

1 **The eFLaG dataset: developing nationally consistent projections of**  
2 **future flows and groundwater based on UKCP18**

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## 30 **Abstract**

31 This paper presents an ‘enhanced future FLOws and Groundwater’ (eFLaG) dataset of  
32 nationally consistent hydrological projections for the UK, based on the latest UK  
33 Climate Projections (UKCP18). The hydrological projections are derived from a range  
34 of river flow models (Grid-to-Grid, PDM, GR4J and GR6J), to provide an indication of  
35 hydrological model uncertainty, as well as groundwater level (Aquimod) and  
36 groundwater recharge (ZOODRM) models. A 12-member ensemble of transient  
37 projections of present and future (up to 2080) daily river flows, groundwater levels and  
38 groundwater recharge were produced using bias corrected data from the UKCP18  
39 Regional (12km) climate ensemble. Projections are provided for 200 river catchments,  
40 54 groundwater level boreholes and 558 groundwater bodies, all sampling across the  
41 diverse hydrological and geological conditions of the UK. An evaluation was carried  
42 out, to appraise the quality of hydrological model simulations against observations and  
43 also to appraise the reliability of hydrological models driven by the RCM ensemble, in  
44 terms of their capacity to reproduce hydrological regimes in the current period. The  
45 dataset was originally conceived as a prototype climate service for drought planning  
46 for the UK water sector, so has been developed with drought, low river flow and low  
47 groundwater level applications as the primary focus. The evaluation metrics show that  
48 river flows and groundwater levels are, for the majority of catchments and boreholes,  
49 well simulated across the flow and level regime, meaning that the eFLaG dataset could  
50 be applied to a wider range of water resources research and management contexts,  
51 pending a full evaluation for the designated purpose.

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## 53 **1. Introduction**

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55 This paper presents an ‘enhanced future FLOws and Groundwater’ (hereafter referred  
56 to as “eFLaG”) dataset of nationally consistent, and spatially coherent, hydrological  
57 (river flow and groundwater) projections for the UK, based on UKCP18 – the latest  
58 climate projections for the UK from the UK Climate Projections programme (Murphy et  
59 al. 2018). eFLaG provides a successor to the Future Flows and Groundwater Levels  
60 (FFGWL) dataset (Prudhomme et al. 2013), which was based on the UKCP09  
61 projections (Murphy et al. 2010).

62 The eFLaG dataset was developed specifically as a demonstration climate service for  
63 use by the water industry for water resources and drought planning, and hence by  
64 design is focused on future projections of drought, low river flows and low groundwater  
65 levels. By providing a consistent dataset of future projections of these variables, eFLaG  
66 can potentially support a wide range of applications across other sectors. The

67 predecessor, FFGWL, has been widely used within the water industry, but also found  
68 very wide application for diverse research purposes (see Section 8).

69 As in FFGWL, in eFLaG the climate projections are used as input to a range of  
70 hydrological models to provide nationally consistent, spatially coherent projections of  
71 river flow and groundwater levels for the 21<sup>st</sup> century. The use of an ensemble of river  
72 flow models also provides information on hydrological model uncertainty. As well as  
73 using an updated set of climate projections, eFLaG capitalises on advances in  
74 national-scale river flow and groundwater modelling since FFGWL, and detailed  
75 evaluation of the applicability of models for drought simulation, notably research under  
76 the NERC Drought and Water Scarcity (DWS) Programme (e.g. Rudd et al. 2017;  
77 Smith et al. 2019).

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### 79 **Previous research on hydrological projections**

80 There is a long history of climate change impact assessment within the UK water  
81 industry and academia, which we do not review in detail here. Watts et al. (2015)  
82 provides an overview of past research (up to around 2013) on climate projections  
83 relevant for the water sector, including for future water resources and drought.  
84 However, as context for eFLaG it is worth considering some key developments since  
85 that review.

86 The original FFGWL did not present an assessment of future drought risk, other than  
87 seasonal river flows (Prudhomme et al. 2012) and groundwater levels (Jackson et al.  
88 2015), which suggested: pronounced decreases in future summer flows; reductions in  
89 annual average groundwater levels; and increases (decreases) in winter (summer)  
90 groundwater levels. Since then, the original FFGWL projections have been used in a  
91 number of hydrological impact studies. Collet et al. (2018) presented a probabilistic  
92 appraisal of future river flow drought (and flood) hazard in the UK, showing hydro-  
93 hazard 'hot-spots' in western Britain and northeast Scotland, especially during the  
94 autumn. Hughes et al. (2021) used the ZOODRM distributed groundwater recharge  
95 model to assess changes in 21<sup>st</sup> century seasonal recharge across river basin districts  
96 and groundwater bodies in the UK based on the FFGWL climate change projections.  
97 The results showed a consistent trend of more recharge being concentrated over fewer  
98 months with increased recharge in winter and decreased recharge in summer.

99 In addition to UKCP09/FFGWL, other datasets have been developed using different  
100 Global Climate Model (GCM)/Regional Climate Model (RCM)/hydrological modelling  
101 chains. One major development has been the use of large ensemble projections of  
102 future climate variables from the Weather@Home RCM (specifically HadRM3P) as  
103 part of the MaRIUS project within the DWS Programme (Guillod et al., 2018). The  
104 MaRIUS projections provide large ensembles (100+) of past, present (1900–2006) and

105 future (2020–2049 and 2070–2099) climate outputs. These were used as inputs to the  
106 national-scale Grid-to-Grid (G2G) hydrological model to provide a similarly large  
107 gridded (1km<sup>2</sup>) dataset of river flow and soil moisture (Bell et al., 2018). Analysis of  
108 these datasets has been conducted for drought (Rudd et al. 2019) and low flows (Kay  
109 et al. 2018), indicating future increases in hydrological drought severity and spatial  
110 extent, and decreases in absolute low flows.

111 A further source of hydro-meteorological projections now available are those from the  
112 EDgE project (End-to-end Demonstrator for improved decision-making for the water  
113 sector in Europe), see Samaniego et al. (2019). EDgE delivered an ensemble  
114 comprising of two GCMs and four ‘impact’ models (gridded land surface and  
115 hydrological models at a 5x5km scale) for the whole of Europe. Visser-Quinn et al.  
116 (2019) analysed future river flow drought risk in this ensemble, using a similar approach  
117 to Collet et al. (2018), and found similar results in terms of the spatial distribution and  
118 magnitude of future changes in droughts, albeit with some differences arising from the  
119 use of different scenarios, GCMs and hydrological models.

120 While such products may be used for climate adaptation research, the most relevant  
121 for eFLaG is the release of UKCP18. To date, relatively few studies using UKCP18  
122 have been published. Kay et al. (2020) made a rapid assessment of UKCP18 impacts  
123 on hydrology compared to UKCP09. More recently, Kay (2021), Kay et al. (2021a,b,c)  
124 and Lane & Kay (2021) provided future assessments of potential changes in seasonal  
125 mean river flows, high flows and low flows using various UKCP18 products with the  
126 G2G hydrological model. They found potential increases in winter mean flows and high  
127 flows, and decreases in summer and low flows, albeit with wide uncertainty ranges. To  
128 date, and to the authors’ knowledge, there have been no published assessments of  
129 future groundwater levels or groundwater recharge using UKCP18.

130 In summary, there have been substantial scientific advances in hydrological  
131 projections for the UK since Watts et al. (2015) and FFGWL, including some research  
132 on future indicators relevant for water resource availability and drought. However,  
133 relatively few datasets have been made available to the community since FFGWL.  
134 While MaRIUS and EDgE provide complementary hydrological datasets, there remains  
135 a need for an accessible dataset based on UKCP18. Existing UKCP18 studies have  
136 been focused on time-slice projections and used a single hydrological model (e.g. Kay  
137 et al., 2021 a,b,c) so there will be significant benefit arising from the eFLaG dataset of  
138 transient projections from a range of hydrological models covering river flows,  
139 groundwater levels and groundwater recharge.

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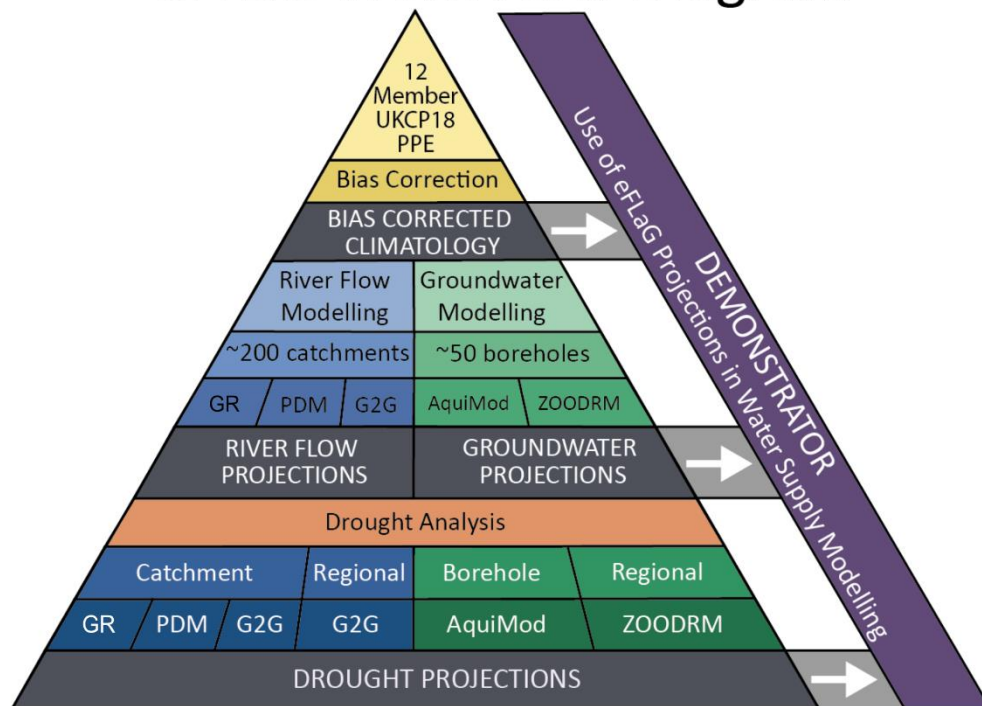
## 2. Outline of dataset and overview of the modelling chain

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144 In the following sections we set out the methodology behind the eFLaG dataset. This  
145 section firstly provides a brief overview of the various stages of the methodology, and  
146 how our method samples the ‘cascade of uncertainty’ (Smith et al. 2019) emerging  
147 from the multiplicity of projections and other modelling choices. While the original  
148 FFGWL methodology provided an initial foundation for eFLaG, much has changed in  
149 the decade since that study was commissioned, and the new UKCP18 projections  
150 differ from UKCP09 (e.g. Kay et al. 2020). . eFLaG therefore required the development  
151 of a new methodology, which is described in detail in the following sections.

152 The whole project workflow is illustrated in Fig 1. eFLaG is driven by the UKCP18  
153 dataset, specifically the ‘Regional’ 12km projections, to which a bias correction is  
154 applied. Section 3 describes the processing of the climate projections, including the  
155 bias correction method. The UKCP18 projections are used as input to three river flow  
156 models (GR, PDM and G2G), one groundwater level model (AquiMod) and one  
157 groundwater recharge model (ZOODRM) to provide simulations for 200 river  
158 catchments, 54 groundwater boreholes and 558 groundwater bodies respectively.  
159 Section 4 provides more detail on how these sites were selected. Details of the  
160 hydrological models and their calibration are given in Section 5. The evaluation of the  
161 models is covered in sections 6 and 7. Fig 1 also illustrates how all of the eFLaG  
162 projections are feeding into a series of water industry demonstrators, in partnership  
163 with UK water providers (specifically, Dwr Cymru/Welsh Water and Thames Water).  
164 These are not discussed in detail in this paper, but these were relevant for the site  
165 selection and as such are mentioned briefly below.

# eFLaG Work Flow Diagram



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167 **Figure 1 Project workflow illustrating the stages of analysis described in this paper**

168 The question of uncertainty in climate impacts modelling is a challenging one that has  
 169 been explored in a whole range of studies, going back as far as climate projections  
 170 have been routinely produced from the 1980s. There are inherent uncertainties at  
 171 every step of the process, from climate emissions scenarios through to climate  
 172 modelling, and on to environmental modelling (in our case hydrological modelling,  
 173 which itself has a vast literature when it comes to uncertainty estimation) and then to  
 174 wider impacts modelling (e.g. in water supply systems). Recently, Smith et al. (2018)  
 175 presented this issue as a 'cascade of uncertainty' (using widely adopted terminology,  
 176 e.g. Wilby and Dessai, 2010). Within eFLaG, as with the majority of climate impact  
 177 applications, it is not possible to sample across all sources of uncertainty. Following  
 178 Smith et al. (2019) we adopted a pragmatic approach to 'crystalising' the uncertainty  
 179 within the available time and resource constraints. In Table 1, we consider the sources  
 180 of uncertainty, and our approach to sampling from them. The focus in eFLaG is on  
 181 uncertainty arising from initial/boundary conditions. Additionally, for the river flow  
 182 simulations, the uncertainty arising from model choice is also accounted for, and within  
 183 this, model structure is accounted for by considering two versions of one of the models.

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187 **Table 1: Sources of uncertainty explored in eFLaG (building on the framework of**  
 188 **Smith et al. 2018)**

<b>Uncertainty Source</b>	<b>Sampling Approach</b>	<b>Details</b>
<b>Emissions Scenarios</b>	One scenario	RCP8.5
<b>Climate Models</b>	One model	Hadley Centre GCM
<b>Initial/Boundary Conditions</b>	12x member PPE (Perturbed Parameter Ensemble)	PPE perturbs the parameters of the climate model (both the RCM, and the GCM within which it is nested)
<b>Temporal/Spatial Downscaling</b>	One method	Hadley Centre RCM, monthly mean bias correction
<b>Model Choice</b>	3x river flow models 2x groundwater models	GR, PDM, G2G Aquimod, ZOODRM
<b>Model Structure</b>	2x model structures for the GR modelling framework	Fixed structure for G2G and PDM, but for GR two different model structures were used (GR4J and GR6J), as discussed in section 4.
<b>Model parameter uncertainty</b>	Not considered in eFLaG	Not considered in eFLaG

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### 191 **3. UKCP Data Processing**

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193 The UKCP18 regional climate projections were created using perturbed-parameter  
 194 runs of the Hadley Centre global climate model (GCM, HadGEM3-GC3.05) and  
 195 regional climate model (RCM, HadREM3-GA705) (Murphy et al. 2018). These provide  
 196 a set of 12 high resolution (12km) spatially consistent climate projections over the UK,  
 197 covering the period Dec 1980-Nov 2080. The 12-member RCM perturbed parameter  
 198 ensemble (PPE) is valuable to represent climate model parameter uncertainty;  
 199 ensemble members are numbered 01–15 excluding 02, 03 and 14 (as there are no  
 200 RCM equivalents for these GCM PPE members, Murphy et al. 2018 section 4.3), and  
 201 01 is the standard parameterisation. However, it is important to note that, as all  
 202 ensemble members are based on the same high emissions scenario (RCP8.5) and  
 203 underlying climate model structure, they do not represent the full climate uncertainty.  
 204 The UKCP18 RCM output was processed to provide the variables needed for

205 hydrological modelling – namely, 1km gridded and catchment-average time-series of  
206 available precipitation (i.e. after the application of a snow module, see below) and  
207 Potential Evapotranspiration (PET), not itself a UKCP18 output but estimated using  
208 available UKCP18 variables as described below.

209 The Hadley Centre climate model uses a simplified 360-day year, consisting of twelve  
210 30-day months. The RCM precipitation and temperature time-series are given for this  
211 360-day calendar, and are therefore not consistent with the 365/6-day observed time-  
212 series. Previously, the FFGWL Climate project inserted five (or six in a leap year) days  
213 of zero rainfall into the RCM time-series so that the observed and RCM data were  
214 using comparable calendars (Prudhomme et al., 2012). However, here the data were  
215 kept in the 360-day format, to avoid modifying the time-series with artificial data.

## 216 **Precipitation**

217 Daily precipitation time-series were available for each of the UKCP18 RCM-PPE  
218 members. However, the RCM data showed biases compared to observed precipitation,  
219 as is common for climate data (Murphy et al., 2018; Teutschbein & Seibert, 2012). The  
220 RCM data was found to substantially over-estimate precipitation for most months, the  
221 exception being for August-October, as shown in Murphy et al. (2018) Fig 4.4. A simple  
222 monthly-mean bias-correction methodology was therefore applied, through the  
223 following steps:

- 224 1. The 1km HadUK-Grid observed rainfall product was averaged to 12km for  
225 consistency with the RCM data (Hollis et al., 2019).
- 226 2. For each month and grid-cell, change factors were calculated between the RCM  
227 simulated precipitation and observation-based HadUK-Grid time-slice mean of  
228 monthly total rainfall over the period 1981-2010. This resulted in bias-correction  
229 factor grids being made for each month and RCM, as shown in Fig. 2.
- 230 3. The change factor grids were then smoothed to prevent spatial discontinuities,  
231 by updating each grid cell using a weighted combination of the original grid-cell  
232 value and neighbouring values, as in Guillod et al. (2018).
- 233 4. To produce bias-corrected precipitation estimates, the RCM simulated  
234 precipitation time-series were multiplied by the bias-correction factor grid for  
235 each month (i.e. all January precipitation was multiplied by the January bias-  
236 correction grids, February precipitation by the February correction grid, etc.).

237 The bias-corrected precipitation products were then downscaled from 12km to 1km  
238 based on the distribution of the Standard Average Annual Rainfall (SAAR) for the  
239 period 1961-1990, as in previous studies (Bell et al., 2007; Kay & Crooks, 2014). This  
240 involved calculating the ratio of the observed SAAR at 1km to the observed SAAR  
241 averaged up to the 12km RCM grid, and then multiplying RCM precipitation values by  
242 this ratio. This introduces further spatial variability related to typical rainfall patterns,  
243 but the total rainfall across the original 12km RCM grid cell remains unchanged.



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## 245 **Accounting for snowmelt processes**

246 A simple snow module was applied to account for snow-melt processes (Bell et al.,  
247 2016). The snow module converted the 1km bias-corrected precipitation into rainfall  
248 plus snowmelt (i.e. available precipitation), based on temperature. This used the  
249 minimum and maximum daily temperatures provided by each RCM ensemble member,  
250 which were first scaled from a 12km resolution to 1km using a lapse rate based on  
251 elevation data. The parameters used in the snow module are given in Supplementary  
252 Info (Table S1).

## 253 **Potential evapotranspiration**

254 Potential evapotranspiration (PET) was not directly available as an RCM output, and  
255 was therefore generated using a range of variables from the RCM-PPE climate time-  
256 series (Table S2). The PET was calculated using the same methodology as the hydro-  
257 PE dataset (Robinson et al. 2022) except for the use of eFLaG bias-corrected  
258 precipitation data within the interception correction component. This produces  
259 Penman-Monteith PET parameterised for short grass. The equation also included  
260 monthly stomatal resistance values, which were adjusted for the future period to  
261 account for the impact of increased carbon dioxide concentrations on stomata (as in  
262 Rudd & Kay, (2016), based on Kruijt et al., (2008)). The PET data were then copied  
263 down from a 12km to 1km resolution by simply setting all 1km grid cells to the value of  
264 the containing 12km grid cell.

## 265 **Outputs**

266 The 1km gridded time-series of 'available precipitation' and PET were then used to  
267 produce the time-series of catchment-averages required for each of the eFLaG river  
268 catchments and groundwater boreholes. For the river catchments, the catchment  
269 average values were derived using the standard UK National River Flow Archive  
270 approach for catchment average rainfalls, as described in NRFA (2021). For the  
271 boreholes, following Mackay et al. (2014a), averages were taken over the  
272 representative aquifer length which was determined as the groundwater flow path  
273 between the borehole and a single discharge point on a river based on the catchment  
274 geometry and hydrogeology. For the grid-based models, ZOODRM and G2G, the  
275 gridded data were used directly.

276 The bias-corrected climate outputs are part of the eFLaG dataset described further in  
277 Section 9. For each river catchment and groundwater borehole, bias-corrected data  
278 are available for the observational period, for the purposes of evaluation of the  
279 hydrological model outputs, and for the future. In addition, the gridded bias-corrected  
280 climatology will be made available as a separate dataset in future.



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284 **Figure 2: Bias-correction grids applied to correct monthly precipitation. Values are**  
285 **correction factors used to modify precipitation, with a value of 0.5 halving precipitation,**  
286 **1 meaning no change to precipitation and 2 doubling precipitation etc. Columns show**  
287 **results from each RCM PPE member, rows show results for each month. Note the**  
288 **column numbers reflect the RCM PPE number (see Sect. 3)**

#### 289 4. Catchment selection

290

291 The UK is fortunate to have one of the densest hydrometric networks in the world, with  
292 a legacy of strong commitment to data quality and completeness. There are more than  
293 1,500 river flow gauging stations with flow records on the UK National River Flow  
294 Archive (NRFA, Dixon et al. 2013 and <https://nrfa.ceh.ac.uk/>) and more than 180  
295 observation boreholes with groundwater level records on the BGS National  
296 Groundwater Level Archive (NGLA). These archives are the principal sources of  
297 validated river flow and groundwater level data at the UK scale. A remit of the NRFA  
298 and NGLA is to archive data that are useful for a wide variety of applications, primarily  
299 focusing on the most strategically important records. However, such catchments are  
300 not always the most relevant for the water industry, and water companies often have  
301 their own sites on which they undertake analysis. Since the eFLaG project aims to  
302 maximise utility for a range of users, the catchment selection strategy considered both  
303 research and industry needs.

304 Detailed site lists and metadata for river flow, groundwater level and groundwater  
305 recharge are catalogued on the dataset held on the Environmental Informatics Data  
306 Centre (EIDC) (Hannaford et al. 2022).

#### 307 River Flows

308 To support selection, a metadatabase was assembled for all NRFA gauging stations  
309 in the UK, primarily using the NRFA's metadata holdings published on the NRFA  
310 website and in the UK Hydrometric Register (Marsh and Hannaford, 2008). Metadata  
311 compiled included membership of key national strategic networks (e.g. near-natural  
312 Benchmark (UKBN2; Harrigan et al. 2018a) and operational monitoring networks),  
313 capitalising on efforts of other projects in quality controlling data and ensuring  
314 catchments are fit for purpose. Selection also considered whether catchments were  
315 used in previous relevant projects that have simulated river flows for drought analysis.  
316 The selection ensured a strong representation of the original FFGWL catchments (with  
317 117 catchments featuring in both) and also overlap with recent modelling endeavours  
318 through the DWS Programme (AboutDrought, 2021) projects 'Historic Droughts',  
319 'IMPETUS' and 'MaRIUS' projects, that used several of the models used by eFLaG  
320 (specifically G2G, GR4J). In this regard we ensured that 165 eFLaG catchments  
321 overlapped with at least one DWS project.

322 Selection also focused on data quality. Longer record lengths were prioritised and  
323 hydrometric quality was evaluated where possible. Given the extent of hydrometric  
324 issues (at low flows especially) it is not possible for all sites to have the highest quality  
325 data, but where decisions were made on similar sites, quality was considered as a  
326 tiebreaker. The selection included 80 Benchmark catchments, but did not seek to focus  
327 entirely on natural catchments given the limited range of variability they capture (being

328 mostly small and clustered in headwaters), and also included large and disturbed sites  
329 known to be important for water industry purposes. Artificial influences are prevalent  
330 across the UK and have been shown to prominently affect flow regimes (e.g.  
331 Rameshwaran et al. 2022) and drought characteristics (Tijdeman et al. 2018) in many  
332 catchments. Hence, the incorporation of a range of Benchmark near-natural  
333 catchments and artificially influenced sites is important for ensuring representativeness  
334 and demonstrating the utility of the different models used, which treat artificial  
335 influences differently (Sect 5). Membership of the Benchmark catchments is  
336 highlighted in the dataset description, and information on artificial influences can be  
337 accessed for all sites on the NRFA website (in station descriptions and 'Factors  
338 Affecting Runoff' codes).

339 Catchment representativeness was also considered, enabling the eFLaG dataset to  
340 sample the hydrological variability of the UK. Representativeness was considered by  
341 comparing the distribution of eFLaG potential selections relative to various catchment  
342 descriptors from the NRFA Hydrometric Register (altitude, area, annual rainfall, Base  
343 Flow Index, land cover and so on).

344 Finally, this activity focused on ensuring water industry relevance. At the national scale,  
345 this was achieved by asking stakeholders at an eFLaG workshop for views on  
346 additional catchments (Durant et al. 2022). In this way, 12 catchments were added.  
347 Similarly, for the regional demonstrators (Dwr Cymru/Welsh Water and  
348 Thames Water), water company teams were consulted to gain a better understanding  
349 of strategically important flow records for water companies in the case study regions,  
350 leading to an additional five catchments.

351 The final eFLaG dataset consists of 200 catchments (Fig. 3a) giving good geographical  
352 coverage and representativeness of the UK.

### 353 **Groundwater Levels**

354 Boreholes were selected to ensure a number of essential criteria were met. Firstly, only  
355 those boreholes with the highest-quality records of groundwater level were considered.  
356 This required regular (at least monthly) and continuous (at least 10 years in length)  
357 records of data from boreholes that are in zones which are not significantly affected by  
358 groundwater abstraction.

359 Secondly, sites were chosen to ensure coverage of the UK's principal aquifers where  
360 possible, enabling the eFLaG dataset to sample the hydrogeological variability of the  
361 UK. This broadly aligns with the requirements of other national-scale assessments of  
362 groundwater resources undertaken as part of the original FFGWL project and the  
363 'Historic Droughts' and 'IMPETUS' projects. Accordingly, the selection aimed to ensure  
364 good coherence with these studies also.

365 Thirdly, as with river flow catchment selection, an additional activity focused on  
366 ensuring water industry relevance, both at the national scale, through consultation with  
367 stakeholders at the eFLaG workshop, and through consultation with key demonstrator  
368 partners (Dwr Cymru/Welsh Water and Thames Water) who identified strategically  
369 important boreholes that would strengthen the outputs for long-term drought risk  
370 assessment to support the water resources planning case study. Through this activity,  
371 several additional boreholes were identified.

372 These selection criteria identified over 70 'candidate' boreholes for the eFLaG project.  
373 A final quality assurance procedure was then undertaken whereby a preliminary  
374 analysis of AquMod's ability to capture low groundwater levels was undertaken at each  
375 borehole via visual inspection of the simulated hydrographs. A final set of 54 boreholes  
376 was selected (Fig. 3b). They represent a significant advance in aquifer coverage  
377 compared to the 24 NGLA boreholes used in FFGWL, 15 of which are used in both.

### 378 **Groundwater Recharge**

379 The gridded groundwater recharge simulations have been aggregated over 558  
380 'groundwater bodies' covering England (Environment Agency, 2021a), Wales (Natural  
381 Resources Wales, 2021) and Scotland (Ó Dochartaigh et al., 2015) (Fig. 3c). These  
382 units were used for two principal reasons. Firstly, they are physically justifiable as they  
383 reflect known hydrogeological characteristics including groundwater recharge and  
384 groundwater flow regimes so that each catchment represents a distinct body of  
385 groundwater that can reasonably be considered in isolation. Secondly, they are  
386 coherent with the licensing areas defined as part of Catchment Abstraction  
387 Management Strategy (Environment Agency 2021b) and management areas for the  
388 implementation of the Water Framework Directive. They are, therefore, directly  
389 relevant to water regulation and the wider water industry.

390



415 discharges). G2G is not calibrated and simulates natural river flows only (i.e. it does  
416 not include artificial influences). The GR suite and PDM do not explicitly include  
417 artificial influences either, but as calibrated models they will implicitly include the net  
418 effect of artificial influences in the simulations. We return to this important distinction in  
419 the results and discussion.

420 For groundwater, eFLaG adopted the lumped, conceptual, Aquimod groundwater  
421 model (Mackay et al. 2014a) to simulate groundwater level time series on a daily time  
422 step at the boreholes identified in Section 4. Aquimod was the groundwater level model  
423 used in FFGWL providing direct comparison. In addition to groundwater levels, the  
424 zooming object oriented distributed recharge model (ZOODRM) (Mansour and  
425 Hughes, 2004) was used to study changes in future groundwater recharge.

426 In the following sub-sections, we describe each of these models in turn, providing  
427 information on the model set-up, calibration and past approaches to evaluation. A  
428 consistent approach was applied to the model application and evaluation across all  
429 these models where possible. However, it is important to emphasise that while some  
430 aspects were common, insofar as possible (e.g. model driving data), it was necessary  
431 to apply different approaches to suit the model in question. Calibration was done  
432 according to past applications and best-practice. Hence, the calibration approach  
433 described below is similar for the GR suite and PDM, but different for Aquimod, and by  
434 its nature G2G requires no specific calibration here. Where calibration was carried out  
435 for the conceptual models, it was undertaken for the full period of record of available  
436 data.

437 Identical approaches to evaluation were adopted across all river flow models, but minor  
438 differences applied with groundwater, as described below.

439 There are two sets of model output in eFLaG, described below – this terminology is  
440 adopted throughout.

- 441 • simobs: observation-driven simulation (i.e. simulations for the observed period,  
442 driven by observational climate datasets, described below). The simobs period  
443 varies between models, but covers at least the January 1961 – December 2018  
444 period.
- 445 • simrcm: UKCP18 RCM-driven simulation (12 ensemble members) (i.e.  
446 simulations driven by the UKCP18 RCM bias-corrected dataset as described in  
447 Section 3). These are available for 1980 to 2080. The simrcm runs from the  
448 observed period could then be evaluated against the simobs data.

449 Common driving data was applied across all models for the simobs runs. Accepted  
450 national-standard observational climate products were used, including:

- 451       • Precipitation and temperature: HadUK-Grid 1km x 1km dataset (Hollis et al.  
452       2019), the national standard gridded meteorological dataset and observational  
453       product associated with UKCP18.
- 454       • Potential Evaporation (PE). MORECS (Hough et al., 1997), an established,  
455       national gridded PE product. Other PE datasets such as CHESS (Robinson et  
456       al., 2017) and more recently the Environment Agency’s PE product  
457       (Environment Agency, 2021c) are available, however the decision to use  
458       MORECS was based on availability of data for the whole of the UK.
- 459

460 For all models, evaluation was undertaken in two stages, which is typical practice for  
461 appraising a model for simulation of climate change impacts:

- 462       1. Evaluation when driven with baseline observed climate data  
463       2. Evaluation when driven with baseline climate model data.

464 Stage 1 involves the use of a range of statistics to assess the performance of model  
465 simulations driven by observed climate data (the simobs runs) against observations of  
466 river flow and groundwater. For Stage 1, a range of metrics are available and widely  
467 used to assess how well rainfall-runoff or groundwater models perform against  
468 observations. Within eFLaG, a range of different metrics were used to assess  
469 performance (Table 3). For river flows, these metrics have a focus on low flow metrics  
470 (e.g. NSE on log-transformed flows), but some do evaluate performance across the  
471 flow regime. For groundwater levels, a generalised NSE score was used which  
472 provides an overall assessment of process realism and fit to groundwater level data.  
473 The simulated and observed Standardized Groundwater level Index (SGI) were also  
474 compared using the NSE ( $NSE_{SGI}$ ) which focusses in on groundwater extremes  
475 including droughts.

476 It is not possible to do a thorough evaluation of the recharge simulations from  
477 ZOODRM, given the difficulty in measuring recharge, particularly at a scale that is  
478 commensurable with a national model. However, past applications of ZOODRM (e.g.  
479 Mansour et al., 2018) have successfully used monthly river flow data as a means to  
480 evaluate ZOODRM’s ability to capture catchment water balances and infer the  
481 accuracy of seasonal recharge simulations (further details provided in model  
482 description). Accordingly, a subset of the river flow metrics relevant to monthly river  
483 flows have been used to evaluate ZOODRM for stage 1.

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490 **Table 3.** Model calibration and evaluation metrics used in eFLaG.

Evaluation Metric	Equation	Focus
<b>Nash-Sutcliffe Efficiency (<math>R^2</math> Efficiency)</b>	$NSE = 1 - \frac{\sum_{i=1}^n (Q_i - q_i)^2}{\sum_{i=1}^n (Q_i - \bar{Q})^2}$ <p><math>Q_i</math> and <math>q_i</math> are observed and modelled flow for day <math>i</math> of a <math>n</math> day record. <math>\bar{Q}</math> is the mean observed flow.</p> $NSE = 1 - \frac{\sum_{i=1}^n (H_i - h_i)^2}{\sum_{i=1}^n (H_i - \bar{H})^2}$ <p><math>H_i</math> and <math>h_i</math> are observed and modelled groundwater level for day <math>i</math> of a <math>n</math> day record. <math>\bar{H}</math> is the mean observed groundwater level.</p>	High Flows/Generalised groundwater levels
<b>Nash-Sutcliffe Efficiency log flows*</b>	$NSE_{log} = 1 - \frac{\sum_{i=1}^n (\log(Q_i) - \log(q_i))^2}{\sum_{i=1}^n (\log(Q_i) - \overline{\log(Q)})^2}$	Low Flows
<b>Nash-Sutcliffe Efficiency square root flows</b>	$NSE_{sqrt} = 1 - \frac{\sum_{i=1}^n (\sqrt{Q_i} - \sqrt{q_i})^2}{\sum_{i=1}^n (\sqrt{Q_i} - \sqrt{\bar{Q}})^2}$	Generalised Flows
<b>Nash-Sutcliffe Efficiency standardised groundwater level index</b>	$NSE_{SGI} = 1 - \frac{\sum_{i=1}^n (SGI_i - sgi_i)^2}{\sum_{i=1}^n (SGI_i - \overline{SGI})^2}$ <p><math>SGI_i</math> and <math>sgi_i</math> are observed and modelled SGI for day <math>i</math> of a <math>n</math> day record. <math>\overline{SGI}</math> is the mean observed SGI.</p>	Groundwater extremes
<b>Modified Kling Gupta Efficiency [square root flows]</b>	$KGE'_{sqrt} = 1 - \sqrt{(r - 1)^2 + (\beta - 1)^2 + (\gamma - 1)^2}$ <p>where <math>r</math> is the correlation coefficient, <math>\beta</math> is the bias ratio <math>\frac{\mu_{\sqrt{q}}}{\mu_{\sqrt{Q}}}</math>, and <math>\gamma</math> is the variability ratio <math>\frac{CV_{\sqrt{q}}}{CV_{\sqrt{Q}}}</math> or <math>\frac{\sigma_{\sqrt{q}}/\mu_{\sqrt{q}}}{\sigma_{\sqrt{Q}}/\mu_{\sqrt{Q}}}</math></p> <p><math>\mu</math>, <math>\sigma</math> and <math>CV</math> are the mean, standard deviation and coefficient of variation of flow (here of the square root of modelled and observed flows as indicated by the suffix)</p>	Generalised flows

<b>Absolute Percent Bias</b>	$absPBIAS = \left  \frac{\sum (q_i - Q_i)}{\sum Q_i} \right  100$	Water Balance
<b>Mean Absolute Percent Error</b>	$MAPE = \left( \frac{1}{n} \sum_{i=1}^n \left  \frac{Q_i - q_i}{Q_i} \right  \right) 100$	Systematic
<b>Absolute Percent Error in Q95</b>	$Q95_{APE} = \left  \frac{Q95 - q95}{Q95} \right  100$	Low Flows
<b>Low Flow Volume</b>	$LFV = 100 \frac{\sum_{p=70}^{95} (\sqrt{q_p} - \sqrt{Q_p})}{\sum_{p=70}^{95} (\sqrt{Q_p})}$ Here $q_p$ and $Q_p$ are the modelled and observed flow $p$ percentiles	Low Flows
<b>Absolute Percent Error in the Mean Annual Minimum on a 30-day moving average*</b>	$MAM30_{APE} = \left  \frac{QMAM30 - qMAM30}{QMAM30} \right  100$ where $QMAM30 = \frac{1}{n} \sum_{j=1}^n \min_j \left( \frac{Q_{j,i-29} + Q_{j,i-28} + Q_{j,i-27} \dots Q_{j,i-1} + Q_{j,i}}{30} \right)$ Here $Q_{j,i}$ is observed flow for day $i$ of hydrological year $j$ for a record of $n$ years	Low Flows

\*1/100<sup>th</sup> of the mean observed flow was added to both modelled and observed flow values during evaluation in order to avoid errors and biases due to very small and zero flows.

491

492

493

494 Sources of quality controlled, long-term observational data for model calibration and  
495 evaluation were the national standard repositories for hydrological data:

- 496
- River Flows: UK National River Flow Archive <https://nrfa.ceh.ac.uk/>
  - 497 • Groundwater Levels: UK National Groundwater Level Archive
  - 498 <https://www2.bgs.ac.uk/groundwater/datainfo/levels/ngla.html>
- 499

500 Stage 2 appraises the performance of the models when driven by the climate model outputs.  
501 That is, it compares the simobs and simrcm runs over the common baseline period. This  
502 assessment cannot use performance metrics based on time-series, as climate models are  
503 not expected to reproduce the sequencing of events seen over the historical period (Kay et al.

504 2015). Instead, the comparison has been done in terms of river flow and groundwater level  
505 duration curves, low flow/level metrics and seasonal recharge values. Thus, comparing the  
506 statistical characteristics of river flows, groundwater levels and groundwater recharge rather  
507 than their day-to-day equivalence (Kay et al. 2015, 2018). When looking at the performance  
508 of an ensemble of climate model runs, the model simulation driven by observed data would  
509 ideally sit within the range covered by the ensemble (assuming an ensemble of sufficient  
510 size). However, it would not necessarily be expected to sit in the middle of the ensemble  
511 range, because the set of weather events that actually occurred within the historical observed  
512 baseline period is just one realisation of what could have occurred within the range of natural  
513 variability (Kay et al. 2018).

514

## 515 **Description of the models and specific setup**

### 516 **GR4J/GR6J**

517 The GR4J and GR6J models come from a suite of hydrological models provided in the  
518 “airGR” modelling suite (Coron et al. 2021) for the R software programme. Both models are  
519 well suited to application across many catchments using the inbuilt automatic parameter  
520 optimisation function. The simple, efficient form of airGR models also make them suitable for  
521 uncertainty and ensemble analyses.

522 GR4J (Génie Rural à 4 paramètres Journalier) is a simple daily lumped conceptual model  
523 with only four free parameters. GR4J has been used for hydro-climate change research  
524 across the globe, and has demonstrated good performance in a diverse set of catchments in  
525 the UK. The model has been applied in the UK for operational seasonal forecasting, as well  
526 as for long-term drought reconstructions nationwide (Harrigan et al. 2018b, Smith et al.  
527 2019).

528 GR6J (Génie Rural à 6 paramètres Journalier) (Pushpalatha et al. 2011) is a six parameter  
529 variant of the GR modelling suite that was developed to improve low flow simulation and  
530 groundwater exchange. Recently, GR6J has increasingly been applied in UK water resources  
531 applications (e.g. Anglian Water Drought Plan, 2021).

532 For eFLaG, it was decided, therefore, that using both GR4J and GR6J would be beneficial.  
533 Both GR4J and GR6J were calibrated using the inbuilt automatic calibration function, with the  
534 modified Kling Gupta Efficiency (KGE, Gupta et al, 2009; Kling et al 2012) as the Error  
535 criterion (‘ErrorCritKGE2’). KGE offers a thorough error criterion as it calculates the  
536 correlation coefficient, the bias and the variability between simulated and observed flows.  
537 KGE values range from  $-\infty$  to 1, with 1 being a perfect fit. The calibration algorithm was  
538 applied to square-root transformed flows in order to place weight evenly across the flow  
539 regime. The airGR snowmelt module “CemaNeige” was not applied, as a simple snow

540 module was applied to the climate data to pre-process the precipitation data into rainfall and  
541 snowmelt based upon temperature (See section 3).

## 542 **Grid-to-Grid**

543 The Grid-to-Grid (G2G) hydrological model is an established area-wide distributed model that  
544 has been used to investigate the spatial coherence and variability of floods and droughts at  
545 catchment, regional and national scales. Model output typically consists of natural river flows  
546 at both gauged and ungauged locations, and can be provided as time-series for specific  
547 locations as well as 1km x 1km grids. The G2G has been used for climate impacts modelling  
548 of floods (Bell et al., 2009, 2012), low flows (Kay et al., 2018) and droughts (Rudd et al., 2019)  
549 and is also used operationally for flood forecasting (Cole and Moore, 2009; Moore et al.,  
550 2006).

551 The G2G is typically configured on a 1km×1km grid using spatial datasets of landscape  
552 properties such as soil type and drainage network, together with a few nationally-applied  
553 model parameters. The model is thus parameterised using national-scale spatial datasets  
554 (e.g. soil grids), rather than via individual catchment calibration. The spatial datasets and  
555 parameters used here are the same as those used in previous studies (Rudd et al., 2019;  
556 Bell et al., 2009, 2012; Kay et al., 2018).

557 The G2G can either be initialised with model water stores set to default or zero values, or  
558 from a states file appropriate to the run start date. In eFLaG the G2G was run for two years  
559 with observed rainfall and PE to provide a 1 January 1963 states file to initialise the  
560 observation-driven G2G model run. The RCM-driven G2G runs were all initialised with a  
561 generic December states file provided by an obs-driven run (for 1 December 1980), then the  
562 first two years of each RCM-driven run were discarded to allow for model spin up. The eFLaG  
563 river flow datasets therefore cover the periods, 1 January 1963 to 31 December 2018  
564 (simobs) and 1 December 1982 to 30 November 2080 (simrcm).

## 565 **PDM**

566 The Probability Distributed Model or PDM (Moore, 2007; UKCEH, 2021) is a simple, very  
567 widely used lumped rainfall-runoff model that can be configured to a variety of catchment flow  
568 regimes. Within the model, a soil water store with a distribution of water absorption capacities  
569 controls runoff production through a saturation excess process; stored water is also lost to  
570 evaporation. In one configuration, all runoff enters a surface store (the fast pathway) while a  
571 groundwater store (the slow pathway) is recharged by soil water drainage. In an alternative  
572 configuration, the runoff is split between the two stores according to a fixed fraction. Water in  
573 the surface- and ground-water stores is routed using a non-linear storage equation (powers  
574 of 1, 2 and 3 were trialled under eFLaG), or, for the surface store, a cascade of two linear  
575 reservoirs, before being combined to produce the modelled flow at the catchment outlet.  
576 Water is conserved within the model, whilst a multiplicative factor (equal to 1 if not required)

577 is applied to the input precipitation. Alternatively, a Groundwater Extension (Moore and Bell,  
578 2002) may be invoked to allow modelling of underflow at the catchment outlet, external  
579 springs, pumped abstractions, and the incorporation of well level data. Multiple hydrological  
580 response zones within a catchment can also be represented (not trialled under eFLaG). PDM  
581 may be thought of as a toolkit of model components representing a range of runoff production  
582 and flow routing behaviours, and with a choice of time-step.

583 Under eFLaG, single zone PDM models were invoked with a daily time-step. The model  
584 stores were initialised using the mean observed flow over the period of record, and the first  
585 two years of model flow discarded to allow for model spin-up. Nineteen different combinations  
586 of the above-mentioned toolkit options were systematically trialled for each catchment.  
587 Parameter estimation was performed using an automatic calibration procedure that applied  
588 a simplex optimisation scheme (Nelder and Mead, 1965) to different combinations of model  
589 parameters in turn during three increasingly aggressive stages. The rainfall factor, or, when  
590 employed, a spring factor (representing net water exchange for the catchment), were used  
591 to achieve zero bias in the modelled flows with respect to observations. Remaining  
592 parameters were estimated so as to optimise the modified Kling-Gupta Efficiency calculated  
593 on either the square root transformed flows, or, to a limited extent, the log transformed flows  
594 (Supplementary info S.2).

## 595 **AquiMod**

596 AquiMod is a lumped conceptual groundwater model that links simplified equations of soil  
597 drainage, unsaturated zone flow, and saturated groundwater flow to simulate daily  
598 groundwater level time series at a specified borehole (Mackay et al., 2014b). Each of these  
599 three components use model parameters that describe site-specific hydrological and  
600 hydrogeological characteristics of the groundwater catchment surrounding the borehole. The  
601 model also has a flexible saturated zone model structure that can be modified to represent  
602 different levels of vertical heterogeneity in hydrogeological properties.

603 For each borehole, the AquiMod parameters and structure were calibrated to achieve the  
604 most efficient simulation of available historical groundwater level data using the Nash-  
605 Sutcliffe Efficiency (NSE), which provides a reliable assessment of overall process realism  
606 and goodness of fit to groundwater level time series; following the approach of Mackay et al.  
607 (2014a) and Jackson et al. (2016), model parameters that could be related to catchment  
608 information (e.g. relating to known land cover and soil type) were fixed. The remaining  
609 parameters were then calibrated, using six different saturated zone model structures  
610 including a one-layer model (fixed hydraulic conductivity and specific yield); two- and three-  
611 layer models with variable hydraulic conductivity and fixed specific yield; two- and three-layer  
612 models with variable hydraulic conductivity and variable specific yield; and a 'cocktail glass  
613 representation of hydraulic conductivity variation with depth (Williams et al., 2006). The

614 optimal structure-parameter combination was obtained for each borehole using the Shuffled  
615 Complex Evolution global optimisation algorithm.

616 The calibrated models were then evaluated for their ability to capture groundwater level  
617 extremes using the Standardized Groundwater level Index, SGI (Bloomfield and Marchant,  
618 2013) as the basis for this evaluation. The SGI is a normalised index, calculated directly from  
619 groundwater level time series, which can be used to identify droughts and provide a  
620 quantitative status of groundwater resources drought events (e.g. Bloomfield et al., 2019).

621

## 622 **ZOODRM**

623 ZOODRM is a distributed recharge calculation model originally developed to estimate  
624 recharge values to drive groundwater models (Mansour and Hughes, 2004). It is applied over  
625 the British Mainland using a 2km square grid. The FAO Drainage and Irrigation Paper 56  
626 (FAO, 1988) approach, modified by Griffiths et al. (2006), is used to calculate potential  
627 recharge. This method removes actual evaporation and soil moisture deficit from rainfall and  
628 calculates potential recharge as a fraction of the excess water using a runoff coefficient value.  
629 The model was driven by daily rainfall and potential evaporation data. The model was  
630 primarily parameterised using available national scale data including data relating to the soil  
631 hydrology (Boorman et al., 1995), vegetation (LCM2000, NERC) and surface topography.  
632 The latter of these was used to route surface water runoff.

633 The runoff coefficient, which defines the proportion of excess soil water that drains overland  
634 via surface runoff, is an unknown parameter which must be calibrated. This was done in two  
635 stages. Firstly, the calibration problem was simplified by defining zones of equal runoff  
636 coefficient. In total 35 zones were used in ZOODRM which were based on UK  
637 hydrogeological and geological maps (DiGMapGB-625, 2008). Then, the runoff coefficient  
638 for each zone was manually calibrated by comparing simulated runoff to observed river flows  
639 minus baseflow which was calculated using a well-established baseflow separation method  
640 (Gustard et al., 1992). This was done using monthly mean flows given that ZOODRM does  
641 not have a sophisticated runoff routing scheme, and it is not expected, therefore, to capture  
642 daily variability in runoff. The comparison to monthly flows does, however, provide a useful  
643 means to evaluate the seasonal water balance of the model which serves as the best  
644 available proxy for the accuracy of the recharge simulations. In total, 41 gauging stations  
645 were used to assess the model performance.

646 The only hydrological process that needs initialisation in the ZOODRM is the soil moisture  
647 deficit. As all simulations start in January, which is a wet month with minimal potential  
648 evaporation, it is assumed that the initial soil moisture deficit is equal to zero. Even so, a  
649 warm up period of one year is used to initialise the model.

650

## 6. Hydrological model evaluation (Stage 1 evaluation)

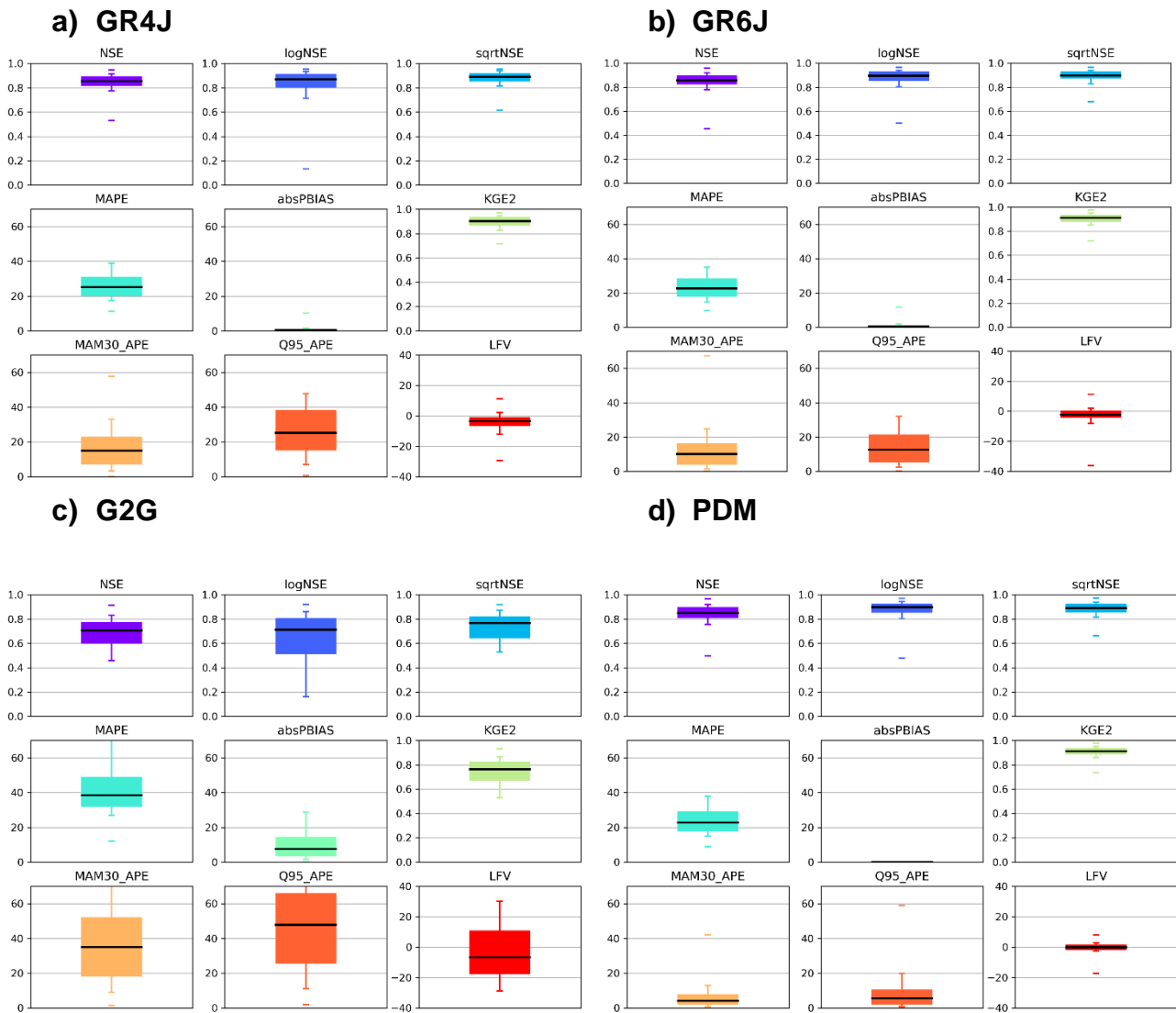
This section provides a brief summary of the outputs of the Stage 1 evaluation. Note that for river flows, model evaluation was undertaken at the same gauged locations and for the same period of time used for model calibration, except G2G which is not specifically calibrated.

### River Flows

Fig. 4 summarises the range of Stage 1 evaluation metrics across all catchments, while Supplementary Figs S2 to S5 provide maps of the evaluation metrics at each catchment. For GR4J, generally there was good performance across performance metrics in most catchments. Some outliers are present in the drought metrics, particularly in the South East and London. For GR6J, we observed good performance across all performance and drought metrics. GR6J generally performs slightly better than GR4J, particularly as shown in low flow catchments in the logNSE metric. For PDM, very good scores are obtained across the 200 sites, especially the low flow/drought indicators (bottom rows).

For G2G, again, good performance was observed overall (medians for NSE/ logNSE/ sqrtNSE/ KGE2  $\geq 0.7$ ). However, the performance was generally lower than for GR or PDM because the G2G is not calibrated to individual catchments, and G2G simulates *natural* flows, whereas the lumped models are calibrated to the observations used for performance assessment. In catchments with a high degree of anthropogenic disturbance, G2G is less able to simulate observed flows, whereas the calibration of the other hydrological models will implicitly account for such artificial impacts, meaning they are inevitably more likely to replicate observed flows, even if these processes are not included explicitly.

This distinction highlights an important benefit of eFLaG: PDM and GR4J/GR6J are calibrated to present-day flows and hence simulated flows are not natural, as they implicitly include artificial impacts. These runs do not, therefore, allow users to separate natural flows and artificial influences in the baseline period, nor to project how they may change relative to each other in future. On the other hand, although not used here, G2G has the capability of including artificial influences separately (e.g. Rameshwaran et al., 2022). We return to this issue in Section 8.



686 **Figure 4: Evaluation results summarised across the different models for all 200 catchments**  
 687 **for the key evaluation metrics outlined in Table 3**

688 In general, the eFLaG dataset shows a very good range of performance comparable with  
 689 previous applications of these models for the UK (e.g. Rudd et al. 2017; Harrigan et al. 2018b;  
 690 Smith et al. 2019). There are some commonalities with these previous studies in terms of  
 691 spatial patterns. Rudd et al. (2017) also noted that G2G performance is likely to reflect the  
 692 fact that simulated flows are natural (hence performance is poorer in the south and east  
 693 where artificial influences are typical greater). Issues with poorer performance in  
 694 groundwater-dominated catchments were highlighted for GR4J by Smith et al. (2019) and the  
 695 present study shows that eFLaG enables some improvement through GR6J. Smith et al.  
 696 (2019) also highlighted how a lack of snowmelt constrained performance in some areas (e.g.  
 697 NE Scotland) while the current results also show improvements in these areas in eFLaG,  
 698 given the inclusion of snowmelt accounting.

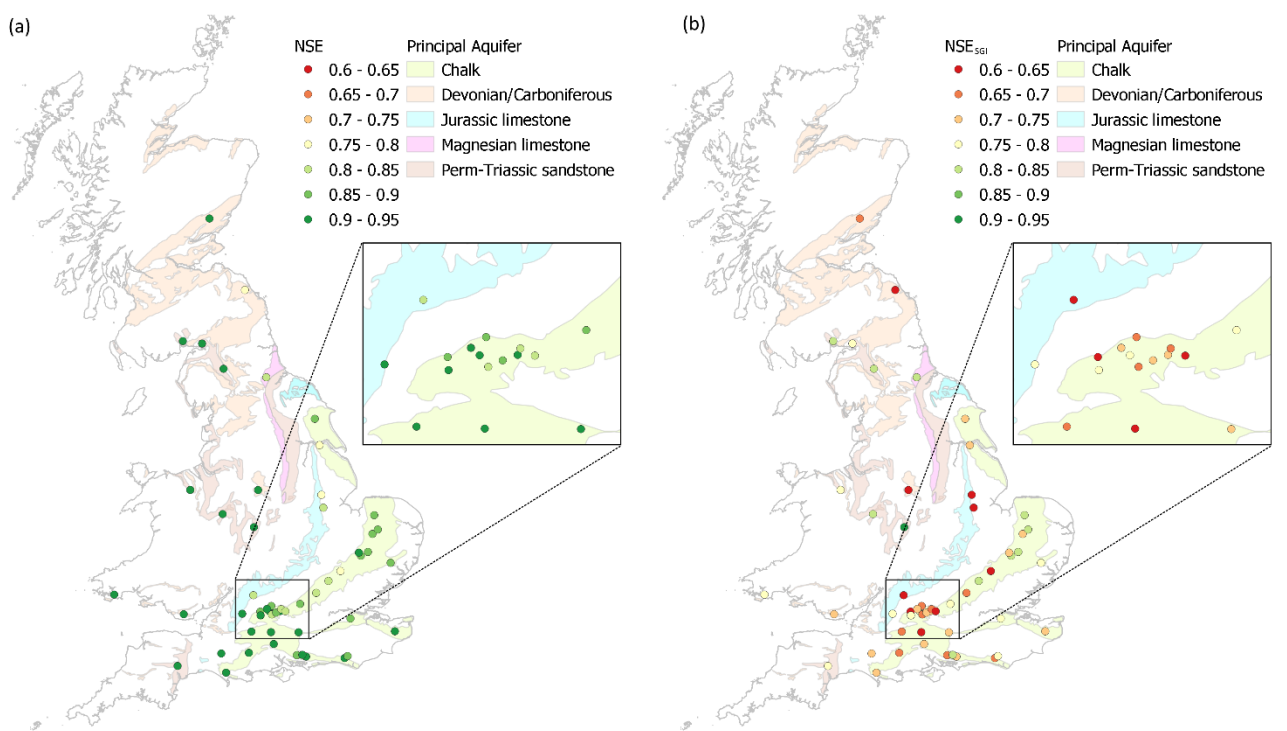
699



700 **Groundwater levels**

701 Fig. 5 summarises the model evaluation results for the 54 AquiMod models used in eFLaG.  
702 The results show that all 54 models demonstrate good overall efficiency in capturing daily  
703 groundwater level dynamics, achieving a  $NSE \geq 0.77$ . All but 11 of the models achieve a  $NSE$   
704  $\geq 0.85$  and 28 of the models achieve a  $NSE \geq 0.90$ . These include all 7 models situated in  
705 the Permo-Triassic sandstone and 4 out of 5 of the models situated in the Devonian and  
706 Carboniferous aquifers. Swan house and Lower Barn Cottage; the only models situated in  
707 the Magnesian limestones and Lower Greensand respectively, achieved a  $NSE$  of 0.82 and  
708 0.86. The Chalk and Jurassic limestones borehole models span the full range of  $NSE$  scores.

709 The results show that all 54 AquiMod models are able to capture the historical SGI time series  
710 efficiently, achieving a  $NSE_{SGI} \geq 0.6$  which indicates that the models effectively capture  
711 groundwater extremes including periods of drought. The majority of models show a lower  
712  $NSE_{SGI}$  compared to the  $NSE$ , although several models show negligible difference. On  
713 average the  $NSE_{SGI}$  is 0.15 less than the  $NSE$ .



714

715 **Figure 5: AquiMod evaluation metric results including  $NSE$  (a) and  $NSE_{SGI}$  (b).**

716

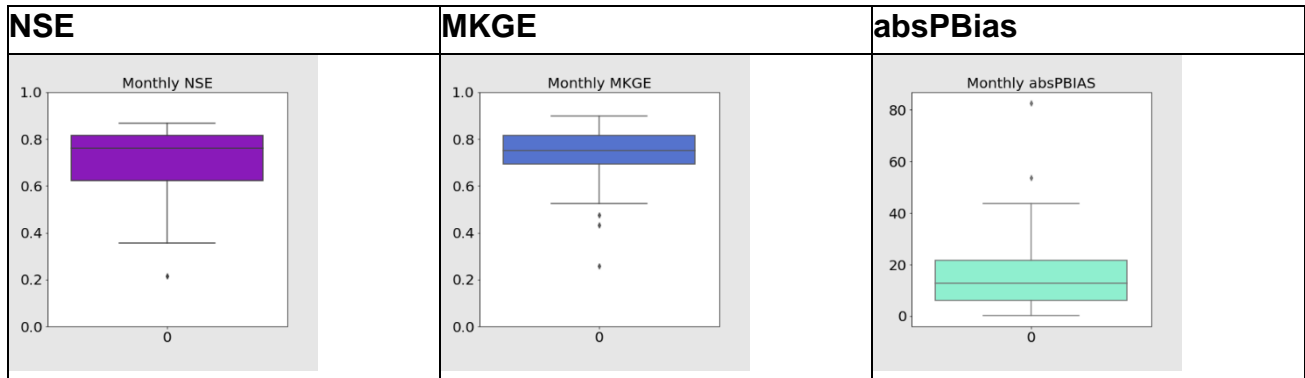
717 **Groundwater recharge**

718 ZOODRM demonstrates an ability to efficiently capture monthly mean river flows as is  
719 reflected by the medians for  $NSE$  and  $KGE2$  which both exceed 0.75 and the median absolute  
720 percent bias which is 12.7% (Fig. 6). Fig. S6 shows the distributed recharge model results at  
721 the 41 gauging stations across the country. The model uses a simplistic overland routing

722 approach, which is implemented to check the water balance at a monthly basis, noting that  
723 large scale spatial recharge values are most commonly used to drive groundwater flow  
724 models using monthly stress periods.

725

726



727

728 **Figure 6: Distributed recharge model ZOODRM evaluation results.**

729

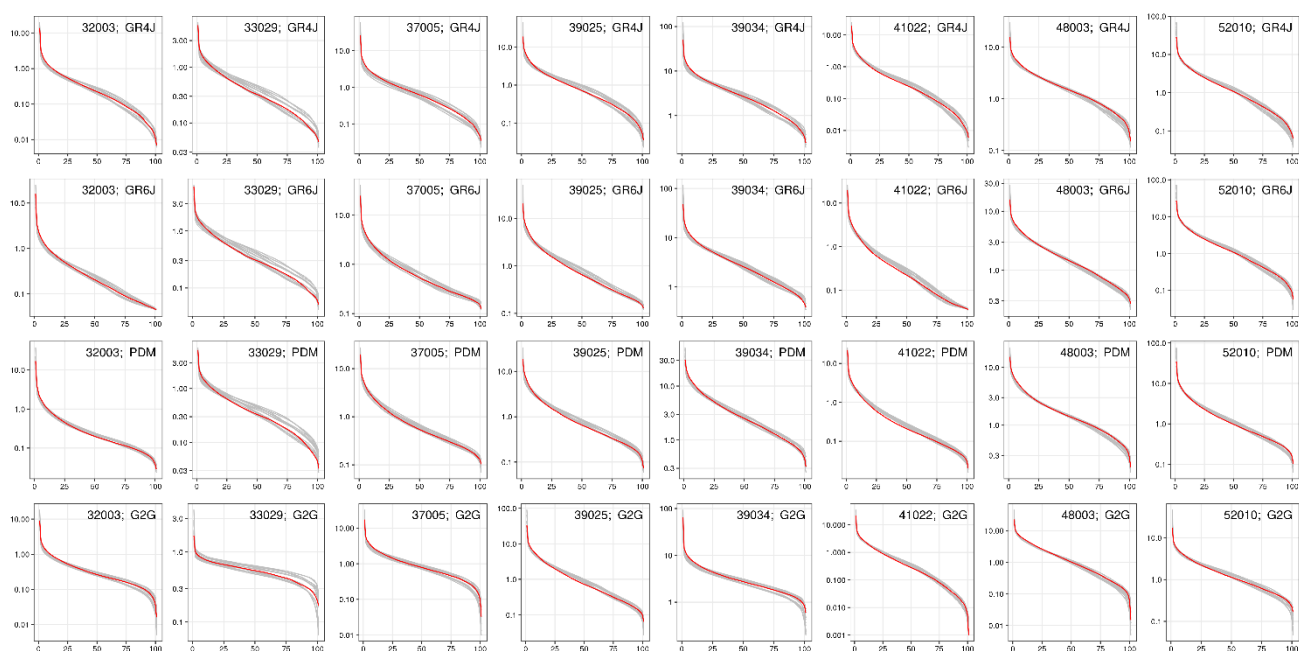
### 730 **7. Evaluation of RCM-based runs in the baseline**

731

732 This section briefly considers the outcomes of the Stage 2 evaluation, focusing firstly on  
733 flow/groundwater duration curves for a subset of eFLaG sites, and then specifically on  
734 representation of particular low flows (low groundwater level) quantiles.

#### 735 **Flow duration curves**

736 Flow duration curves (FDCs) summarise the entirety of the flow regime from high to low flows  
737 by including all river flows and expressing them in terms of the percentage of time a given  
738 flow is exceeded. Fig.7 and Figs. S7 to S9 provides a perspective on the ability of the RCM-  
739 driven river flow simulations (simrcm) to replicate the range and frequency of flows based on  
740 the observation climate-driven river flow simulations (simobs). FDCs are shown for a common  
741 baseline period of 1989-2018



742  
 743 **Figure 7 -- Flow duration curves (FDCs) comparing the baseline flow regime in the 12 RCM**  
 744 **ensemble members (simrcm, grey lines) to simulated observed (simobs, red line), 1989-2018.**  
 745 **FDCs are featured for four hydrological models (GR4J, GR6J, PDM, G2G; rows) and eight**  
 746 **catchments in southern and eastern England (32003 Harpers Brook, 33029 Stringside, 37005**  
 747 **Colne, 39025 Enborne, 39034 Evenlode, 41022 Lod, 48003 Fal, 52010 Brue; columns). The y-**  
 748 **axis represents river flows (cumecs) on a logarithmic scale.**

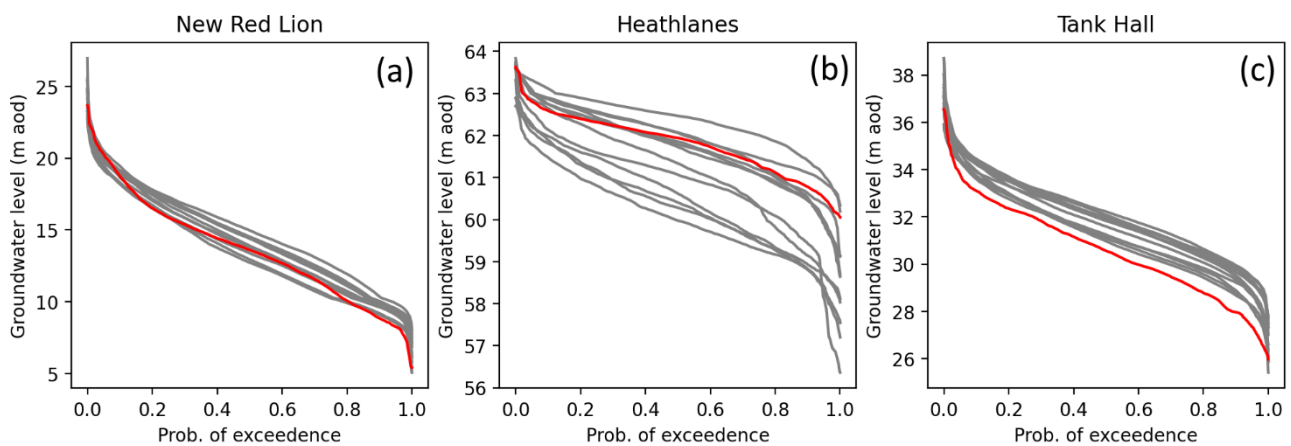
749  
 750 The close correspondence between FDCs derived from the RCM ensemble members and  
 751 model observations suggests that the RCM ensemble is performing well in replicating flows  
 752 across the regime This is consistent across most UK catchments, illustrated by the  
 753 representative subset of 32 catchments featured in Fig. 7 and Figs.S7 to S9. The model  
 754 observations are usually within the range of values from the 12 ensemble members  
 755 throughout the flow regime. There are some catchments for which the RCM ensemble is  
 756 more likely to overestimate the lowest half of the flow regime (exceedance probabilities of 50-  
 757 100), most notably for the Stringside (33029; Fig.7), Dove (28046; Fig.S7), Frome (53006;  
 758 Fig. S8), and Lud (29003; Fig. S7).

759 For certain catchments such as the Stringside (33029; Fig. 7) and Lud (29003; Fig. S7),  
 760 although there appears to be greater RCM uncertainty in river flows than for other  
 761 catchments, the differences tend to be exaggerated in smaller, drier catchments with lower  
 762 flows across the flow regime. The logarithmic y-axis is also a contributing factor to this, and  
 763 also accounts for the seemingly larger RCM uncertainty in low flows than high flows across  
 764 all catchments. These findings are also consistent across the four hydrological models, with  
 765 no systematic differences identified for a given hydrological model. In some exceptional  
 766 circumstances, there are examples of certain models in specific catchments in which the  
 767 lowest river flows derived from the RCM ensemble are much lower than those in the model

768 observations (e.g. 23004 South Tyne (Fig. S7) and 67018 Welsh Dee (Fig. S8) for GR6J,  
769 33029 Stringsides (Fig. 7) for G2G).

### 770 Groundwater level duration curves

771 Overall, an analysis of the groundwater level duration curves (GLDCs) at all boreholes  
772 (Figs.S10-S15) shows close correspondence between the simrcm and simobs runs whereby  
773 the simobs GLDC typically lies within the range of the simrcm GLDCs. However, there are  
774 some different behaviours across the boreholes which are summarised in Fig. 8. Fig.8a  
775 shows the GLDCs for the New Red Lion borehole situated in the Lincolnshire Limestone, the  
776 results of which are representative of most boreholes where the majority of simobs GLDCs  
777 falls within the range of the simrcm GLDCs. Several of the boreholes show a relatively high  
778 degree a variability across the simrcm runs in comparison to the simobs including the  
779 Heathlanes borehole situated in the Permo-Triassic Sandstone (Fig. 8b). These appear to be  
780 associated with boreholes which are known to respond relatively slowly to climate due to local  
781 hydrogeological conditions. For example, Heathlanes is known to be representative of a  
782 relatively low hydraulic diffusivity aquifer. For some boreholes there are areas of the GLDCs  
783 where the simobs GLDC does not lie within the range of the simrcm GLDC. In the most  
784 extreme cases, systematic biases across almost the entire GLDC can be seen (e.g. Fig. 8c).



785

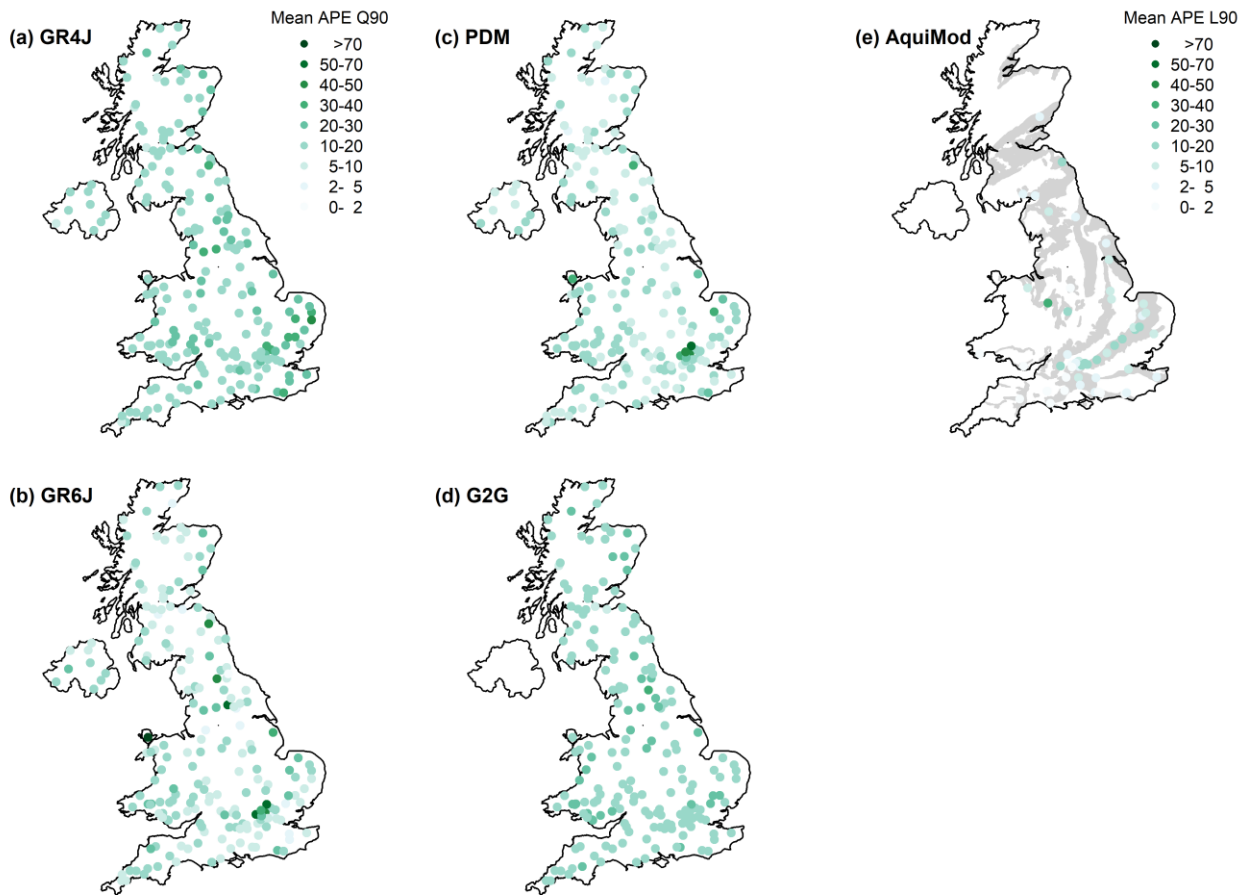
786 **Figure 8 – Groundwater level duration curves (GLDCs) for the period 1989-2018 using the**  
787 **simrcm (grey lines) simobs (red line) simulations. GLDCs are featured for three boreholes in**  
788 **different hydrogeological settings which show contrasting behaviour: (a) New Red Lion,**  
789 **(Lincolnshire Limestone), (b) Heathlanes (Permo-Triassic sandstone, Shropshire), (c) Tank**  
790 **Hall (Chalk).**

791

### 792 Low river flows and groundwater levels

793 Replication of observed low river flows and groundwater levels over a baseline period  
794 provides an indication of how well the simrcm runs are performing at the lower part of the  
795 river flow and groundwater level regime, and therefore enhances confidence in future low

796 flow and level projections. Figs 9a-d show the difference between the simobs and simrcm  
 797 90% exceedance flow (Q90) over the 1989-2018 baseline period reported as absolute  
 798 percentage error (APE) at each of the 200 catchments for all four river flow models.



799

800 **Figure 9 -- Comparison of simobs and simrcm runs for river flows and groundwater levels**  
 801 **exceeded 90% of the time (Q90 and L90 respectively) between 1989 and 2018. Colour scale**  
 802 **indicates the mean of 12 absolute percent errors (APEs) between Q90/L90 in model**  
 803 **observations and Q90/L90 in each of 12 ensemble members. Results are presented for: (a)**  
 804 **GR4J; (b) GR6J; (c) PDM; (d) G2G; (e) AquiMod. Note: AquiMod levels are expressed as a**  
 805 **percentage of the simobs range in groundwater levels to remove the influence of aquifer**  
 806 **storage. Figures S16 to S18 feature the equivalent baseline assessment for Q30/L30, Q50/L50**  
 807 **and Q70/L70.**

808 Overall, there is a reasonable agreement between the simobs and simrcm Q90 values across  
 809 all four models. Mean APEs are less than 20% for most catchments across the four  
 810 hydrological models. Modelled low flows for GR6J, G2G and particularly PDM are especially  
 811 well replicated in catchments across the UK, with mean APEs higher (20-50%) in GR4J river  
 812 flows for catchments in East Anglia and parts of northern England and south Wales. The  
 813 lumped catchment models GR6J and PDM struggle to capture low flows in groundwater-  
 814 influenced catchments of the east Chilterns north of London, with APEs of up to 70%.  
 815 Considering the natural flows simulated by G2G and the prevalence of artificial influences on

816 rivers further south and east in the UK, mean APEs are reasonable in this region and are  
817 actually higher in more natural parts of Wales and northern England.

818 Mean APEs at a range of other flow quantiles demonstrate similar patterns (Figs S16 to S18).  
819 Mean APEs of Q30 for the vast majority of catchments for all four hydrological models are  
820 less than 20% (Fig. S16). Mean APEs of Q50 (Fig. S17) and Q70 (Fig. S18) are also  
821 reasonable in most catchments and models, though higher mean APEs (20-50%) are  
822 apparent for both of these flow quantiles in East Anglia for GR4J, in parts of northern England  
823 for G2G, and in groundwater-influenced parts of the Chilterns for PDM. Mean APEs are  
824 similarly higher in GR6J flows at Q50 in East Anglia and at Q70 in the groundwater-influenced  
825 Chilterns. Whilst this analysis is primarily an assessment of the ability of the RCM ensemble  
826 to replicate flows across the regime, it is clear that the hydrological model calibrations also  
827 have a role in influencing the outcomes.

828 Fig. 9e shows the difference between the simobs and simrcm 90% exceedance groundwater  
829 level (L90) over the 1989-2018 baseline period reported as absolute percentage error (APE)  
830 relative to the simobs range in groundwater levels at each of the 54 boreholes. The use of  
831 the range in groundwater level as a reference removes the influence that the aquifer storage  
832 has on groundwater variability across the boreholes. There is good agreement between the  
833 simobs and simrcm L90 values across the boreholes. Mean APEs are less than 20% for all  
834 of the boreholes except for the Heathlanes borehole in the Permo-Triassic Sandstone where  
835 Mean APE exceeds 30%.

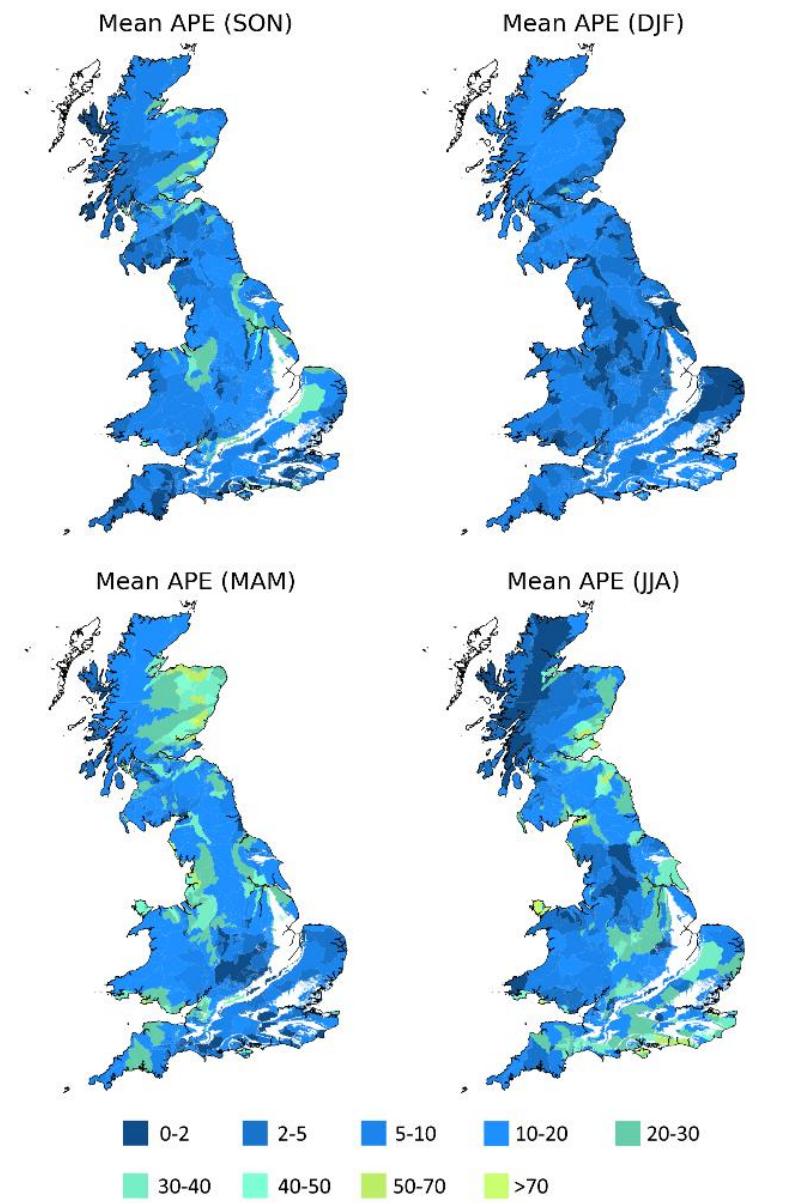
836 Mean APEs at a range of other groundwater level quantiles demonstrate similar patterns  
837 (Figs S16 to S18). Mean APEs of L30 do not exceed 5% for the majority of boreholes. The  
838 mean APE's typically become larger for most boreholes as the level quantile reduces towards  
839 L90. Heathlanes consistently has the highest mean APE for all level quantiles.

#### 840 **Seasonal groundwater recharge**

841 Fig. 10 provides a comparison of simobs and simrcm runs for seasonal average groundwater  
842 recharge between 1989 and 2018 generated by ZOODRM. During the winter months (DJF),  
843 when groundwater recharge is highest, the simrcm simulations show good correspondence  
844 with simobs simulations where the mean APE is less than 20% for all, but seven of the  
845 groundwater bodies. During the summer months (JJA), when groundwater recharge is  
846 lowest, the majority of groundwater bodies still show mean APE of less than 20%, but over  
847 200 of them show errors exceeding 20%. These larger errors are typically associated with  
848 groundwater bodies that have lower than average recharge for this time of year. For MAM,  
849 the majority of groundwater bodies with errors that exceed 20% are also associated with  
850 those GW bodies with below-average recharge for that time of year. There are also some  
851 additional areas with significant recharge that show errors exceeding 20% including  
852 groundwater bodies in eastern-central Scotland, north-west and south-west England. For  
853 autumn (SON), the simrcm simulations show good correspondence with simobs simulation

854 where the majority (>80%) of groundwater bodies show a mean APE of less than 20%. The  
855 majority those with larger errors are situated on the east coast of Scotland and England, north  
856 Wales and Cheshire.

857



858

859 **Figure 10 -- Comparison of simobs and simrcm runs for seasonal average groundwater**  
860 **recharge between 1989 and 2018 generated by ZOODRM. Colour scale indicates the mean of**  
861 **12 absolute percent errors (APEs) between simobs and simrcm.**

862

863

864

## 8. Applications and limitations

### Applications

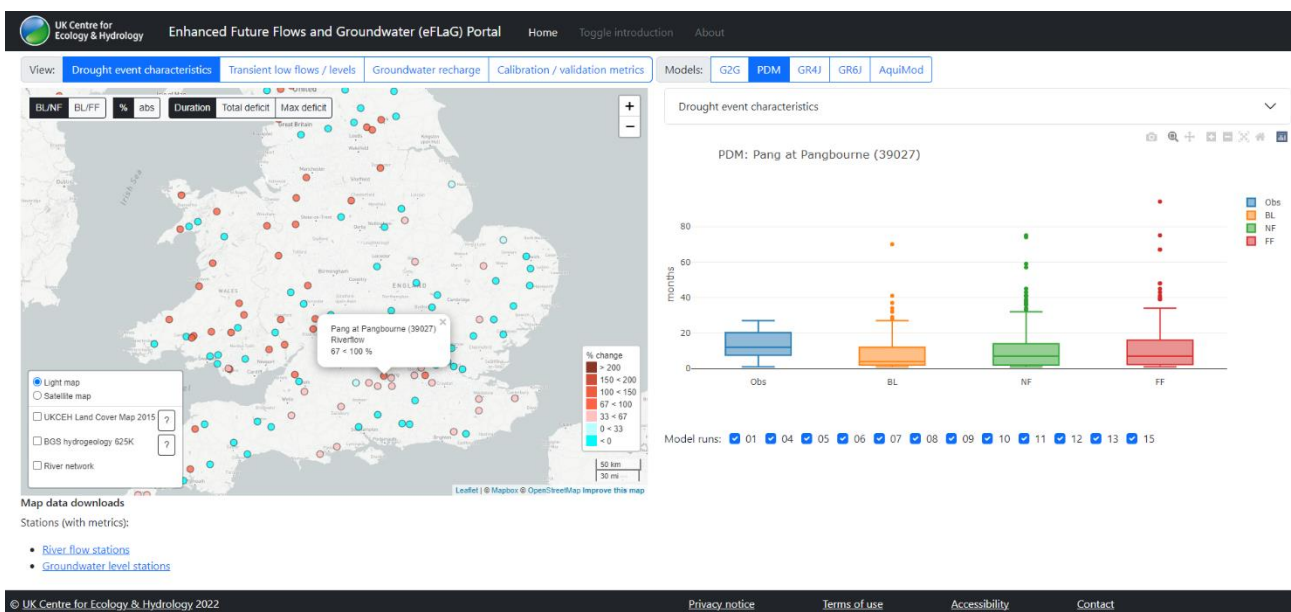
The eFLaG dataset is presented as a nationally consistent dataset of future river flow, groundwater and groundwater recharge, using the latest available climate projections, from UKCP18. In this article, we have described the dataset and its evaluation against observational hydrological datasets, to give some confidence in the use of eFLaG as a dataset that can be used to assess the potential impacts on climate change on UK hydrology for a very wide range of applications.

The eFLaG dataset was developed specifically as a demonstration climate service for use by the water industry for water resources and drought planning, and hence by design is focused on future projections of drought, low river flows and low groundwater levels. We therefore present eFLaG primarily as a dataset for this purpose. Ongoing work is underway to demonstrate the utility of eFLaG for future drought projections (Parry et al. submitted; Tanguy et al. submitted) and for future drought/water resources planning in practice (Counsell et al. in prep.). The predecessor product, FFGWL, has been widely used within the water industry to provide insight into the future evolution of river flows and groundwater levels through the 21st century to support water resources management plans, and also supported significant academic water resource planning studies (e.g. Borgeomo et al. 2015; Huskova et al. 2016).

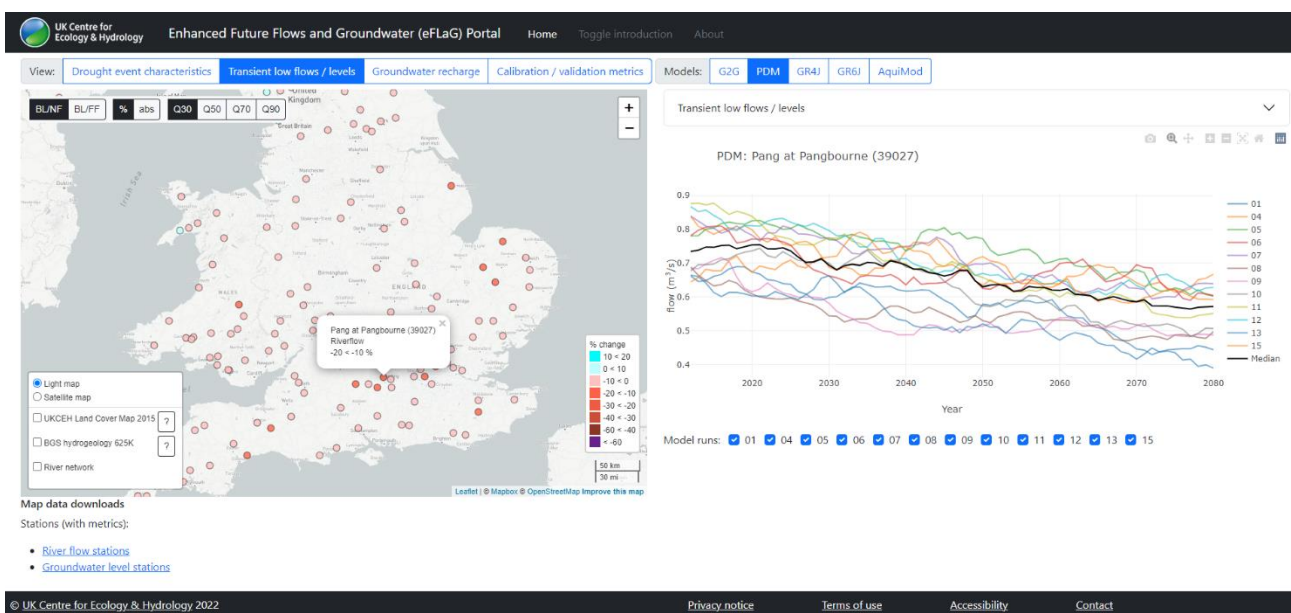
To provide users with a platform for accessing eFLaG datasets, and all the evaluation approaches outlined here, an interactive web application has been developed, the eFLaG Portal (<https://eip.ceh.ac.uk/hydrology/eflag/>). The Portal provides a user friendly front-end for accessing eFLaG results, with several examples shown in Fig 11. The figure demonstrates how eFLaG data can be used to project future drought characteristics for various timeslices, and also how low flow characteristics change through the 21<sup>st</sup> century, based on the analysis conducted in Parry et al. (submitted).



900



901



902

903 **Figure 11: screenshots from the eFLaG Portal. Top: map showing percentage change in**  
904 **drought duration between baseline and near future for eFLaG catchments nationally, using**  
905 **PDM; boxplots showing % changes (using PDM) for a river in southern England (the river**  
906 **Pang) for three timeslices, with boxplots showing range of RCM uncertainty; other drought**  
907 **characteristics available on other tabs. Bottom: map showing percentage change in a low flow**  
908 **metric (Q90) between baseline and near-future for eFLaG catchments nationally, using PDM;**  
909 **with time series showing transient projections of Q90 in moving windows through to the 2080s**  
910 **for the river Pang, each colour representing different RCM runs, black representing median.**  
911 **For all outputs, models other than PDM can be selected using the tabs at the top.**

912

913 By providing a consistent dataset of future river flows, groundwater levels and groundwater  
914 recharge, eFLaG can potentially support a wide range of applications across other sectors.  
915 The FFGWL product also found very wide application for diverse research purposes (for:  
916 water quality, e.g. Charlton et al. 2018; hydroecology, e.g. Royan et al. 2016; groundwater  
917 recharge, Hughes et al., 2021; groundwater level reconstruction, Jackson et al., 2016). For  
918 eFLaG, the good simulation of river flows and groundwater behaviours across much of the  
919 hydrological range suggests that this product could also find application in a whole range of  
920 impact studies, subject to additional evaluation for the purposes in mind. While not validated  
921 specifically for floods, the encouraging evaluation outputs for higher flow percentiles suggests  
922 users can analyse high flow metrics and variability (e.g. frequency of flows above a  
923 threshold), even if not annual maximum peak flows.

924 As with FFGWL, there are a number of advantages of using eFLaG for future projections: it  
925 is a spatially coherent dataset, meaning that future changes in hydrological variables can be  
926 compared between catchments, boreholes and aquifers at the regional-to-national scale. This  
927 is a key benefit for both research as well as practical water resources planning. Spatially  
928 coherent projections are needed to address the spatio-temporal dynamics of droughts (e.g.  
929 Tanguy et al. 2021) and how these may change in future and what this may mean for water  
930 resources planning – where, in practice, water resources management plans often involve  
931 transfers between regions (e.g. Murgatroyd et al. 2021). Tanguy et al. (submitted) address  
932 the changing future spatial coherence of droughts using eFLaG.

933 Another key benefit of eFLaG is that transient time series (daily data from 1980 to 2080) allow  
934 users to can explore the future evolution of river flow and groundwater variability on  
935 interannual and decadal timescales, rather than just using ‘Change Factor’ approaches that  
936 compare between future time slices and the baseline.

937 The use of an ensemble of outputs enables users to consider uncertainty in driving data (via  
938 the 12 member RCM ensemble) as well as, for river flows, hydrological model uncertainty. In  
939 addition, different models provide different benefits: G2G performs less well against  
940 observations than the (calibrated) lumped catchment models, but does enable the  
941 characterisation of natural flows, which is vital for some uses (e.g. in providing naturalised  
942 river flows for regionalisation or as a baseline for assessing impacts, as common in regulatory  
943 and hydroecology applications e.g. Terrier et al. 2021). Moreover, abstractions and  
944 discharges can be added to the naturalised runs, as demonstrated by Rameshwaran et al.  
945 2022. This opens up the possibility of projecting the evolution of future naturalised and  
946 impacted river flows separately – a follow-up study on this topic is underway by the authors.

947 Furthermore, G2G’s response to rainfall may be less tailored to the present-day climate than  
948 the calibrated models, as noted in the limitations section. The eFLaG hydrological model  
949 ensemble therefore includes models that may be beneficial for different applications  
950 according to the particular needs of end-users.

951 **Limitations and guidance**

952 Users of the eFLaG dataset should be aware of its limitations. While the evaluation shows  
953 encouraging results at the national scale, there are inevitably some catchments and  
954 boreholes where the evaluation (either Stage 1, Stage 2 or both) indicates poorer quality  
955 simulations. Users must be aware of this, and should consult all the provided evaluation  
956 metrics when considering which catchments to use (and which models to use) in their  
957 analyses.

958 Users must also be aware that while there is some consideration of uncertainty through the  
959 adoption of the RCM PPE, and the use of a multiple models for river flows, there are many  
960 other sources of uncertainty not sampled in eFLaG. While the PPE gives a range of 12  
961 outcomes, it is only one UKCP18 product and one emissions scenario, so does not sample  
962 the full range of outcomes in UKCP18. The emissions scenario, RCP8.5, is considered to be  
963 a pessimistic scenario (Hausfather & Peters, 2020), so this should be borne in mind, and the  
964 eFLaG projections (along with other uses of the UKCP18 Regional projections) can arguably  
965 be seen as akin to a 'worst case' for planning (Arnell et al. 2021). Future work should position  
966 eFLaG against the wider range of UKCP18 outcomes.

967 Furthermore, only one bias correction approach is used. Although we use a range of river  
968 flow models, clearly other hydrological models could provide different outcomes than the set  
969 used here, and we have only used one groundwater level model and recharge model  
970 respectively so have not considered model uncertainty for groundwater. We have also not  
971 considered other sources of uncertainty in the hydrological modelling (e.g. parametric  
972 uncertainty, as in e.g. Smith et al. 2019), nor the impacts of different observational driving  
973 climate datasets (e.g. different formulations of Potential Evapotranspiration, as in e.g. Tanguy  
974 et al. 2018). These studies demonstrate these can be significant sources of uncertainty, but  
975 it was beyond scope to consider within the resources available to eFLaG given the high  
976 number of existing runs – future studies should address this.

977 The eFLaG modelling framework adopted the approach of calibrating using a full period-of-  
978 record, rather than using a split sample approach. Given the length of record, this is unlikely  
979 to be too significant (as shown for GR4J in the UK by Harrigan et al. 2018) relative to using  
980 split sampling, but at the same time, uncertainties inevitably remain about future projections  
981 well outside the calibration period, not least given likely non-stationarities in catchment  
982 properties. It should also be born in mind that strong performance of a model as indicated by  
983 good metric values is not necessarily a reliable indicator of a models ability to reproduce  
984 trends in hydrological signatures such as those describing low flows (Todorović et al. 2022).

985 Following on from this, one important limitation of this study – in common with the original  
986 Future Flows product (Prudhomme et al. 2012), and indeed a great majority of climate  
987 projections in hydrology – is the lack of explicit modelling of human disturbances. This is  
988 simply unavoidable as large-scale datasets of artificial influences have only recently been

989 made available in the UK, and only for England (e.g. Rameshwaran et al. 2022). This  
990 especially applies for the lumped catchment models and groundwater level model. As such  
991 processes are not represented, they will simply be accounted for implicitly during calibration.  
992 Of course, this is unrealistic as artificial influences are likely to change in future and such  
993 non-stationarity could be locally significant. However, it should be borne in mind that the  
994 purpose of eFLaG is to model future river flow characteristics based on current catchment  
995 conditions, rather than truly chart future river flow trajectories in these catchments. For most  
996 practical applications, assuming current artificial influences and projecting forwards in time is  
997 entirely reasonable, especially in the absence of any informed understanding of how artificial  
998 influences will change.

999 There are also considerations for end users when applying the projections directly in impact  
1000 assessments. Notably, the HadREM3-GA705 climate model that underpins the UKCP18  
1001 RCM outputs is run on a 360-day calendar year. The eFLaG projections do not modify this  
1002 calendar when producing the meteorological, hydrological and hydrogeological variables and  
1003 it is therefore the responsibility of the end user to deal with this in an appropriate way. There  
1004 are a number of ways of doing this (e.g. Prudhomme et al. 2012; Dobor et al. 2015) and in  
1005 general, there is no agreed optimal approach. Where this is performed as a post-processing  
1006 step by the user (as with the eFLaG datasets), it is likely that the best approach will depend  
1007 on the impact or systems modelling being undertaken.

1008 Finally, eFLaG only provides projections for a subset of the UK gauging station network (200  
1009 catchments from some 1200 on the NRFA). This is an inevitable constraint, as with the  
1010 original FFGWL product (300 locations). While we have tried to sample UK hydrology to give  
1011 users as much scope as possible, there will still be a need to transpose projections to sites  
1012 of interest for some users. One of the benefits of eFLaG is that gridded river flow and recharge  
1013 models are used. While these gridded datasets are not yet openly available, current follow-  
1014 up initiatives are looking to exploit them for providing projections at ungauged locations.

1015

## 1016 **9. Data Availability**

1017

1018 The eFLag dataset is associated with a Digital Object Identifier. This must be referenced fully  
1019 for every use of the eFLag data as: <https://doi.org/10.5285/1bb90673-ad37-4679-90b9-0126109639a9>  
1020

1021

1022 All eFLaG files are available through the UKCEH Environmental Informatics Data Centre:  
1023 <https://catalogue.ceh.ac.uk/documents/1bb90673-ad37-4679-90b9-0126109639a9>  
1024

1024

1025 The data are stored as .csv files in the folder structure shown in the Guidance note available  
1026 at Hannaford et al. (2022). In total there are 3304 files: one for each variable, model and

1027 catchment/borehole combination. They can be broadly split into two groups of files (Table 4),  
 1028 simobs and simrcm, as follows.

1029 simobs

1030 For the meteorological data, the simobs files contain date-indexed, observation-driven  
 1031 simulations (sim) data for precipitation with snowmelt and potential evaporation. For river  
 1032 flows and groundwater levels the simobs files contain date-indexed, observation-driven  
 1033 simulations (sim) and associated observations (obs) if they exist.

1034 simrcm

1035 For the meteorological data, the simrcm files contain date-indexed, RCM-driven simulations  
 1036 for the twelve RCMs used in eFLaG for both precipitation with snowmelt and potential  
 1037 evaporation. For river flows and groundwater levels the simrcm files contain date-indexed,  
 1038 RCM-driven simulations for the twelve RCMs used in eFLaG.

1039 **Table 4.** eFLaG dataset structure information

	Data	Name of file	Years available
simobs	Daily meteorology (precipwsnow (mm d <sup>-1</sup> ) + PET (mm d <sup>-1</sup> ))	<i>ukcp18_simobs_[nrfa-station-number/borehole-name].csv</i>	Jan 1961 – Dec 2018
	Daily river flow (m <sup>3</sup> s <sup>-1</sup> )	<i>modelname_simobs_nrfa-station-number.csv</i>	Jan 1963 – Dec 2018
	Daily groundwater levels (m AOD)	<i>AquiMod_simobs_borehole-name.csv</i>	Jan 1962 – Dec 2018
	Daily groundwater recharge (mm d <sup>-1</sup> )	<i>zoodrm_simobs_groundwater-body-name.csv</i>	Jan 1962 – Dec 2018
simrcm	Daily meteorology (precipwsnow (mm d <sup>-1</sup> ) + PE mm d <sup>-1</sup> )	<i>ukcp18_simobs_nrfa-station-number.csv</i>	Dec 1980 – Nov 2080
	Daily river flow (m <sup>3</sup> s <sup>-1</sup> )	<i>modelname_simrcm_nrfa-station-number.csv</i>	Dec 1982 – Nov 2080
	Daily groundwater levels (m AOD)	<i>AquiMod_simrcm_borehole-name.csv</i>	Jan 1982 – Nov 2080
	Daily groundwater recharge (mm d <sup>-1</sup> )	<i>zoodrm_simrcm_groundwater-body-name.csv</i>	Jan 1981 – Nov 2080

1040

1041 where *modelname* is G2G, PDM, GR4J, GR6J. NRFA station numbers and borehole names are given  
 1042 in the eFLaG\_Station\_Metadata.xlsx workbook.

1043

1044 **Conditions of Use**

1045 The eFLaG dataset is available under a licensing condition agreement. For non-commercial  
 1046 use, the products are available free of charge. For commercial use, the data might be made

1047 available conditioned to a fee to be agreed with UKCEH and NERC BGS licensing teams,  
1048 owners of the IPR of the datasets and products.

1049

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1056 events that helped shape eFLaG. JM, MM, MA and CJ publish with the permission of the  
1057 Executive Director, British Geological Survey (UKRI).

1058

## 1059 **Author Contributions**

1060 JH led the study and the river flow components, JM led the groundwater level and  
1061 groundwater recharge components. AK and RL created the bias-corrected climate input data.  
1062 Site selection was carried out by SP, TC and JM. Hydrological simulations were run by KS  
1063 and TC (GR models), AR, AK and VB (G2G model) and JW, RM, SC and SW (PDM). JM and  
1064 MM produced the groundwater level and groundwater recharge simulations. SP and TC led  
1065 on evaluation and flow regime/drought analysis. CC, MD, MS, AW carried out the  
1066 demonstrator work and water industry engagement that helped design and shape eFLaG. ST  
1067 led on data management and portal development. JH led the preparation of the manuscript  
1068 with input from all authors. All authors contributed to the direction of the study and delivery of  
1069 the dataset.

1070

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