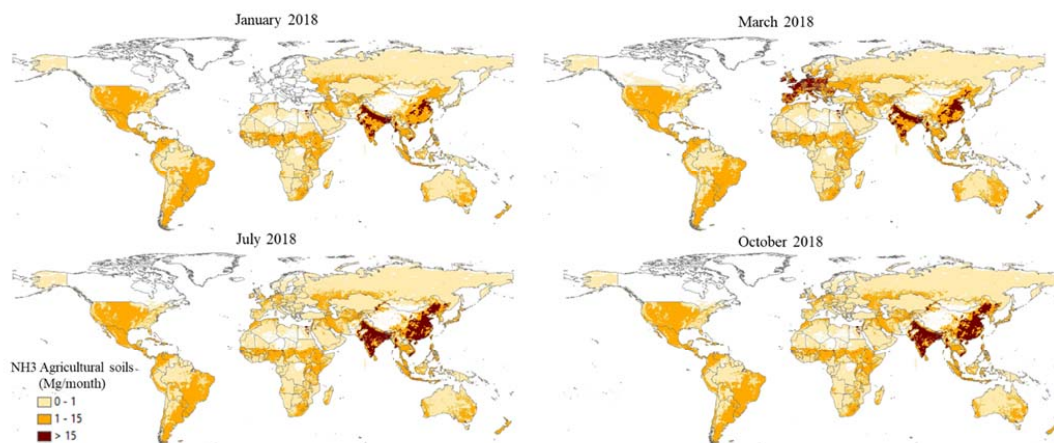
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Figure 12 – PM₁₀ monthly emission maps from agricultural residues burning in 2018 from HTAP_v3.



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2 **Figure 13 – NH₃ monthly emission maps from agricultural soils in 2018 from HTAP_v3.**

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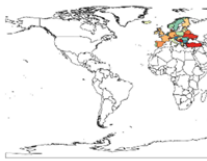



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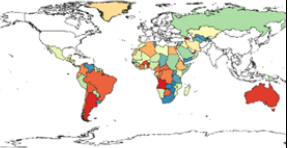
7 **Table 1 – Overview of data input to the HTAP_v3 emission mosaic.**

Data source	CAMS-REG-v5.1	US EPA	ECCC	REASv3.2.1
Type of data source	Country inventories as emission time series by sector and country and emission gridmaps as .csv files.	Country inventories	Country inventories as emission time series by sector and country and emission gridmaps as .NetCDF files.	Country inventories as emission gridmaps as text files.
Sectors coverage	All sectors, excluding international shipping and aviation (international and domestic).	All sectors, excluding international shipping and aviation (international and domestic).	All sectors, excluding agricultural waste burning, international shipping and aviation (international and domestic).	All sectors, excluding brake and tyre wear, domestic shipping, waste, agricultural waste burning, international shipping and aviation (international and domestic).
Temporal coverage	2000-2018	2002-2017	2000-2016	2000-2015+trends from MEIC over China for 2016, 2017, 2018
Temporal resolution	Annual emission gridmaps + monthly profiles	Monthly emission gridmaps	Monthly emission gridmaps	Monthly emission gridmaps



Spatial resolution	0.1°x0.1°	0.1°x0.1°	0.1°x0.1°	0.1°x0.1° (The original spatial resolution of REASv3.2.1 is 0.25°x0.25°. Assuming that emissions are equally distributed in the 0.25° cell, REASv3.2.1 data were converted to 0.1° cell and provided to HTAP v3)
Substances	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC
Geocoverage				
References				http://meicmodel.org

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Data source	CAPSS-KU	JAPAN (PM2.5EI and J-STREAM)	EDGARv6.1
Type of data source	Country inventory as emission time series by sector and country and emission gridmaps as .NetCDF files.	Country inventory as emission time series by sector and country and emission gridmaps as .NetCDF files.	Country inventory as emission time series by sector and country and emission gridmaps as .NetCDF files.
Sectors coverage	All sectors, excluding international shipping and aviation (international and domestic).	All sectors, excluding international shipping, domestic shipping and aviation (international and domestic).	All sectors, including international shipping and aviation (international and domestic)
Temporal coverage	2000-2018	2000-2017	2000-2018
Temporal resolution	Annual emission gridmaps + monthly profiles	Monthly emission gridmaps	Monthly emission gridmaps
Spatial resolution	0.1°x0.1°	0.1°x0.1°	0.1°x0.1°
Substances	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC	SO ₂ , NO _x , CO, NMVOC, NH ₃ , PM ₁₀ , PM _{2.5} , BC, OC
Geocoverage			
References			https://edgar.jrc.ec.europa.eu/dataset_ap61

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1 Table 2 – Definition of HTAP_v3 sectors and correspondence to IPCC codes.

HTAP_v3 main sectors	HTAP_v3 detailed sectors	Sector description	IPCC 1996 codes	IPCC 2006 codes
HTAP_1: International Shipping	HTAP_1: International Shipping	International water-born navigation.	1C2	1.A.3.d.i
HTAP_2: Aviation	HTAP_2.1: Domestic Aviation	Civil Aviation.	1A3a _{ii}	1.A.3.a.ii
	HTAP_2.2: International Aviation	International Aviation.	1A3a _i	1.A.3.a.i
HTAP_3: Energy	HTAP_3: Energy	Power generation.	1A1a	1.A.1.a
HTAP_4: Industry	HTAP_4.1: Industry	Industrial non-power large-scale combustion emissions and emissions of industrial processes. It includes: manufacturing, mining, metal, cement, chemical and fossil fuel fires.	1A2 + 2 + 5B	1A2 + 2 (excluding 2.D.3 + 2.E + 2.F + 2.G) + 7A
	HTAP_4.2: Fugitive	It includes oil and gas exploration and production and transmission, including evaporative emissions (mainly NMVOC).	1B + 1A1b + 1A1c _i + 1A1c _{ii} + 1A5b _{iii}	1.B + 1.A.1.b + 1.A.1.c.i + 1.A.1.c.ii + 1.A.5.b.i.i
	HTAP_4.3: Solvents	Solvents and product use.	3	2D3 + 2E + 2F + 2G
HTAP_5: Ground Transport	HTAP_5.1: Road Transport	Road Transport, combustion and evaporative emissions only.	1A3b (excluding resuspension)	1.A.3.b (excluding resuspension)
	HTAP_5.2: Brake and Tyre wear	Re-suspended dust from pavements or tyre and brake wear from road transport.	1A3b (resuspension only)	1.A.3.b (resuspension only)
	HTAP_5.3: Domestic shipping	Domestic shipping: inland waterways + domestic shipping.	1A3d ₂	1.A.3.d.ii
	HTAP_5.4: Other ground transport	Ground transport by pipelines and other ground transport of mobile machinery.	1A3c + 1A3e	1.A.3.c + 1.A.3.e.ii
HTAP_6: Residential	HTAP_6: Residential	Small-scale combustion, including heating, cooling, lighting, cooking and auxiliary engines, to equip residential, commercial buildings, service institutes, and agricultural facilities and fisheries.	1A4 + 1A5	1.A.4 + 1.A.5
HTAP_7: Waste	HTAP_7: Waste	Solid waste disposal and wastewater treatment.	6	4
HTAP_8: Agriculture	HTAP_8.1: Agricultural waste burning	Agricultural waste burning (excluding Savannah burning).	4F	3.C.1.b
	HTAP_8.2: Agriculture livestock	Livestock emissions, including manure management.	4B	3.A.2
	HTAP_8.3: Agriculture crops	Emissions from crops, fertilisers, and all agricultural soils activities.	4C + 4D	3.C.2 + 3.C.3 + 3.C.4 + 3.C.7

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2 **Table 3 – Overview of pollutant and sector provided by each inventory in HTAP_v3. Cells with N/A indicate**
 3 **that the emissions for those sectors were not provided and/or used in HTAP_v3 for a specific inventory,**
 4 **while gapfilled with the corresponding information from EDGARv6.1. The other cells represent the data**
 5 **availability for each sector and inventory. The color codes used for the pollutants refer to the data source:**
 6 **black color represents pollutant emissions provided by a specific inventory, red color indicates emissions**
 7 **gapfilled using EDGARv6.1 and violet color indicates combinations of sectors-pollutants available for**
 8 **specific regional inventories but not in EDGAR, which typically represent minor sources of emissions**
 9 **included in officially reported inventories. These minor sources are included in the HTAP_v3 mosaic.**

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Data provider	REAS v3.2.1	CAPS S-KU	JAPAN	ECCC	US EPA	CAMS-REG-v5.1	EDGARv 6.1
HTAP_1: International Shipping	N/A	N/A	N/A	N/A	N/A	N/A	All substances
HTAP_2.1: Domestic Aviation	N/A	N/A	N/A	N/A	N/A	N/A	All substances
HTAP_2.2: International Aviation	N/A	N/A	N/A	N/A	N/A	N/A	All substances
HTAP_3: Energy	All substances	All substances	BC, OC, NOx, NH3, CO, PM2.5, PM10, NMVOC, SO2	All substances	All substances	All substances	All substances
HTAP_4.1: Industry	All substances	All substances	BC, OC, NOx, NH3, CO, PM2.5, PM10, NMVOC, SO2	All substances	All substances	All substances	All substances
HTAP_4.2: Fugitive	All substances	BC, OC, NOx, NH3, CO, PM2.5, PM10, NMVOC, SO2	BC, OC, NOx, NH3, CO, PM2.5, PM10, NMVOC, SO2	All substances	All substances	All substances	All substances
HTAP_4.3: Solvents	NMVOC, NH3, PM10, PM2.5	NMVOC, NH3, PM10, PM2.5	NMVOC, NH3, PM10, PM2.5	NMVOC, NH3, PM10, PM2.5	CO, NOx, OC, NMVOC, NH3, PM10, PM2.5, SO2	NOx, NH3, CO, PM2.5, PM10, NMVOC, SO2	All substances
HTAP_5.1: Road Transport	All substances	All substances	All substances	All substances	All substances	All substances	All substances
HTAP_5.2: Brake and Tyre wear	N/A	BC, OC, NOx	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10



		NH ₃ , CO, PM _{2.5} , PM ₁₀ , NMVOC, OC, SO ₂			0, NMVOC		
HTAP_5.3: Domestic shipping	N/A	All substances	N/A	All substances	All substances	BC, OC, NO _x , NH ₃ , CO, PM _{2.5} , PM ₁₀ , NMVOC, SO ₂	All substances
HTAP_5.4: Other ground transport	All substances	All substances	BC, OC, NO _x , NH ₃ , CO, PM _{2.5} , PM ₁₀ , NMVOC, SO ₂	All substances	All substances	All substances	All substances
HTAP_6: Residential	All substances	All substances	BC, OC, NO _x , NH ₃ , CO, PM _{2.5} , PM ₁₀ , NMVOC, SO ₂	All substances	All substances	All substances	All substances
HTAP_7: Waste	N/A	All substances	All substances	All substances	All substances	All substances	All substances
HTAP_8.1: Agricultural waste burning	N/A	All substances	All substances	N/A	All substances	All substances	All substances
HTAP_8.2: Agriculture livestock	NH ₃ , NMVOC, OC, NO _x , PM ₁₀ , PM _{2.5}	NH ₃ , NMVOC, OC, NO _x , PM ₁₀ , PM _{2.5}	NH ₃ , NMVOC, NO _x , PM ₁₀ , PM _{2.5}	NH ₃ , NMVOC, NO _x , PM ₁₀ , PM _{2.5} , BC, OC	NH ₃ , NMVOC, NO _x , PM ₁₀ , PM _{2.5} , BC, OC	NH ₃ , NMVOC, NO _x , PM ₁₀ , PM _{2.5} , OC	NH ₃ , NMVOC, NO _x , PM ₁₀ , PM _{2.5}
HTAP_8.3: Agriculture_crops	NH ₃ , NO _x , PM ₁₀ , PM _{2.5}	NH ₃ , NO _x , PM ₁₀ , PM _{2.5} , BC, OC	NH ₃ , NO _x , PM ₁₀ , PM _{2.5}	NH ₃ , NO _x , PM ₁₀ , PM _{2.5} , BC, NMVOC, OC	NH ₃ , NO _x , PM ₁₀ , PM _{2.5} , BC, OC, CO, NMVOC, SO ₂	NH ₃ , NO _x , PM ₁₀ , PM _{2.5} , SO ₂ , CO, OC, NMVOC	NH ₃ , NO _x , PM ₁₀ , PM _{2.5}

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