



1 **eFLaG: enhanced future FLoWs and Groundwater. A**  
2 **national dataset of hydrological projections based on**  
3 **UKCP18.**

4 **Jamie Hannaford<sup>1,2</sup>, Jonathan D. Mackay<sup>3,4</sup>, Matthew Ascott<sup>5</sup>, Victoria A.**  
5 **Bell<sup>1</sup>, Thomas Chitson<sup>1\*</sup>, Steven Cole<sup>1</sup>, Christian Counsell<sup>6</sup>, Mason Durant<sup>6</sup>,**  
6 **Christopher R. Jackson<sup>3</sup>, Alison L. Kay<sup>1</sup>, Rosanna A. Lane<sup>1</sup>, Majdi**  
7 **Mansour<sup>3</sup>, Robert Moore<sup>1</sup>, Simon Parry<sup>1</sup>, Alison C. Rudd<sup>1</sup>, Michael**  
8 **Simpson<sup>6\*\*</sup>, Katie Facer-Childs<sup>1</sup>, Stephen Turner<sup>1</sup>, John R. Wallbank<sup>1</sup>,**  
9 **Steven Wells<sup>1</sup>, Amy Wilcox<sup>6</sup>**

10 <sup>1</sup>UK Centre for Ecology & Hydrology, Maclean Building, Benson Lane, Crowmarsh Gifford,  
11 Wallingford, Oxon, OX10 8BB, UK

12  
13 <sup>2</sup>Irish Climate Analysis and Research UnitS (ICARUS), Maynooth University, Ireland

14 <sup>3</sup>British Geological Survey, Keyworth, Nottingham, NG12 5GG, UK

15 <sup>4</sup>School of Geography, Earth and Environmental Sciences, University of Birmingham,  
16 Edgbaston, B15 2TT, UK

17 <sup>5</sup>British Geological Survey, Maclean Building, Benson Lane, Crowmarsh Gifford,  
18 Wallingford, Oxon, OX10 8BB, UK

19 <sup>6</sup>HR Wallingford, Howbery Park, Crowmarsh Gifford, OX10 8BA

20 \*Now at: British Antarctic Survey, High Cross, Madingley Rd, Cambridge CB3 0ET

21 \*\*Now at: Natural Environment Research Council, Polaris House, Swindon, Wilts, SN2 1EU

22 Corresponding authors:

23 Jamie Hannaford [jaha@ceh.ac.uk](mailto:jaha@ceh.ac.uk)

24 Jonathan MacKay [joncka@bgs.ac.uk](mailto:joncka@bgs.ac.uk)

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## 29 **Abstract**

30 This paper presents an ‘enhanced future FLOws and Groundwater’ (eFLaG) dataset of  
31 nationally consistent hydrological projections for the UK, based on the latest UK Climate  
32 Projections (UKCP18). The hydrological projections are derived from a range of river flow  
33 models (Grid-to-Grid, PDM, GR4J and GR6J), to provide an indication of hydrological model  
34 uncertainty, as well as groundwater level (AquiMod) and groundwater recharge (ZOODRM)  
35 models. A 12-member ensemble of transient projections of present and future (up to 2080) daily  
36 river flows, groundwater levels and groundwater recharge were produced using bias corrected  
37 data from the UKCP18 Regional (12km) climate ensemble. Projections are provided for 200  
38 river catchments, 54 groundwater level boreholes and 558 groundwater bodies, all sampling  
39 across the diverse hydrological and geological conditions of the UK. An evaluation was carried  
40 out, to appraise the quality of hydrological model simulations against observations and also to  
41 appraise the reliability of hydrological models driven by the RCM ensemble, in terms of their  
42 capacity to reproduce hydrological regimes in the current period. The dataset was originally  
43 conceived as a prototype climate service for drought planning for the UK water sector, so has  
44 been developed with drought, low river flow and low groundwater level applications as the  
45 primary focus. The evaluation metrics show that river flows and groundwater levels are, for the  
46 majority of catchments and boreholes, well simulated across the flow and level regime, meaning  
47 that the eFLaG dataset could be applied to a wider range of water resources research and  
48 management contexts, pending a full evaluation for the designated purpose.

49

## 50 **1. Introduction**

51

52 This paper presents an ‘enhanced future FLOws and Groundwater’ (hereafter referred to as  
53 “eFLaG”) dataset of nationally consistent, and spatially coherent, hydrological (river flow and  
54 groundwater) projections for the UK, based on UKCP18 – the latest climate projections for the  
55 UK from the UK Climate Projections programme (Murphy et al. 2018). eFLaG provides a  
56 successor to the Future Flows and Groundwater Levels (FFGWL) dataset (Prudhomme et al.  
57 2013), which was based on the UKCP09 projections (Murphy et al. 2010).

58 The eFLaG dataset was developed specifically as a demonstration climate service for use by  
59 the water industry for water resources and drought planning, and hence by design is focused on  
60 future projections of drought, low river flows and low groundwater levels. By providing a  
61 consistent dataset of future projections of these variables, eFLaG can potentially support a wide  
62 range of applications across other sectors. The predecessor, FFGWL, has been widely used  
63 within the water industry, but also found very wide application for diverse research purposes  
64 (see Section 8).



65 As in FFGWL, in eFLaG the climate projections are used as input to a range of hydrological  
66 models to provide nationally consistent, spatially coherent projections of river flow and  
67 groundwater levels for the 21<sup>st</sup> century. The use of an ensemble of river flow models also  
68 provides information on hydrological model uncertainty. As well as using an updated set of  
69 climate projections, eFLaG capitalises on advances in national-scale river flow and  
70 groundwater modelling since FFGWL, and detailed evaluation of the applicability of models  
71 for drought simulation, notably research under the NERC Drought and Water Scarcity (DWS)  
72 Programme (e.g. Rudd et al. 2017; Smith et al. 2019).

73

#### 74 **Previous research on hydrological projections**

75 There is a long history of climate change impact assessment within the UK water industry and  
76 academia, which we do not review in detail here. Watts et al. (2015) provides an overview of  
77 past research (up to around 2013) on climate projections relevant for the water sector, including  
78 for future water resources and drought. However, as context for eFLaG it is worth considering  
79 some key developments since that review.

80 The original FFGWL did not present an assessment of future drought risk, other than seasonal  
81 river flows (Prudhomme et al. 2012) and groundwater levels (Jackson et al. 2015), which  
82 suggested: pronounced decreases in future summer flows; reductions in annual average  
83 groundwater levels; and increases (decreases) in winter (summer) groundwater levels. Since  
84 then, the original FFGWL projections have been used in a number of hydrological impact  
85 studies. Collet et al. (2018) presented a probabilistic appraisal of future river flow drought (and  
86 flood) hazard in the UK, showing hydro-hazard ‘hot-spots’ in western Britain and northeast  
87 Scotland, especially during the autumn. Hughes et al. (2021) used the ZOODRM distributed  
88 groundwater recharge model to assess changes in 21<sup>st</sup> century seasonal recharge across river  
89 basin districts and groundwater bodies in the UK based on the FFGWL climate change  
90 projections. The results showed a consistent trend of more recharge being concentrated over  
91 fewer months with increased recharge in winter and decreased recharge in summer.

92 In addition to UKCP09/FFGWL, other datasets have been developed using different Global  
93 Climate Model (GCM)/Regional Climate Model (RCM)/hydrological modelling chains. One  
94 major development has been the use of large ensemble projections of future climate variables  
95 from the Weather@Home RCM (specifically HadRM3P) as part of the MaRIUS project within  
96 the DWS Programme (Guillod et al., 2018). The MaRIUS projections provide large ensembles  
97 (100+) of past, present (1900–2006) and future (2020–2049 and 2070–2099) climate outputs.  
98 These were used as inputs to the national-scale Grid-to-Grid (G2G) hydrological model to  
99 provide a similarly large gridded (1km<sup>2</sup>) dataset of river flow and soil moisture (Bell et al.,  
100 2018). Analysis of these datasets has been conducted for drought (Rudd et al. 2019) and low  
101 flows (Kay et al. 2018), indicating future increases in hydrological drought severity and spatial  
102 extent, and decreases in absolute low flows.



103 A further source of hydro-meteorological projections now available are those from the EDgE  
104 project (End-to-end Demonstrator for improved decision-making for the water sector in  
105 Europe), see Samaniego et al. (2019). EdGE delivered an ensemble comprising of two GCMs  
106 and four ‘impact’ models (gridded land surface and hydrological models at a 5x5km scale) for  
107 the whole of Europe. Visser-Quinn et al. (2019) analysed future river flow drought risk in this  
108 ensemble, using a similar approach to Collet et al. (2018), and found similar results in terms of  
109 the spatial distribution and magnitude of future changes in droughts, albeit with some  
110 differences arising from the use of different scenarios, GCMs and hydrological models.

111 While such products may be used for climate adaptation research, the most relevant for eFLaG  
112 is the release of UKCP18. To date, relatively few studies using UKCP18 have been published.  
113 Kay et al. (2020) made a rapid assessment of UKCP18 impacts on hydrology compared to  
114 UKCP09. More recently, Kay (2021), Kay et al. (2021a,b,c) and Lane & Kay (2021) provided  
115 future assessments of potential changes in seasonal mean river flows, high flows and low flows  
116 using various UKCP18 products with the G2G hydrological model. They found potential  
117 increases in winter mean flows and high flows, and decreases in summer and low flows, albeit  
118 with wide uncertainty ranges. To date, and to the authors’ knowledge, there have been no  
119 published assessments of future groundwater levels or groundwater recharge using UKCP18.

120 In summary, there have been substantial scientific advances in hydrological projections for the  
121 UK since Watts et al. (2015) and FFGWL, including some research on future indicators relevant  
122 for water resource availability and drought. However, relatively few datasets have been made  
123 available to the community since FFGWL. While MaRIUS and EdGE provide complementary  
124 hydrological datasets, there remains a need for an accessible dataset based on UKCP18.  
125 Existing UKCP18 studies have been focused on time-slice projections and used a single  
126 hydrological model (e.g. Kay, 2021, a,b,c) so there will be significant benefit arising from the  
127 eFLaG dataset of transient projections from a range of hydrological models covering river  
128 flows, groundwater levels and groundwater recharge.

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## 130 **2. Outline of dataset and overview of the modelling chain**

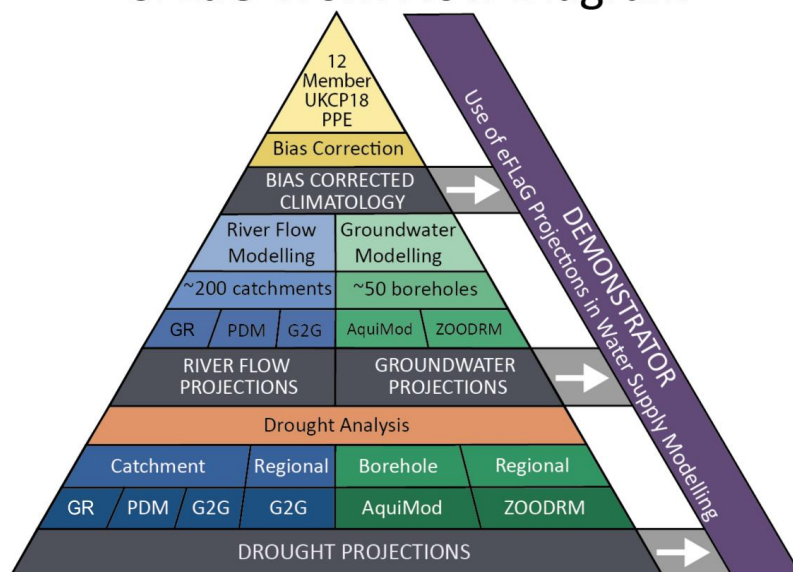
131

132 In the following sections we set out the methodology behind the eFLaG dataset. This section  
133 firstly provides a brief overview of the various stages of the methodology, and how our method  
134 samples the ‘cascade of uncertainty’ (Smith et al. 2019) emerging from the multiplicity of  
135 projections and other modelling choices. While the original FFGWL methodology provided an  
136 initial foundation for eFLaG, much has changed in the decade since that study was  
137 commissioned, and the new UKCP18 projections differ from UKCP09 (e.g. Kay et al 2020). .  
138 eFLaG therefore required the development of a new methodology, which is described in detail  
139 in the following sections.



140 The whole project workflow is illustrated in Fig 1. eFLaG is driven by the UKCP18 dataset,  
141 specifically the ‘Regional’ 12km projections, to which a bias correction is applied. Section 3  
142 describes the processing of the climate projections, including the bias correction method. The  
143 UKCP18 projections are used as input to three river flow models (GR, PDM and G2G), one  
144 groundwater level model (AquiMod) and one groundwater recharge model (ZOODRM) to  
145 provide simulations for 200 river catchments, 54 groundwater boreholes and 558 groundwater  
146 bodies respectively. Section 4 provides more detail on how these sites were selected. Details  
147 of the hydrological models and their calibration are given in Section 5. The evaluation of the  
148 models is covered in sections 6 and 7. Fig 1 also illustrates how all of the eFLaG projections  
149 are feeding into a series of water industry demonstrators, in partnership with UK water  
150 providers (specifically, Dwr Cymru/Welsh Water and Thames Water). These are not discussed  
151 in detail in this paper, but these were relevant for the site selection and as such are mentioned  
152 briefly below.

## eFLaG Work Flow Diagram



153

154 **Figure 1 Project workflow illustrating the stages of analysis described in this paper**

155 The question of uncertainty in climate impacts modelling is a challenging one that has been  
156 explored in a whole range of studies, going back as far as climate projections have been  
157 routinely produced from the 1980s. There are inherent uncertainties at every step of the process,  
158 from climate emissions scenarios through to climate modelling, and on to environmental  
159 modelling (in our case hydrological modelling, which itself has a vast literature when it comes  
160 to uncertainty estimation) and then to wider impacts modelling (e.g. in water supply systems).



161 Recently, Smith et al. (2018) presented this issue as a ‘cascade of uncertainty’ (using widely  
162 adopted terminology, e.g. Wilby and Dessai, 2010). Within eFLaG, as with the majority of  
163 climate impact applications, it is not possible to sample across all sources of uncertainty.  
164 Following Smith et al. (2019) we adopted a pragmatic approach to ‘crystalising’ the uncertainty  
165 within the available time and resource constraints. In Table 1, we consider the sources of  
166 uncertainty, and our approach to sampling from them. The focus in eFLaG is on uncertainty  
167 arising from initial/boundary conditions. Additionally, for the river flow simulations, the  
168 uncertainty arising from model choice is also accounted for, and within this, model structure is  
169 accounted for by considering two versions of one of the models.

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173 **Table 1: Sources of uncertainty explored in eFLaG**

Uncertainty Source	Sampling Approach	Details
Emissions Scenarios	One scenario	RCP8.5
Climate Models	One model	Hadley Centre GCM
Initial/Boundary Conditions	12x member PPE	PPE perturbs the parameters of the climate model (both the RCM, and the GCM within which it is nested)
Temporal/Spatial Downscaling	One method	Hadley Centre RCM, monthly mean bias correction
Model Choice	3x river flow models	GR, PDM, G2G
Model Structure	2x model structures for the GR modelling framework	Fixed structure for G2G and PDM, but for GR two different model structures were used (GR4J and GR6J), as discussed in section 4.

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175

### 176 **3. UKCP Data Processing**

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178 The regional climate projections were created using perturbed-parameter runs of the Hadley  
179 Centre global climate model (GCM) and regional climate models (HadGEM3-GC3.05 and  
180 HadREM3-GA705 respectively). These provide a set of 12 high resolution (12km) spatially  
181 consistent climate projections over the UK, covering the period Dec 1980-Nov 2080. The 12-  
182 member perturbed parameter ensemble (PPE) is valuable to represent climate model parameter



183 uncertainty. However, it is important to note that, as all ensemble members are based on the  
184 same high emissions scenario (RCP8.5) and underlying climate model structure, they do not  
185 represent the full climate uncertainty. The UKCP18 RCM output was processed to provide the  
186 variables needed for hydrological modelling – namely, 1km gridded and catchment-average  
187 time-series of available precipitation (i.e. after the application of a snow module, see below)  
188 and Potential Evapotranspiration (PET), not itself a UKCP18 output but estimated using  
189 available UKCP18 variables as described below.

190 The Hadley Centre climate model uses a simplified 360-day year, consisting of twelve 30-day  
191 months. The RCM precipitation and temperature time-series are given for this 360-day  
192 calendar, and are therefore not consistent with the 365/6-day observed time-series. Previously,  
193 the FFGWL Climate project inserted five (or six in a leap year) days of zero rainfall into the  
194 RCM time-series so that the observed and RCM data were using comparable calendars  
195 (Prudhomme et al., 2012). However, here the data were kept in the 360-day format, to avoid  
196 modifying the time-series with artificial data.

### 197 **Precipitation**

198 Daily precipitation time-series were available for each of the UKCP18 RCM-PPE members.  
199 However, the RCM data showed biases compared to observed precipitation, as is common for  
200 climate data (Murphy et al., 2018; Teutschbein & Seibert, 2012). A simple monthly-mean bias-  
201 correction methodology was therefore applied, through the following steps:

- 202 1. The 1km HadUK-Grid observed rainfall product was averaged to 12km for consistency  
203 with the RCM data (Hollis et al., 2019).
- 204 2. For each month and grid-cell, change factors were calculated between the RCM  
205 simulated precipitation and observation-based HadUK-Grid time-slice mean of monthly  
206 total rainfall over the period 1981-2010. This resulted in bias-correction factor grids  
207 being made for each month and RCM, as shown in Fig 2.
- 208 3. The change factor grids were then smoothed to prevent spatial discontinuities, by  
209 updating each grid cell using a weighted combination of the original grid-cell value and  
210 neighbouring values, as in Guillod et al. (2018).
- 211 4. To produce bias-corrected precipitation estimates, the RCM simulated precipitation  
212 time-series were multiplied by the bias-correction factor grid for each month (i.e. all  
213 January precipitation was multiplied by the January bias-correction grids, February  
214 precipitation by the February correction grid, etc.).

215 The bias-corrected precipitation products were then downscaled from 12km to 1km based on  
216 the distribution of the Standard-period Average Annual Rainfall (SAAR), as in previous studies  
217 (Bell et al., 2007; Kay & Crooks, 2014).

218

219



## 220 **Accounting for snowmelt processes**

221 A simple snow module was applied to account for snow-melt processes (Bell et al., 2016). The  
222 snow module converted the 1km bias-corrected precipitation into rainfall plus snowmelt (i.e.  
223 available precipitation), based on temperature. This used the minimum and maximum daily  
224 temperatures provided by each RCM ensemble member, which were first scaled from a 12km  
225 resolution to 1km using a lapse rate based on elevation data. The parameters used in the snow  
226 module are given in Supplementary Info (Table S1).

## 227 **Potential evapotranspiration**

228 Potential evapotranspiration (PET) was not directly available as an RCM output, and was  
229 therefore generated using a range of variables from the RCM-PPE climate time-series (Table  
230 S2). The calculation for PET was based on the CHES method (Robinson et al., 2016), with  
231 some details, in particular an interception correction, introduced from the MORECS method  
232 (Hough et al., 1997) – as Robinson et al. (2021), except with the bias-corrected precipitation  
233 used within the interception correction. The equation also included monthly stomatal resistance  
234 values, which were adjusted for the future period to account for the impact of increased carbon  
235 dioxide concentrations on stomata (as in Rudd & Kay, (2016), based on Kruijt et al., (2008)).  
236 The PET data were then copied down from a 12km to 1km resolution.

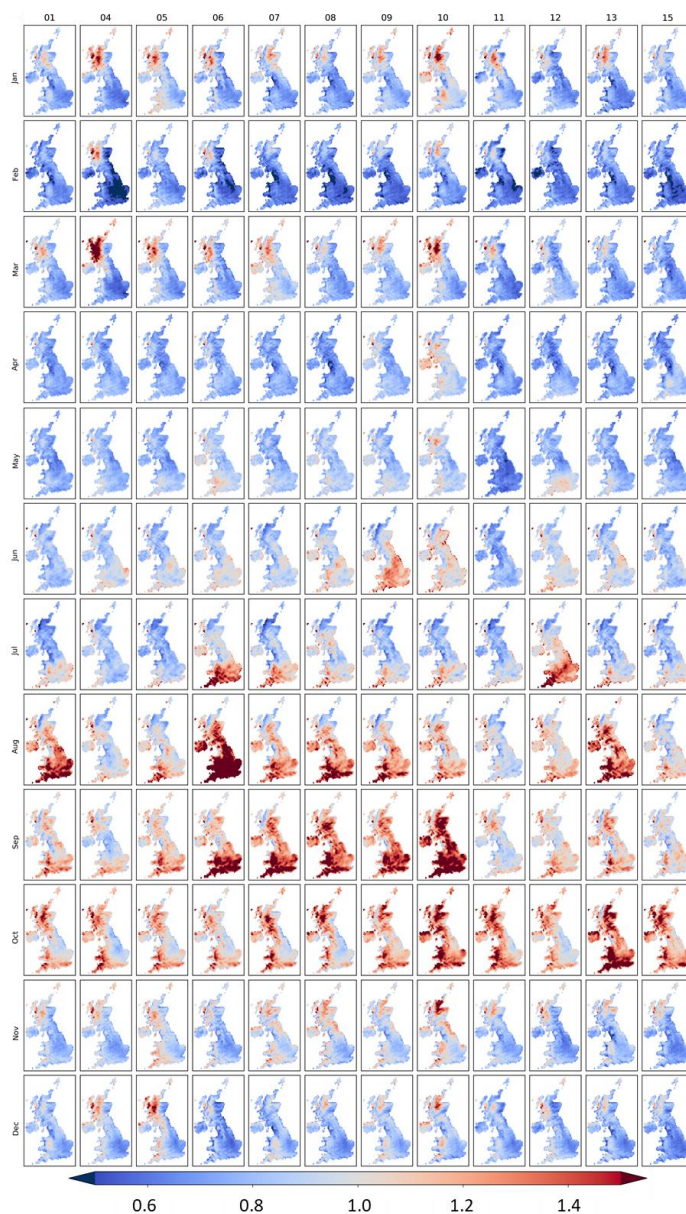
## 237 **Outputs**

238 The 1km gridded time-series of ‘available precipitation’ and PET were then used to produce  
239 the time-series of catchment-averages required for each of the eFLaG river catchments and  
240 groundwater boreholes. For the river catchments, the catchment average values were derived  
241 using the standard UK National River Flow Archive approach for catchment average rainfalls,  
242 as described in NRFA (2021). For the boreholes, following Mackay et al. (2014a), averages  
243 were taken over the representative aquifer length which was determined as the groundwater  
244 flow path between the borehole and a single discharge point on a river based on the catchment  
245 geometry and hydrogeology. For the grid-based models, ZOODRM and G2G, the gridded data  
246 were used directly.

247 The bias-corrected climate outputs are part of the eFLaG dataset described further in Section 9.  
248 For each river catchment and groundwater borehole, bias-corrected data are available for the  
249 observational period, for the purposes of evaluation of the hydrological model outputs, and for  
250 the future. In addition, the gridded bias-corrected climatology will be made available as a  
251 separate dataset in future.

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253

254 **Figure 2: Bias-correction grids applied to correct monthly precipitation. Values are**  
255 **correction factors used to modify precipitation, with a value of 0.5 halving precipitation,**  
256 **1 meaning no change to precipitation and 2 doubling precipitation etc. Columns show**  
257 **results from each RCM PPE member, rows show results for each month.**

258

259



#### 260 4. Catchment selection

261

262 The UK is fortunate to have one of the densest hydrometric networks in the world, with a legacy  
263 of strong commitment to data quality and completeness. There are more than 1,500 river flow  
264 gauging stations with flow records on the UK National River Flow Archive (NRFA, Dixon et  
265 al. 2013 and <https://nrfa.ceh.ac.uk/>) and more than 180 observation boreholes with groundwater  
266 level records on the BGS National Groundwater Level Archive (NGLA). These archives are  
267 the principal sources of validated river flow and groundwater level data at the UK scale. A remit  
268 of the NRFA and NGLA is to archive data that are useful for a wide variety of applications,  
269 primarily focusing on the most strategically important records. However, such catchments are  
270 not always the most relevant for the water industry, and water companies often have their own  
271 sites on which they undertake analysis. Since the eFLaG project aims to maximise utility for a  
272 range of users, the catchment selection strategy considered both research and industry needs.

273 Detailed site lists and metadata for river flow, groundwater level and groundwater recharge are  
274 catalogued on the EIDC dataset (Hannaford et al. 2022).

#### 275 River Flows

276 To support selection, a metadatabase was assembled for all NRFA gauging stations in the UK,  
277 primarily using the NRFA's metadata holdings published on the NRFA website and in the UK  
278 Hydrometric Register (Marsh and Hannaford, 2008). Metadata compiled included membership  
279 of key national strategic networks (e.g. near-natural Benchmark (UKBN2; Harrigan et al.  
280 2018a) and operational monitoring networks), capitalising on efforts of other projects in quality  
281 controlling data and ensuring catchments are fit for purpose. Selection also considered whether  
282 catchments were used in previous relevant projects that have simulated river flows for drought  
283 analysis. The selection ensured a strong representation of the original FFGWL catchments (with  
284 117 catchments featuring in both) and also overlap with recent modelling endeavours through  
285 the DWS Programme (AboutDrought, 2021) projects 'Historic Droughts', 'IMPETUS' and  
286 'MaRIUS' projects, that used several of the models used by eFLaG (specifically G2G, GR4J).  
287 In this regard we ensured that 165 eFLaG catchments overlapped with at least one DWS project.

288 Selection also focused on data quality. Longer record lengths were prioritised and hydrometric  
289 quality was evaluated where possible. Given the extent of hydrometric issues (at low flows  
290 especially) it is not possible for all sites to have the highest quality data, but where decisions  
291 were made on similar sites, quality was considered as a tiebreaker. The selection included 80  
292 Benchmark catchments, but did not seek to focus entirely on natural catchments given the  
293 limited range of variability they capture (being mostly small and clustered in headwaters), and  
294 also included large and disturbed sites known to be important for water industry purposes.

295 Catchment representativeness was also considered, enabling the eFLaG dataset to sample the  
296 hydrological variability of the UK. Representativeness was considered by comparing the  
297 distribution of eFLaG potential selections relative to various catchment descriptors from the



298 NRFA Hydrometric Register (altitude, area, annual rainfall, Base Flow Index, land cover and  
299 so on).

300 Finally, this activity focused on ensuring water industry relevance. At the national scale, this  
301 was achieved by asking stakeholders at an eFLaG workshop for views on additional catchments  
302 (Durant et al. 2022). In this way, 12 catchments were added. Similarly, for the regional  
303 demonstrators (Dwr Cymru/Welsh Water and Thames Water), water company teams were  
304 consulted to gain a better understanding of strategically important flow records for water  
305 companies in the case study regions, leading to an additional five catchments.

306 The final eFLaG dataset consists of 200 catchments (Fig. 3a) giving good geographical  
307 coverage and representativeness of the UK.

### 308 **Groundwater Levels**

309 Boreholes were selected to ensure a number of essential criteria were met. Firstly, only those  
310 boreholes with the highest-quality records of groundwater level were considered. This required  
311 regular (at least monthly) and continuous (at least 10 years in length) records of data from  
312 boreholes that are in zones which are not significantly affected by groundwater abstraction.

313 Secondly, sites were chosen to ensure coverage of the UK's principal aquifers where possible,  
314 enabling the eFLaG dataset to sample the hydrogeological variability of the UK. This broadly  
315 aligns with the requirements of other national-scale assessments of groundwater resources  
316 undertaken as part of the original FFGWL project and the 'Historic Droughts' and 'IMPETUS'  
317 projects. Accordingly, the selection aimed to ensure good coherence with these studies also.

318 Thirdly, as with river flow catchment selection, an additional activity focused on ensuring water  
319 industry relevance, both at the national scale, through consultation with stakeholders at the  
320 eFLaG workshop, and through consultation with key demonstrator partners (Dwr  
321 Cymru/Welsh Water and Thames Water) who identified strategically important boreholes that  
322 would strengthen the outputs for long-term drought risk assessment to support the water  
323 resources planning case study. Through this activity, several additional boreholes were  
324 identified.

325 These selection criteria identified over 70 'candidate' boreholes for the eFLaG project. A final  
326 quality assurance procedure was then undertaken whereby a preliminary analysis of AquMod's  
327 ability to capture low groundwater levels was undertaken at each borehole via visual inspection  
328 of the simulated hydrographs. A final set of 54 boreholes was selected (Fig. 3b). They represent  
329 a significant advance in aquifer coverage compared to the 24 NGLA boreholes used in FFGWL,  
330 15 of which are used in both.

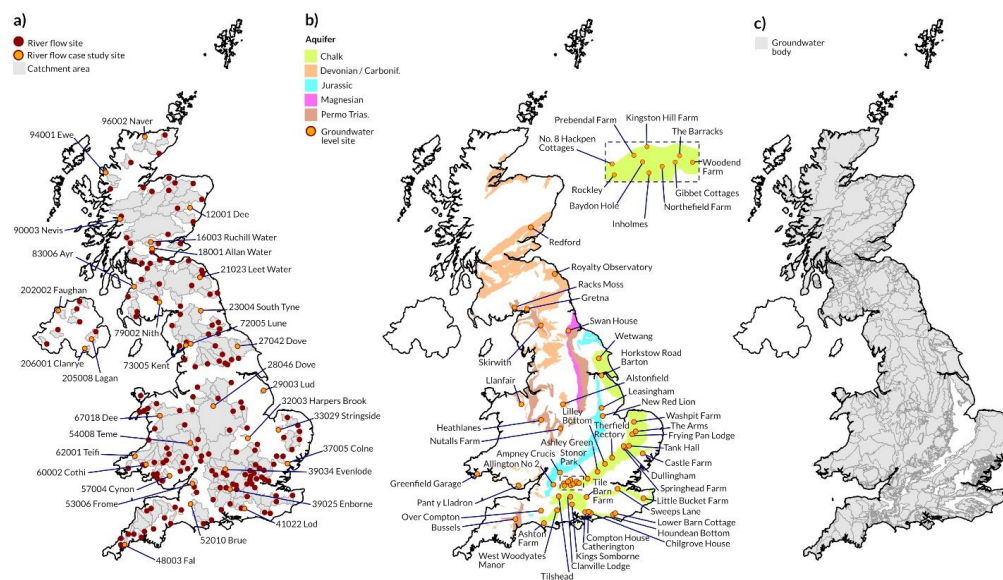
### 331 **Groundwater Recharge**

332 The gridded groundwater recharge simulations have been aggregated over 558 'groundwater  
333 bodies' covering England (Environment Agency, 2021a), Wales (Natural Resources Wales,



334 2021) and Scotland (Ó Dochartaigh et al., 2015) (Fig. 3c). These units were used for two  
335 principal reasons. Firstly, they are physically justifiable as they reflect known hydrogeological  
336 characteristics including groundwater recharge and groundwater flow regimes so that each  
337 catchment represents a distinct body of groundwater that can reasonably be considered in  
338 isolation. Secondly, they are coherent with the licensing areas defined as part of Catchment  
339 Abstraction Management Strategy (Environment Agency 2021b) and management areas for the  
340 implementation of the Water Framework Directive. They are, therefore, directly relevant to  
341 water regulation and the wider water industry.

342



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344

345 **Figure 3 a) Map of the 200 eFLaG catchments - highlighting those used as Case Study sites; b) Map of 54**  
346 **eFLaG boreholes and principal UK Aquifers including The Chalk, Devonian and Carboniferous aquifers**  
347 **(Devonian/Carbonif.), Jurassic limestones (Jurassic), Magnesian limestones (Magnesian) and Permo-**  
348 **Triassic sandstones (Permo Trias.); c) Map of 558 groundwater bodies. Inset of Figure 3b shows the**  
349 **Berkshire downs where there are a high number of boreholes.**

350

## 351 5. Hydrological and groundwater model ensemble setup

352

353 Creation of an enhanced Future Flow and Groundwater (eFLaG) dataset is underpinned by  
354 hydrological and groundwater models used to transform rainfall and potential evaporation (PE)  
355 to river flow, soil moisture, groundwater levels and recharge. The approach builds on that



356 employed under FFGWL (Prudhomme et al. 2013) whilst exploiting developments in  
357 hydrological modelling for droughts since that time.

358 For modelling of river flows, eFLaG used two lumped catchment models, PDM (Moore 2007)  
359 and the GR suite (Perrin et al. 2003), and one distributed grid-based hydrological model, Grid-  
360 to-Grid (G2G; Bell et al. 2009). PDM was used in FFWGL and therefore provides some  
361 comparability with that project. Embracing a range of different model structures and spatial  
362 representations can provide insights into how assessments of future river flows (and hence,  
363 drought or low flow risk under climate change) is sensitive to hydrological model choice. For  
364 groundwater, eFLaG adopted the lumped, conceptual, Aquimod groundwater model (Mackay  
365 et al. 2014a) to simulate groundwater level time series on a daily time step at the boreholes  
366 identified in Section 4. Aquimod was the groundwater level model used in FFGWL providing  
367 direct comparison. In addition to groundwater levels, the zooming object oriented distributed  
368 recharge model (ZOODRM) (Mansour and Hughes, 2004) was used to study changes in future  
369 groundwater recharge.

370 In the following sub-sections, we describe each of these models in turn, providing information  
371 on the model set-up, calibration and past approaches to evaluation. A consistent approach was  
372 applied to the model application and evaluation across all these models where possible.  
373 However, it is important to emphasise that while some aspects were common, insofar as  
374 possible (e.g. model driving data), it was necessary to apply different approaches to suit the  
375 model in question. Calibration was done according to past applications and best-practice.  
376 Hence, the calibration approach described below is similar for the GR suite and PDM, but  
377 different for Aquimod, and by its nature G2G requires no specific calibration here. Identical  
378 approaches to evaluation were adopted across all river flow models, but minor differences  
379 applied with groundwater, as described below.

380 There are two sets of model output in eFLaG, described below – this terminology is adopted  
381 throughout.

- 382 • simobs: observation-driven simulation (i.e. simulations for the observed period, driven  
383 by observational climate datasets, described below). The simobs period varies between  
384 models, but covers at least the January 1961 – December 2018 period.
- 385 • simrcm: UKCP18 RCM-driven simulation (12 ensemble members) (i.e. simulations  
386 driven by the UKCP18 RCM bias-corrected dataset as described in Section 3). These  
387 are available for 1980 to 2080. The simrcm runs from the observed period could then  
388 be evaluated against the simobs data.

389 Common driving data was applied across all models for the simobs runs. Accepted national-  
390 standard observational climate products were used, including:

- 391 • Precipitation and temperature: HadUK-Grid 1km x 1km dataset (Hollis et al. 2019), the  
392 national standard gridded meteorological dataset and observational product associated  
393 with UKCP18.



394 • Potential Evaporation (PE). MORECS (Hough et al., 1997), an established, national  
395 gridded PE product. Other PE datasets such as CHESS (Robinson et al., 2017) and more  
396 recently the Environment Agency’s PE product (Environment Agency, 2021c) are  
397 available, however the decision to use MORECS was based on availability of data for  
398 the whole of the UK.

399

400 For all models, evaluation was undertaken in two stages, which is typical practice for appraising  
401 a model for simulation of climate change impacts:

- 402 1. Evaluation when driven with baseline observed climate data
- 403 2. Evaluation when driven with baseline climate model data.

404 Stage 1 involves the use of a range of statistics to assess the performance of model simulations  
405 driven by observed climate data (the simobs runs) against observations of river flow and  
406 groundwater. For Stage 1, a range of metrics are available and widely used to assess how well  
407 rainfall-runoff or groundwater models perform against observations. Within eFLaG, a range of  
408 different metrics were used to assess performance (Table 3). For river flows, these metrics have  
409 a focus on low flow metrics (e.g. NSE on log-transformed flows), but some do evaluate  
410 performance across the flow regime. For groundwater levels, a generalised NSE score was used  
411 which provides an overall assessment of process realism and fit to groundwater level data. The  
412 simulated and observed Standardized Groundwater level Index (SGI) were also compared using  
413 the NSE ( $NSE_{SGI}$ ) which focusses in on groundwater extremes including droughts.

414 It is not possible to do a thorough evaluation of the recharge simulations from ZOODRM, given  
415 the difficulty in measuring recharge, particularly at a scale that is commensurable with a  
416 national model. However, past applications of ZOODRM (e.g. Mansour et al., 2018) have  
417 successfully used monthly river flow data as a means to evaluate ZOODRM’s ability to capture  
418 catchment water balances and infer the accuracy of seasonal recharge simulations (further  
419 details provided in model description). Accordingly, a subset of the river flow metrics relevant  
420 to monthly river flows have been used to evaluate ZOODRM for stage 1.

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433 **Table 3.** Model calibration and evaluation metrics used in eFLaG.

Evaluation Metric	Equation	Focus
<b>Nash-Sutcliffe Efficiency (<math>R^2</math> Efficiency)</b>	$NSE = 1 - \frac{\sum_{i=1}^n (Q_i - q_i)^2}{\sum_{i=1}^n (Q_i - \bar{Q})^2}$ <p><math>Q_i</math> and <math>q_i</math> are observed and modelled flow for day <math>i</math> of a <math>n</math> day record. <math>\bar{Q}</math> is the mean observed flow.</p> $NSE = 1 - \frac{\sum_{i=1}^n (H_i - h_i)^2}{\sum_{i=1}^n (H_i - \bar{H})^2}$ <p><math>H_i</math> and <math>h_i</math> are observed and modelled groundwater level for day <math>i</math> of a <math>n</math> day record. <math>\bar{H}</math> is the mean observed groundwater level.</p>	High Flows/Generalised groundwater levels
<b>Nash-Sutcliffe Efficiency log flows*</b>	$NSE_{log} = 1 - \frac{\sum_{i=1}^n (\log(Q_i) - \log(q_i))^2}{\sum_{i=1}^n (\log(Q_i) - \log(\bar{Q}))^2}$	Low Flows
<b>Nash-Sutcliffe Efficiency square root flows</b>	$NSE_{sqrt} = 1 - \frac{\sum_{i=1}^n (\sqrt{Q_i} - \sqrt{q_i})^2}{\sum_{i=1}^n (\sqrt{Q_i} - \sqrt{\bar{Q}})^2}$	Generalised Flows
<b>Nash-Sutcliffe Efficiency standardised groundwater level index</b>	$NSE_{SGI} = 1 - \frac{\sum_{i=1}^n (SGI_i - sgi_i)^2}{\sum_{i=1}^n (SGI_i - \overline{SGI})^2}$ <p><math>SGI_i</math> and <math>sg_i</math> are observed and modelled SGI for day <math>i</math> of a <math>n</math> day record. <math>\overline{SGI}</math> is the mean observed SGI.</p>	Groundwater extremes
<b>Modified Kling Gupta Efficiency [square root flows]</b>	$KGE'_{sqrt} = 1 - \sqrt{(r - 1)^2 + (\beta - 1)^2 + (\gamma - 1)^2}$ <p>where <math>r</math> is the correlation coefficient, <math>\beta</math> is the bias ratio <math>\frac{\mu_{\sqrt{q}}}{\mu_{\sqrt{Q}}}</math>, and <math>\gamma</math> is the variability ratio <math>\frac{CV_{\sqrt{q}}}{CV_{\sqrt{Q}}}</math> or <math>\frac{\sigma_{\sqrt{q}}/\mu_{\sqrt{q}}}{\sigma_{\sqrt{Q}}/\mu_{\sqrt{Q}}}</math></p> <p><math>\mu</math>, <math>\sigma</math> and <math>CV</math> are the mean, standard deviation and coefficient of variation of flow (here of the square root of modelled and observed flows as indicated by the suffix)</p>	Generalised flows
<b>Absolute Percent Bias</b>	$absPBIAS = \left  \frac{\sum (q_i - Q_i)}{\sum Q_i} \right  100$	Water Balance



<b>Mean Absolute Percent Error</b>	$MAPE = \left( \frac{1}{n} \sum_{i=1}^n \left  \frac{Q_i - q_i}{Q_i} \right  \right) 100$	Systematic
<b>Absolute Percent Error in Q95</b>	$Q95_{APE} = \left  \frac{Q95 - q95}{Q95} \right  100$	Low Flows
<b>Low Flow Volume</b>	$LFV = 100 \frac{\sum_{p=70}^{95} (\sqrt{q_p} - \sqrt{Q_p})}{\sum_{p=70}^{95} (\sqrt{Q_p})}$ Here $q_p$ and $Q_p$ are the modelled and observed flow $p$ percentiles	Low Flows
<b>Absolute Percent Error in the Mean Annual Minimum on a 30-day moving average*</b>	$MAM30_{APE} = \left  \frac{QMAM30 - qMAM30}{QMAM30} \right  100$ where $QMAM30$ $= \frac{1}{n} \sum_{j=1}^n \min_j \left( \frac{Q_{j,i-29} + Q_{j,i-28} + Q_{j,i-27} \dots Q_{j,i-1} + Q_{j,i}}{30} \right)$ Here $Q_{j,i}$ is observed flow for day $i$ of hydrological year $j$ for a record of $n$ years	Low Flows
*1/100 <sup>th</sup> of the mean observed flow was added to both modelled and observed flow values during evaluation in order to avoid errors and biases due to very small and zero flows.		

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437 Sources of quality controlled, long-term observational data for model calibration and evaluation were  
 438 the national standard repositories for hydrological data:

- 439 • River Flows: UK National River Flow Archive <https://nrfa.ceh.ac.uk/>
- 440 • Groundwater Levels: UK National Groundwater Level Archive  
 441 <https://www2.bgs.ac.uk/groundwater/datainfo/levels/ngla.html>

442

443 Stage 2 appraises the performance of the models when driven by the climate model outputs. That is, it  
 444 compares the simobs and simrcm runs over the common baseline period. This assessment cannot use  
 445 performance metrics based on time-series, as climate models are not expected to reproduce the  
 446 sequencing of events seen over the historical period (Kay et al. 2015). Instead, the comparison has  
 447 been done in terms of river flow and groundwater level duration curves, low flow/level metrics and  
 448 seasonal recharge values. Thus, comparing the statistical characteristics of river flows, groundwater  
 449 levels and groundwater recharge rather than their day-to-day equivalence (Kay et al. 2015, 2018).  
 450 When looking at the performance of an ensemble of climate model runs, the model simulation driven  
 451 by observed data would ideally sit within the range covered by the ensemble (assuming an ensemble





452 of sufficient size). However, it would not necessarily be expected to sit in the middle of the ensemble  
453 range, because the set of weather events that actually occurred within the historical observed baseline  
454 period is just one realisation of what could have occurred within the range of natural variability (Kay  
455 et al. 2018).

456

## 457 **Description of the models and specific setup**

### 458 **GR4J/GR6J**

459 The GR4J and GR6J models come from a suite of hydrological models provided in the “airGR”  
460 modelling suite (Coron et al. 2021) for the R software programme. Both models are well suited to  
461 application across many catchments using the inbuilt automatic parameter optimisation function. The  
462 simple, efficient form of airGR models also make them suitable for uncertainty and ensemble analyses.

463 GR4J (Génie Rural à 4 paramètres Journalier) is a simple daily lumped conceptual model with only  
464 four free parameters. GR4J has been used for hydro-climate change research across the globe, and has  
465 demonstrated good performance in a diverse set of catchments in the UK. The model has been applied  
466 in the UK for operational seasonal forecasting, as well as for long-term drought reconstructions  
467 nationwide (Harrigan et al. 2018b, Smith et al. 2019).

468 GR6J (Génie Rural à 6 paramètres Journalier) (Pushpalatha et al. 2011) is a six parameter variant of  
469 the GR modelling suite that was developed to improve low flow simulation and groundwater exchange.  
470 Recently, GR6J has increasingly been applied in UK water resources applications (e.g. Anglian Water  
471 Drought Plan, 2021).

472 For eFLaG, it was decided, therefore, that using both GR4J and GR6J would be beneficial. Both GR4J  
473 and GR6J were calibrated using the inbuilt automatic calibration function, with the modified Kling  
474 Gupta Efficiency (KGE, Gupta et al, 2009; Kling et al 2012) as the Error criterion (‘ErrorCritKGE2’).  
475 KGE offers a thorough error criterion as it calculates the correlation coefficient, the bias and the  
476 variability between simulated and observed flows. KGE values range from  $-\infty$  to 1, with 1 being a  
477 perfect fit. The calibration algorithm was applied to square-root transformed flows in order to place  
478 weight evenly across the flow regime. The airGR snowmelt module “CemaNeige” was not applied, as  
479 a simple snow module was applied to the climate data to pre-process the precipitation data into rainfall  
480 and snowmelt based upon temperature (See section 3).

### 481 **Grid-to-Grid**

482 The Grid-to-Grid (G2G) hydrological model is an established area-wide distributed model that has  
483 been used to investigate the spatial coherence and variability of floods and droughts at catchment,  
484 regional and national scales. Model output typically consists of natural river flows at both gauged and  
485 ungauged locations, and can be provided as time-series for specific locations as well as 1km x 1km  
486 grids. The G2G has been used for climate impacts modelling of floods (Bell et al., 2009, 2012), low



487 flows (Kay et al., 2018) and droughts (Rudd et al., 2019) and is also used operationally for flood  
488 forecasting (Cole and Moore, 2009; Moore et al., 2006).

489 The G2G is typically configured on a 1km×1km grid using spatial datasets of landscape properties  
490 such as soil type and drainage network, together with a few nationally-applied model parameters. The  
491 model is thus parameterised using national-scale spatial datasets (e.g. soil grids), rather than via  
492 individual catchment calibration. The spatial datasets and parameters used here are the same as those  
493 used in previous studies (Rudd et al., 2019; Bell et al., 2009, 2012; Kay et al., 2018).

494 The G2G can either be initialised with model water stores set to default or zero values, or from a states  
495 file appropriate to the run start date. In eFLaG the G2G was run for two years with observed rainfall  
496 and PE to provide a 1 January 1963 states file to initialise the observation-driven G2G model run. The  
497 RCM-driven G2G runs were all initialised with a generic December states file provided by an obs-  
498 driven run (for 1 December 1980), then the first two years of each RCM-driven run were discarded to  
499 allow for model spin up. The eFLaG river flow datasets therefore cover the periods, 1 January 1963 to  
500 31 December 2018 (simobs) and 1 December 1982 to 30 November 2080 (simrcm).

## 501 **PDM**

502 The Probability Distributed Model or PDM (Moore, 2007; UKCEH, 2021) is a simple, very widely  
503 used lumped rainfall-runoff model that can be configured to a variety of catchment flow regimes. A  
504 brief summary follows but full details are available in Supplementary info S.2.

505 Within the model, a soil water store with a distribution of water absorption capacities controls runoff  
506 production through a saturation excess process; stored water is also lost to evaporation. In one  
507 configuration, all runoff enters a surface store (the fast pathway) while a groundwater store (the slow  
508 pathway) is recharged by soil water drainage. In an alternative configuration, the runoff is split between  
509 the two stores according to a fixed fraction. Water in the surface- and ground-water stores is routed  
510 using a non-linear storage equation (powers of 1, 2 and 3 were trialled under eFLaG), or, for the surface  
511 store, a cascade of two linear reservoirs, before being combined to produce the modelled flow at the  
512 catchment outlet. Water is conserved within the model, whilst a multiplicative factor (equal to 1 if not  
513 required) is applied to the input precipitation. Alternatively, a Groundwater Extension (Moore and  
514 Bell, 2002) may be invoked to allow modelling of underflow at the catchment outlet, external springs,  
515 pumped abstractions, and the incorporation of well level data. Multiple hydrological response zones  
516 within a catchment can also be represented (not trialled under eFLaG). PDM may be thought of as a  
517 toolkit of model components representing a range of runoff production and flow routing behaviours,  
518 and with a choice of time-step.

519 Under eFLaG, single zone PDM models were invoked with a daily time-step. The model stores were  
520 initialised using the mean observed flow over the period of record, and the first two years of model  
521 flow discarded to allow for model spin-up. Nineteen different combinations of the above-mentioned  
522 toolkit options were systematically trialled for each catchment. Parameter estimation was performed  
523 using an automatic calibration procedure that applied a simplex optimisation scheme (Nelder and



524 Mead, 1965) to different combinations of model parameters in turn. The rainfall factor, or, when  
525 employed, a spring factor (representing net water exchange for the catchment), were used to achieve  
526 zero bias in the modelled flows with respect to observations. Remaining parameters were estimated so  
527 as to optimise the modified Kling-Gupta Efficiency calculated on either the square root transformed  
528 flows, or, to a lesser extent, the log transformed flows. Each calibration began from multiple different  
529 initial parameter choices, with model parameters and performance metrics output at three increasingly  
530 aggressive calibration stages. This produced a total of 138 candidate PDM model calibrations per  
531 catchment. Final selection among these candidates first excluded any models deemed unphysical, such  
532 as those containing extreme model parameter values, or using the Groundwater Extension for  
533 inappropriate catchments. The best remaining candidate was then selected according to a weighted  
534 sum of the modified Kling-Gupta Efficiency calculated on square root ( $KGE'_{sqr}$ ) and log ( $KGE'_{log}$ )  
535 transformed flows, with weights of 0.8 and 0.2 respectively.

### 536 **AquiMod**

537 AquiMod is a lumped conceptual groundwater model that links simplified equations of soil drainage,  
538 unsaturated zone flow, and saturated groundwater flow to simulate daily groundwater level time series  
539 at a specified borehole (Mackay et al., 2014b). Each of these three components use model parameters  
540 that describe site-specific hydrological and hydrogeological characteristics of the groundwater  
541 catchment surrounding the borehole. The model also has a flexible saturated zone model structure that  
542 can be modified to represent different levels of vertical heterogeneity in hydrogeological properties.

543 For each borehole, the AquiMod parameters and structure were calibrated to achieve the most efficient  
544 simulation of available historical groundwater level data using the Nash-Sutcliffe Efficiency (NSE),  
545 which provides a reliable assessment of overall process realism and goodness of fit to groundwater  
546 level time series; following the approach of Mackay et al. (2014a) and Jackson et al. (2016), model  
547 parameters that could be related to catchment information (e.g. relating to known land cover and soil  
548 type) were fixed. The remaining parameters were then calibrated, using six different saturated zone  
549 model structures including a one-layer model (fixed hydraulic conductivity and specific yield); two-  
550 and three-layer models with variable hydraulic conductivity and fixed specific yield; two- and three-  
551 layer models with variable hydraulic conductivity and variable specific yield; and a 'cocktail glass  
552 representation of hydraulic conductivity variation with depth (Williams et al., 2006). The optimal  
553 structure-parameter combination was obtained for each borehole using the Shuffled Complex  
554 Evolution global optimisation algorithm.

555 The calibrated models were then evaluated for their ability to capture groundwater level extremes using  
556 the Standardized Groundwater level Index, SGI (Bloomfield and Marchant, 2013) as the basis for this  
557 evaluation. The SGI is a normalised index, calculated directly from groundwater level time series,  
558 which can be used to identify droughts and provide a quantitative status of groundwater resources  
559 drought events (e.g. Bloomfield et al., 2019).

560



## 561 **ZOODRM**

562 ZOODRM is a distributed recharge calculation model originally developed to estimate recharge values  
563 to drive groundwater models (Mansour and Hughes, 2004). It is applied over the British Mainland  
564 using a 2km square grid. The FAO Drainage and Irrigation Paper 56 (FAO, 1988) approach, modified  
565 by Griffiths et al. (2006), is used to calculate potential recharge. This method removes actual  
566 evaporation and soil moisture deficit from rainfall and calculates potential recharge as a fraction of the  
567 excess water using a runoff coefficient value. The model was driven by daily rainfall and potential  
568 evaporation data. The model was primarily parameterised using available national scale data including  
569 data relating to the soil hydrology (Boorman et al., 1995), vegetation (LCM2000, NERC) and surface  
570 topography. The latter of these was used to route surface water runoff.

571 The runoff coefficient, which defines the proportion of excess soil water that drains overland via  
572 surface runoff, is an unknown parameter which must be calibrated. This was done in two stages. Firstly,  
573 the calibration problem was simplified by defining zones of equal runoff coefficient. In total 35 zones  
574 were used in ZOODRM which were based on UK hydrogeological and geological maps (DiGMapGB-  
575 625, 2008). Then, the runoff coefficient for each zone was manually calibrated by comparing simulated  
576 runoff to observed river flows minus baseflow which was calculated using a well-established baseflow  
577 separation method (Gustard et al., 1992). This was done using monthly mean flows given that  
578 ZOODRM does not have a sophisticated runoff routing scheme, and it is not expected, therefore, to  
579 capture daily variability in runoff. The comparison to monthly flows does, however, provide a useful  
580 means to evaluate the seasonal water balance of the model which serves as the best available proxy for  
581 the accuracy of the recharge simulations. In total, 41 gauging stations were used to assess the model  
582 performance.

583 The only hydrological process that needs initialisation in the ZOODRM is the soil moisture deficit. As  
584 all simulations start in January, which is a wet month with minimal potential evaporation, it is assumed  
585 that the initial soil moisture deficit is equal to zero. Even so, a warm up period of one year is used to  
586 initialise the model.

587

## 588 **6. Hydrological model evaluation (Stage 1 evaluation)**

589

590 This section provides a brief summary of the outputs of the Stage 1 evaluation. Note that for river  
591 flows, model evaluation was undertaken at the same gauged locations and for the same period of time  
592 used for model calibration, except G2G which is not specifically calibrated.

### 593 **River Flows**

594 Fig. 4 summarises the range of Stage 1 evaluation metrics across all catchments, while Supplementary  
595 Figs S2 to S5 provide maps of the evaluation metrics at each catchment. For GR4J, generally there  
596 was good performance across performance metrics in most catchments. Some outliers are present in



597 the drought metrics, particularly in the South East and London. For GR6J, we observed good  
598 performance across all performance and drought metrics. GR6J generally performs slightly better than  
599 GR4J, particularly as shown in low flow catchments in the logNSE metric. For PDM, very good scores  
600 are obtained across the 200 sites, especially the low flow/drought indicators (bottom rows).

601 For G2G, again, good performance was observed overall (medians for NSE/ logNSE/ sqrtNSE/ KGE2  
602  $\geq 0.7$ ). However, the performance was generally lower than for GR or PDM because the G2G is not  
603 calibrated to individual catchments, and G2G simulates *natural* flows, whereas the lumped models are  
604 calibrated to the observations used for performance assessment. In catchments with a high degree of  
605 anthropogenic disturbance, G2G is less able to simulate observed flows, whereas the calibration of the  
606 other hydrological models will implicitly account for such artificial impacts, to a degree.

607 This distinction highlights an important benefit of eFLaG: PDM and GR4J/GR6J are calibrated to  
608 present-day flows and hence simulated flows are not natural, as they implicitly include artificial  
609 impacts. These runs do not, therefore, allow users to separate natural flows and artificial influences in  
610 the baseline period, nor to project how they may change relative to each other in future. On the other  
611 hand, although not used here, G2G has the capability of including artificial influences separately (e.g.  
612 Rameshwaran et al., 2022), and specifically modelling their future evolution. Furthermore, G2G's  
613 response to rainfall may be less tailored to the present-day climate than the calibrated models. The  
614 eFLaG hydrological model ensemble therefore includes models that may be beneficial for different  
615 applications according to the particular needs of end-users.

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630 **Figure 4: Evaluation results summarised across the different models for the key evaluation metrics**

631 In general, the eFLaG dataset shows a very good range of performance comparable with previous  
632 applications of these models for the UK (e.g. Rudd et al. 2017; Harrigan et al. 2018b; Smith et al.  
633 2019). There are some commonalities with these previous studies in terms of spatial patterns. Rudd et  
634 al. (2017) also noted that G2G performance is likely to reflect the fact that simulated flows are natural  
635 (hence performance is poorer in the south and east where artificial influences are typical greater).  
636 Issues with poorer performance in groundwater-dominated catchments were highlighted for GR4J by  
637 Smith et al. (2019) and the present study shows that eFLaG enables some improvement through GR6J.  
638 Smith et al. (2019) also highlighted how a lack of snowmelt constrained performance in some areas  
639 (e.g. NE Scotland) while the current results also show improvements in these areas in eFLaG, given  
640 the inclusion of snowmelt accounting.

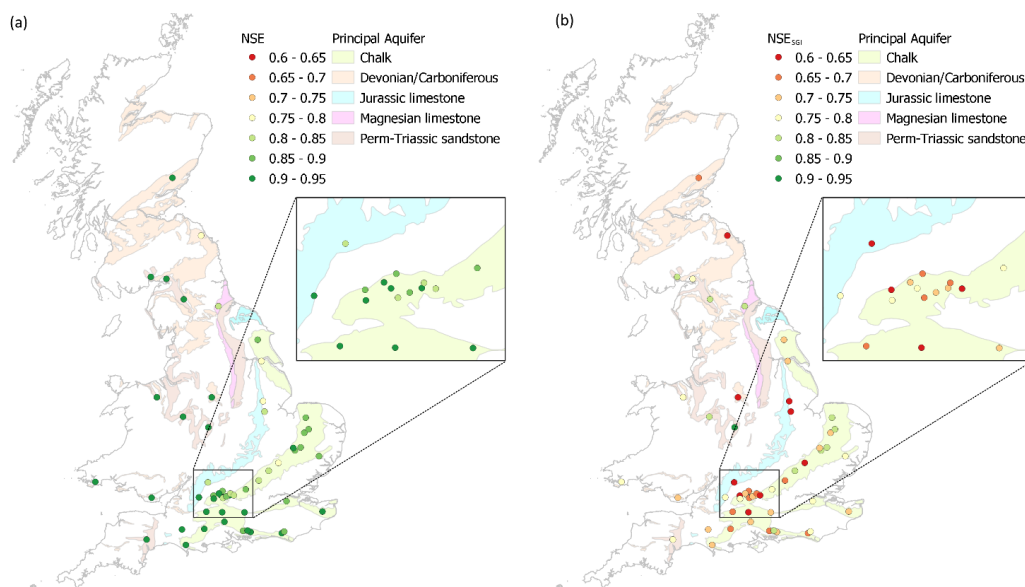
#### 641 **Groundwater levels**

642 Fig. 5 summarises the model evaluation results for the 54 AquMod models used in eFLaG. The results  
643 show that all 54 models demonstrate good overall efficiency in capturing daily groundwater level



644 dynamics, achieving a  $NSE \geq 0.77$ . All but 11 of the achieve a  $NSE \geq 0.85$  and 28 of the models  
645 achieve a  $NSE \geq 0.90$ . These include all 7 models situated in the Permo-Triassic sandstone and 4 out  
646 of 5 of the models situated in the Devonian and Carboniferous aquifers. Swan house and Lower Barn  
647 Cottage; the only models situated in the Magnesian limestones and Lower Greensand respectively,  
648 achieved a NSE of 0.82 and 0.86. The Chalk and Jurassic limestones borehole models span the full  
649 range of NSE scores.

650 The results show that all 54 AquiMod models are able to capture the historical SGI time series  
651 efficiently, achieving a  $NSE_{SGI} \geq 0.6$  which indicates that the models effectively capture groundwater  
652 extremes including periods of drought. The majority of models show a lower  $NSE_{SGI}$  compared to the  
653 NSE, although several models show negligible difference. On average the  $NSE_{SGI}$  is 0.15 less than the  
654 NSE.



655  
656 **Figure 5: AquiMod evaluation metric results including SGI (a) and  $SGI_{NSE}$  (b).**

657

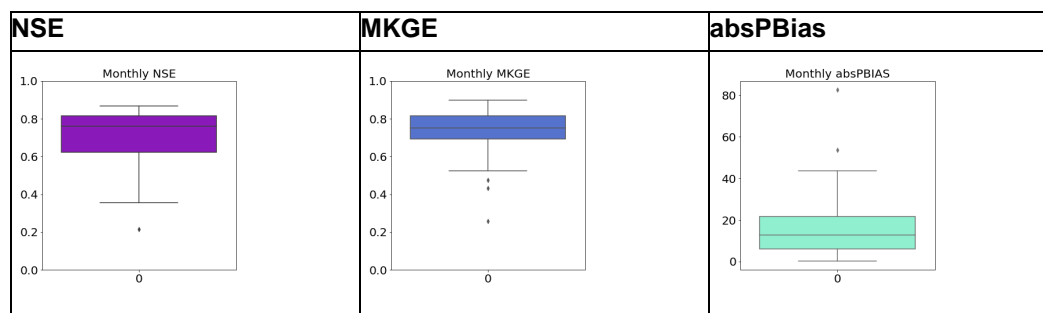
### 658 **Groundwater recharge**

659 ZOODRM demonstrates an ability to efficiently capture monthly mean river flows as is reflected by  
660 the medians for NSE and KGE2 which both exceed 0.75 and the median absolute percent bias which  
661 is 12.7% (Fig. 6). Fig. S6 shows the distributed recharge model results at the 41 gauging stations across  
662 the country. The model uses a simplistic overland routing approach, which is implemented to check  
663 the water balance at a monthly basis, noting that large scale spatial recharge values are most commonly  
664 used to drive groundwater flow models using monthly stress periods.

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667

668 **Figure 6: Distributed recharge model ZOODRM evaluation results.**

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## 670 **7. Evaluation of RCM-based runs in the baseline**

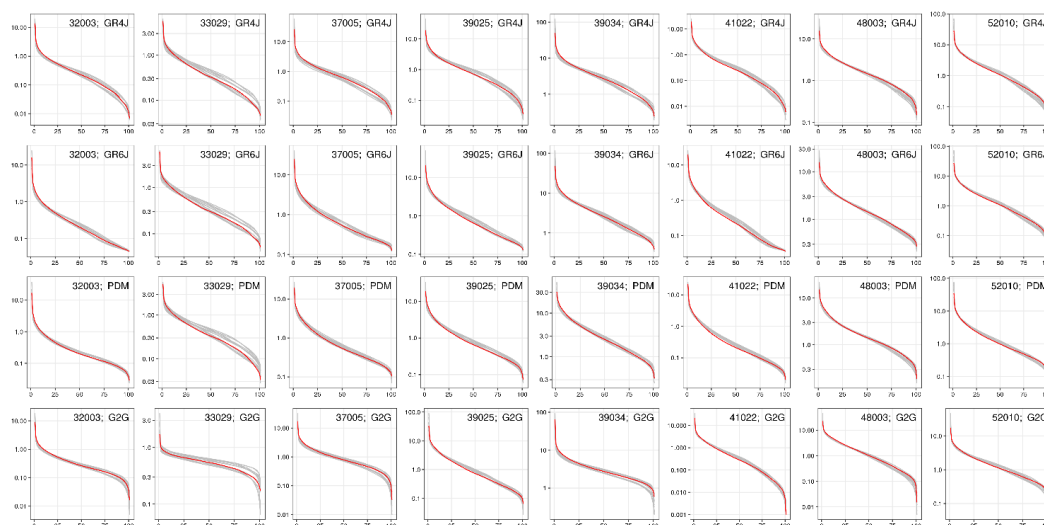
671

672 This section briefly considers the outcomes of the Stage 2 evaluation, focusing firstly on  
673 flow/groundwater duration curves for a subset of eFLaG sites, and then specifically on representation  
674 of particular low flows (low groundwater level) quantiles.

### 675 **Flow duration curves**

676 Flow duration curves (FDCs) summarise the entirety of the flow regime from high to low flows by  
677 including all river flows and expressing them in terms of the percentage of time a given flow is  
678 exceeded. Fig.7 and Figs. S7 to S9 provides a perspective on the ability of the RCM-driven river flow  
679 simulations (simrcm) to replicate the range and frequency of flows based on the observation climate-  
680 driven river flow simulations (simobs). FDCs are shown for a common baseline period of 1989-2018





681  
682 **Figure 7 -- Flow duration curves (FDCs) comparing the baseline flow regime in the 12 RCM**  
683 **ensemble members (simrcm, grey lines) to simulated observed (simobs, red line), 1989-2018.**  
684 **FDCs are featured for four hydrological models (GR4J, GR6J, PDM, G2G; rows) and eight**  
685 **catchments in southern and eastern England (32003 Harpers Brook, 33029 Stringsides, 37005**  
686 **Colne, 39025 Enborne, 39034 Evenlode, 41022 Lod, 48003 Fal, 52010 Brue; columns). The y-axis**  
687 **represents river flows (cumecs) on a logarithmic scale.**

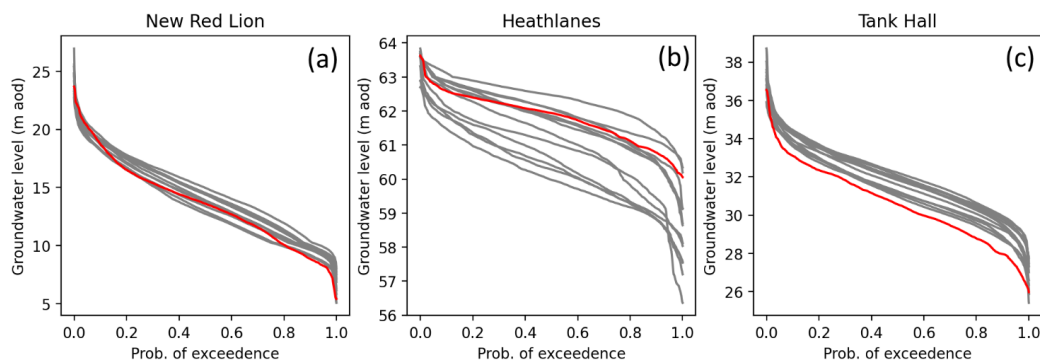
688 The close correspondence between FDCs derived from the RCM ensemble members and model  
689 observations suggests that the RCM ensemble is performing well in replicating flows across the regime  
690 This is consistent across most UK catchments, illustrated by the representative subset of 32 catchments  
691 featured in Fig. 7 and Figs.S7 to S9. The model observations are usually within the range of values  
692 from the 12 ensemble members throughout the flow regime. There are some catchments for which the  
693 RCM ensemble is more likely to overestimate the lowest half of the flow regime (exceedance  
694 probabilities of 50-100), most notably for the Stringsides (33029; Fig.7), Dove (28046; Fig.re S7),  
695 Frome (53006; Fig. S8), and Lud (29003; Fig. S7).

696 For certain catchments such as the Stringsides (33029; Fig. 7) and Lud (29003; Fig. S7), although there  
697 appears to be greater RCM uncertainty in river flows than for other catchments, the differences tend  
698 to be exaggerated in smaller, drier catchments with lower flows across the flow regime. The  
699 logarithmic y-axis is also a contributing factor to this, and also accounts for the seemingly larger RCM  
700 uncertainty in low flows than high flows across all catchments. These findings are also consistent  
701 across the four hydrological models, with no systematic differences identified for a given hydrological  
702 model. In some exceptional circumstances, there are examples of certain models in specific catchments  
703 in which the lowest river flows derived from the RCM ensemble are much lower than those in the  
704 model observations (e.g. 23004 South Tyne (Fig. S7) and 67018 Welsh Dee (Fig. S8) for GR6J, 33029  
705 Stringsides (Fig. 7) for G2G).



## 706 Groundwater level duration curves

707 Overall, an analysis of the groundwater level duration curves (GLDCs) at all boreholes (Figs.S10-S15)  
708 shows close correspondence between the simrcm and simobs runs whereby the simobs GLDC typically  
709 lies within the range of the simrcm GLDCs. However, there are some different behaviours across the  
710 boreholes which are summarised in Fig. 8. Fig.8a shows the GLDCs for the New Red Lion borehole  
711 situated in the Lincolnshire Limestone, the results of which are representative of most boreholes where  
712 the majority of simobs GLDCs falls within the range of the simrcm GLDCs. Several of the boreholes  
713 show a relatively high degree a variability across the simrcm runs in comparison to the simobs  
714 including the Heathlanes borehole situated in the Permo-Triassic Sandstone (Fig. 8b). These appear to  
715 be associated with boreholes which are known to respond relatively slowly to climate due to local  
716 hydrogeological conditions. For example, Heathlanes is known to be representative of a relatively low  
717 hydraulic diffusivity aquifer. For some boreholes there are areas of the GLDCs where the simobs  
718 GLDC does not lie within the range of the simrcm GLDC. In the most extreme cases, systematic biases  
719 across almost the entire GLDC can be seen (e.g. Fig. 8c).



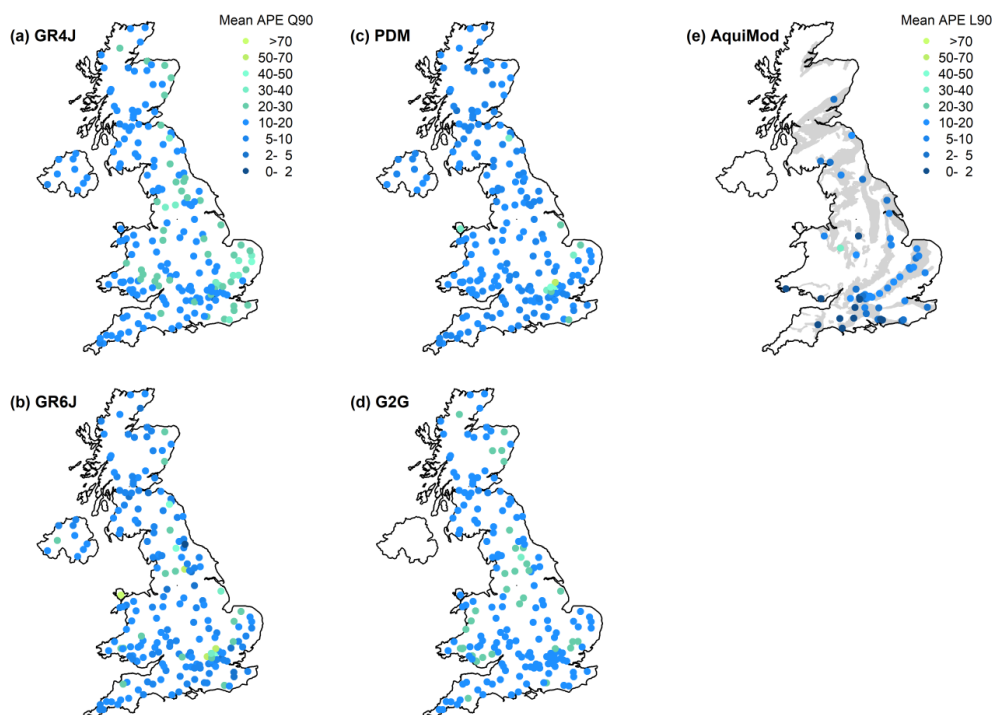
720

721 **Figure 8 – Groundwater level duration curves (GLDCs) for the period 1989-2018 using the**  
722 **simrcm (grey lines) simobs (red line) simulations. GLDCs are featured for three boreholes in**  
723 **different hydrogeological settings which show contrasting behaviour: (a) New Red Lion,**  
724 **(Lincolnshire Limestone), (b) Heathlanes (Permo-Triassic sandstone, Shropshire), (c) Tank**  
725 **Hall (Chalk).**

726

## 727 Low river flows and groundwater levels

728 Replication of observed low river flows and groundwater levels over a baseline period provides an  
729 indication of how well the simrcm runs are performing at the lower part of the river flow and  
730 groundwater level regime, and therefore enhances confidence in future low flow and level projections.  
731 Figs 9a-d show the difference between the simobs and simrcm 90% exceedance flow (Q90) over the  
732 1989-2018 baseline period reported as absolute percentage error (APE) at each of the 200 catchments  
733 for all four river flow models.



734

735 **Figure 9 -- Comparison of simobs and simrcm runs for river flows and groundwater levels**  
736 **exceeded 90% of the time (Q90 and L90 respectively) between 1989 and 2018. Colour scale**  
737 **indicates the mean of 12 absolute percent errors (APEs) between Q90/L90 in model observations**  
738 **and Q90/L90 in each of 12 ensemble members. Results are presented for: (a) GR4J; (b) GR6J;**  
739 **(c) PDM; (d) G2G; (e) AquiMod. Note: AquiMod levels are expressed as a percentage of the**  
740 **simobs range in groundwater levels to remove the influence of aquifer storage. Figures S16 to**  
741 **S18 feature the equivalent baseline assessment for Q30/L30, Q50/L50 and Q70/L70.**

742 Overall, there is a reasonable agreement between the simobs and simrcm Q90 values across all four  
743 models. Mean APEs are less than 20% for most catchments across the four hydrological models.  
744 Modelled low flows for GR6J, G2G and particularly PDM are especially well replicated in catchments  
745 across the UK, with mean APEs higher (20-50%) in GR4J river flows for catchments in East Anglia  
746 and parts of northern England and south Wales. The lumped catchment models GR6J and PDM  
747 struggle to capture low flows in groundwater-influenced catchments of the east Chilterns north of  
748 London, with APEs of up to 70%. Considering the natural flows simulated by G2G and the prevalence  
749 of artificial influences on rivers further south and east in the UK, mean APEs are reasonable in this  
750 region and are actually higher in more natural parts of Wales and northern England.

751 Mean APEs at a range of other flow quantiles demonstrate similar patterns (Figs S16 to S18). Mean  
752 APEs of Q30 for the vast majority of catchments for all four hydrological models are less than 20%  
753 (Fig. S16). Mean APEs of Q50 (Fig. S17) and Q70 (Fig. S18) are also reasonable in most catchments



754 and models, though higher mean APEs (20-50%) are apparent for both of these flow quantiles in East  
755 Anglia for GR4J, in parts of northern England for G2G, and in groundwater-influenced parts of the  
756 Chilterns for PDM. Mean APEs are similarly higher in GR6J flows at Q50 in East Anglia and at Q70  
757 in the groundwater-influenced Chilterns. Whilst this analysis is primarily an assessment of the ability  
758 of the RCM ensemble to replicate flows across the regime, it is clear that the hydrological model  
759 calibrations also have a role in influencing the outcomes.

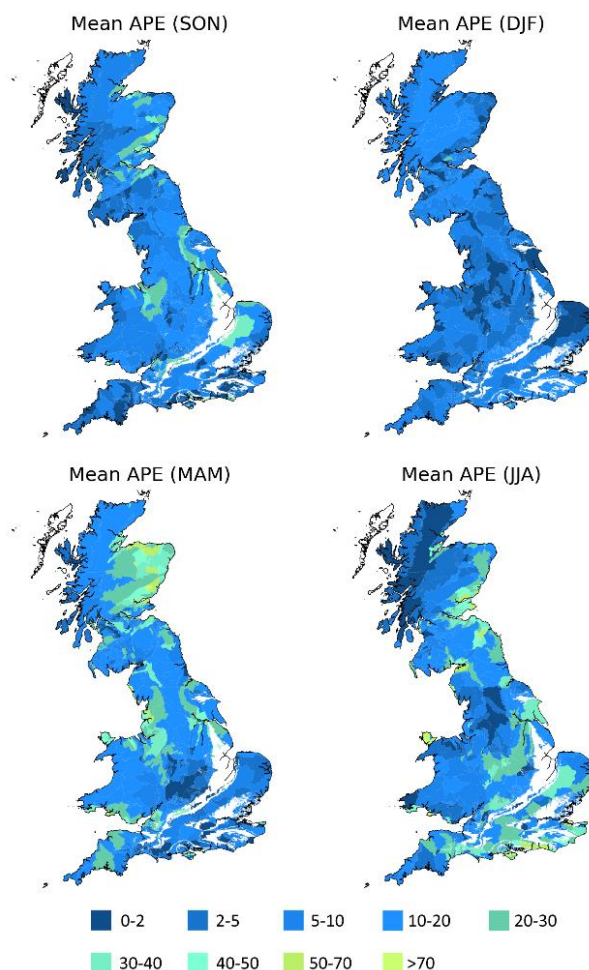
760 Fig. 9e shows the difference between the simobs and simrcm 90% exceedance groundwater level (L90)  
761 over the 1989-2018 baseline period reported as absolute percentage error (APE) relative to the simobs  
762 range in groundwater levels at each of the 54 boreholes. The use of the range in groundwater level as  
763 a reference removes the influence that the aquifer storage has on groundwater variability across the  
764 boreholes. There is good agreement between the simobs and simrcm L90 values across the boreholes.  
765 Mean APEs are less than 20% for all of the boreholes except for the Heathlanes borehole in the Permo-  
766 Triassic Sandstone where Mean APE exceeds 30%.

767 Mean APEs at a range of other groundwater level quantiles demonstrate similar patterns (Figs S16 to  
768 S18). Mean APEs of L30 do not exceed 5% for the majority of boreholes. The mean APE's typically  
769 become larger for most boreholes as the level quantile reduces towards L90. Heathlanes consistently  
770 has the highest mean APE for all level quantiles.

#### 771 **Seasonal groundwater recharge**

772 Fig. 10 provides a comparison of simobs and simrcm runs for seasonal average groundwater recharge  
773 between 1989 and 2018 generated by ZOODRM. During the winter months (DJF), when groundwater  
774 recharge is highest, the simrcm simulations show good correspondence with simobs simulations where  
775 the mean APE is less than 20% for all, but seven of the groundwater bodies. During the summer months  
776 (JJA), when groundwater recharge is lowest, the majority of groundwater bodies still show mean APE  
777 of less than 20%, but over 200 of them show errors exceeding 20%. These larger errors are typically  
778 associated with groundwater bodies that have lower than average recharge for this time of year. For  
779 MAM, the majority of groundwater bodies with errors that exceed 20% are also associated with those  
780 GW bodies with below-average recharge for that time of year. There are also some additional areas  
781 with significant recharge that show errors exceeding 20% including groundwater bodies in eastern-  
782 central Scotland, north-west and south-west England. For autumn (SON), the simrcm simulations show  
783 good correspondence with simobs simulation where the majority (>80%) of groundwater bodies show  
784 a mean APE of less than 20%. The majority those with larger errors are situated on the east coast of  
785 Scotland and England, north Wales and Cheshire.

786



787

788 **Figure 10 -- Comparison of simobs and simrcm runs for seasonal average groundwater recharge**  
789 **between 1989 and 2018 generated by ZOODRM. Colour scale indicates the mean of 12 absolute**  
790 **percent errors (APEs) between simobs and simrcm.**

791

792

## 793 **8. Conclusion and limitations**

794

795 The eFLaG dataset is presented as a nationally consistent dataset of future river flow, groundwater and  
796 groundwater recharge, using the latest available climate projections, from UKCP18. In this article, we  
797 have described the dataset and its evaluation against observational hydrological datasets, to give some  
798 confidence in the use of eFLaG as a dataset that can be used to assess the potential impacts on climate  
799 change on UK hydrology for a very wide range of applications.



800 The eFLaG dataset was developed specifically as a demonstration climate service for use by the water  
801 industry for water resources and drought planning, and hence by design is focused on future projections  
802 of drought, low river flows and low groundwater levels. We therefore present eFLaG primarily as a  
803 dataset for this purpose. Ongoing work is underway to demonstrate the utility of eFLaG for future  
804 drought projections (Parry et al. in prep.) and for future drought/water resources planning in practice  
805 (Counsell et al. in prep.). The predecessor product, FFGWL, has been widely used within the water  
806 industry to provide insight into the future evolution of river flows and groundwater levels through the  
807 21st century to support water resources management plans, and also supported significant academic  
808 water resource planning studies (e.g. Borgeomo et al. 2015; Huskova et al. 2016).

809 By providing a consistent dataset of future river flows, groundwater levels and groundwater recharge,  
810 eFLaG can potentially support a wide range of applications across other sectors. The FFGWL product  
811 also found very wide application for diverse research purposes (for: water quality, e.g. Charlton et al.  
812 2018; hydroecology, e.g. Royan et al. 2016; groundwater recharge, Hughes et al., 2021; groundwater  
813 level reconstruction, Jackson et al., 2016). For eFLaG, the good simulation of river flows and  
814 groundwater behaviours across much of the hydrological range suggests that this product could also  
815 find application in a whole range of impact studies, subject to additional evaluation for the purposes  
816 in mind. While not validated specifically for floods, the encouraging evaluation outputs for higher flow  
817 percentiles suggests users can analyse high flow metrics and variability (e.g. frequency of flows above  
818 a threshold), even if not annual maximum peak flows.

819 As with FFGWL, there are a number of advantages of using eFLaG for future projections: it is a  
820 spatially coherent dataset, meaning that future changes in hydrological variables can be compared  
821 between catchments, boreholes and aquifers at the regional-to-national scale. This is a key benefit for  
822 both research as well as practical water resources planning. Spatially coherent projections are needed  
823 to address the spatio-temporal dynamics of droughts (e.g. Tanguy et al. 2021) and how these may  
824 change in future and what this may mean for water resources planning – where, in practice, water  
825 resources management plans often involve transfers between regions (e.g. Murgatroyd et al. 2021).  
826 Another key benefit of eFLaG is that transient time series (daily data from 1980 to 2080) allow users  
827 to can explore the future evolution of river flow and groundwater variability on interannual and decadal  
828 timescales, rather than just using ‘Change Factor’ approaches that compare between future time slices  
829 and the baseline.

830 The use of an ensemble of outputs enables users to consider uncertainty in driving data (via the 12  
831 member RCM ensemble) as well as, for river flows, hydrological model uncertainty. In addition,  
832 different models provide different benefits: G2G performs less well against observations than the  
833 (calibrated) lumped catchment models, but does enable the characterisation of natural flows, which is  
834 vital for some uses (and against which artificial influences can be modelled separately in future).

835 Users of the eFLaG dataset should be aware of its limitations. While the evaluation shows encouraging  
836 results at the national scale, there are inevitably some catchments and boreholes where the evaluation  
837 (either Stage 1, Stage 2 or both) indicates poorer quality simulations. Users must be aware of this, and



838 should consult all the provided evaluation metrics when considering which catchments to use (and  
839 which models to use) in their analyses.

840 Users must also be aware that while there is some consideration of uncertainty through the adoption  
841 of the RCM PPE, and the use of a multiple models for river flows, there are many other sources of  
842 uncertainty not sampled in eFLaG. While the PPE gives a range of 12 outcomes, it is only one UKCP18  
843 product and one emissions scenario, so does not sample the full range of outcomes in UKCP18.  
844 Furthermore, only one bias correction approach is used. Although we use a range of hydrological  
845 models, clearly other hydrological models could provide different outcomes than the set used here,  
846 and we have also not considered other sources of uncertainty in the hydrological modelling (e.g.  
847 parametric uncertainty, as in e.g. Smith et al. 2019), nor the impacts of different observational driving  
848 climate datasets (e.g. different formulations of Potential Evapotranspiration, as in e.g. Tanguy et al.  
849 2018).

850 Finally, eFLaG only provides projections for a subset of the UK gauging station network (200  
851 catchments from some 1200 on the NRFA, for example). This is an inevitable constraint, as with the  
852 original FFGWL product (300 locations). While we have tried to sample UK hydrology to give users  
853 as much scope as possible, there will still be a need to transpose projections to sites of interest for some  
854 users. One of the benefits of eFLaG is that gridded river flow and recharge models are used. While  
855 these gridded datasets are not made available here, future initiatives will be looking to exploit them  
856 for providing projections at ungauged locations.

857

## 858 **9. Data Availability**

859

860 The eFLaG dataset is associated with a Digital Object Identifier. This must be referenced fully for every  
861 use of the eFLaG data as: <https://doi.org/10.5285/1bb90673-ad37-4679-90b9-0126109639a9>

862

863 All eFLaG files are available through the UKCEH Environmental Informatics Data Centre:

864 <https://catalogue.ceh.ac.uk/documents/1bb90673-ad37-4679-90b9-0126109639a9>

865

866 The data are stored as .csv files in the folder structure shown in the Guidance note available at  
867 Hannaford et al. (2022). In total there are 3304 files: one for each variable, model and  
868 catchment/borehole combination. They can be broadly split into two groups of files (Table 4), simobs  
869 and simrcm, as follows.

870 simobs

871 For the meteorological data, the simobs files contain date-indexed, observation-driven simulations  
872 (sim) data for precipitation with snowmelt and potential evaporation. For river flows and groundwater  
873 levels the simobs files contain date-indexed, observation-driven simulations (sim) and associated  
874 observations (obs) if they exist.



875 simrcm  
 876 For the meteorological data, the simrcm files contain date-indexed, RCM-driven simulations for the  
 877 twelve RCMs used in eFLaG for both precipitation with snowmelt and potential evaporation. For river  
 878 flows and groundwater levels the simrcm files contain date-indexed, RCM-driven simulations for the  
 879 twelve RCMs used in eFLaG.

880 **Table 4.** eFLaG dataset structure information

	Data	Name of file	Years available
simobs	Daily meteorology (precipwsnow (mm d <sup>-1</sup> ) + PET (mm d <sup>-1</sup> ))	<i>ukcp18_simobs_[nrfa-station-number/borehole-name].csv</i>	Jan 1961 – Dec 2018
	Daily river flow (m <sup>3</sup> s <sup>-1</sup> )	<i>modelname_simobs_nrfa-station-number.csv</i>	Jan 1963 – Dec 2018
	Daily groundwater levels (m AOD)	<i>AquiMod_simobs_borehole-name.csv</i>	Jan 1962 – Dec 2018
	Daily groundwater recharge (mm d <sup>-1</sup> )	<i>zoodrm_simobs_groundwater-body-name.csv</i>	Jan 1962 – Dec 2018
simrcm	Daily meteorology (precipwsnow (mm d <sup>-1</sup> ) + PE mm d <sup>-1</sup> )	<i>ukcp18_simobs_nrfa-station-number.csv</i>	Dec 1980 – Nov 2080
	Daily river flow (m <sup>3</sup> s <sup>-1</sup> )	<i>modelname_simrcm_nrfa-station-number.csv</i>	Dec 1982 – Nov 2080
	Daily groundwater levels (m AOD)	<i>AquiMod_simrcm_borehole-name.csv</i>	Jan 1982 – Nov 2080
	Daily groundwater recharge (mm d <sup>-1</sup> )	<i>zoodrm_simrcm_groundwater-body-name.csv</i>	Jan 1981 – Nov 2080

881

882 where *modelname* is G2G, PDM, GR4J, GR6J. NRFA station numbers and borehole names are given in the  
 883 eFLaG\_Station\_Metadata.xlsx workbook.

884

#### 885 **Conditions of Use**

886 The eFLaG dataset is available under a licensing condition agreement. For non-commercial use, the  
 887 products are available free of charge. For commercial use, the data might be made available  
 888 conditioned to a fee to be agreed with UKCEH and NERC BGS licensing teams, owners of the IPR of  
 889 the datasets and products.

890

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 897 CJ publish with the permission of the Executive Director, British Geological Survey (UKRI).





898

#### 899 **Author Contributions**

900 JH led the study and the river flow components, JM led the groundwater level and groundwater  
901 recharge components. AK and RL created the bias-corrected climate input data. Site selection was  
902 carried out by SP, TC and JM. Hydrological simulations were run by KS and TC (GR models), AR,  
903 AK and VB (G2G model) and JW, RM, SC and SW (PDM). JM and MM produced the groundwater  
904 level and groundwater recharge simulations. CC, MD, MS, AW carried out the demonstrator work and  
905 water industry engagement that helped design and shape eFLaG. ST led on data management. JH led  
906 the preparation of the manuscript with input from all authors. All authors contributed to the direction  
907 of the study and delivery of the dataset.

908

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