



1 **HTAP\_v3 emission mosaic: a global effort to tackle air quality**  
2 **issues by quantifying global anthropogenic air pollutant**  
3 **sources**

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31 **Abstract.** This study, performed under the umbrella of the Task Force on Hemispheric Transport of Air Pollution  
32 (TF-HTAP), responds to the need of the global and regional atmospheric modelling community of having a mosaic  
33 emission inventory of air pollutants that conforms to specific requirements: global coverage, long time series,  
34 spatially distributed emissions with high time resolution, and a high sectoral resolution. The mosaic approach of  
35 integrating official regional emission inventories based on locally reported data, with a global inventory based on  
36 a globally consistent methodology, allows modellers to perform simulations of a high scientific quality while also  
37 ensuring that the results remain relevant to policymakers.

38 HTAP\_v3, an ad-hoc global mosaic of anthropogenic inventories, has been developed by integrating official  
39 inventories over specific areas (North America, Europe, Asia including Japan and Korea) with the independent  
40 Emissions Database for Global Atmospheric Research (EDGAR) inventory for the remaining world regions. The  
41 results are spatially and temporally distributed emissions of SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>, Black  
42 Carbon (BC), and Organic Carbon (OC), with a spatial resolution of 0.1 x 0.1 degree and time intervals of months



1 and years covering the period 2000-2018 (DOI 10.5281/zenodo.7516361,  
2 [https://edgar.jrc.ec.europa.eu/dataset\\_htap\\_v3](https://edgar.jrc.ec.europa.eu/dataset_htap_v3)). The emissions are further disaggregated to 16 anthropogenic  
3 emitting sectors. This paper describes the methodology applied to develop such an emission mosaic, reports on  
4 source allocation, differences among existing inventories, and best practices for the mosaic compilation. One of  
5 the key strengths of the HTAP\_v3 emission mosaic is its temporal coverage, enabling the analysis of emission  
6 trends over the past two decades. The development of a global emission mosaic over such long time series  
7 represents a unique product for global air quality modelling and for better-informed policy making, reflecting the  
8 community effort expended by the TF-HTAP to disentangle the complexity of transboundary transport of air  
9 pollution.

## 10 1 Introduction

11 Common international efforts have procured an agreement to reduce global air pollutant emissions. For this  
12 purpose, the United Nations Economic Commission for Europe (UNECE) Convention on Long Range  
13 Transboundary Air Pollution (CLRTAP) and the Task Force on Hemispheric Transport of Air Pollution (TF-  
14 HTAP) have been instrumental in developing the understanding of intercontinental transport of air pollution and  
15 thus contributing to the reduction of key pollutants in Europe and North America.

16 The success of CLRTAP is based on meeting strict reduction targets for pollutant releases. Therefore, evaluating  
17 the resulting implications of these reductions requires an ongoing improvement of global emission inventories in  
18 terms of emission updating and of methodological refinements. These aspects are instrumental to gain  
19 understanding of transboundary air pollution processes and drivers and to measure the effectiveness of emissions  
20 reduction and air quality mitigation policies. New guidance is available to achieve further emission reductions  
21 across all emitting sectors. For example, the 2019 establishment of the Task Force for International Cooperation  
22 on Air Pollution, which is intended to promote international collaboration for preventing and reducing air pollution  
23 and improving air quality globally (UNECE, 2021). As part of the ongoing effort by CLRTAP to reduce emissions  
24 and to set out more effective and accountable mitigation measures, the 2005 Gothenburg Protocol (UNECE, 2012)  
25 has been revised, including the review of the obligations in relation to emission reductions and mitigation  
26 measures (e.g., black carbon and ammonia) and the review of the progress towards achieving the environmental  
27 and health objectives of the Protocol.

28  
29 The Task Force on Hemispheric Transport of Air Pollution (TF-HTAP) of the Convention has a mandate to  
30 promote the scientific understanding of the intercontinental transport of air pollution to and from the UNECE area  
31 (<https://unece.org/geographical-scope>), to quantify its impacts on human health, vegetation and climate, and to  
32 identify emission mitigation options that will shape future global policies.

33 This paper describes and discusses a consistent global emission inventory of air pollutants emitted by  
34 anthropogenic activities. This important tool has been developed to assess the contribution of anthropogenic air  
35 pollution emission sources within and outside the UNECE-area through atmospheric modelling. This inventory  
36 has been compiled based on officially reported emissions, and an independent global inventory where officially  
37 reported emissions are not used. This harmonised emissions “mosaic” dataset, hereafter referred to as the  
38 HTAP\_v3, contains annual and monthly:

- 39 - emission time series (from 2000 to 2018) of SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>, BC, OC by  
40 emitting sector and country, and
- 41 - spatially distributed emissions on a global grid with spatial spacing of 0.1x0.1 degree.

42 A global emission mosaic that gathers officially reported air pollutant emissions is essential since it aggregates  
43 the officially available information on air pollution as input for atmospheric modellers and policy evaluations.  
44 The creation of a global emission mosaic requires the harmonisation of several data sources, detailed analysis of  
45 contributing sectors for the different input inventories, development of data quality control procedures, and a  
46 robust and consistent gap-filling methodology when lacking information. The development of HTAP\_v3 builds  
47 upon the previous experience of the HTAPv1 (Janssens-Maenhout et al., 2012) and HTAPv2.2 (Janssens-  
48 Maenhout et al., 2015) global inventories. HTAP\_v3, as requested by the TF-HTAP modelling community,  
49 provides a more refined sectoral disaggregation compared to the previous HTAP emission mosaics. It also  
50 includes tools ([https://edgar.jrc.ec.europa.eu/htap\\_tool/](https://edgar.jrc.ec.europa.eu/htap_tool/)) that allows the extraction of emission data over selected  
51 domains (detailed later in section 4).

52 This paper describes the development the HTAP\_v3 database as a global anthropogenic air pollutant emissions  
53 inventory mosaic for the period 2000-2018. The HTAP\_v3 mosaic has been composed by integrating official,  
54 spatially distributed emissions data from CAMS-REF-v5.1 (Kuenen et al., 2022), US EPA (U.S. Environmental  
55 Protection Agency, 2021a, b), Environment and Climate Change Canada (ECCC) (NPRI, 2017), REAS, CAPSS-  
56 KU, and JAPAN (<https://www.env.go.jp/air/osen/pm/info.html>) (Kurokawa and Ohara, 2020; Chatani et al.,



1 2018; Chatani et al., 2020) inventories. As the information gathered from the official reporting covers only part  
2 of the globe, HTAP\_v3 has been completed using emissions from the Emissions Database for Global Atmospheric  
3 Research (EDGAR) version 6.1 ([https://edgar.jrc.ec.europa.eu/dataset\\_ap61](https://edgar.jrc.ec.europa.eu/dataset_ap61)).

4 One of the key strengths of the HTAP\_v3 emission mosaic is the wide temporal coverage of the emissions,  
5 spanning the 2000-2018 period, enabling the analysis of emission trends over the past two decades. The  
6 development of a global emission mosaic over such long time series represents a unique product for air quality  
7 modelling and for better-informed policy making, reflecting the effort of the TF-HTAP community to improve  
8 understanding of the transboundary transport of air pollution. The year 2000 was chosen as the baseline year since  
9 it often represents the year from which complete datasets of annual air pollutant emissions can be generated. It  
10 also represents a turning point for several emerging economies (e.g., China) and the strengthening of mitigation  
11 measures in historically developed regions (e.g., EU, USA, etc.).

12 The two previous HTAP emission mosaics had limited temporal coverage. HTAPv1 covered the period 2000-  
13 20005 with annual resolution ([https://edgar.jrc.ec.europa.eu/dataset\\_htap\\_v1](https://edgar.jrc.ec.europa.eu/dataset_htap_v1), (Janssens-Maenhout et al., 2012)),  
14 while HTAPv2.2 covered two recent years (2008 and 2010), but with monthly resolution (Janssens-Maenhout et  
15 al., 2015) ([https://edgar.jrc.ec.europa.eu/dataset\\_htap\\_v2](https://edgar.jrc.ec.europa.eu/dataset_htap_v2)). However, the needs of the TF-HTAP modelling  
16 community are continuously evolving to both foster forward-looking air quality science and produce more fit-for-  
17 purpose analyses in support of efficient policy making. HTAP\_v3 therefore not only covers the time period of the  
18 previous HTAP phases, but also extends it forward by almost a decade, to provide the most up-to-date picture of  
19 global air pollutant emission trends. Another distinguishing feature of the HTAPv3 mosaic is a considerably  
20 higher sectoral resolution than previous iterations of the HTAP mosaic inventories (section 2.2), enabling more  
21 policy-relevant use of the inventory.

22 The methodology and data sources for the HTAP\_v3 emission mosaic are described in section 2. The long-time  
23 coverage of two decades, allows comprehensive trend analysis (see section 3), the HTAP\_v3 data format and  
24 data-set access are presented in section 4 and conclusions are provided in section 5.

25

## 26 **2 HTAP\_v3 emission mosaic overview: data sources, coverage, and methodology**

### 27 **2.1 Data input**

28 The HTAP\_v3 mosaic is a database of monthly- and sector-specific global air pollutant emission gridmaps  
29 developed by integrating spatially explicit regional information from recent officially-reported national or regional  
30 emission inventories. Data from six main regional inventories were integrated into HTAP\_v3, which covered only  
31 North America, Europe, and a portion of Asia (including Japan, China, India, and South Korea) (Fig.1). The  
32 geographical domain covered by each of these inventories is depicted in Fig. 1, while further details on each  
33 contributing inventory are presented in section 2.3. The emissions for all other countries, international shipping  
34 and aviation (international and domestic) have been retrieved from the Emissions Database for Global  
35 Atmospheric Research (EDGARv6.1, [https://edgar.jrc.ec.europa.eu/dataset\\_ap61](https://edgar.jrc.ec.europa.eu/dataset_ap61)) as represented by the grey  
36 areas in Fig.1. Depending on the pollutant, more than half of global emissions are provided by region-specific  
37 inventories, while the remaining contribution is derived from the EDGAR global inventory as reported in the bar  
38 graph of Fig.1, where the share of each individual inventory to global emissions is represented. For all pollutants,  
39 the Asian domain is contributing most to global emissions, hence the importance of having accurate emission  
40 inventories for this region.

41 Table 1 provides an overview of all data providers, in terms of geographical and temporal coverage, data format,  
42 and sectoral and pollutant data availability. Table 2 defines the HTAP\_v3 sectors and corresponding IPCC codes.  
43 Table 3 further details the sector-pollutant data availability for each inventory and the gap-filling approach  
44 required for some sectors and pollutants.

### 45 **2.2 Pollutant, spatial, temporal and sectoral coverage**

46 The HTAP\_v3 emission mosaic helps to address the transboundary role of air pollutants by providing a key input  
47 for atmospheric modellers and supporting the evaluation of environmental impact analyses for poor air quality.  
48 For this reason, HTAP\_v3 provides global 0.1 x 0.1 degree emission gridmaps for all air pollutants and specifically  
49 for acidifying and eutrophying gases (such as SO<sub>2</sub>, NH<sub>3</sub>, NO<sub>x</sub>), ozone precursors (NMVOC, CO, NO<sub>x</sub>), and  
50 primary particulate matter (PM<sub>10</sub>, PM<sub>2.5</sub>, BC, OC).



1 Emissions from each officially-reported inventory were submitted to HTAP on 0.1 x 0.1 degree regional gridmaps.  
2 Spatial allocation was performed to these gridmaps for each sector by each inventory group using the best  
3 available set of subsector spatial surrogate fields used by each group (e.g., <https://www.cmascenter.org/sa-tools>).  
4 EDGARv6.1 global gridmaps are also on a 0.1 x 0.1 degree grid.

5 Compared to the two previous HTAP emission mosaics, HTAP\_v3 input emission gridmaps were provided with  
6 monthly time distributions to better reflect the regional seasonality of sector specific emissions (e.g., household,  
7 power generation, and agricultural activities). Information on emission peaks over certain months of the year is  
8 also a useful information for the development of territorial policies to mitigate localised emission sources in space  
9 and time (e.g., emissions from residential heating over winter months, agricultural residue burning, etc.).

10 The HTAP\_v3 mosaic provides emissions for gaseous and particulate matter air pollutants arising from all  
11 anthropogenic emitting sectors except for wildfires and savannah burning, which represent major sources of  
12 particulate matter and CO emissions. Wildfires and savannah burning are not included in the current mosaic since  
13 community efforts are ongoing to tackle these sources specifically. Modellers can find these additional sources  
14 on several publicly available global wildfire emission datasets compiled based on the best available scientific  
15 knowledge, such as the Global Fire Emission Database (GFED, <https://www.globalfiredata.org/>) or the Global  
16 Wildfire Information System (GWIS, <https://gwis.jrc.ec.europa.eu/>). When using satellite retrieved emissions  
17 from fires, they should be treated with caution to avoid double counting the emissions released by e.g. agricultural  
18 crop residue burning activities.

19 HTAP\_v3 provides emissions at higher sectoral disaggregation than previous HTAP experiments<sup>1</sup> to better  
20 understand drivers of emission trends and the effectiveness of sector-specific policy implementation. Emissions  
21 from 16 sectors are provided by the HTAP\_v3 mosaic, namely: International Shipping; Domestic Shipping;  
22 Domestic Aviation; International Aviation; Energy; Industry; Fugitives; Solvent Use; Road Transport; Brake and  
23 Tyre Wear; Other Ground Transport; Residential; Waste; Agricultural Waste Burning; Livestock; and Agricultural  
24 Crops. Further details on the sector definitions as well as their correspondence with the IPCC codes (IPCC, 1996,  
25 2006) are provided in Table 2. The selection of the number of sectors was constrained by the sectoral  
26 disaggregation of the input inventories (see Table S1). Table 3 provides the complete overview of the emission  
27 data provided by each inventory group indicating the pollutants covered for each sector and eventual gap-filling  
28 information included using the EDGARv6.1 data.

## 29 2.3 Inventory overviews

30 In the following sub-sections, details are provided on each officially-reported inventory used to construct the  
31 HTAP\_v3 emission mosaic.

### 32 2.3.1 CAMS-REG-v5.1 inventory

33 The CAMS-REG-v5.1 emission inventory was developed to support air pollutant and greenhouse gas modelling  
34 activities at the European scale. The inventory builds largely on the official reported data to the UN Framework  
35 Convention on Climate Change (UNFCCC) for greenhouse gases (for CO<sub>2</sub> and CH<sub>4</sub>), and the Convention on  
36 Long-Range Transboundary Air Pollution (CLRTAP) for air pollutants. For the latter, data are collected for NO<sub>x</sub>,  
37 SO<sub>2</sub>, CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub> and PM<sub>2.5</sub>, including all major air pollutants. For each of these pollutants, the  
38 emission data are collected at the sector level at which these are reported for the time series 2000-2018 for each  
39 year and country. The CAMS-REG inventory covers UNECE-Europe, extending eastward until 60°E, therefore  
40 including the European part of Russia. For some non-EU countries, the reported data are found to be partially  
41 available or not available at all. In other cases, the quality of the reported data is found to be insufficient, i.e. with  
42 important data gaps or following different formats or methods. In this case, emission data from the IIASA GAINS  
43 model instead (IIASA, 2018) are used. This model is the main tool used to underpin pan-European and EU level  
44 air quality policies such as the UNECE Convention on Long Range Transboundary Air Pollution (UNECE, 2012)  
45 and the EU National Emission reduction Commitments Directive (European Commission, 2016).

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<sup>1</sup>HTAPv1 covered 10 broad emission sectors (Aircraft, Ships, Energy, Industry Processes, Ground Transport, Residential, Solvents, Agriculture, Agriculture Waste Burning, and Waste), while even broader sectoral emissions were provided in HTAPv2.2 (Air, Ships, Energy, Industry, Transport, Residential (including waste), and Agriculture (only for NH<sub>3</sub>)).



1 After collecting all the emission data from reporting and GAINS, the source sectors are harmonised, distinguishing  
2 around 250 different subsectors. For each detailed sector, a speciation is applied to the PM<sub>2.5</sub> and PM<sub>10</sub> emissions,  
3 distinguishing elemental carbon (representing BC in the HTAP\_v3 inventory), organic carbon and other non-  
4 carbonaceous emissions for both the coarse (2.5-10 µm) and fine (<2.5 µm) mode.

5 A consistent spatial resolution is applied across the entire domain, where a specific proxy is selected for each  
6 subsector to spatially distribute emissions, including for instance the use of point source emissions, e.g., from the  
7 European Pollutant Release and Transfer Register (E-PRTR), complemented with additional data from the  
8 reporting of EU Large Combustion Plants (European Commission, 2001) and the Platts/WEPP commercial  
9 database for power plants (Platts, 2017). Road transport emissions are spatially disaggregated using information  
10 from OSM (Open Street Map, 2017), combined with information on traffic intensity in specific road segments  
11 from OTM (OpenTransportMap, 2017). Agricultural livestock emissions are spatially distributed using global  
12 gridded livestock numbers (FAO, 2010). Furthermore, CORINE land cover (Copernicus Land Monitoring  
13 Service, 2016) and population density are other key spatial distribution proxies.

14 After having spatially distributed the data, the ~ 250 different source categories are aggregated to fit with the  
15 HTAP\_v3 sector classification (Table S1). CAMS-REG-v5.1 is an update of an earlier version, CAMS-REG-v4.2  
16 and based on the 2020 submissions to cover the years 2000-2018. A detailed description of the CAMS-REG-v4.2  
17 inventory is provided in Kuenen et al. (2022).

18 The data are provided as gridded annual totals at a resolution of 0.05°x0.1° (lat-lon). Along with the grids,  
19 additional information is available including height profiles as well as temporal profiles to break down the annual  
20 emissions into hourly data (monthly profiles, day-of-the-week profiles and hourly profiles for each day).  
21 Furthermore, the CAMS-REG inventory provides dedicated speciation profiles for NMVOC per year, country  
22 and sector.

### 23 2.3.2 US EPA inventory

24 Emissions estimates for the United States were based primarily on estimates produced for the EPA's Air Quality  
25 Time Series Project (EQUATES), which generated a consistent set of modelled emissions, meteorology, air  
26 quality, and pollutant deposition for the United States spanning the years 2002 through 2017  
27 (<https://www.epa.gov/cmaq/equates>). For each sector, a consistent methodology was used to estimate emissions  
28 for each year in the 16-year period, in contrast to the evolving methodologies applied in the triennial U.S. National  
29 Emissions Inventories (NEIs) produced over that span. The HTAPv3 time series was extended back one year to  
30 2001 and forward one year to 2018 using country, sector, and pollutant specific trends from EDGARv6.1.

31

32 Emissions estimates were calculated for more than 8000 Source Classification Codes grouped into 101 sectors  
33 and then aggregated to the 16 HTAP\_v3 emission sectors. The 2017 NEI (U.S. Environmental Protection Agency,  
34 2021a) served as the base year for the time series. For each sector, emissions estimates were generated for  
35 previous years using one of four methods: 1) applying new methods to create consistent emissions for all years,  
36 2) scaling the 2017 NEI estimates using annual sector-specific activity data and technology information at the  
37 county level, 3) using annual emissions calculated consistently in previous NEIs and interpolating to fill missing  
38 years, and 4) assuming emissions were constant at 2017 levels constant. The assumption of constant emissions  
39 was applied to a very limited number of sources. Foley et al. (submitted, 2022) provides a detailed explanation of  
40 the assumptions used for each sector.

41 Emissions from electric generating units were estimated for individual facilities, combining available hourly  
42 emissions data for units with continuous emissions monitors (CEMs) and applying regional fuel-specific profiles  
43 to units without CEMS. On-road transport and non-road mobile emissions were estimated using emission factors  
44 from the MOVES v3 model (U.S. Environmental Protection Agency, 2021b). A complete MOVES simulation  
45 was completed only for the NEI years with national adjustment factors applied for years plus or minus one from  
46 the NEI year. For California, emission factors for all on-road sources for all years were based on the California  
47 Air Resources Board Emission Factor Model (EMFAC) ([https://ww2.arb.ca.gov/our-work/programs/mobile-  
48 source-emissions-inventory/](https://ww2.arb.ca.gov/our-work/programs/mobile-source-emissions-inventory/)). New non-road emissions estimates for Texas were provided by the Texas  
49 Commission on Environmental Quality. Emissions from oil and gas exploration and production were calculated  
50 using point source specific data and the EPA Oil and Gas Tool (U.S. Environmental Protection Agency, 2021a),  
51 incorporating year-specific spatial, temporal, and speciation profiles. Residential wood combustion estimates  
52 were developed with an updated methodology incorporated into the 2017 NEI and scaled backward to previous



1 years using a national activity as a scaling factor. Solvent emissions were estimated using the Volatile Chemical  
2 Product (VCPy) framework of Seltzer et al. (2021). Emissions from livestock waste were calculated with revised  
3 annual animal counts to address missing data and methodological changes over the period. Emissions for  
4 agricultural burning were developed using a new suite of activity data with the same methodology and input data  
5 sets for all years except 2002. County-level estimates were only available for 2002 because activity data based  
6 on satellite information was not yet available. Emissions for forest wildfires, prescribed burns, grass and  
7 rangeland fires were also calculated in EQUATES but not included in the HTAP\_v3 data. For EQUATES,  
8 fugitive dust emissions (e.g., unpaved road dust, coal pile dust, dust from agricultural tilling) were reduced to  
9 account for precipitation and snow cover by grid cell. For use in HTAP\_v3, however, no meteorological  
10 adjustments (which decrease annual PM<sub>10</sub> emissions by about 75% on average) were applied to fugitive dust  
11 emissions. Wind-blown fugitive dust emissions are not included in the estimates for other regions in the HTAP\_v3  
12 mosaic

13 Non-point source emissions were allocated spatially based on a suite of activity surrogates (e.g. population, total  
14 road miles, housing, etc.), many of which are sector specific. The spatial allocation factors were calculated for  
15 the 0.1 degree grid used by EDGARv6.1 with no intermediate re-gridding. The spatial allocation factors for all  
16 sectors were held constant for the entire time series except for oil and gas sectors which were year-specific.  
17 Depending on the sector, either 2017-based or 2014-based surrogates were developed for the same sectors as in  
18 the EQUATES.

19 Emissions from the US EPA inventory were provided from 2002-2017 (Table 1). Emissions for the year 2018  
20 were estimated applying country, sector and pollutant specific trends from EDGAR and for the 2000 and 2001 to  
21 complete the entire time series. Table S1 provides an overview about the US EPA inventory sector mapping to  
22 the HTAP\_v3 sectors.

### 23 2.3.3 Environment and Climate Change Canada (ECCC) inventory

24 The Canadian emissions inventory data were obtained from 2018-released edition of Canada's Air Pollutant  
25 Emissions Inventory (APEI) originally compiled by the Pollutant Inventories and Reporting Division (PIRD) of  
26 Environment and Climate Change Canada (ECCC) (APEI, 2018). This inventory contains a comprehensive and  
27 detailed estimate of annual emissions of seven criteria air pollutants (SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOC, NH<sub>3</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>)  
28 at the national and provincial/territorial level for each year for the period from 1990 to 2016. The APEI inventory  
29 was developed based on a bottom-up approach for facility-level data reported to the National Pollutant Release  
30 Inventory (NPRI) (NPRI, 2017), as well as an in-house top-down emission estimates based on source-specific  
31 activity data and emissions factors. In general, methodologies used to estimate Canadian emissions are consistent  
32 with those developed by the U.S. EPA (EPA, 2009) or those recommended in the European emission inventory  
33 guidebook (EMEP/EEA, 2013). These methods are often further adjusted by PIRD to reflect the Canadian climate,  
34 fuels, technologies and practices.

35 To prepare emissions in the desired HTAP classification, the APEI sector emissions were first mapped to the  
36 United Nations Economic Commission for Europe (UNECE) Nomenclature for Reporting (NFR) categories,  
37 which involved dividing the sector emissions into their combustion and process components. The NFR categories  
38 were then mapped to the HTAP 16 sector categories provided in the sector disaggregation scheme guide. Table  
39 S1 provides an overview of ECCC sector mapping to the HTAP\_v3 sectors.

40 The HTAP-grouped APEI inventory emissions files were further processed by the Air Quality Policy-Issue  
41 Response (REQA) Section of ECCC to prepare the air-quality-modelling version of inventory files in the standard  
42 format (i.e., FF10 format) supported by the U.S EPA emissions processing framework. To process emissions into  
43 gridded, speciated and total monthly values, a widely-used emissions processing system called the Sparse Matrix  
44 Operator Kernel Emissions (SMOKE) model, version 4.7 (UNC, 2019) was used. As part of the preparation for  
45 SMOKE processing, a gridded latitude-longitude North American domain at 0.1 x 0.1 degree resolution was  
46 defined with 920 columns and 450 rows covering an area of -142W to -50W and 40N to 85N. The point-source  
47 emissions in the APEI include latitude and longitude information so those sources were accurately situated in the  
48 appropriate grid cell in the Canadian HTAP gridded domain. However, to allocate provincial-level non-point  
49 source emissions into this domain, a set of gridded spatial surrogate fields was generated for each province from  
50 statistical proxies, such as population, road network, dwellings, crop distributions, etc. Over 80 different surrogate  
51 ratio files were created using the 2011 Canadian census data obtained from Statistics Canada website  
52 (<https://www12.statcan.gc.ca/census-recensement/2011/index-eng.cfm>) and other datasets, such as the Canadian  
53 National Road Network (<https://open.canada.ca/data/en/dataset/3d282116-e556-400c-9306-ca1a3cada77f>).



1 To map the original APEI inventory species to the HTAP's desired list of species, PM speciation profiles from  
2 the SPECIATE version 4.5 database (EPA, 2016) were used to calculate source-type-specific EC and OC  
3 emissions. As a final step in SMOKE processing, the monthly emissions values were estimated using a set of  
4 sector-specific temporal profiles developed and recommended by the U.S. EPA (Sassi, 2021). For the point  
5 sources the NPRI annually reported monthly emissions proportions were applied. Emissions for the years 2017  
6 and 2018 were calculated by applying sector- and pollutant-specific trends from EDGAR.

#### 7 **2.3.4 REASv3.2.1 inventory**

8 The Regional Emission inventory in ASia (REAS) series have been developed for providing historical trends of  
9 emissions in the Asian region including East, Southeast, and South Asia. REASv3.2.1, the version used in  
10 HTAP\_v3, runs from 1950 to 2015. REASv3.2.1 includes emissions of SO<sub>2</sub>, NO<sub>x</sub>, CO, NMVOCs, NH<sub>3</sub>, CO<sub>2</sub>,  
11 PM<sub>10</sub>, PM<sub>2.5</sub>, BC, and OC from major anthropogenic sources: fuel combustion in power plant, industry, transport,  
12 and domestic sectors; industrial processes; agricultural activities; evaporation; and others.

13 Emissions from stationary fuel combustion and non-combustion sources are traditionally calculated using activity  
14 data and emission factors, including the effects of control technologies. For fuel consumption, the amount of  
15 energy consumption for each fuel type and sector was obtained from the International Energy Agency World  
16 Energy Balances for most countries and province-level tables in the China Energy Statistical Yearbook were used  
17 for China. Other activity data such as the amount of emissions produced from industrial processes were obtained  
18 from related international and national statistics. For emission factors, those without effects of abatement measures  
19 were set and then, effects of control measures were considered based on temporal variations of their introduction  
20 rates. Default emission factors and settings of country- and region-specific emission factors and removal  
21 efficiencies were obtained from scientific literature studies as described in Kurokawa and Ohara (2020) and  
22 references therein.

23 Emissions from road transport were calculated using vehicle numbers, annual distance travelled, and emission  
24 factors for each vehicle type. The number of registered vehicles were obtained from national statistics in each  
25 country and the World Road Statistics. For emission factors, year-to-year variation were considered by following  
26 procedures: (1) Emission factors of each vehicle type in a base year were estimated; (2) Trends of the emission  
27 factors for each vehicle type were estimated considering the timing of road vehicle regulations in each country  
28 and the ratios of vehicle production years; (3) Emission factors of each vehicle type during the target period were  
29 calculated using those of base years and the corresponding trends.

30 In REASv3.2.1, only large power plants were treated as point sources. For emissions from cement, iron, and steel  
31 plants, grid allocation factors were developed based on positions, production capacities, and start and retire years  
32 for large plants. Gridded emission data of EDGARv4.3.2 were used for grid allocation factors for the road  
33 transport sector. Rural, urban, and total population data were used to allocation emissions from the residential  
34 sector. For other sources, total population were used for proxy data.

35 For temporal distribution, if data for monthly generated power and production amounts of industrial products were  
36 available, monthly emissions were estimated by allocating annual emissions to each month using the monthly data  
37 as proxy. For the residential sector, monthly variation of emissions was estimated using surface temperature in  
38 each grid cell. If there is no appropriate proxy data, annual emissions were distributed to each month based on  
39 number of dates in each month.

40 Monthly gridded emission data sets at 0.25° × 0.25° resolution for major sectors and emission table data for major  
41 sectors and fuel types in each country and region during 1950-2015 are available in text format from a data  
42 download site of REAS (<https://www.nies.go.jp/REAS/>). Table S1 provides an overview about the REASv3.2.1  
43 sector mapping to the HTAP\_v3 sectors.

44 More details of the methodology of REASv3.2.1 are available in Kurokawa and Ohara (2020) and its supplement.  
45 (Note that REASv3.2.1 is the version after error corrections of REASv3.2 of Kurokawa and Ohara (2020)). Details  
46 of the error corrections are described in the data download site of REAS.) Table S1 provides an overview about  
47 the REASv3.2.1 sector mapping to the HTAP\_v3 sectors.

48 The MEIC inventory (<http://meicmodel.org>, 2021) is not currently included in the HTAP\_v3 mosaic. Since the  
49 REAS inventory only includes emissions until 2015, the REAS-based HTAP\_v3 mosaic is only complete until  
50 this year. Emissions beyond 2015 were extrapolated using trends derived from a combination of MEIC and  
51 EDGAR. To extend the Chinese emission estimates to most recent years, MEIC data were used to adjust sector  
52 and pollutant specific trend for China for the years 2016 and 2017 (refer to Table S2 for the mapping sectors of  
53 MEIC and HTAP\_v3). Then, the 2018 data were calculated based on the 2015-2017 trend. For all the other



1 countries belonging to the REAS domain, the emissions were extended beyond 2015 applying the sector-, country-  
2 , and pollutant-specific trends from EDGAR.

### 3 **2.3.5 CAPSS-KU inventory**

4 In the Republic of Korea, the National Air Emission Inventory and Research Center (NAIR) estimates annual  
5 emissions of the air pollutants CO, NO<sub>x</sub>, SO<sub>x</sub>, TSP, PM<sub>10</sub>, PM<sub>2.5</sub>, BC, VOCs, and NH<sub>3</sub> via the Clean Air Policy  
6 Support System (CAPSS). The CAPSS inventory is divided into four source-sector levels (high, medium, low and  
7 detailed) based on the European Environment Agency's (EEA) CORE Inventory of AIR emissions  
8 (EMEP/CORINAIR). For activity data, various national- and regional-level statistical data collected from 150  
9 domestic institutions are used. For large point sources, emissions are estimated directly using real-time stack  
10 measurements. For small point, area and mobile sources, indirect calculation methods using activity data, emission  
11 factors, and control efficiency are used.

12 Even though CAPSS (Clean Air Policy Support System) has been estimating annual emissions since 1999, some  
13 inconsistencies exist in the time series because of the data and methodological changes over the period. For  
14 example, emissions of PM<sub>2.5</sub> were initiated from the year 2011 and not from 1999. Therefore, in the CAPSS  
15 emission inventory, PM<sub>2.5</sub> emissions were calculated from 2011, and post-2011 the PM<sub>10</sub> to PM<sub>2.5</sub> emission ratio  
16 was used to calculate the emissions from 2000 to 2010. These limitations make it difficult to compare and analyse  
17 emissions inter-annually. To overcome these limitations, re-analysis of the annual emissions of pollutants was  
18 conducted using upgrades of the CAPSS inventory, such as missing source addition and emission factor updates.

19 The biomass combustion and fugitive dust sector emissions from 2000 to 2014 were estimated and added in the  
20 inventory, which are newly calculated emission sources from 2015. As for the on-road mobile sector, new  
21 emission factors using 2016 driving conditions were applied from the year 2000 to 2015. Since the emissions from  
22 the combustion of imported anthracite coal were calculated only from 2007, the coal use statistics of imported  
23 anthracite from 2000 to 2006 were collected to estimate emissions for those years.

24 After all the adjustments, a historically re-constructed emissions inventory using the latest emission estimation  
25 method and data was developed. Table S1 provides an overview about the CAPSS sector mapping to the  
26 HTAP\_v3 sectors.

### 27 **2.3.6 JAPAN inventory (PM2.5EI and J-STREAM)**

28 The Japanese emission inventory contributing to the HTAP\_v3 mosaic is jointly developed by the Ministry of the  
29 Environment, Japan (MOEJ) for emissions arising from mobile sources and by the National Institute of  
30 Environmental Studies (NIES) for estimating emissions from fixed sources.

31 The mobile source emissions data for the HTAP\_5.1, 5.2, and 5.4 sectors are based on the air pollutant emission  
32 inventory named "PM2.5 Emission Inventory (PM2.5EI), <https://www.env.go.jp/air/osen/pm/info.html>).  
33 PM2.5EI has been developed for the years 2012 and 2015, while for 2018 is currently under development. Almost  
34 all anthropogenic sources are covered, but emissions from vehicles are estimated in particular detail based on  
35 JATOP (Shibata, 2021). The emission factor of automobiles is constructed by MOEJ as a function of the average  
36 vehicle speed over several kilometres in a driving cycle that simulates driving on a real road. Emission factors are  
37 organized by 7 types of vehicles, 2 fuel types, 5 air pollutants, and regulation years, and have been implemented  
38 since 1997 as a project of MOEJ. By using these emission factors and giving the average vehicle speed on the  
39 road to be estimated, it is possible to estimate the air pollutant emissions per kilometre per vehicle. The hourly  
40 average vehicle speed of trunk roads, which account for 70% of Japan's traffic volume, is obtained at intervals of  
41 several kilometres nationwide every five years, so the latest data for the target year is used. For narrow roads, the  
42 average vehicle speed by prefecture measured by probe information is applied. It is 20 km/h in Tokyo, but slightly  
43 faster in other prefectures. Starting emission is defined as the difference between the exhaust amount in the  
44 completely cold state and the warm state in the same driving cycle and is estimated by the times the engine started  
45 in a day. Chassis dynamometer tests are performed in a well-prepared environment, so for more realistic emissions  
46 estimates, temperature correction factor, humidity correction factor, deterioration factor, DPF regeneration factor,  
47 and soak time correction factor are used. In addition to running and starting emissions, evaporative emissions  
48 from gasoline vehicles and non-exhaust particles such as road dust (including brake wear particles) and tire wear  
49 particles are combined to provide a vehicle emissions database with a spatial resolution of approximately 1 km ×  
50 1 km (30° latitude, 45° longitude), and a temporal resolution of an hour by month, including weekdays and  
51 holidays.



1 Off-road vehicle emissions are estimated separately for 17 types of construction machinery, industrial machinery  
2 (forklifts), and 5 types of agricultural machinery. In all cases, emission factors by type and regulatory year per  
3 workload are used, as researched by the MOEJ. Although not as precise as automobiles, the off-road database is  
4 provided with the same temporal and spatial resolution as the automobile database.

5 Emissions from stationary sources in Japan are derived from the emission inventory developed in the Japan's  
6 Study for Reference Air Quality Modelling (J-STREAM) model intercomparison project (Chatani et al., 2018;  
7 Chatani et al., 2020). In this emission inventory, emissions from stationary combustion sources are estimated by  
8 multiplying emission factors and activities including energy consumption, which is available in the comprehensive  
9 energy statistics. Large stationary sources specified by the air pollution control law need to report emissions to  
10 the government every three years. The emission factors and their annual variations were derived from the  
11 emissions reported by over 100,000 sources (Chatani et al., 2020). For fugitive VOC emissions, MOEJ maintains  
12 a special emission inventory to check progress on regulations and voluntary actions targeting 30% reduction of  
13 fugitive VOC emissions starting from 2000. VOC emissions estimated in this emission inventory are used.  
14 Emissions from agricultural sources are consistent with the emissions estimated in the national greenhouse gas  
15 emission inventory (Center for Global Environmental Research et al., 2022). Emissions of all the stationary  
16 sources are divided into prefecture, city, and grid (approximately 1 x 1 km) levels based on spatial proxies specific  
17 to each source. Emissions for the year 2018 were estimated applying sector- and pollutant-specific trends from  
18 EDGAR. Table S1 provides an overview about the Japanese inventory sector mapping to the HTAP\_v3 sectors.

#### 19 2.4 Gap-filling methodology with EDGARv6.1

20 EDGAR is a globally consistent emission inventory of air pollutant and greenhouse gases developed and  
21 maintained by the Joint Research Centre of the European Commission (<https://edgar.jrc.ec.europa.eu/>). The  
22 EDGAR methodology used to compute GHG and air pollutant emissions has been described in detail in several  
23 publications (Janssens-Maenhout et al., 2019; Crippa et al., 2018) and summarised here after. In EDGAR, air  
24 pollutant emissions are computed making use of international statistics as activity data (e.g., International Energy  
25 Balance data, Food and Agriculture Organisation statistics, US Commodity statistics, etc.), region- and/or country-  
26 specific emission factors by pollutant/sector and technology and abatement measures, following Eq. 1:

$$27 \quad EM_{i(C,t,x)} = \sum_{j,k} AD_{i(C,t)} * TECH_{i,j(C,t)} * EOP_{i,j,k(C,t)} * EF_{i(C,t,x)} * (1 - RED)_{i,j,k(C,t,x)}$$

28 (Eq. 1)

29 where EM are the emissions from a given sector *i* in a country *C* accumulated during a year *t* for a chemical  
30 compound *x*, AD the country-specific activity data quantifying the human activity for sector *i*, TECH the mix of  
31 *j* technologies (varying between 0 and 1), EOP the mix of *k* (end-of-pipe) abatement measures (varying between  
32 0 and 1) installed with a share *k* for each technology *j*, and EF the uncontrolled emission factor for each sector *i*  
33 and technology *j* with relative reduction (RED) by abatement measure *k*. Emission factors are typically derived  
34 from the EMEP/EEA Guidebooks (EMEP/EEA, 2013, 2019, 2016), the AP-42(EPA, 2009) inventory and  
35 scientific literature.

36 Annual country and sector specific air pollutant emissions are then disaggregated into monthly values (Crippa et  
37 al., 2020) and subsequently spatially distributed making use of detailed proxy data (Janssens-Maenhout et al.,  
38 2019; Crippa et al., 2021).

39 As the most comprehensive and globally consistent emission database, the latest update of the EDGAR air  
40 pollutant emissions inventory, EDGARv6.1 ([https://edgar.jrc.ec.europa.eu/dataset\\_ap61](https://edgar.jrc.ec.europa.eu/dataset_ap61)), is used in the HTAP\_v3  
41 mosaic to complete missing information from the officially reported inventories, as reported in Table 3.  
42 EDGARv6.1 includes important updates to estimate air pollutant emissions such as the improvement of road  
43 transport emission estimates for many world regions, the inclusion of agricultural NMVOC emissions, revised  
44 monthly emission profiles (Crippa et al., 2020) and updated spatial proxies to distribute national emissions by  
45 sector over the globe (Crippa et al., 2021). EDGARv6.1 also includes new international shipping proxies and their  
46 monthly distribution based on the STEAM model (Jalkanen et al., 2012; Johansson et al., 2017). In the  
47 Supplementary Material (section S2), the assessment of EDGAR emission data is reported in comparison with  
48 global and regional inventories.

### 49 3 Results

#### 50 3.1 Annual time series analysis: trends and regional and sectoral contributions



1 Having a consistent set of global annual emission inventories for a two-decade period allows the investigation of  
2 global emissions trends for the inventory pollutants and regional and sectoral contributions. Figure 2 presents  
3 annual time series (2000-2018) of the global emissions of the nine air pollutants included in the HTAP\_v3 mosaic  
4 separated into the actual contributions of 12 regions. Figure 3 shows the corresponding relative contributions of  
5 (a) 16 sectors and (b) 12 regions to the 2018 global emissions of these same pollutants. We can then consider each  
6 pollutant in turn.

7 Global SO<sub>2</sub> emissions declined from 99.4 to 72.9 Mt over the past two decades. This decreasing pattern is found  
8 for several world regions with the fastest decline in Eastern Asia, where after the year 2005 SO<sub>2</sub> emissions began  
9 to decrease steadily. This is consistent with the use of cleaner fuels with lower sulphur content and the  
10 implementation of desulphurisation techniques in power plants and industrial facilities in China in accordance  
11 with the 11th Five-Year Plan (FYP, 2006–2010 (Planning Commission, 2008)) and the 12th Five-Year Plan (FYP,  
12 2011–2015 (Hu, 2016)) (Sun et al., 2018). Similarly, industrialised regions, such as North America and Europe,  
13 are characterised by a continuous decreasing trend in SO<sub>2</sub> emission, which had started well before the year 2000  
14 due to the implementation of environmental and air quality legislation (EEA, 2022). Increasing SO<sub>2</sub> emissions,  
15 on the other hand, are found for Southern Asia (+112% compared to 2000), South-East Asia and developing  
16 Pacific (+62%), and Africa (+40%). These increases mostly arise from the energy, industry, and (partly)  
17 residential sectors, and reflect the need for emerging and developing economies to mitigate these emissions.  
18 Emissions estimated using satellite retrievals and model inversions confirm the trends provided by the HTAP\_v3  
19 mosaic (Liu et al., 2018). SO<sub>2</sub> is mostly emitted by power generation and industrial activities, which in 2018  
20 represent 42% and 26%, respectively, of the global total. Despite measures in some specific sea areas to mitigate  
21 sulphur emissions, globally they have been rising steadily with increasing activity. International shipping  
22 represents 13.8% of global SO<sub>2</sub> emissions in 2018, and it is 41% higher compared to the 2000 levels (Fig. 3).

23 Global NO<sub>x</sub> emissions increased from 110.4 Mt in 2000 to 117.4 Mt in 2018 as a result of the increase in energy-  
24 and industry-related activities for most of the world regions (in particular over the Asian domain). The strongest  
25 decreases are found for North America (-63%), Europe (-42%), Asia-Pacific Developed (-32%) and to a lower  
26 extent for Eurasia (-6%). Comparable spatio-temporal patterns are found by satellite OMI data and ground based  
27 measurements of NO<sub>2</sub> concentrations (Jamali et al., 2020). NO<sub>x</sub> is mainly produced at high combustion  
28 temperatures (e.g., power and industrial activities, 38% of the global total), but also by transportation (27% of the  
29 global total) and international shipping (14% of the global total).

30 CO is mostly emitted by incomplete combustion processes from residential combustion, transportation and the  
31 burning of agricultural residues. Globally, CO emissions showed little change over the past two decades (502.7  
32 Mt in 2000 vs. 499.8 in 2018), but different regional trends are present. Historically industrialised regions have  
33 reduced their emissions over the years (-42% in Europe and -62% in North America), while CO emissions  
34 increased in Africa by 45% and in Southern Asia by 49%. Road transport CO emissions halved over the past two  
35 decades (-55%), while the emissions from all other sectors increased. These results are consistent with MOPITT  
36 satellite retrievals, which mostly show the same trends over the different regional domains over the past decades  
37 (Yin et al., 2015).

38 NMVOC emissions increased from 115.2 Mt in 2000 to 146 Mt in 2018. These emissions are mostly associated  
39 with the use of solvents (25% of the 2018 global total), fugitive emissions (23%), road transportation (including  
40 both combustion and evaporative emissions, 15%) and small-scale combustion activities (19%). The most  
41 prominent increases in the emissions at the global level are found for the energy (+98%) and solvents (+81%). In  
42 2018, NMVOC emissions from solvents were 3.7 and 3.5 times higher than in 2000 in China and India,  
43 respectively, while a rather stable trend in found for US and Europe.

44 Global NH<sub>3</sub> emissions increased from 47.4 Mt in 2000 to 58.9 Mt in 2018 due to enhanced emissions from  
45 agricultural activities. In particular, NH<sub>3</sub> emissions strongly increased in Africa (+60.5%), South-East Asia and  
46 developing Pacific (48.3%), Southern Asia (+38.7%), and Latin America and Caribbean (+41.1%).

47 Particulate matter emissions showed little change over the past two decades at the global level, whereas regional  
48 emission increases are found for Africa (e.g., +47.0% for PM<sub>10</sub>), Latin America and Caribbean (+39.1%), Middle  
49 East (48.3%), and Southern Asia (+56%), mostly associated with increases in agricultural waste burning and the  
50 livestock, energy, and waste sectors. By contrast, Eastern Asia (-39.5%), Europe (24.3%), and Asia-Pacific  
51 Developed (-36.8%) significantly decreased their PM<sub>10</sub> emissions over the past two decades due to the continuous  
52 implementation of reduction and abatement measures for the energy, industry, road transport and residential  
53 sectors ((Crippa et al., 2016). As shown in Fig. 3, the relative contribution of North America to global PM<sub>10</sub> is  
54 quite high compared to other substances due to fugitive dust emissions (e.g., unpaved road dust, coal pile dust,  
55 dust from agricultural tilling) which have not been adjusted for meteorological conditions (e.g., rain, snow) and  
56 near-source settling and mitigation (e.g., tree wind breaks) because these removal mechanisms are better



1 addressed by the chemical transport models. Additional uncertainty may be therefore introduced for these  
2 emissions, depending on the modelling assumptions of each official inventory. Similarly, particulate matter  
3 speciation into its carbonaceous components it is often challenging and subjected to higher level of uncertainty,  
4 for instance because different definitions are used for PM in inventories, including condensable emissions or not  
5 (Denier van der Gon et al., 2015). Attempts to improve the accuracy of such emissions (e.g. BC and OC emissions  
6 over the European domain) are ongoing.

### 7 3.2 Emission maps

8 Spatially distributed emission data describe where emissions take place, as input for local, regional and global air  
9 quality modelling. As noted in section 2.2, nationally aggregated air pollutant emissions are spatially distributed  
10 over the corresponding national territory using spatial proxy data which are believed to provide a relatively good  
11 representation of where emissions takes place. Depending on the emitting sector, air pollutants can be associated  
12 with the spatial distributions of point sources (e.g., in the case of power plant or industrial activities), road  
13 networks (e.g., for transportation related emissions), settlement areas (e.g., for small-scale combustion emissions),  
14 crop and livestock distribution maps, ship tracks etc. Using reliable and up-to-date spatial information to distribute  
15 national emissions is therefore relevant, although challenging. Multiple assumptions are often made by inventory  
16 compilers when developing their inventories, which may result in differences when analysing spatially distributed  
17 emissions provided by different inventory compilers over the same geographical domain.

18 One key goal of the HTAP\_v3 mosaic is to collate in one inventory the most accurate spatially-distributed  
19 emissions for all air pollutants at the global level, based on the best available local information. Point sources  
20 related with emissions from power plant and industrial facilities represent one the most critical spatial information  
21 to be retrieved, and their misallocation can significantly affect the characterisation of local air quality. This  
22 challenge is also present in the HTAP\_v3 mosaic. For example, the REASv3.2.1 inventory is still using limited  
23 information to distribute emissions from these two sectors especially for industrial plants. Depending on the  
24 region, point source information could be limited compared to datasets used in inventories of North America and  
25 Europe. To overcome this issue, the participation of national emission inventory developers not only from China,  
26 but also India and other Asian countries is recommended. The impact can be seen in Fig. 4, which shows the  
27 global map of SO<sub>2</sub> emissions in 2018 based on the HTAP\_v3 mosaic compilation, where information about the  
28 magnitude and the type of emission sources for the different regions can be retrieved. The energy and industry  
29 sectors contribute a large fraction of SO<sub>2</sub> emissions (Fig. 3a), but the spatial distribution of these emissions is  
30 qualitatively different in North America and Europe than in Asia (i.e., more “spotty”, less smooth and widely  
31 distributed). Ship tracks cover the entire geographical marine domain, consistent with emissions from the STEAM  
32 model (Jalkanen et al., 2012; Johansson et al., 2017) included in the EDGARv6.1 database, although showing  
33 marked emissions over the Mediterranean Sea, Asian domain, Middle East and North American coasts.  
34 Furthermore, emissions from power plant and industrial activities, as well as small-scale combustion are  
35 prominent over the Asian domain, Eastern Europe, and some African regions.

36 Sector-specific case studies are presented in the maps of Figs. 5-8. Figure 5 shows the comparison of annual NO<sub>x</sub>  
37 emissions for the year 2000 and 2018. The road transport sector is a key source of NO<sub>x</sub> emissions (cf. Fig. 3a),  
38 and this contribution is reflected in the visible presence of road networks in the maps. Decreasing emissions are  
39 found for industrialised regions (USA, Europe, Japan) thanks to the introduction of increasingly restrictive  
40 legislation on vehicle emissions since the 1990s, whereas a steep increase is found for emerging economies and  
41 in particular India, China, and the Asian domain. Figure 6 shows the different spatial allocation of PM<sub>10</sub> emissions  
42 from the residential sector during the month of January 2018, with higher emission intensities evident in the  
43 Northern Hemisphere (cold season) and the lower values in the Southern Hemisphere (warm season). Figures 7  
44 and 8 show the spatio-temporal allocation of agriculture-related emissions, and specifically, PM<sub>10</sub> emissions from  
45 agricultural waste burning and NH<sub>3</sub> emissions from agricultural soil activities.

### 46 3.3 Monthly temporal distribution

#### 47 3.3.1 Monthly variability by region

48 The magnitude of air pollutant emissions varies by month because of the seasonality of different anthropogenic  
49 activities and their geographical location (e.g., Northern vs. Southern Hemisphere regions). Figures 9 and 10 (and  
50 S3.1, S3.2 and S3.3) show the monthly distribution of regional emissions for those pollutants and sectors for  
51 which higher variability is expected. The year 2015 was chosen since it is the last year for which all of the official  
52 data providers have data. Figure 9 shows monthly NH<sub>3</sub> emissions by region from three agricultural activities  
53 (agricultural waste burning, livestock, and crops). These sectors display the largest variability by month, reflecting  
54 the seasonal cycle and the region-specific agricultural practices, such as fertilisation, crop residue burning, manure  
55 and pasture management, animal population changes, etc. In Figure 10, NO<sub>x</sub> emissions from residential activities



1 show a particular monthly distribution, with the highest emissions occurring during the cold months shifted for  
2 the Northern and Southern Hemispheres. By contrast, regions in the equatorial zone do not show a marked monthly  
3 profile even for residential activities. The energy sector also follows monthly-seasonal cycles related to the  
4 demand for power generation, which is also correlated with ambient temperature and local day length. Transport-  
5 related emissions do not show a large variation by month, whereas daily and weekly cycles for transport-related  
6 emissions, which are typically more relevant, are beyond the temporal resolution of this work.

7 Although a spatio-temporal variability of the HTAP\_v3 emissions is found in these figures, a more in-depth  
8 analysis reveals that with the exception of few regions and sectors (e.g., Canada, USA and regions gap-filled with  
9 EDGAR), no inter-annual variability of the monthly profiles is present, meaning that the majority of official  
10 inventories assume the same monthly distribution of the emissions for the past two decades (refer to Figs. S3.4-  
11 S3.9). This is different from the approach used for example by EDGAR (Crippa et al., 2020), ECCC for Canada,  
12 and U.S. EPA for the USA, where year-dependent monthly profiles are used for specific sectors, in particular for  
13 residential, power generation, and agricultural activities. Further analysis has shown that for the European domain  
14 regional rather than country-specific monthly profiles are applied. Improvements of the monthly distribution could  
15 be achieved through the implementation of the state-of-the-art information on this topic (Crippa et al., 2020;  
16 Guevara et al., 2021).

17

### 18 3.3.2 Spatially-distributed monthly emissions

19 An important added value of HTAP\_v3 comes from the availability of monthly gridmaps that reflect the  
20 seasonality of the emissions for different world regions. Access to spatially distributed monthly emissions is  
21 essential to design effective mitigation actions, providing information on hot spots of emissions and critical  
22 periods of the year when emissions are highest.

23 Figure 11 shows mid-season PM<sub>2.5</sub> monthly emissions arising from the residential sector in 2018. The global map  
24 shows higher emissions in the Northern Hemisphere during January, while the opposite pattern is found for the  
25 Southern Hemisphere in July. Agriculture is an important activity characterised by strong seasonal patterns, as  
26 shown in Figs. 12 and 13. Figure 12 shows PM<sub>10</sub> monthly emission maps from agricultural residue burning in  
27 2018 from HTAP\_v3, highlighting higher emissions over certain months of the year related with specific burning  
28 practices of agricultural residues for different world regions. For example, during the month of April, intense  
29 burning of crop residues is found in Africa (Nigeria, Ethiopia, Sudan, South Africa, etc.), South America (Brazil,  
30 Argentina, Colombia, etc.), Northern India, and South-Eastern Asia (e.g., Vietnam, Thailand, Indonesia,  
31 Philippines, etc.). Figure 13 represents the yearly variability of NH<sub>3</sub> emissions from agricultural soils activities,  
32 mostly related with fertilisation. During the month of March and April, intense agricultural soils activities are  
33 found over Europe and North America compared to other months, while during the month of October the highest  
34 emissions are for this sector are found in China, India, several countries of the Asian domain, but also in USA,  
35 Australia, and Latin America. These results are consistent with satellite based observations performed using  
36 Cross-track Infrared Sounder (Shephard et al., 2020).

## 37 3.4 Vertical distribution of the emissions

### 38 3.4.1 Aircraft emissions

39 In EDGAR6.1 the emissions are provided at three effective altitude levels (landing/take-off, ascent/descent, and  
40 cruising). The spatial proxy for the aviation sector is derived from International Civil Aviation Organization  
41 (ICAO, 2015) which specifies a typical flight pattern with landing/take-off cycle within few km of the airport,  
42 followed by climb-out/descending phase during the first 100 km and the last 100km of a flight and finally the  
43 remaining part from 101 km until the last 101 km as the cruise phase. Routes and airport locations are taken from  
44 the Airline Route Mapper of ICAO (2015). In HTAP\_v3, aircraft emissions are provided as domestic and  
45 international, but with no information about altitude ranges. We recommend modellers to use the corresponding  
46 EDGARv6.1 data ([https://edgar.jrc.ec.europa.eu/dataset\\_ap61](https://edgar.jrc.ec.europa.eu/dataset_ap61)) including the vertical distribution of the  
47 emissions.

### 48 49 3.4 Speciation of NMVOC emissions

50 For emission data to be useful for modellers, total NMVOC emissions must be decomposed into emissions of  
51 individual NMVOC species. As the chemical mechanisms used by models can differ with respect to the NMVOC  
52 species they include, it is not practical to provide an NMVOC speciation which is usable by all models. Instead,  
53 a speciation is provided here for the set of 25 NMVOCs defined by Huang et al. (2017) and the corresponding  
54 data are made available on the HTAP\_v3 website. The absolute values of 25-category speciated NMVOC



1 emissions were obtained for all countries for the 28 EDGAR sectors from here:  
2 [https://edgar.jrc.ec.europa.eu/dataset\\_ap432\\_VOC\\_spec](https://edgar.jrc.ec.europa.eu/dataset_ap432_VOC_spec). The absolute NMVOC emissions of each species from  
3 each sector in this dataset were remapped to the HTAP\_v3 sectors following the mapping from Table 2, then  
4 converted to a speciation by dividing by the total emissions of each individual species for the four world regions  
5 defined by Huang et al. (2017): Asia; Europe; North America; and Other. The resulting NMVOC speciation is  
6 provided in the supplementary material to this paper for the 25 NMVOC species, 4 world regions, and 15 emitting  
7 NMVOC sectors<sup>2</sup> following the HTAP\_v3 sector classification (including 13 sectors defined over the 4 world  
8 regions, and the two international sectors: international shipping and international aviation). The list of countries  
9 comprising each region is also provided in the supplement.

10

#### 11 4 Data availability

12 The HTAP\_v3 emission mosaic data can be freely accessed and cited using DOI 10.5281/zenodo.7516361. All  
13 data can be also accessed through the EDGAR website at the following link:  
14 [https://edgar.jrc.ec.europa.eu/dataset\\_htap\\_v3](https://edgar.jrc.ec.europa.eu/dataset_htap_v3).

15 Data are made available in the following formats:

- 16 • Monthly gridmaps of emissions (in Mg/month) at 0.1x0.1degree resolution: there is one .NetCDF file  
17 per year and substance that includes the emissions for each sector for the 12 months.
- 18 • Monthly gridmaps of emission fluxes (in kg/m2/s) at 0.1x0.1degree resolution: there is one .NetCDF file  
19 per year and substance that includes the emission fluxes for each sector the emission fluxes for the 12  
20 months.
- 21 • Annual gridmaps of emissions (in Mg/year at 0.1x0.1degree resolution: there is one .NetCDF file per  
22 year and substance that includes the emissions for each sector.
- 23 • Annual gridmaps of emission fluxes (in kg/m2/s) at 0.1x0.1degree resolution: there is one .NetCDF file  
24 per year and substance that includes the emission fluxes for each sector.

25 The full set of HTAP\_v3 data is quite large, requiring substantial network bandwidth and time for download, and  
26 substantial storage space. To make it easier for users to query and use the data, additional products are available.  
27 For global modellers who may not require such high spatial resolution, gridmaps at 0.5x0.5 degree resolution are  
28 made available following the abovementioned specifications of the higher spatial resolution data. Furthermore, to  
29 allow regional modellers to download only the data for the regions they need, the JRC EDGAR group has also  
30 developed an interface to allow the users of the HTAP\_v3 mosaic to extract emission data over arbitrarily specified  
31 geographical domains. The HTAP tool is accessible after creation of an ECAS account  
32 (<https://webgate.ec.europa.eu/cas/login>) and it is available at: [https://edgar.jrc.ec.europa.eu/htap\\_tool/](https://edgar.jrc.ec.europa.eu/htap_tool/).

33

#### 34 5 Conclusions

35 The global air pollution mosaic inventory HTAP\_v3 presented and discussed in this paper is a state-of-the-art tool  
36 for addressing the present status and the recent evolution of a set of policy-relevant air pollutants. The inventory  
37 derives from the harmonization and blending of six regional inventories, gapfilled using the most recent release  
38 of EDGAR (EDGARv6.1). By directly incorporating the best available local information, including the spatial  
39 distribution of emissions, the HTAP\_v3 mosaic inventory can be used for policy-relevant studies at both regional  
40 and global levels. As such, the HTAP\_v3 mosaic inventory provides a complement to globally consistent emission  
41 inventories such as EDGAR. The global and regional trends of air pollutant emissions in the HTAP\_v3 mosaic  
42 are comparable with other commonly available global emission datasets.

43 By providing consistent times series for almost two decades, HTAP\_v3 allows an evaluation of the impact and  
44 success of the pollution control measures deployed across various regions of the world since 2000. Similarly, its  
45 finer sectoral resolution is suitable for understanding how and where technological changes have resulted in  
46 emissions reductions, suggesting possible pathways for strengthening appropriate policy actions.

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<sup>2</sup> No speciation profile is provided for the ‘tyre and brake wear sector’ not being a source of NMVOC emissions.



1 All these features make HTAP\_v3 a tool of certain interest for policy makers active in the air quality regulatory  
2 efforts. Indeed, although HTAP\_v3 provides a picture of a world where most pollutants are following a steady or  
3 decreasing path, several areas of the world show an opposite emission increasing trend, while wide portions of  
4 the world remain subject to unsatisfactory levels of ambient air pollution.

5 Similar to its predecessor (e.g. HTAP\_v2.2 mosaic inventory), we expect that this new HTAP\_v3 mosaic  
6 inventory will be used as a basis for global assessments of long-range, transboundary transport of air pollution  
7 under the Task Force on Hemispheric Transport of Air Pollution, while also providing a convenient and useful  
8 tool for regional modellers seeking the best available regional emissions with a consistent gap-filling  
9 methodology. Future versions of the HTAP mosaic inventory will incorporate new regional emission inventories  
10 and updated datasets as time and resources permit.

11

#### 12 **Author contributions.**

13 MC and DG developed the mosaic gathering input from all data providers. The co-chairs of the TF-HTAP (TK,  
14 TB, RW and JaKa) fostered the dialogue with international institutions contributing to this work with their data.  
15 PM, RM, JR, JZ, DN, MS, MDM, RW provided data for Canada, JuKu, SC, TM provided data for Japan, JeKu  
16 provided data for Europe, J-HW, JK provided data for Korea, TK, GP provided data for USA, JiKi provided data  
17 for Asia. The JRC EDGAR group (MC, ES, DG, EP, MM, FM, ES, MB, FP) lead the drafting of the publication  
18 with input from colleagues contributing to the HTAP\_v3 mosaic. SJS and HS performed detailed data comparison  
19 among available emission inventories.

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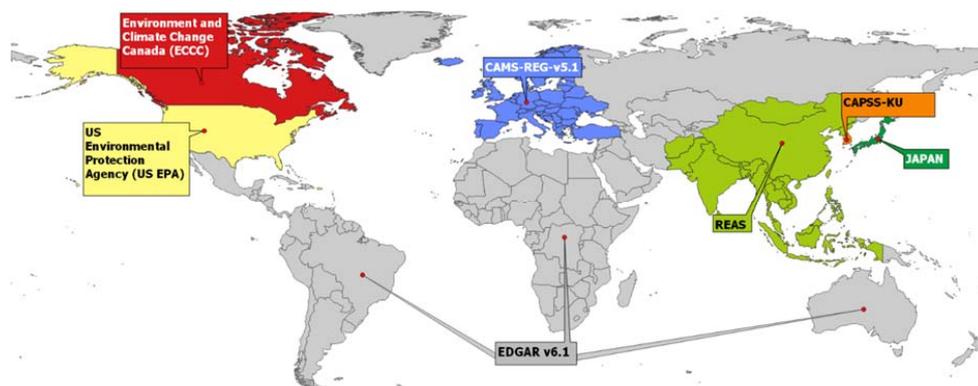
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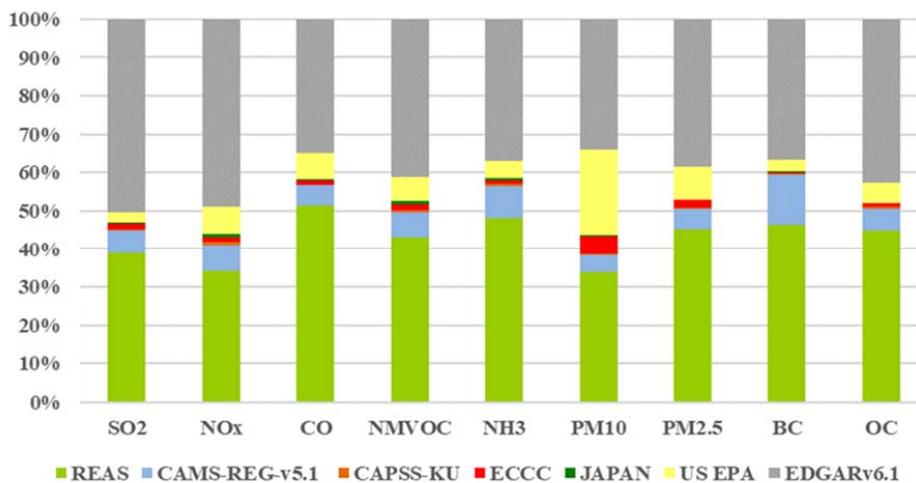
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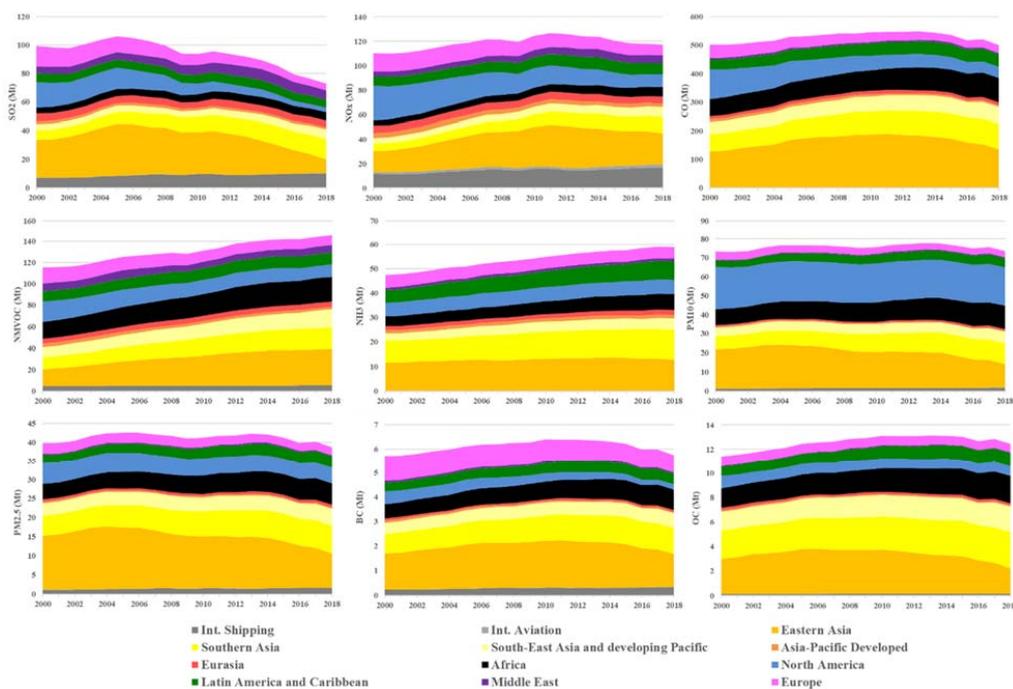


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Figure 1 – Overview of the HTAP\_v3 mosaic data providers. Data from officially reported emission gridmaps were collected from the US Environmental Protection Agency, Environment and Climate Change Canada, CAMS-REG-v5.1 for Europe, REASv3.2.1 for most of the Asian domain, CAPSS-KU for Korea and JAPAN (PM2.5EI and J-STREAM) for Japan. The share of the total emissions covered by each data provider is reported in the bar chart at the bottom.

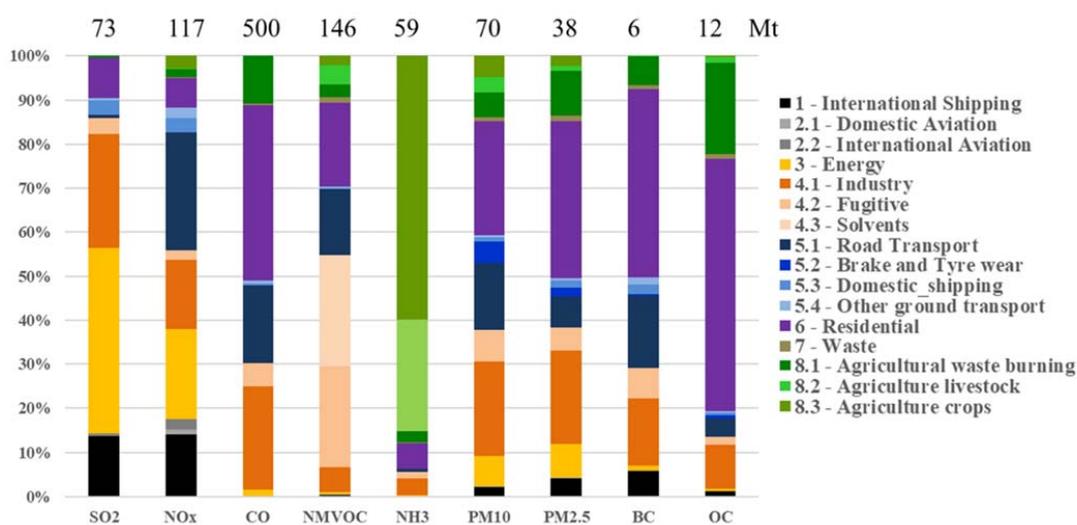


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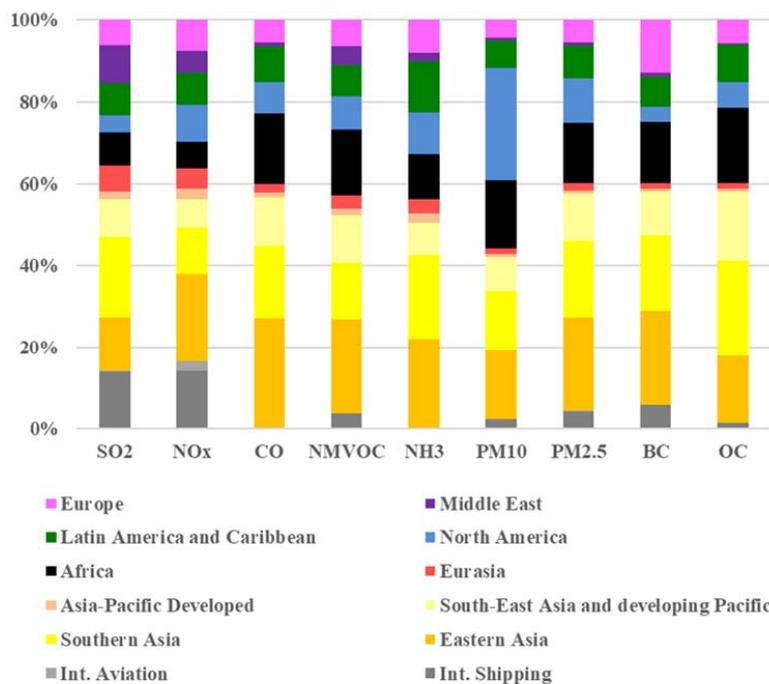


1 **Figure 2 – Time series of gaseous and particulate matter pollutants from HTAP\_v3 by aggregated regions.**  
 2 **Regional grouping follows the Intergovernmental Panel on Climate Change Sixth Assessment Report**  
 3 **(IPCC AR6) definitions. Table S3 provides information on the regional belonging of each country to the**  
 4 **IPCC AR6 regions.**

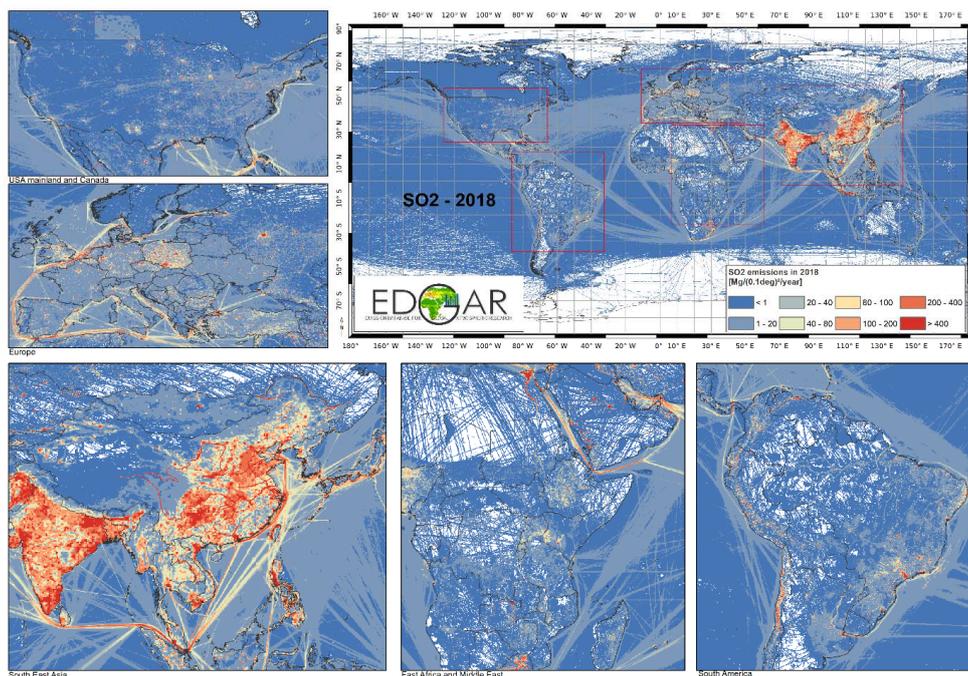
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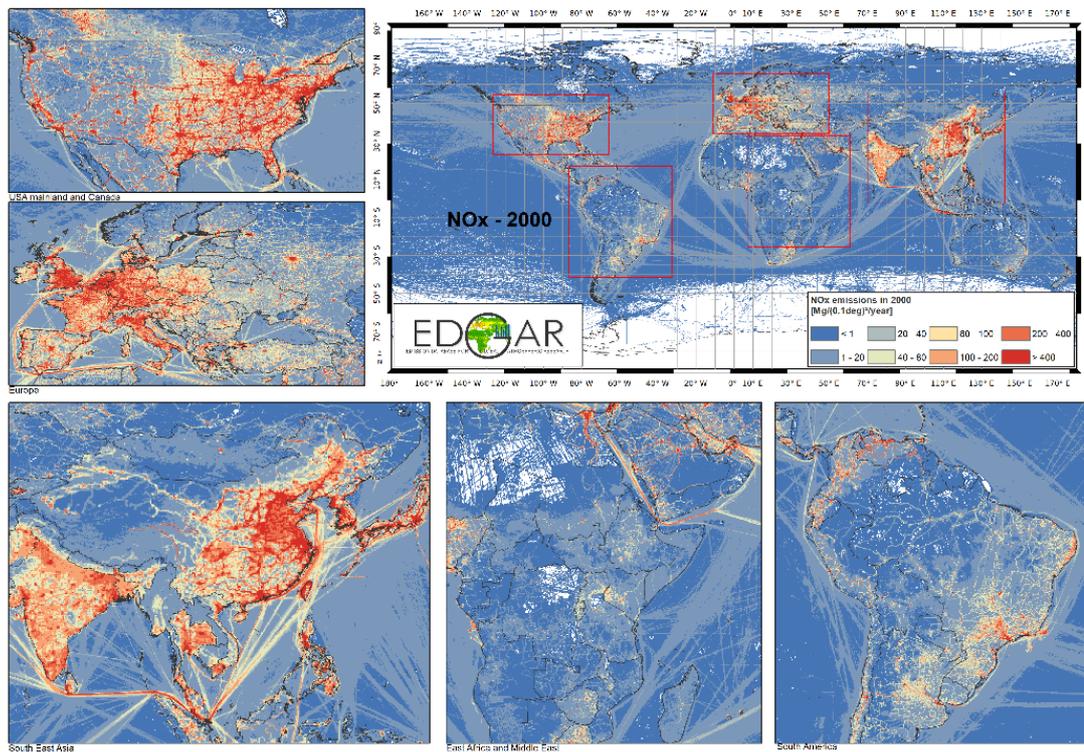


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2 **Figure 3 - 2018 sectoral (panel a) and regional (panel b) breakdown of air pollutant emissions from**  
3 **HTAP\_v3. At the top of each bar in panel a, total emissions for each pollutant are reported (in Mt).**

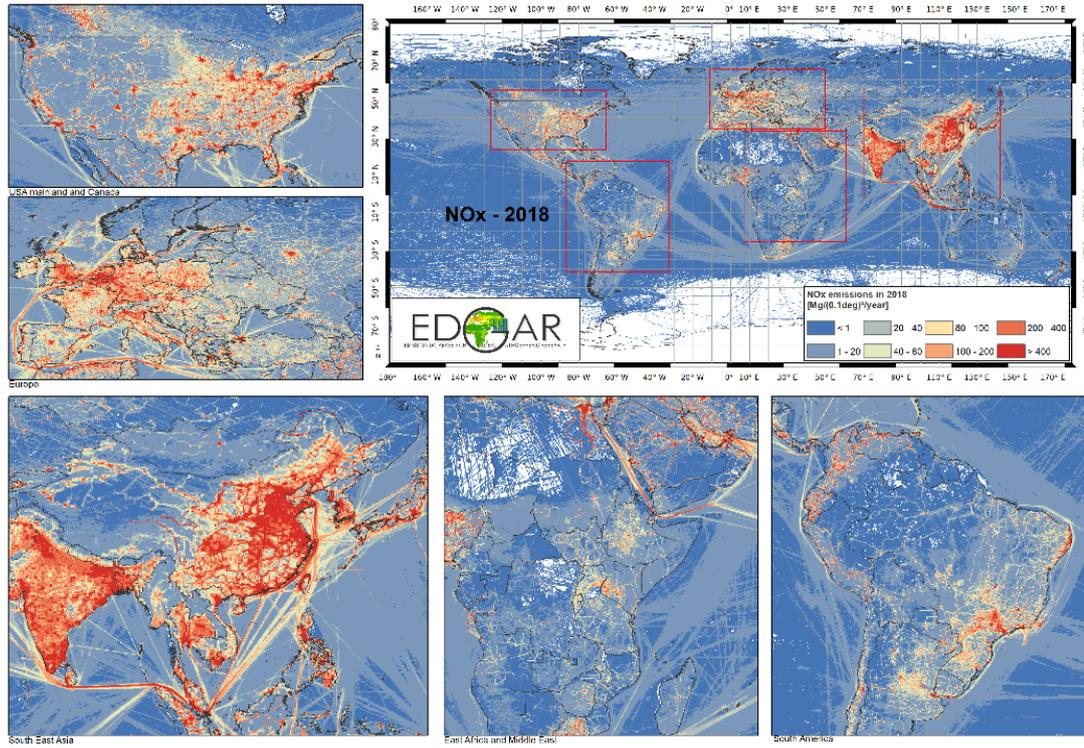


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Figure 4 – HTAP\_v3 mosaic: SO<sub>2</sub> emission gridmaps for the year 2018.



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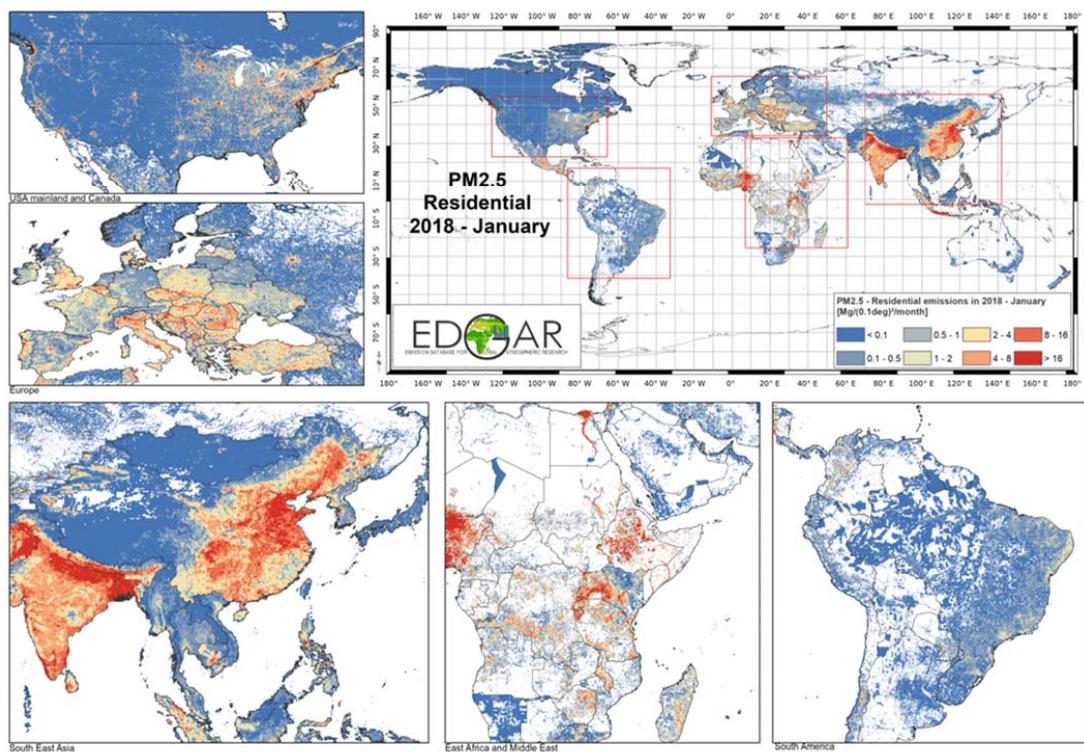


1

2 **Figure 5 – HTAP\_v3 mosaic: NO<sub>x</sub> emission gridmaps in 2000 (top panel) and 2018 (bottom panel).**

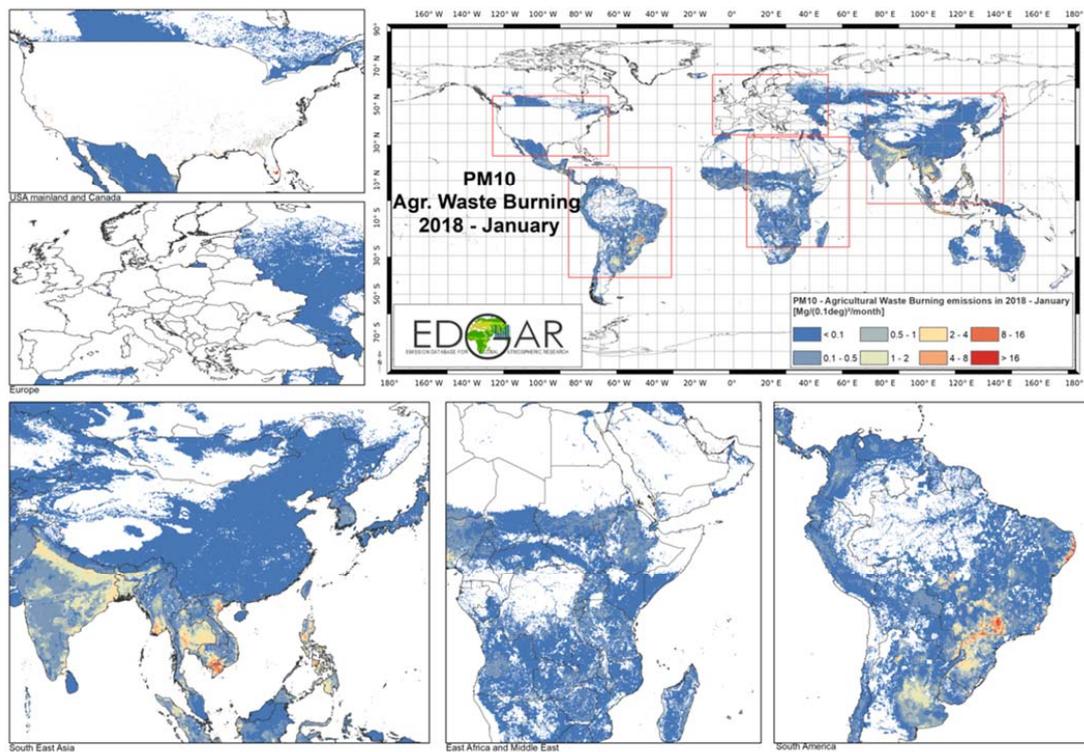


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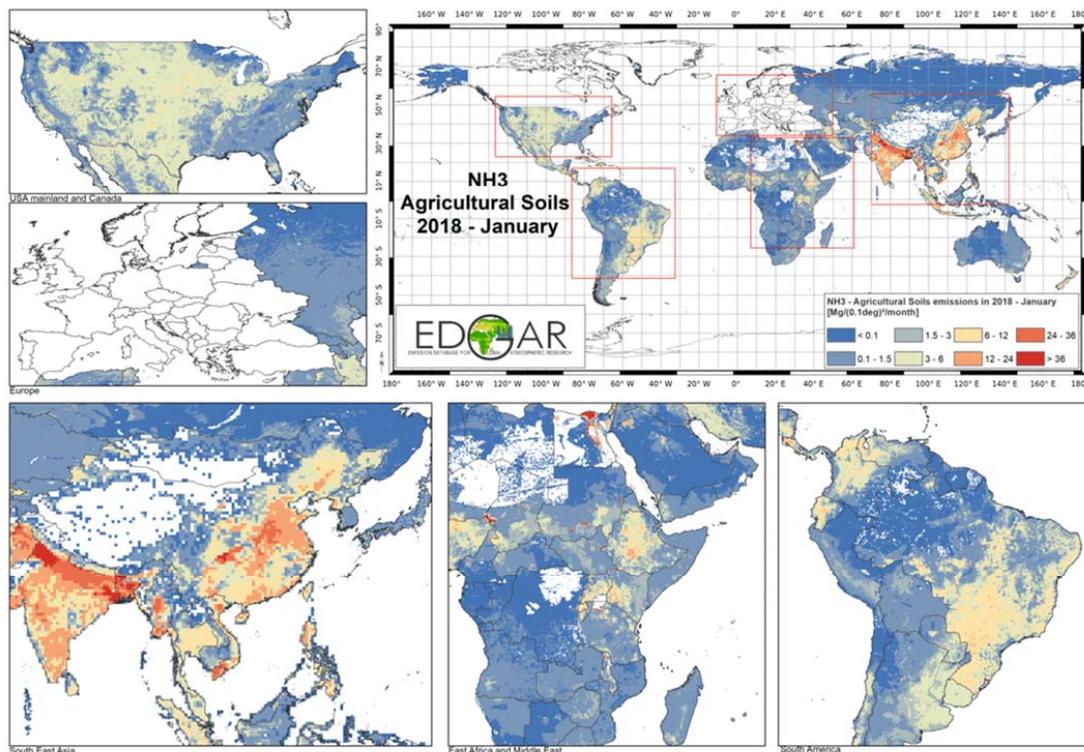
2

3 **Figure 6 – HTAP\_v3 mosaic: PM<sub>2.5</sub> emissions from residential activities in January 2018.**



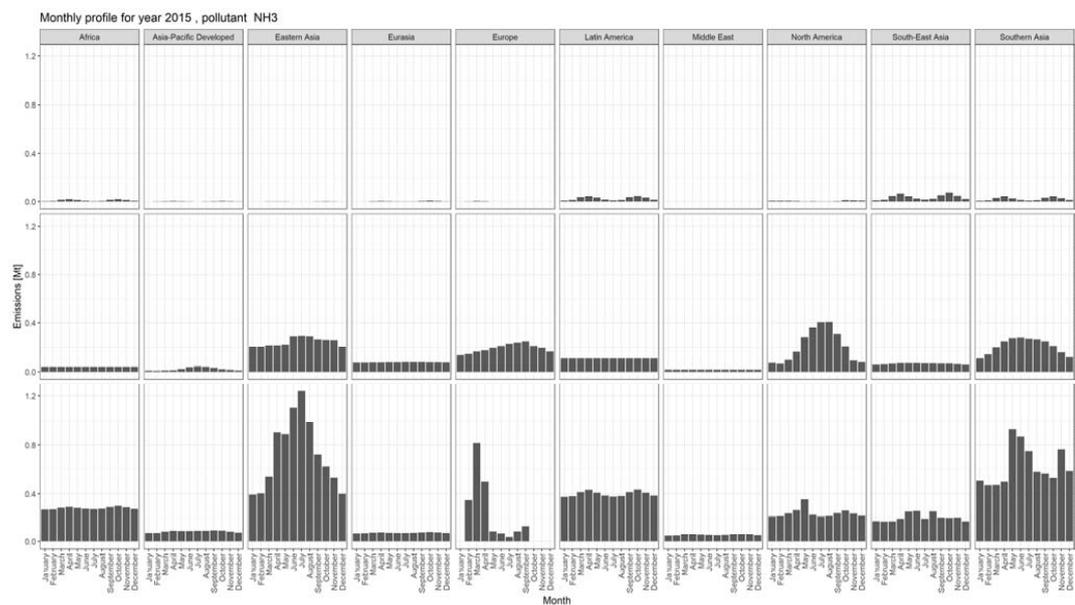
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2 **Figure 7 – HTAP\_v3 mosaic: PM<sub>10</sub> emissions from agricultural waste burning in January 2018.**



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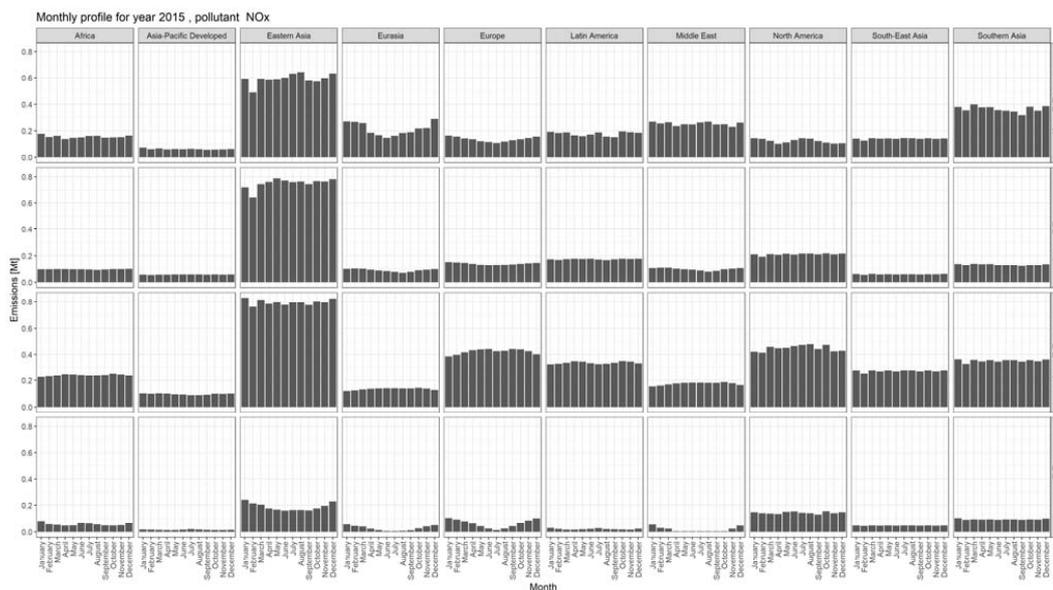
Figure 8 – HTAP\_v3 mosaic: NH<sub>3</sub> emissions from agricultural soils activities in January 2018.



4



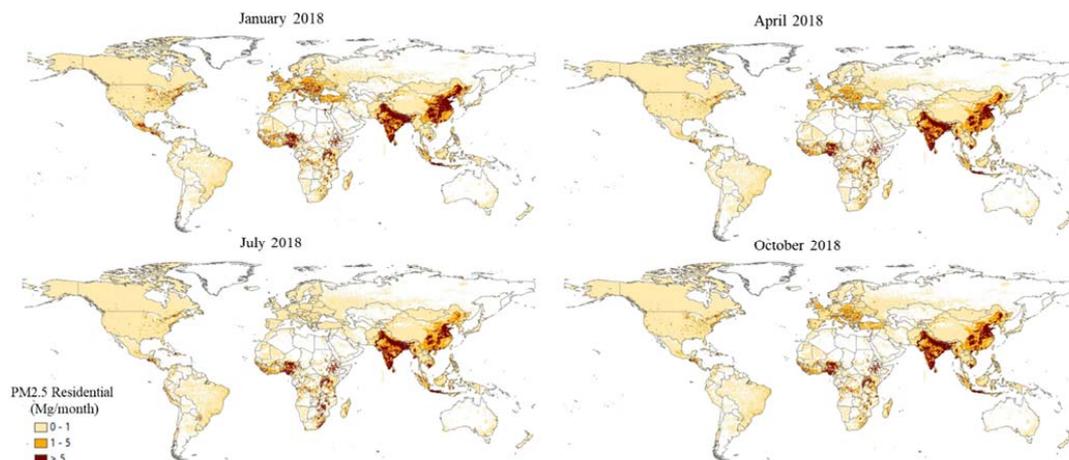
1 **Figure 9 – Monthly variability of NH<sub>3</sub> emissions for agriculture related activities for the different world**  
2 **regions in 2015.**



3  
4 **Figure 10 – Monthly variability of NO<sub>x</sub> emissions for relevant emission sectors for the different world**  
5 **regions in 2015.**  
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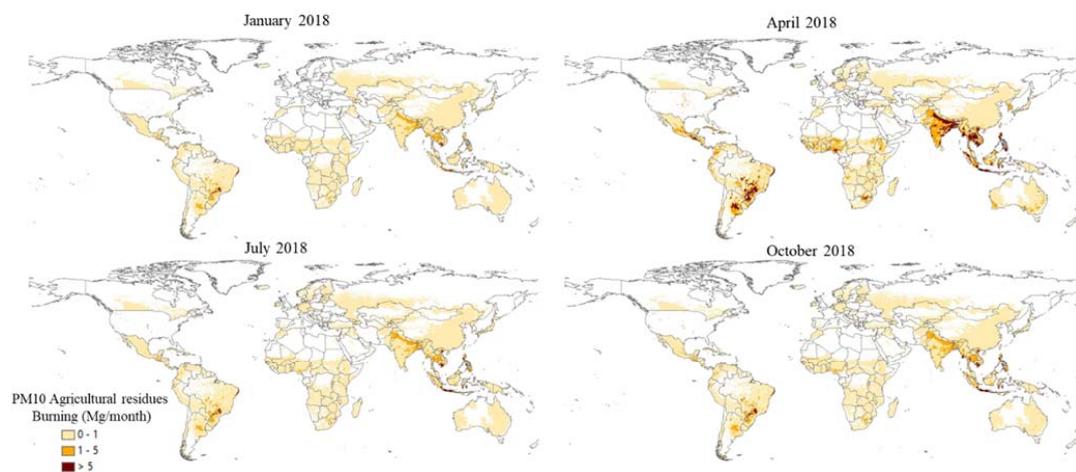


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**Figure 11 – PM<sub>2.5</sub> monthly emission maps from the residential sector in 2018 from HTAP\_v3.**

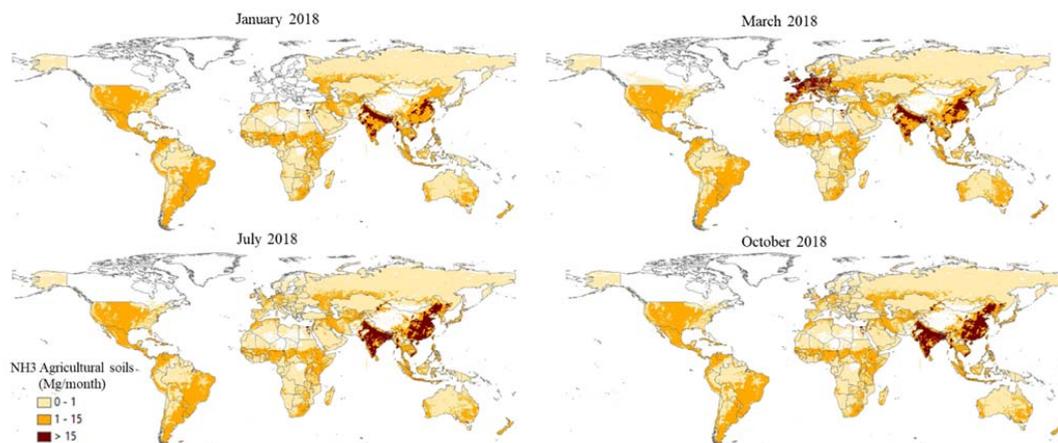
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**Figure 12 – PM<sub>10</sub> monthly emission maps from agricultural residues burning in 2018 from HTAP\_v3.**



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2 **Figure 13 – NH<sub>3</sub> monthly emission maps from agricultural soils in 2018 from HTAP\_v3.**

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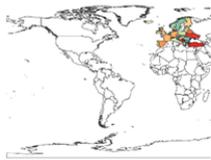
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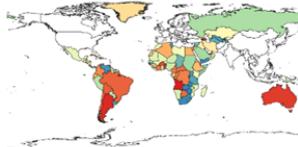
7 **Table 1 – Overview of data input to the HTAP\_v3 emission mosaic.**

Data source	CAMS-REG-v5.1	US EPA	ECCC	REASv3.2.1
Type of data source	Country inventories as emission time series by sector and country and emission gridmaps as .csv files.	Country inventories	Country inventories as emission time series by sector and country and emission gridmaps as .NetCDF files.	Country inventories as emission gridmaps as text files.
Sectors coverage	All sectors, excluding international shipping and aviation (international and domestic).	All sectors, excluding international shipping and aviation (international and domestic).	All sectors, excluding agricultural waste burning, international shipping and aviation (international and domestic).	All sectors, excluding brake and tyre wear, domestic shipping, waste, agricultural waste burning, international shipping and aviation (international and domestic).
Temporal coverage	2000-2018	2002-2017	2000-2016	2000-2015+trends from MEIC over China for 2016, 2017, 2018
Temporal resolution	Annual emission gridmaps + monthly profiles	Monthly emission gridmaps	Monthly emission gridmaps	Monthly emission gridmaps



<b>Spatial resolution</b>	0.1°x0.1°	0.1°x0.1°	0.1°x0.1°	0.1°x0.1° (The original spatial resolution of REASv3.2.1 is 0.25°x0.25°. Assuming that emissions are equally distributed in the 0.25° cell, REASv3.2.1 data were converted to 0.1° cell and provided to HTAP v3)
<b>Substances</b>	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC
<b>Geocoverage</b>				
<b>References</b>				<a href="http://meicmodel.org">http://meicmodel.org</a>

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<b>Data source</b>	CAPSS-KU	JAPAN (PM2.5EI and J-STREAM)	EDGARv6.1
<b>Type of data source</b>	Country inventory as emission time series by sector and country and emission gridmaps as .NetCDF files.	Country inventory as emission time series by sector and country and emission gridmaps as .NetCDF files.	Country inventory as emission time series by sector and country and emission gridmaps as .NetCDF files.
<b>Sectors coverage</b>	All sectors, excluding international shipping and aviation (international and domestic).	All sectors, excluding international shipping, domestic shipping and aviation (international and domestic).	All sectors, including international shipping and aviation (international and domestic)
<b>Temporal coverage</b>	2000-2018	2000-2017	2000-2018
<b>Temporal resolution</b>	Annual emission gridmaps + monthly profiles	Monthly emission gridmaps	Monthly emission gridmaps
<b>Spatial resolution</b>	0.1°x0.1°	0.1°x0.1°	0.1°x0.1°
<b>Substances</b>	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	SO <sub>2</sub> , NO <sub>x</sub> , CO, NMVOC, NH <sub>3</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC
<b>Geocoverage</b>			
<b>References</b>			<a href="https://edgar.jrc.ec.europa.eu/dataset_ap61">https://edgar.jrc.ec.europa.eu/dataset_ap61</a>

2



1 Table 2 – Definition of HTAP\_v3 sectors and correspondence to IPCC codes.

HTAP_v3 main sectors	HTAP_v3 detailed sectors	Sector description	IPCC 1996 codes	IPCC 2006 codes
HTAP_1: International Shipping	HTAP_1: International Shipping	International water-born navigation.	1C2	1.A.3.d.i
HTAP_2: Aviation	HTAP_2.1: Domestic Aviation	Civil Aviation.	1A3a <sub>ii</sub>	1.A.3.a.ii
	HTAP_2.2: International Aviation	International Aviation.	1A3a <sub>i</sub>	1.A.3.a.i
HTAP_3: Energy	HTAP_3: Energy	Power generation.	1A1a	1.A.1.a
HTAP_4: Industry	HTAP_4.1: Industry	Industrial non-power large-scale combustion emissions and emissions of industrial processes. It includes: manufacturing, mining, metal, cement, chemical and fossil fuel fires.	1A2 + 2 + 5B	1A2 + 2 (excluding 2.D.3 + 2.E + 2.F + 2.G) + 7A
	HTAP_4.2: Fugitive	It includes oil and gas exploration and production and transmission, including evaporative emissions (mainly NMVOC).	1B + 1A1b + 1A1c <sub>i</sub> + 1A1c <sub>ii</sub> + 1A5b <sub>iii</sub>	1.B + 1.A.1.b + 1.A.1.c.i + 1.A.1.c.ii + 1.A.5.b.i.i
	HTAP_4.3: Solvents	Solvents and product use.	3	2D3 + 2E + 2F + 2G
HTAP_5: Ground Transport	HTAP_5.1: Road Transport	Road Transport, combustion and evaporative emissions only.	1A3b (excluding resuspension)	1.A.3.b (excluding resuspension)
	HTAP_5.2: Brake and Tyre wear	Re-suspended dust from pavements or tyre and brake wear from road transport.	1A3b (resuspension only)	1.A.3.b (resuspension only)
	HTAP_5.3: Domestic shipping	Domestic shipping: inland waterways + domestic shipping.	1A3d <sub>2</sub>	1.A.3.d.ii
	HTAP_5.4: Other ground transport	Ground transport by pipelines and other ground transport of mobile machinery.	1A3c + 1A3e	1.A.3.c + 1.A.3.e.ii
HTAP_6: Residential	HTAP_6: Residential	Small-scale combustion, including heating, cooling, lighting, cooking and auxiliary engines, to equip residential, commercial buildings, service institutes, and agricultural facilities and fisheries.	1A4 + 1A5	1.A.4 + 1.A.5
HTAP_7: Waste	HTAP_7: Waste	Solid waste disposal and wastewater treatment.	6	4
HTAP_8: Agriculture	HTAP_8.1: Agricultural waste burning	Agricultural waste burning (excluding Savannah burning).	4F	3.C.1.b
	HTAP_8.2: Agriculture livestock	Livestock emissions, including manure management.	4B	3.A.2
	HTAP_8.3: Agriculture crops	Emissions from crops, fertilisers, and all agricultural soils activities.	4C + 4D	3.C.2 + 3.C.3 + 3.C.4 + 3.C.7

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2 **Table 3 – Overview of pollutant and sector provided by each inventory in HTAP\_v3. Cells with N/A indicate**  
 3 **that the emissions for those sectors were not provided and/or used in HTAP\_v3 for a specific inventory,**  
 4 **while gapfilled with the corresponding information from EDGARv6.1. The other cells represent the data**  
 5 **availability for each sector and inventory. The color codes used for the pollutants refer to the data source:**  
 6 **black color represents pollutant emissions provided by a specific inventory, red color indicates emissions**  
 7 **gapfilled using EDGARv6.1 and violet color indicates combinations of sectors-pollutants available for**  
 8 **specific regional inventories but not in EDGAR, which typically represent minor sources of emissions**  
 9 **included in officially reported inventories. These minor sources are included in the HTAP\_v3 mosaic.**

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Data provider	REAS v3.2.1	CAPS S-KU	JAPAN	ECCC	US EPA	CAMS-REG-v5.1	EDGARv 6.1
<b>HTAP_1: International Shipping</b>	N/A	N/A	N/A	N/A	N/A	N/A	All substances
<b>HTAP_2.1: Domestic Aviation</b>	N/A	N/A	N/A	N/A	N/A	N/A	All substances
<b>HTAP_2.2: International Aviation</b>	N/A	N/A	N/A	N/A	N/A	N/A	All substances
<b>HTAP_3: Energy</b>	All substances	All substances	BC, OC, NOx, <b>NH3</b> , CO, PM2.5, PM10, NMVOC, SO2	All substances	All substances	All substances	All substances
<b>HTAP_4.1: Industry</b>	All substances	All substances	BC, OC, NOx, <b>NH3</b> , CO, PM2.5, PM10, NMVOC, SO2	All substances	All substances	All substances	All substances
<b>HTAP_4.2: Fugitive</b>	All substances	BC, OC, NOx, <b>NH3</b> , CO, PM2.5, PM10, NMVOC, SO2	BC, OC, NOx, <b>NH3</b> , CO, PM2.5, PM10, NMVOC, SO2	All substances	All substances	All substances	All substances
<b>HTAP_4.3: Solvents</b>	NMVOC, <b>NH3</b> , <b>PM10</b> , <b>PM2.5</b>	NMVOC, <b>NH3</b> , <b>PM10</b> , <b>PM2.5</b>	NMVOC, <b>NH3</b> , <b>PM10</b> , <b>PM2.5</b>	NMVOC, <b>NH3</b> , <b>PM10</b> , <b>PM2.5</b>	<b>CO</b> , NOx, OC, NMVOC, <b>NH3</b> , PM10, PM2.5, SO2	NOx, <b>NH3</b> , CO, PM2.5, PM10, NMVOC, SO2	All substances
<b>HTAP_5.1: Road Transport</b>	All substances	All substances	All substances	All substances	All substances	All substances	All substances
<b>HTAP_5.2: Brake and Tyre wear</b>	N/A	BC, OC, <b>NOx</b> ,	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10	BC,OC,P M2.5,PM10



		NH <sub>3</sub> , CO, PM <sub>2.5</sub> , PM <sub>10</sub> , NMVOC, SO <sub>2</sub>			0, NMVOC		
<b>HTAP_5.3: Domestic shipping</b>	N/A	All substances	N/A	All substances	All substances	BC, OC, NO <sub>x</sub> , NH <sub>3</sub> , CO, PM <sub>2.5</sub> , PM <sub>10</sub> , NMVOC, SO <sub>2</sub>	All substances
<b>HTAP_5.4: Other ground transport</b>	All substances	All substances	BC, OC, NO <sub>x</sub> , NH <sub>3</sub> , CO, PM <sub>2.5</sub> , PM <sub>10</sub> , NMVOC, SO <sub>2</sub>	All substances	All substances	All substances	All substances
<b>HTAP_6: Residential</b>	All substances	All substances	BC, OC, NO <sub>x</sub> , NH <sub>3</sub> , CO, PM <sub>2.5</sub> , PM <sub>10</sub> , NMVOC, SO <sub>2</sub>	All substances	All substances	All substances	All substances
<b>HTAP_7: Waste</b>	N/A	All substances	All substances	All substances	All substances	All substances	All substances
<b>HTAP_8.1: Agricultural waste burning</b>	N/A	All substances	All substances	N/A	All substances	All substances	All substances
<b>HTAP_8.2: Agriculture livestock</b>	NH <sub>3</sub> , NMVOC, OC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>	NH <sub>3</sub> , NMVOC, OC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>	NH <sub>3</sub> , NMVOC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>	NH <sub>3</sub> , NMVOC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	NH <sub>3</sub> , NMVOC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	NH <sub>3</sub> , NMVOC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , OC	NH <sub>3</sub> , NMVOC, NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>
<b>HTAP_8.3: Agriculture_crops</b>	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, NMVOC, OC	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , BC, OC, CO, NMVOC, SO <sub>2</sub>	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub> , SO <sub>2</sub> , CO, OC, NMVOC	NH <sub>3</sub> , NO <sub>x</sub> , PM <sub>10</sub> , PM <sub>2.5</sub>

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