



1 **Integrated ecohydrological hydrometric and stable water isotope data**  
2 **of a drought-sensitive mixed land use lowland catchment**

3

4 **Doerthe Tetzlaff<sup>1,2</sup>, Aaron Smith<sup>1</sup>, Lukas Kleine<sup>1,2</sup>, David Dubbert<sup>1</sup>, Jonas Freymüller<sup>1</sup>**  
5 **Hauke Daempfling<sup>1</sup> and Chris Soulsby<sup>3</sup>**

6 <sup>1</sup>IGB Leibniz Institute of Freshwater Ecology and Inland Fisheries Berlin, Berlin, Germany

7 <sup>2</sup>Humboldt University Berlin, Berlin, Germany

8 <sup>3</sup>Northern Rivers Institute, School of Geosciences, University of Aberdeen, UK

9

10 **Corresponding author:**

11 Doerthe Tetzlaff; [doerthe.tetzlaff@igb-berlin.de](mailto:doerthe.tetzlaff@igb-berlin.de)

12

13

14 **Abstract**

15 Data from long-term experimental catchments are the foundation of hydrological sciences and  
16 are crucial to benchmark process understanding, observe trends and natural cycles, and are  
17 prerequisites for testing predictive models. Integrated data sets which capture all compartments  
18 of our landscapes are particularly important in times of land use and climate change. Here, we  
19 present ecohydrological data measured at multiple spatial scales which allows differentiation of  
20 “blue” water fluxes (which maintain streamflow generation and groundwater recharge) and  
21 “green” water fluxes (which sustain vegetation growth). There are two particular unique aspects  
22 to this data set though: a) we measured water stable isotopes in the different landscape  
23 compartments (that is in precipitation, surface water, soil, ground- and plant water); and b)  
24 conducted this monitoring during the extreme drought of 2018 in Central Europe. Stable water  
25 isotopes are so useful in hydrology as they provide “fingerprints” of the pathways water took  
26 when moving through a catchment. Thus, isotopes allow to evaluate the dynamic relationships  
27 between water storage changes and fluxes, which is fundamental to understanding how  
28 catchments respond to hydroclimate perturbations or abrupt land use conversion. Second, as  
29 we provide the data until 2020 one can also investigate recovery of water stores and fluxes after  
30 extreme droughts. Last but not least: lowland headwaters are often understudied systems  
31 despite them providing important ecosystem services such as groundwater and drinking water  
32 provision and management for forestry and agriculture.

33



## 34 **1. Introduction**

35

36 Progress in scientific hydrology and provision of an evidence base for sustainable land and water  
37 management are only possible due to detailed, long-term observational data collected from long-  
38 term experimental watersheds (Hewlett et al., 1969; Robinson et al., 2013). Such experimental  
39 “outdoor laboratories” are invaluable scientific resources given the complexity of increasing  
40 pressures on water supplies (e.g. Cosgrove and Loucks, 2015), land use change (Neill et al.,  
41 2021) and the uncertain effects and non-stationarity of projected climate change (Milly et al.,  
42 2015).

43 Ecohydrology adopts an interdisciplinary approach to investigating interlinkages between the  
44 structure and function of ecological systems and the partitioning, flux and storage of fresh water  
45 (Guswa et al., 2020). Recent advances in monitoring and modeling have created manifold  
46 opportunities to address urgent ecohydrological questions on the importance of interlinkages of  
47 processes across the critical zone (CZ) - the dynamic, life-sustaining near-surface of the terrestrial  
48 earth that extends between the top of vegetation canopies, through the soil and into groundwater  
49 (Grant & Dietrich, 2017). Within the CZ concept, vegetation plays a central and dynamic role in  
50 partitioning incoming precipitation into “blue” water fluxes (streamflow generation and  
51 groundwater recharge) and “green” water fluxes which maintain vegetation growth (Evaristo et  
52 al., 2015).

53 To enhance ecohydrological process understanding in catchment systems, robust, multi-scale  
54 integrated data sets are required (Tetzlaff et al., 2021). In this regard, water stable isotopes and  
55 other tracers can help identify sources and pathways of water in the landscape and across the  
56 CZ to elucidate how different land use affects water partitioning between green and blue water  
57 fluxes (Dubbert and Werner, 2019; Tetzlaff et al., 2015). Importantly, water stable isotopes have  
58 enhanced the characterization of the celerity of hydrological fluxes in different CZ compartments,  
59 as well as quantifying the velocity of water particles and associated mixing relationships in the  
60 subsurface (Benettin et al., 2015; Birkel et al., 2011). Evaluating the dynamic relationships  
61 between water storage changes and fluxes is fundamental to understanding how catchments  
62 respond to hydroclimate perturbations, such as anomalous dry or wet periods, or abrupt land use  
63 conversion. This provides a more nuanced and integrated understanding of how key  
64 ecohydrological couplings may be at risk during long-term changes in blue and green water  
65 partitioning resulting from climate and land use change (Orth and Destouni, 2018). Such  
66 integrated understanding is important in the context of projected increases in air temperature,  
67 aridity, and in precipitation patterns, which may cause more variability in water availability  
68 threatening the sustainability of important ecosystem services (Okruszko et al., 2011). As an  
69 increase in drought frequency and severity is expected across Europe as the 21<sup>st</sup> century  
70 progresses, the development of effective and evidence-based amelioration measures to underpin  
71 sustainable and integrated land and water management policies for changing climatic conditions  
72 is urgently needed (Samaniego et al., 2018).



73 Consequently, integrated ecohydrological and stable isotope data sets targeted at understanding  
74 the effects of different types of environmental change have outstanding potential, not least  
75 because interdisciplinary environmental research tends to give unanticipated insights (Burt,  
76 1994). Such integrated data streams allow identification and quantification of the linkages  
77 between rainfall, soil moisture, groundwater and runoff generation, facilitating deeper  
78 understanding of flood and drought risk in different types of landscapes and under different land  
79 use management (Huntingford et al., 2014).

80 Water resources in the extensive, glacially formed, lowland landscape of northern Europe,  
81 including the North German Plain (NGP) sustain food production (Gutzler et al., 2015; Barkmann  
82 et al., 2017) and water supplies to large cities like Berlin. Interestingly, such lowlands catchments  
83 are still relatively understudied compared to more upland headwater landscapes with stronger  
84 topographic controls on drainage of surface and subsurface water (Devito et al., 2005). In low  
85 elevation catchments across the NGP, streams are usually groundwater-dominated, but the  
86 temporal and spatial heterogeneities in the hydrological functioning of these catchments are still  
87 not fully understood (Boulton and Hancock, 2006). For example, there is still a limited evidence  
88 base for quantifying how drought affects groundwater recharge and stream flow generation in  
89 lowland areas in Central Europe, including the cessation of flow during the summer (Germer et  
90 al., 2011).

91 To help address these knowledge gaps, here, we present a comprehensive set of ecohydrological  
92 hydrometric and stable water isotope data of two years of data for the Demnitzer Mill Creek  
93 catchment, NE Germany. The data set is unique in its integrative characteristics; that the different  
94 compartments of the CZ were sampled across a mesoscale catchment in terms of their isotopic  
95 signature and supporting ecohydrological data. By coincidence, these first two years, of what will  
96 be a long-term study, captured the changing impacts of a prolonged drought period (2018-2020)  
97 with a strong negative rainfall anomaly that became the most severe regional drought so far in  
98 the 21<sup>st</sup> century (Kleine et al., 2021a). The data allow the effects of droughts (and their  
99 persistence) on water storage, fluxes and age dynamics in the CZ to be investigated (Smith et al.,  
100 2022). Our objective here is to provide this high spatio-temporal resolution ecohydrological  
101 dataset to improve understanding of the storages and flow pathways of both blue and green water  
102 across processes at the larger catchment scale in lowland catchments. We are continuing these  
103 observations to assess long-term climatic trends at the drought sensitive region of NE Germany.

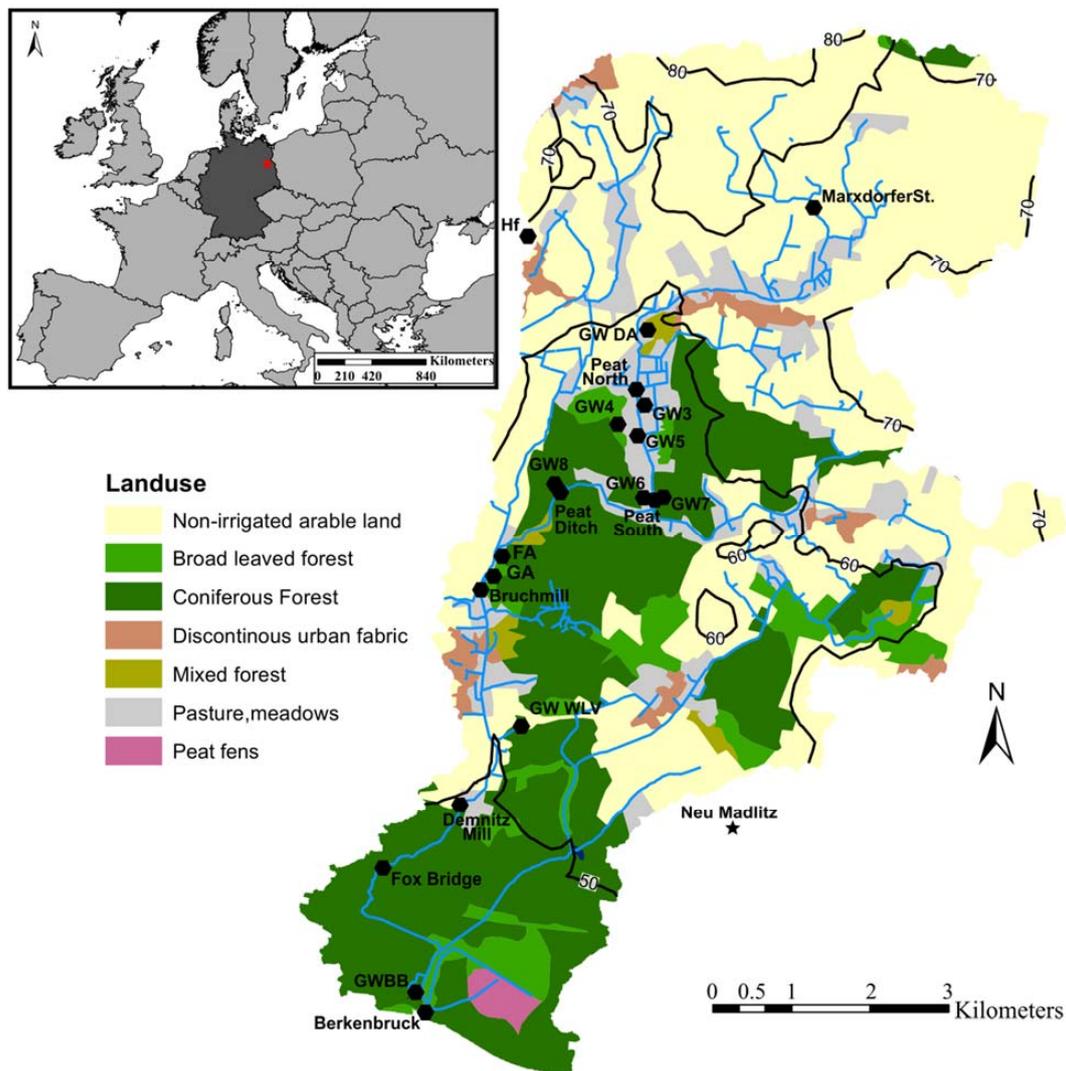
104

105



106 **2. Site description**

107



108

109 **Figure 1: The Demnitzer Mill Creek catchment and its location within Europe and**  
110 **Germany. Hexagonal points (●) are measurement locations in the catchment and the star**  
111 **(★) are meteorological measurements by the German Weather service (DWD Deutscher**  
112 **Wetterdienst).**

113

114



115 **Table 1** - Overview of the properties of the Demnitzer Millcreek catchment at the catchment  
116 outlet. Overview includes physiological characteristics, landuse, and geology.

<b>Area (km<sup>2</sup>)</b>	66.39	<b>Topographic Relief (m)</b>	50.23
<b>Runoff Ratio</b>	0.10	<b>Mean Slope (%)</b>	1.98
<b>Landuse (%)</b>			
Mixed Forest	1.0	<b>Geology (%)</b>	
Conifer Forest	29.2	Base moraine	35.5
Broadleaf Forest	6.0	End moraine	2.3
Peat	0.7	Deposits of glacial valleys	6.9
Pasture	10.2	Peat Fen	5.9
Agricultural/arable land	50.4	Periglacial/fluvial deposits	16.3
Urban	2.5	Glacial/fluvial deposits	31.1
		Sandy peat fen	2.0

117  
118 The data presented here were monitored in the Demnitzer Millcreek catchment (DMC) located in  
119 NE Germany (52°23'N, 14°15'E; Figure 1). The DMC is a lowland drought-sensitive area south  
120 east of Berlin, the German capital, and situated in the NGP. The region has high socio-economic  
121 significance through the provision of numerous ecosystem services; including food security,  
122 timber production, groundwater recharge and river flow generation which sustains drinking water  
123 supplies for Berlin (Kleine et al., 2021a). The original motivation behind establishing DMC as an  
124 observatory in 1990 was to investigate the impact of agricultural pollutants on surface water  
125 quality (Gelbrecht et al., 2000, 2005).

126 The hydroclimate is temperate with warm, humid summers (Kottek et al., 2006). Mean annual  
127 precipitation and air temperature are 567 mm yr<sup>-1</sup> and 9.6°C, respectively (DWD, 2020, for 2006-  
128 2015). Seasonal contrasts are characterized by higher summer precipitation, mainly from high  
129 intensity, convective events; and slightly lower precipitation during frequent, frontal rainfall events  
130 in winter. The landscape was shaped by the last glaciation (Weichselian); soils are predominantly  
131 sandy and formed on glacial and fluvial deposits (Kleine et al., 2021b). The catchment is  
132 dominated by groundwater and likely had little surface runoff before human intervention.  
133 Previously, numerous peat fens and freshwater lakes in hollows existed, but these were drained  
134 during a long historic evolution of anthropogenic management (Nützmänn et al., 2011). Land use  
135 is currently dominated by farming and forestry (Kleine et al., 2020; Smith et al., 2020c). The  
136 catchment is also relatively sparsely populated, and has recently experienced recolonization of  
137 beaver (Smith et al., 2020a), wolf (Vogel, 2014) and even sporadic sighting of elk (Martin, 2014).  
138 Maintenance of crucial ecosystem services in the landscape is dependent on sufficient seasonal  
139 precipitation input to sustain adequate soil moisture levels in the rooting zone to support crop and  
140 tree growth (Drastig et al., 2011); and acceptable groundwater recharge to sustain groundwater-  
141 surface water exchanges. However, high (~90 %) proportions of evapotranspiration, particularly  
142 from forested areas and poor water retention in the widespread sandy soils (Smith et al., 2021),  
143 result in catchment drought sensitivity (Kleine et al., 2020). Further, increased flow disconnections  
144 and fragmentation of the stream network occurs during droughts (Kleine et al., 2021a; Smith et  
145 al., 2021).

146



147 **3. Data and instrumentation overview**

148 **3.1 Instrumentation overview**

149 A fully automatic weather station (AWS) was installed and has been operated in Hasenfelde (Hf,  
150 Figure 1) since April 2018, including radiation, air temperature, relative humidity, precipitation and  
151 ground heat flux every 15-minutes. A modified autosampler (ISCO 3700, Teledyne Isco, Lincoln,  
152 USA) was installed nearby to collect daily samples of precipitation to supplement the AWS.  
153 Weekly cumulative precipitation was additionally collected at four locations nested from north to  
154 south in the catchment: Marxdorfer St., Demnitz Mill, Bruchmill, and Berkenbruck (Figure 1&2)  
155 from July 2018 to April 2020. Measurements of throughfall were collected under the canopy at  
156 Forest A at five locations (Forest A1-5) within a 10m square fenced area. Throughfall was  
157 collected using simple rain gauges (Rain gauge kit, S. Brannan & Sons, Cleator Moor, UK;  
158 <https://doi.org/10.18728/igb-fred-623.0>)

159 Soil moisture and temperature profiles were established at Forest A (FA) and Grass A (GA) in  
160 June 2018 with 18-sensors per site (SMT-100, Umwelt-Geräte-Technik GmbH, Müncheberg,  
161 Germany). The sensors were distributed equally at soil depths of 20, 60, and 100cm at each site  
162 (i.e. three sensors per depth), measuring every 15-minutes (<https://doi.org/10.18728/igb-fred-623.0>).  
163

164 Sap flow measurements were established in 12 trees at Forest A including Scots Pine (*Pinus*  
165 *sylvestris*), European Oak (*Quercus robur*), common hazel (*Corylus avellana*), and Red Oak  
166 (*Quercus rubra*). Measurements were conducted using 2-4 radially installed thermal dissipation-  
167 based sap flow sensors (TDP probes, Dynamix Inc., Houston, TX, USA). Sap flow measurements  
168 were recorded every 15 minutes (<https://doi.org/10.18728/igb-fred-623.0>).

169 Stream water level was established at four locations within the catchment; Peat North, Bruchmill,  
170 Demnitz Mill, and Berkenbruck (<https://doi.org/10.18728/igb-fred-623.0>).The water level was  
171 established by IGB Leibniz Institute of Freshwater Ecology and Inland Fisheries and recorded  
172 with divers (Micro 10m and Baro) at Peat North and Demnitz Mill, and at Bruchmill (Van Essen  
173 Instruments). The divers utilized at each site include an internal atmospheric pressure correction  
174 (AquiLite ATP 10, AquiTronic Umweltmeßtechnik GmbH, Kirchheim/Teck, Germany). Water level  
175 measurements began at Demnitz Mill in 1986, and in January and June 2018 for Peat North and  
176 Bruchmill, respectively. Water level has been recorded since 1982 at Berkenbruck using pressure  
177 transducers and was established and collected by the Landesamt für Umwelt. Channel stability  
178 at Demnitz Mill and Berkenbruck has permitted rating curve development to translate water level  
179 measurements to discharge. Stream water level at Bruchmill was supplemented with daily stream  
180 water samples for stable water isotope analysis collected from an autosampler (ISCO 3700,  
181 Teledyne Isco, Lincoln, USA). The autosampler was established in December 2018  
182 (<https://doi.org/10.18728/igb-fred-623.0>).

183 Groundwater level divers were installed at five locations throughout the catchment in 2001 (GW3,  
184 GW4, GW5, GW7, and GW8) (Figure 1&2). Groundwater level at each site was measured every  
185 four hours with an AquiLite ATP-10 diver (AquiTronic Umweltmeßtechnik GmbH, Kirchheim/Teck,



186 Germany) with internal correction for atmospheric pressure ([https://doi.org/10.18728/igb-fred-](https://doi.org/10.18728/igb-fred-623.0)  
187 [623.0](https://doi.org/10.18728/igb-fred-623.0))

188

### 189 **3.2 Isotope sampling overview**

190 Manual sampling from different locations and different water cycle / landscape compartments  
191 supplemented the autosamplers installed for precipitation at Hasenfelde and for stream water at  
192 Bruchmill. Samples were taken from the weekly cumulative precipitation and throughfall (Forest  
193 A) for each location (Figure 2). Further, monthly samples of soil water were taken at 6 depths  
194 (2.5, 7.5, 15, 30, 60, 90 cm) in triplicate for Forest A and Grass A. This was complemented by  
195 synoptic, spatially distributed sampling of the upper 30cm in 2019. Samples were placed in a  
196 sterile zip-lock bag (CB400-420siZ, Weber Packaging GmbH, Güglingen, Germany) and  
197 analyzed using the direct water vapour equilibrium method (Wassenaar et al., 2008). Weekly grab  
198 samples of stream water were taken at all nested stream water locations (eight locations).  
199 Groundwater isotopes were sampled at six groundwater wells, including two with continuous  
200 groundwater level measurement (GW3, GW8). Groundwater levels at the other sites (GW DA,  
201 GW6, GW WLV, GW BB) were periodically recorded. Vegetation isotopic sampling was  
202 conducted by taking twig samples from different vegetation in Forest A and samples of the non-  
203 green stem of the grass at site Grass A. Vegetation samples were stored at -20°C after sampling  
204 until analysis. Reference for all isotope samples is <https://doi.org/10.18728/igb-fred-623.0>.

205 A layer of paraffin was added to the bottom of all autosampler containers to prevent evaporation  
206 and fractionation from collected water. Autosamplers are emptied each week. Collected weekly  
207 precipitation, throughfall, stream water, and groundwater were sealed and refrigerated until  
208 isotopic analysis.

209 All liquid water samples ( $P_{iso}$ ,  $THR_{iso}$ ,  $Q_{iso}$ ,  $GW_{iso}$ ) were filtered (0.2  $\mu$ m, cellulose acetate, Lab  
210 Logistics Group GmbH, Meckenheim, Germany) and cooled before being analyzed using Cavity  
211 Ring-Down Spectroscopy (CRDS, L2130-i, Picarro, Inc., CA, USA). Additionally, the CDRS was  
212 used for the analysis of the to direct liquid-water equilibrium method for soil water. Vegetation  
213 samples were extracted in January 2020 using the cryogenic extraction method given in Dubbert  
214 et al. (2013, 2014) and analyzed with the CDRS.

215



216 **Table 2 – Site locations in DMC, including site name, coordinates, data collected, start**  
 217 **and end dates, and resolution. N/A indicates not applicable, P is precipitation, GW is**  
 218 **groundwater level, THR is throughfall, Ts is soil temperature, va is wind speed/direction,**  
 219 **Ta is air temperature, Pa is air pressure, RH is relative humidity, NR is net radiation, Sap**  
 220 **is sap flow, and subscript iso indicates isotopic sampling. AWS indicates measurements**  
 221 **of P, va, Ta, Pa, RH, and NR**

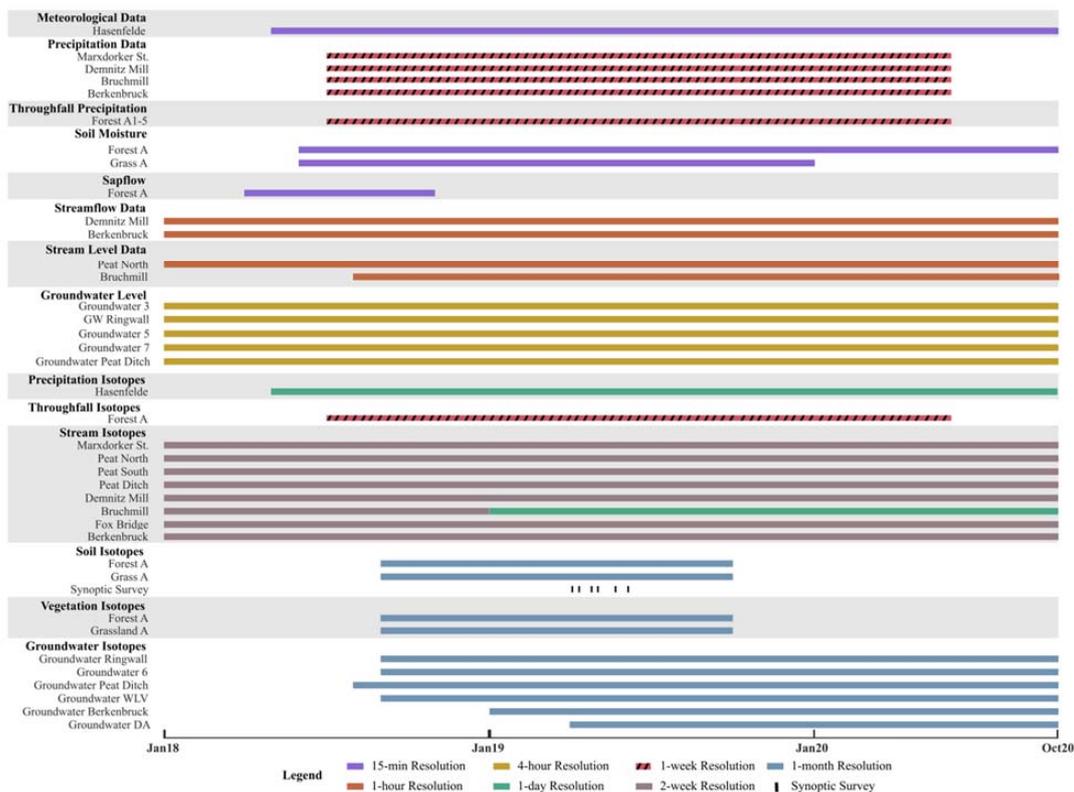
Site Name	ID	Location (UTM 33N)		Data Type	Installation/Start Date	Discontinued/End Date	Resolution	
		Latitude	Longitude				Temporal	Spatial
Marxdo rfer St.	Marxdo rfer St.	5810076	449773	P, P <sub>iso</sub> , Q <sub>iso</sub> , T <sub>s</sub>	Jan 10, 2018 (Q <sub>iso</sub> ) Jul 9, 2018 (P&P <sub>iso</sub> ) Aug 16, 2019 (T <sub>s</sub> )	Jun 2, 2020 (P&P <sub>iso</sub> ) Jul 11, 2020 (T <sub>s</sub> )	Weekly (P, P <sub>iso</sub> & Q <sub>iso</sub> ) 15-min (T <sub>s</sub> )	T <sub>s</sub> (5cm)
Hasenfelde	Hf	5809705	446068	P, P <sub>iso</sub> , va, Ta, Pa, RH, NR, T <sub>s</sub>	Mar 17, 2018 (AWS) Jul 12, 2018 (P <sub>iso</sub> ) Aug 16, 2019 (T <sub>s</sub> )	Jul 11, 2020 (T <sub>s</sub> )	15-min (AWS & T <sub>s</sub> ) Daily (P <sub>iso</sub> )	AWS (2m) T <sub>s</sub> (5cm)
Ground water DA	GW DA	5808335	447527	GW <sub>iso</sub>	Apr 16, 2019	N/A	Monthly	N/A
Peat North	PN	5807703	447474	Q <sub>iso</sub>	Jan 10, 2018	N/A	Weekly	N/A
Ground water 3	GW3	5807499	447582	GW	Jan 10, 2001	N/A	4-hour	N/A
Ground water Ringwa II	GW4	5807247	447233	GW, GW <sub>iso</sub>	Feb 22, 2001 (GW) Sep 11, 2018 (GW <sub>iso</sub> )	N/A	4-hour (GW) Monthly (GW <sub>iso</sub> )	N/A
Ground water 5	GW5	5807099	447490	GW	Jan 10, 2001	N/A	4-hour	N/A
Peat Ditch	Peat Ditch	5806364	446487	Q <sub>iso</sub>	Mar 21, 2018	N/A	Weekly (Q <sub>iso</sub> )	N/A
Ground water Peat Ditch	GW8	5806320	446488	GW, GW <sub>iso</sub>	Jan 10, 2001 (GW) Aug 15, 2018 (GW <sub>iso</sub> )	N/A	4-hour (GW) Monthly (GW <sub>iso</sub> )	N/A
Ground water 7	GW7	5806307	447726	GW	Feb 22, 2001 (GW)	N/A	4-hour (GW)	N/A
Ground water 6	GW6	5806274	447678	GW <sub>iso</sub>	Sep 11, 2018	N/A	Monthly	N/A
Peat South	Peat South	5806262	447712	Q <sub>iso</sub> , T <sub>s</sub>	Jan 10, 2018 (Q <sub>iso</sub> ) Aug 16, 2019 (T <sub>s</sub> )	Jul 11, 2020 (T <sub>s</sub> )	Weekly (Q <sub>iso</sub> ) 15-min (T <sub>s</sub> )	T <sub>s</sub> (5cm)
Forest A	FA	5805520	445731	Sap, SM, SM <sub>iso</sub> , THR, THR <sub>iso</sub> , T <sub>s</sub>	Apr 21, 2018 (Sap) Jun 15, 2018 (SM & T <sub>s</sub> ) Oct 18, 2018 (SM <sub>iso</sub> ) Jul 11, 2018 (THR & THR <sub>iso</sub> )	Nov 1, 2018 (Sap) N/A (SM) Jul 16, 2019 (SM <sub>iso</sub> ) May 19, 2020 (THR & THR <sub>iso</sub> )	15-min (Sap) 15-min (SM & T <sub>s</sub> ) Monthly (SM <sub>iso</sub> ) Weekly (THR & THR <sub>iso</sub> )	12 Trees (Sap) SM & T <sub>s</sub> (6 sites, 20, 60, 100cm depths)



								THR & THR <sub>iso</sub> (5 sites)
Grass A	GA	5805125	445495	SM, SM <sub>iso</sub> , T <sub>s</sub>	Jun 15, 2018 (SM & T <sub>s</sub> ) Oct 18, 2018 (SM <sub>iso</sub> )	Jul 16, 2019 (SM <sub>iso</sub> ) Jan 7, 2020 (SM & T <sub>s</sub> )	15-min (SM & T <sub>s</sub> ) Monthly (SM <sub>iso</sub> )	SM & T <sub>s</sub> (6 sites, 20, 60, 100cm depths)
Bruchm ill	Bruchm ill	5805088	445459	P, P <sub>iso</sub> , Q <sub>iso</sub>	Jan 10, 2018 (Q <sub>iso</sub> -weekly) Dec 28, 2018 (Q <sub>iso</sub> -daily) Jul 9, 2018 (P&P <sub>iso</sub> )	Dec 28, 2018 (Q <sub>iso</sub> -weekly) Jun 2, 2020 (P&P <sub>iso</sub> )	Weekly (P & P <sub>iso</sub> ) Daily (Q <sub>iso</sub> )	N/A
Ground water WLV	GW WLV	5803322	445982	GW <sub>iso</sub>	Sep 20, 2018	N/A	Monthly	N/A
Demnitz Mill	Demnitz Mill	5802298	445188	P, P <sub>iso</sub> , Q, Q <sub>iso</sub>	Jan 10, 2018 (Q <sub>iso</sub> ) Jul 9, 2018 (P&P <sub>iso</sub> ) Feb 22, 2011 (Q)	Jun 2, 2020 (P&P <sub>iso</sub> )	Weekly (P, P <sub>iso</sub> & Q <sub>iso</sub> ) 4-hour (Q)	N/A
Fox Bridge	Fox Bridge	5801469	444189	Q <sub>iso</sub>	Jan 10, 2018	N/A	Weekly	N/A
Ground water Berken bruck	GW BB	5799862	444611	GW <sub>iso</sub>	Jan 21, 2019	N/A	Monthly	N/A
Berken bruck	Berken bruck	5799604	444737	P, P <sub>iso</sub> , Q, Q <sub>iso</sub> , T <sub>s</sub>	Nov 1, 1982 (Q) Jan 10, 2018 (Q <sub>iso</sub> ) Jul 9, 2018 (P&P <sub>iso</sub> ) Aug 16, 2019 (T <sub>s</sub> )	Jun 2, 2020 (P&P <sub>iso</sub> ) Jul 11, 2020 (T <sub>s</sub> )	Daily (Q) Weekly (P, P <sub>iso</sub> & Q <sub>iso</sub> ) 15-min (T <sub>s</sub> )	T <sub>s</sub> (5cm)

222

223



224  
 225  
 226  
 227

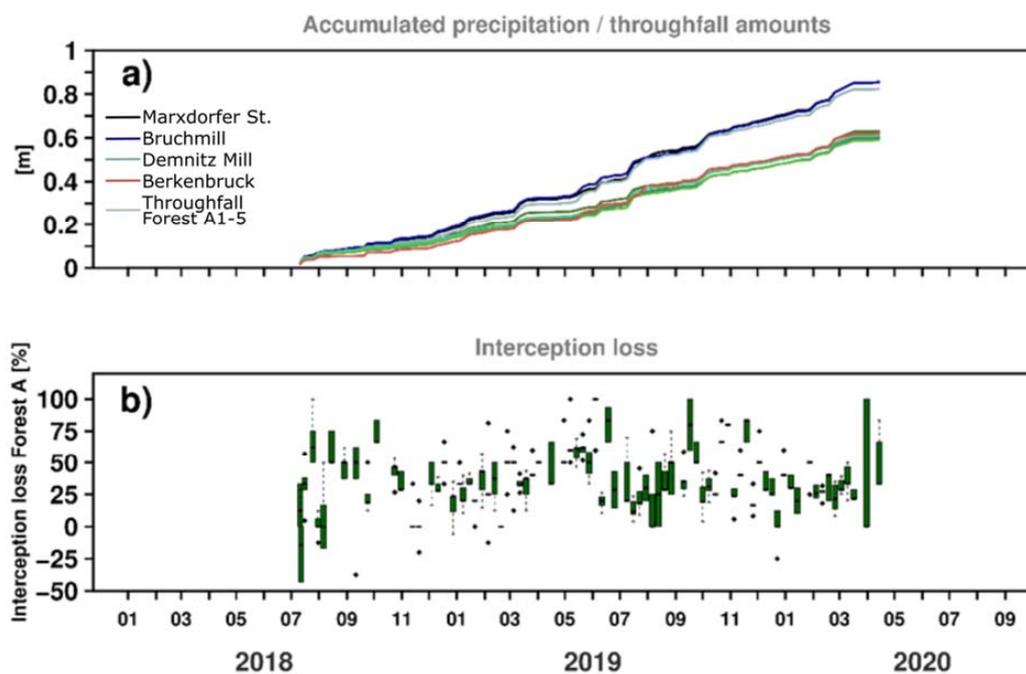
**Figure 2: Spatial data availability and temporal resolution (colour code) of the measurements within the Demnitzer Millcreek Catchment including, meteorological, soil, vegetation, stream, and groundwater hydrological and isotope data sets.**



#### 228 4. Precipitation and throughfall amount

229

230 Monitoring for precipitation commenced in the 2018 summer drought when low rainfall inputs  
231 continued through the following winter (Figure 3a). Large rainfall events (>20 mm/d) were  
232 relatively rare and mostly summer convective storms. Even by summer 2020, most months had  
233 below average rainfall. Throughfall at the Forest A site typically was 70-90 % of incident rainfall,  
234 with higher interception losses in low intensity summer storms and lowest in winter or high  
235 intensity summer storms. Heterogeneity in throughfall was marked (Figure 3b), emphasizing the  
236 importance of the forest canopy in redistributing net rainfall to the forest floor.



237

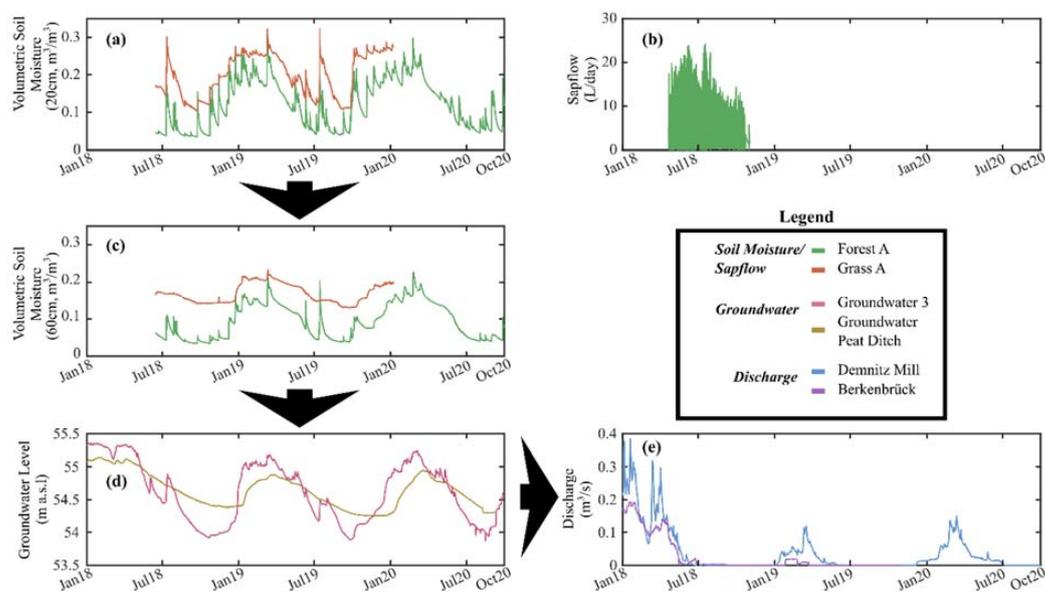
238 **Figure 3: (a) Cumulative precipitation and throughfall at multiple locations throughout**  
239 **the catchment. Throughfall was collected weekly at Forest A with (b) five samplers (1-5)**  
240 **distributed throughout the 10m square fenced region.**



241 **5. Catchment response data**

242 Rainfall fluxes mostly drove short term soil moisture variations (Figure 4a, c); which were more  
243 responsive in the upper soil layers (at 20 cm) than deeper layers. Variability was also more  
244 sensitive under forested land cover, where soils are sandier, more structured and effective rainfall  
245 is lower due to interception losses. Seasonality in evapotranspiration (usefully indexed by sapflow  
246 in Figure 4b) modulated the effects of rainfall on soil moisture storage. Seasonal soil moisture  
247 dynamics also governed groundwater recharge and variation in groundwater levels, which had an  
248 annual range of ~1.5 m at well G3 and ~1m at the peat ditch well (Figure 4d). Despite clear winter  
249 recharge and spring drawdown in each well, peak winter and summer levels were lower in 2019  
250 and 2020 compared to 2018 indicating the cumulative “memory effects” of the drought. This was  
251 also evident in the stream hydrograph with very low discharge peaks in 2019 and 2020, which  
252 also had prolonged periods where flow ceased in the summer, particularly at Berkenbrück. Thus,  
253 despite winter soil moisture replenishment, this was insufficient to match long-term groundwater  
254 recharge. These different correlations underline the added value of simultaneous data from long-  
255 term study sites on transpiration, soil water, groundwater and stream flow as droughts develop  
256 (Smith et al., 2022).

257



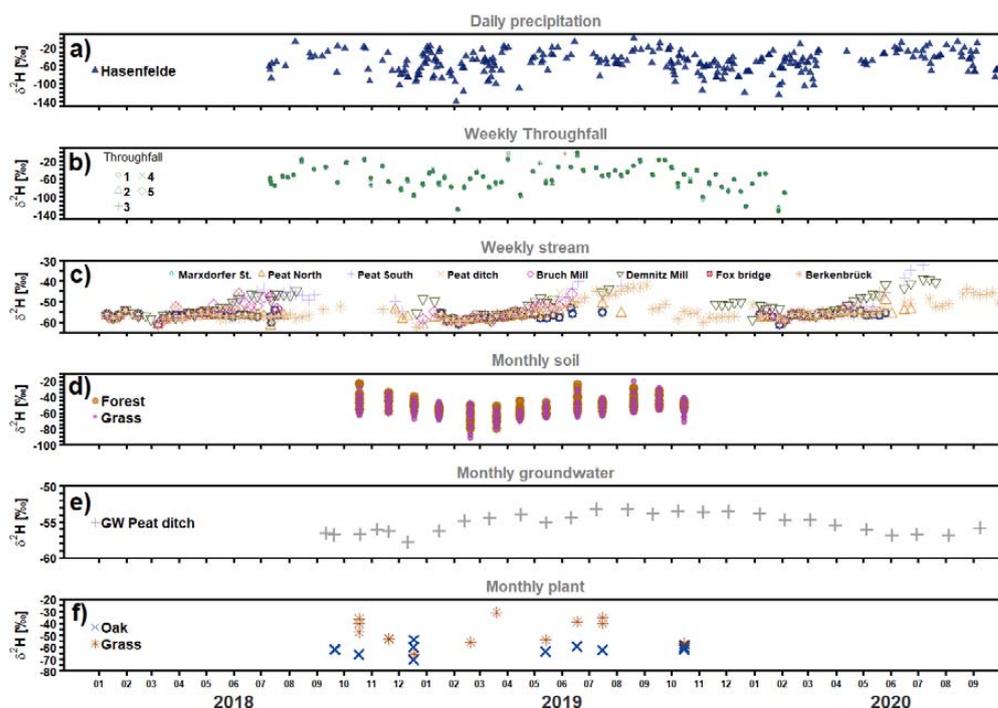
258  
259

260 **Figure 4: (a) Shallow and (c) deep soil moisture, (b) sapflow, (d) groundwater levels and**  
261 **(e) discharge within the Demnitzer Millcreek catchment. Arrows show connections**  
262 **between layers and fluxes. \*Groundwater 3 is within the wetland and Groundwater Peat**  
263 **Ditch is outside the wetland (near Forest A and Grass A, Fig. 1).**



## 264 6. Stable water isotopes

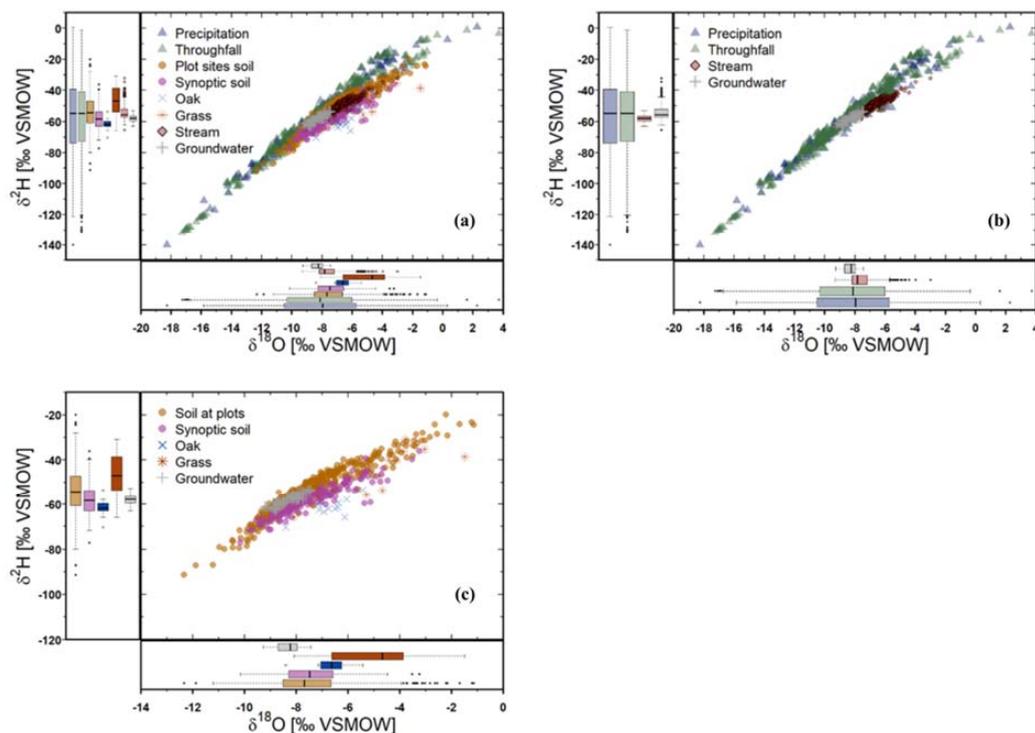
265 Stable water isotope signatures in precipitation showed high day-to-day variability superimposed  
266 on strong seasonality; with more depleted values in winter and more enriched values in summer  
267 (Fig 5a). Interestingly, weekly throughfall signatures were very similar to the (weekly and daily)  
268 precipitation signal showing no strong signs of evaporative fractionation during canopy storage  
269 (Fig 5b). This likely reflects the high intensity nature of most summer rainfall, which affords limited  
270 opportunity for canopy evaporation. Streamwater signatures at all nested sites showed similar  
271 seasonality but much more damping in the signal (Fig. 5c). Groundwater was most damped, and  
272 similar in composition to streamflow during winter (Fig 5d). In summer, sites downstream of  
273 Marxdorfer Strasse showed evidence of evaporative fractionation from either the channel network  
274 or riparian soils and plotted below the meteoric water line before stream flow ceased. Monthly soil  
275 water samples showed higher variability in isotopic composition under forest than under grass,  
276 mainly reflecting soil characteristics with more retentive, loamy and wetter soils at the grassland  
277 site buffering the effects of rainfall inputs. At both sites, seasonal variation in isotopic composition  
278 tracked precipitation, though in deeper soil water samples were more damped. Vegetation  
279 samples from the oaks showed higher variation than from grass.  
280



281  
282 **Figure 5: Time series of deuterium ( $\delta^2\text{H}$ ) in (a) precipitation, (b) throughfall (Forest A), (c)**  
283 **stream water, (d) soil water, (e) groundwater and (f) plant samples at various locations in**  
284 **the catchment.**



285  
286



287  
288  
289  
290  
291

**Figure 6: Dual isotope space ( $\delta^2\text{H}$ - $\delta^{18}\text{O}$ ) plots for (a) all measured isotopic datasets, (b) precipitation, throughfall, stream, and groundwater, and (c) soil (multiple depths), synoptic soil survey (upper 30cm), vegetation, and groundwater.**

292 Differences in the isotope dynamics of different critical zone compartments are shown in dual  
293 isotope space in Figure 6a. The damping of precipitation in groundwater and streamflow is  
294 apparent, as is the fractionation of more enriched summer stream flow samples (Figure 6b).  
295 The role of the soil in partitioning water is apparent from the overlap between deeper soil horizons  
296 and groundwater which were both more weighted to winter precipitation – when recharge is  
297 greatest (Fig 6c). Xylem water in oaks and grass tended to show the effects of fractionation,  
298 which was most marked in the oaks and may point to different soil water sources of root uptake.



299 **7. Data availability**

300 All data presented in this paper are available from the IGB open data repository FRED  
301 <https://fred.igb-berlin.de/data/package/622> (Tetzlaff et al., 2022). The data is published with  
302 detailed metadata (<https://doi.org/10.18728/igb-fred-623.0>) and contact information for any further  
303 questions. There is a readme section per each dataset. We also included a digital elevation model,  
304 shapefile of the catchment boundary and the station locations.

305

306 **8. Summary**

307 The integrated data set presented in this paper is unique because: (1) it captures complicated  
308 ecohydrological dynamics over two years during an exceptional drought (in 2018/2019) in Central  
309 Europe; (2) the different compartments of the critical zone were monitored through stable water  
310 isotope data and complimentary ecohydrological data for contrasting land use and (3) multi-scale,  
311 nested catchment time series were derived. In total data from 49 time series / data sets are  
312 available. The data are quality controlled. We included meteorological data and precipitation and  
313 throughfall amount. Catchment response data include stream discharge at the catchment outlet  
314 and another nested site, and stream level data at two further sites; soil moisture from multiple  
315 depths at two locations (two different landuses), groundwater level data at five locations and  
316 sapflow measurements from one forest location. Stable water isotope data include precipitation  
317 water, throughfall, streamwater at eight sites, soil water isotopes from two sites plus spatially  
318 distributed samples of upper soils, vegetation samples at two locations and groundwater at six  
319 locations. Data continue to be collected and updated data sets will be published based on  
320 available resources.

321 As such, these data provide an excellent, integrated ecohydrological perspective on the drought  
322 response of a lowland agricultural landscape. Such data are of course important in their own right,  
323 but are equally invaluable for challenging environmental models as constraints on internal model  
324 function that can be used to increase confidence in the use of models in projecting the impacts of  
325 future change. Integrated data like the ones summarised here are also important for a range of  
326 scientific questions that are growing in importance as the effects of climate change become more  
327 apparent. These include understanding how do droughts develop and propagate through  
328 components of hydrological systems and compartments of the critical zone? What are the effects  
329 of land cover on this propagation and how does it affect water cycling in vegetation? How long  
330 does recovery of different system components take once rainfall anomalies become positive?  
331 How resilient are different critical zone compartments or entire landscapes against climate  
332 extremes such as droughts? Hopefully, this data set will be used by scientists to increase  
333 understanding on critical issues such as what are the water footprints of alternative land uses and  
334 how can these be reduced whilst maintaining societal needs. This will help to contribute to the  
335 development of more sustainable and resilient land and water management policies that will be  
336 needed in the face of increased longevity and frequency of droughts.

337



338 **Author contributions:** AS and LK prepared the data sets. Datasets were collected by LK and  
339 JF. Isotope data were analysed by DD. DT, CS, AS prepared the manuscript with contributions  
340 from all co-authors.

341

342 **Competing interests:** The authors declare that they have no conflict of interest.

343

344 **Disclaimer:** any reference to specific equipment types or manufacturers is for informational  
345 purposes and does not represent product endorsement. IGB is an equal opportunity provider.

346

347 **Acknowledgements:**

348 We acknowledge the BMBF (funding code 033W034A) which supported the stable isotope 650  
349 laboratory at IGB. Funding for DT was also received through the Einstein 652 Research Unit  
350 “Climate and Water under Change” from the Einstein Foundation Berlin and 653 Berlin  
351 University Alliance.

352

353 **References:**

354

- 355 Barkmann, T., Siebert, R., & Lange, A. (2017). Land-use experts' perception of regional climate  
356 change: An empirical analysis from the North German Plain. *Climatic Change*, 144(2), 287–  
357 301. <https://doi.org/10.1007/s10584-017-2041-x>
- 358 Benettin, P., Bailey, S. W., Campbell, J. L., Green, M. B., Rinaldo, A., Likens, G. E., McGuire, K.  
359 J., & Botter, G. (2015). Linking water age and solute dynamics in streamflow at the Hubbard  
360 Brook Experimental Forest, NH, USA. *Water Resources Research*, 51(11), 9256–9272.  
361 <https://doi.org/10.1002/2015WR017552>
- 362 Birkel, C., Tetzlaff, D., Dunn, S. M., & Soulsby, C. (2011). Using lumped conceptual rainfall–  
363 runoff models to simulate daily isotope variability with fractionation in a nested mesoscale  
364 catchment. *Advances in Water Resources*, 34(3), 383–394.  
365 <https://doi.org/10.1016/j.advwatres.2010.12.006>
- 366 Boulton, A. J., Hancock, P. J., Boulton, A. J., & Hancock, P. J. (2006). Rivers as groundwater-  
367 dependent ecosystems: A review of degrees of dependency, riverine processes and  
368 management implications. *Australian Journal of Botany*, 54(2), 133–144.  
369 <https://doi.org/10.1071/BT05074>
- 370 Burt, T. P. (1994). Long-term study of the natural environment—Perceptive science or mindless  
371 monitoring? *Progress in Physical Geography: Earth and Environment*, 18(4), 475–496.  
372 <https://doi.org/10.1177/030913339401800401>
- 373 Cosgrove, W. J., & Loucks, D. P. (2015). Water management: Current and future challenges  
374 and research directions. *Water Resources Research*, 51(6), 4823–4839.  
375 <https://doi.org/10.1002/2014WR016869>
- 376 Devito, K., Creed, I., Gan, T., Mendoza, C., Petrone, R., Silins, U., & Smerdon, B. (2005). A  
377 framework for broad-scale classification of hydrologic response units on the Boreal Plain: Is  
378 topography the last thing to consider? *Hydrological Processes*, 19(8), 1705–1714.  
379 <https://doi.org/10.1002/hyp.5881>
- 380 Drastig, K., Prochnow, A., Baumecker, M., Berg, W., & Brunsch, R. (2011). Agricultural Water  
381 Management in Brandenburg. *DIE ERDE – Journal of the Geographical Society of Berlin*,  
382 142(1–2), 119–140.



- 383 Dubbert, M., & Werner, C. (2019). Water fluxes mediated by vegetation: Emerging isotopic  
384 insights at the soil and atmosphere interfaces. *New Phytologist*, 221(4), 1754–1763.  
385 <https://doi.org/10.1111/nph.15547>
- 386 Evaristo, J., Jasechko, S., & McDonnell, J. J. (2015). Global separation of plant transpiration  
387 from groundwater and streamflow. *Nature*, 525(7567), 91–94.  
388 <https://doi.org/10.1038/nature14983>
- 389 Gelbrecht, J., Driescher, E., & Exner, H. J. (2000). Long-term investigations on nutrient input  
390 from catchment of the brook Demnitzer Mühlenfließ and restoration measures to reduce  
391 nonpoint pollution. In *Berichte des IGB* (Vol. 10, pp. 151–160). Institute Freshwater Ecology  
392 and Inland Fisheries. [http://www.mendeley.com/documents/?uuid=e2bdb786-9532-4313-  
393 9dfb-17bc6f2c8e76](http://www.mendeley.com/documents/?uuid=e2bdb786-9532-4313-9dfb-17bc6f2c8e76)
- 394 Gelbrecht, J., Lengsfeld, H., Pöthig, R., & Opitz, D. (2005). Temporal and spatial variation of  
395 phosphorus input, retention and loss in a small catchment of NE Germany. *Journal of  
396 Hydrology*, 304(1), 151–165. <https://doi.org/10.1016/j.jhydrol.2004.07.028>
- 397 Germer, S., Kaiser, K., Bens, O., & Hüttli, R. F. (2011). Water Balance Changes and Responses  
398 of Ecosystems and Society in the Berlin-Brandenburg Region – a Review. *DIE ERDE –  
399 Journal of the Geographical Society of Berlin*, 142(1–2), 65–95.
- 400 Grant, G. E., & Dietrich, W. E. (2017). The frontier beneath our feet. *Water Resources  
401 Research*, 53(4), 2605–2609. <https://doi.org/10.1002/2017WR020835>
- 402 Guswa, A. J., Tetzlaff, D., Selker, J. S., Carlyle-Moses, D. E., Boyer, E. W., Bruen, M., Cayuela,  
403 C., Creed, I. F., van de Giesen, N., Grasso, D., Hannah, D. M., Hudson, J. E., Hudson, S. A.,  
404 Iida, S., Jackson, R. B., Katul, G. G., Kumagai, T., Llorens, P., Lopes Ribeiro, F., ... Levia, D.  
405 F. (2020). Advancing ecohydrology in the 21st century: A convergence of opportunities.  
406 *Ecohydrology*, 13(4), e2208. <https://doi.org/10.1002/eco.2208>
- 407 Gutzler, C., Helming, K., Balla, D., Dannowski, R., Deumlich, D., Glemnitz, M., Knierim, A.,  
408 Mirschel, W., Nendel, C., Paul, C., Sieber, S., Stachow, U., Starick, A., Wieland, R., Wurbs,  
409 A., & Zander, P. (2015). Agricultural land use changes – a scenario-based sustainability  
410 impact assessment for Brandenburg, Germany. *Ecological Indicators*, 48, 505–517.  
411 <https://doi.org/10.1016/j.ecolind.2014.09.004>
- 412 Hewlett, J. D., Lull, H. W., & Reinhart, K. G. (1969). In Defense of Experimental Watersheds.  
413 *Water Resources Research*, 5(1), 306–316. <https://doi.org/10.1029/WR005i001p00306>
- 414 Huntingford, C., Marsh, T., Scaife, A. A., Kendon, E. J., Hannaford, J., Kay, A. L., Lockwood,  
415 M., Prudhomme, C., Reynard, N. S., Parry, S., Lowe, J. A., Screen, J. A., Ward, H. C.,  
416 Roberts, M., Stott, P. A., Bell, V. A., Bailey, M., Jenkins, A., Legg, T., ... Allen, M. R. (2014).  
417 Potential influences on the United Kingdom's floods of winter 2013/14. *Nature Climate  
418 Change*, 4(9), 769–777. <https://doi.org/10.1038/nclimate2314>
- 419 Kleine, L., Tetzlaff, D., Smith, A., Wang, H., & Soulsby, C. (2020). Using water stable isotopes  
420 to understand evaporation, moisture stress, and re-wetting in catchment forest and grassland  
421 soils of the summer drought of 2018. *Hydrology and Earth System Sciences*, 24(7), 3737–  
422 3752. <https://doi.org/10.5194/hess-24-3737-2020>
- 423 Kleine, L., Tetzlaff, D., Smith, A., Goldammer, T., & Soulsby, C. (2021a). Using isotopes to  
424 understand landscape-scale connectivity in a groundwater-dominated, lowland catchment  
425 under drought conditions. *Hydrological Processes*, 35(5), e14197.  
426 <https://doi.org/10.1002/hyp.14197>
- 427 Kleine L, Tetzlaff D, Smith A, Dubbert M, Soulsby C. (2021b) Modelling ecohydrological  
428 feedbacks in forest and grassland plots under a prolonged drought anomaly in central  
429 Europe 2018-2020. *Hydrological Processes*, <https://doi.org/10.1002/hyp.14325>
- 430 Kottek, M., Grieser, J., Beck, C., Rudolf, B., & Rubel, F. (2006). World Map of the Köppen-  
431 Geiger Climate Classification Updated. *Meteorologische Zeitschrift*, 15, 259–263.  
432 <https://doi.org/10.1127/0941-2948/2006/0130>



- 433 Martin, I. (2014). Zum Vorkommen von Elchen (*Alces alces*) in Brandenburg. *Nationalpark-*  
434 *Jahrbuch Unteres Odertal*, 11, 73–78.
- 435 Milly, P. C. D., Betancourt, J., Falkenmark, M., Hirsch, R. M., Kundzewicz, Z. W., Lettenmaier,  
436 D. P., Stouffer, R. J., Dettinger, M. D., & Krysanova, V. (2015). On Critiques of “Stationarity is  
437 Dead: Whither Water Management?” *Water Resources Research*, 51(9), 7785–7789.  
438 <https://doi.org/10.1002/2015WR017408>
- 439 Neill AJ, Birkel C, Maneta MP, Tetzlaff D, Soulsby C. (2021) Structural changes to forests  
440 during regeneration affect water flux partitioning, water ages and hydrological connectivity:  
441 Insights from tracer-aided ecohydrological modeling. *Hydrology and Earth System Sciences*  
442 (*HESS*). <https://hess.copernicus.org/articles/25/4861/2021/>
- 443 Nützmann, G., Wolter, C., Venohr, M., & Pusch, M. (2011). Historical Patterns of Anthropogenic  
444 Impacts on Freshwaters in the Berlin-Brandenburg Region. *DIE ERDE – Journal of the*  
445 *Geographical Society of Berlin*, 142(1–2), 41–64.
- 446 Okruszko, T., Duel, H., Acreman, M., Grygoruk, M., Flörke, M., & Schneider, C. (2011). Broad-  
447 scale ecosystem services of European wetlands—Overview of the current situation and  
448 future perspectives under different climate and water management scenarios. *Hydrological*  
449 *Sciences Journal*, 56(8), 1501–1517. <https://doi.org/10.1080/02626667.2011.631188>
- 450 Orth, R., & Destouni, G. (2018). Drought reduces blue-water fluxes more strongly than green-  
451 water fluxes in Europe. *Nature Communications*, 9(1), 3602. [https://doi.org/10.1038/s41467-](https://doi.org/10.1038/s41467-018-06013-7)  
452 [018-06013-7](https://doi.org/10.1038/s41467-018-06013-7)
- 453 Robinson, M., J. C. Rodda, and J. V. Sutcliffe (2013), Long-term environmental monitoring in  
454 the UK: Origins and achievements of the Plynlimon catchment study, *Trans. Inst. Br. Geogr.*,  
455 38, 451–463, doi:10.1111/j.1475-5661.2012.00534
- 456 Samaniego, L., Thober, S., Kumar, R., Wanders, N., Rakovec, O., Pan, M., Zink, M., Sheffield,  
457 J., Wood, E. F., & Marx, A. (2018). Anthropogenic warming exacerbates European soil  
458 moisture droughts. *Nature Climate Change*, 8(5), 421–426. [https://doi.org/10.1038/s41558-](https://doi.org/10.1038/s41558-018-0138-5)  
459 [018-0138-5](https://doi.org/10.1038/s41558-018-0138-5)
- 460 Smith, A., Tetzlaff, D., Gelbrecht, J., Kleine, L., & Soulsby, C. (2020). Riparian wetland  
461 rehabilitation and beaver re-colonization impacts on hydrological processes and water quality  
462 in a lowland agricultural catchment. *Science of The Total Environment*, 699, 134302.  
463 <https://doi.org/10.1016/j.scitotenv.2019.134302>
- 464 Smith, A., Tetzlaff, D., Kleine, L., Maneta, M. P., & Soulsby, C. (2020). Isotope-aided modelling  
465 of ecohydrologic fluxes and water ages under mixed land use in Central Europe: The 2018  
466 drought and its recovery. *Hydrological Processes*, 34(16), 3406–3425.  
467 <https://doi.org/10.1002/hyp.13838>
- 468 Smith, A., Tetzlaff, D., Kleine, L., Maneta, M., & Soulsby, C. (2021). Quantifying the effects of  
469 land use and model scale on water partitioning and water ages using tracer-aided  
470 ecohydrological models. *Hydrology and Earth System Sciences*, 25(4), 2239–2259.  
471 <https://doi.org/10.5194/hess-25-2239-2021>
- 472 Smith AA, Tetzlaff D, Maneta M, Soulsby C. (2022) Critical zone response times and water age  
473 relationships under variable catchment wetness states: insights using a tracer-aided  
474 ecohydrological model. *Water Resources Research*, <https://doi.org/10.1029/2021WR030584>
- 475 Tetzlaff, D., Buttle, J., Carey, S. K., van Huijgevoort, M. H. J., Laudon, H., McNamara, J. P.,  
476 Mitchell, C. P. J., Spence, C., Gabor, R. S., & Soulsby, C. (2015). A preliminary assessment  
477 of water partitioning and ecohydrological coupling in northern headwaters using stable  
478 isotopes and conceptual runoff models. *Hydrological Processes*, 29(25), 5153–5173.  
479 <https://doi.org/10.1002/hyp.10515>
- 480 Tetzlaff, D., Buttle, J., Carey, S. K., Kohn, M. J., Laudon, H., McNamara, J. P., Smith, A.,  
481 Sprenger, M., & Soulsby, C. (2021). Stable isotopes of water reveal differences in plant – soil  
482 water relationships across northern environments. *Hydrological Processes*, 35(1), e14023.  
483 <https://doi.org/10.1002/hyp.14023>



- 484 Tetzlaff D, Smith A, Kleine L, Dubbert D, Freymueller J, Soulsby C. (2022) An integrated  
485 ecohydrological hydrometric and stable water isotope data set of a drought-sensitive mixed  
486 land use lowland catchment. <https://fred.igb-berlin.de/data/package/622>.  
487 Vogel, C. (2014). Der Wolf in Brandenburg – Leben mit einem Rückkehrer. *Nationalpark-*  
488 *Jahrbuch Unteres Odertal*, 11, 54–58.