

1 1 km-resolution maps reveal increases in above- and belowground  
2 forest biomass carbon pools in China over the past 20 years

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## 18 **Abstract**

19 To quantify the ecological consequences of recent nationwide restoration efforts in China,  
20 spatially-explicit information on forest biomass carbon stock changes over the past 20 years is  
21 critical. However, long-term biomass tracking at the national scale remains challenging as it  
22 requires continuous and high-resolution monitoring. Here, we characterize the changes in  
23 forests' above- and belowground biomass carbon (AGBC and BGBC) in China between 2002  
24 and 2021 at 1 km spatial resolution by integrating multiple types of remote sensing observations  
25 with intensive field measurements through regression and machine learning approaches. On  
26 average,  $8.6 \pm 0.6$  and  $2.2 \pm 0.1$  PgC were stored in above- and belowground live forests in  
27 China. Over the last 20 years, the total forest biomass carbon pool in China has increased at a  
28 rate of  $114.5 \pm 16.3$  TgC/yr (approximately 1.1%/yr). The most pronounced forest biomass  
29 carbon stock gains occurred in central to southern China, including the southern Loess Plateau,  
30 Qinling Mountains, southwest karsts and southeast forests. While the combined use of multi-  
31 source remote sensing data provides a powerful tool to assess the forest biomass carbon changes,  
32 future research is also needed to explore the drivers of the observed woody biomass trends, and  
33 to evaluate the degree to which biomass gains will translate into biodiverse, healthy ecosystems  
34 and thus are sustainable.

35 **Key words:** Aboveground biomass carbon pool; Belowground biomass carbon; Long-term  
36 continuous mapping; China

## 37 **1 Introduction**

38 Forest biomass carbon stock contributes to over 90% of the global vegetation biomass carbon  
39 pool (Ma et al., 2021). As a net outcome of carbon gains from photosynthesis and carbon losses  
40 from respiration, mortality and disturbances, forest biomass carbon stock (approximately 50%  
41 of biomass) is a critical indicator of ecosystem function and ecosystem services, such as carbon  
42 sequestration, wood production and resource allocation (Kumar and Mutanga, 2017). Accurate  
43 forest biomass carbon stock monitoring over space and time is thus essential for assessing  
44 ecosystem management strategies and mitigation policies (Kumar and Mutanga, 2017).

45 In recent decades, remote sensing tools have been integral in our efforts to map aboveground  
46 biomass (AGB) or carbon stock (AGBC). By combining satellite imagery (e.g., MODIS) and  
47 airborne LiDAR signals, forest cover and canopy height can be mapped across large spatial  
48 scales (Hu et al., 2016; Saatchi et al., 2011; Su et al., 2016; Tong et al., 2020; Xu et al., 2021).

49 Apart from optical images and LiDAR signals, microwaves can provide more detailed insights  
50 into subcanopy forest structure and AGBC due to their ability to penetrate the canopy. Active  
51 microwave techniques, i.e., Synthetic Aperture Radar (SAR) backscatters, facilitate high-  
52 resolution (e.g., 100 m) AGB mapping, but the temporal coverage is limited (Cartus et al., 2012;  
53 Bouvet et al., 2018). Conversely, vegetation optical depth (VOD) retrieved from multiple  
54 passive microwave sensors can be used to produce long-term continuous AGB maps (Frappart  
55 et al., 2020; Liu et al., 2011; Liu et al., 2015), yet at a coarse spatial resolution (e.g., 0.25°).  
56 Because different remote sensing techniques have their advantages and pitfalls, combining

57 these techniques and complementing them with direct ground measurements is integral to  
58 maximizing the accuracy and precision of biomass carbon estimations across space and time.

59 Another source of uncertainty in vegetation biomass carbon stocks is the extent of biomass that  
60 is stored belowground as roots. While AGBC mapping is facilitated by a suite of emerging  
61 remote sensing techniques, investigating the spatiotemporal variation in belowground biomass  
62 carbon pool (BGBC) remains challenging despite the large contribution of roots to total carbon  
63 storage (Huang et al., 2021; Ma et al., 2021). To map BGBC, the commonly-used approach is  
64 to combine aboveground biomass information with vegetation type-specific ratios of BGB to  
65 AGB (i.e., root-shoot ratio, or RSR) (Xu et al., 2021; Saatchi et al., 2011). Because field studies  
66 indicate a near-linear relationship between log-transformed BGB and AGB (Enquist Brian and  
67 Niklas Karl, 2002), BGB variations at large scales have often been approximated using this  
68 relationship (Spawn et al., 2020). To capture the complex relationship between BGB and biotic  
69 or abiotic variables (e.g., stand age, heat and water availability), machine learning algorithms  
70 have been applied to map BGB (Huang et al., 2021) and root-mass fractions (Ma et al., 2021)  
71 globally. However, the reference plots were unevenly distributed across the world, limited in  
72 developing countries, leading to some uncertainties in BGB and BGBC estimation within those  
73 regions (Huang et al., 2021).

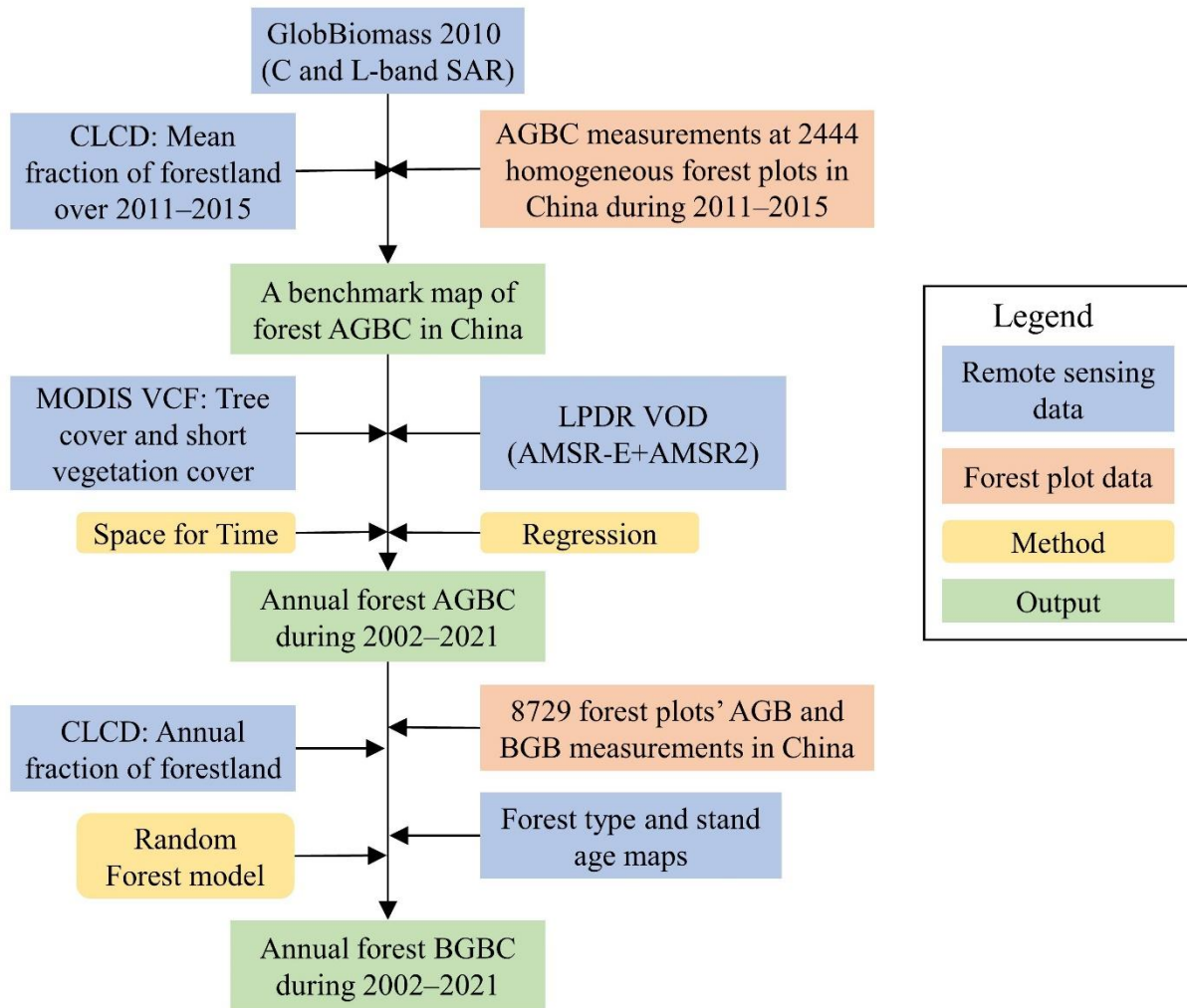
74 China has been implementing national-scale afforestation and reforestation programs since the  
75 late 1990s (Lu et al., 2018), promoting vegetation cover and carbon storage in the Loess Plateau  
76 and the southwest karst regions, etc. (Chen et al., 2019a; Niu et al., 2019; Tong et al., 2018). A

77 spatial understanding of forest biomass trends can help evaluate the efficiency of ecological  
78 restoration programs. High quality, high resolution and long-term continuous woody biomass  
79 monitoring in China has remained challenging (Zhang et al., 2019; Huang et al., 2019).

80 In this study, by integrating multi-source remote sensing data with large quantities of plot  
81 measurements, we produced 1 km resolution above- and belowground forest biomass carbon  
82 pool maps for China during the past 20 years (2002–2021). This dataset could provide new  
83 insights into forest carbon stock changes in China over the past two decades.

## 84 **2 Materials and methods**

85 To map above- and belowground forest biomass carbon stock in China during 2002–2021, we  
86 1) calibrated a SAR-based high-resolution forest aboveground biomass map in China based on  
87 massive field measurements of AGBC during 2011–2015; 2) extended the AGBC time series to  
88 2002–2021 by referring to the tree and short vegetation cover retrieved from optical remote  
89 sensing; 3) calibrated the AGBC time series in some specific areas using a long-term integrated  
90 microwave-based VOD dataset; and 4) mapped forestlands' BGBC through a random forest  
91 model developed based on the in-situ records in published literature. The basic procedure is  
92 shown in Figure 1 and described below.



93

94 **Figure 1.** Workflow of forest biomass carbon pool monitoring in China during 2002–2021.  
 95 AGBC, BGBC: aboveground and belowground biomass carbon; VCF: vegetation continuous  
 96 fields; LPDR VOD: global land parameter data record- vegetation optical depth; CLCD: China  
 97 Land Cover Dataset

98 **2.1 A benchmark map of forest aboveground biomass carbon (AGBC) in China**

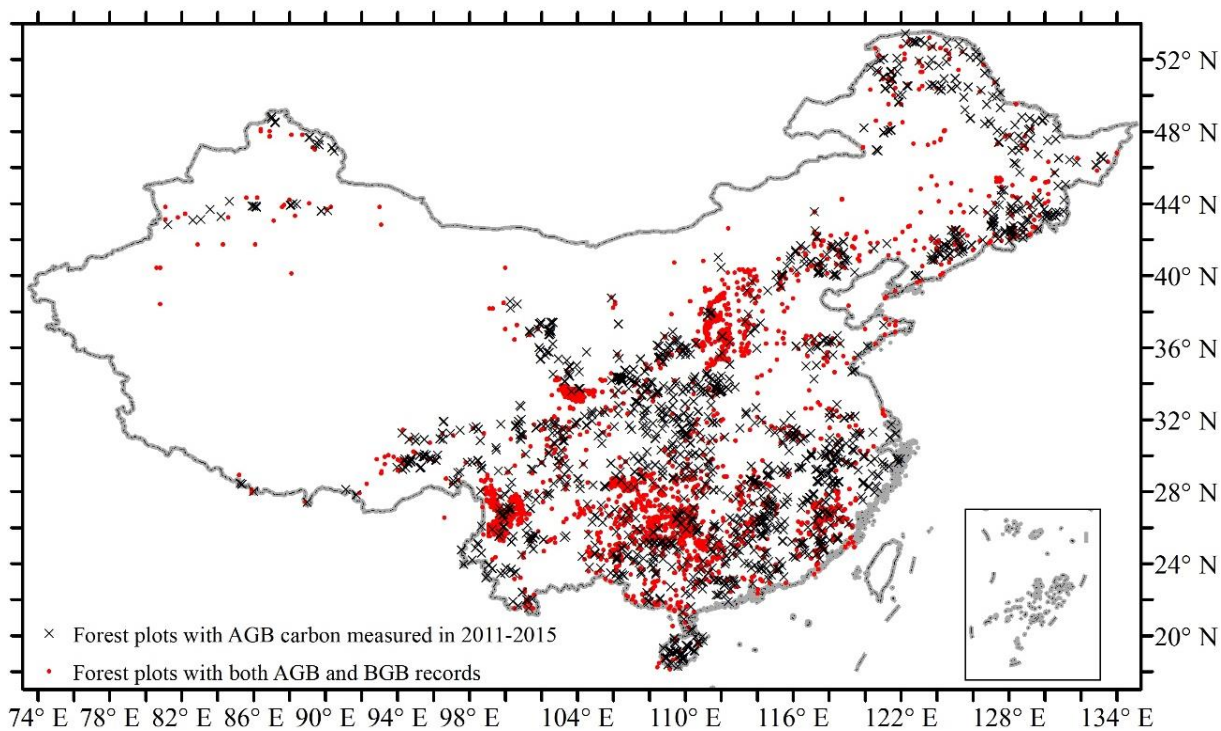
99 By combining multiple satellite observations of SAR backscatter, including the L-band ALOS  
 100 PALSAR and C-band Envisat ASAR around the year 2010, the first global high-resolution (100  
 101 m) forest AGB dataset, GlobBiomass 2010, was published through the European Space Agency  
 102 (ESA)’s Data User Element project (Santoro et al., 2021), whose relative root mean square error

103 (RMSE) was below 30% (Mialon et al., 2020). Apart from GlobBiomass 2010, another high-  
104 resolution (30 m) forest AGB for China was produced by relating the ICESat GLAS (LiDAR)-  
105 derived footprint AGB to various variables derived from Landsat optical images (Huang et al.,  
106 2019). Because the ICESat data in 2006 were applied as the training target of the random forest  
107 model, Huang's dataset refers to the AGB status in 2006. According to a recent validation study,  
108 GlobBiomass and Huang's AGB performed the best among all existing AGB datasets in China  
109 (Chang et al., 2021). Mean forest canopy heights and tree coverage are also good indicators of  
110 the spatial pattern of forest biomass. The high-resolution (30 m) forest canopy height map for  
111 China was developed by interpolating the ICESat-2 and GEDI data in 2019 through a neural  
112 network (Liu et al., 2022), while the tree cover map at the same resolution was derived from  
113 cloud-free growing season composite Landsat 7 data in around 2010 (Hansen et al., 2013). We  
114 resampled GlobBiomass from 100 m resolution ( $1/1125^\circ$ ) to  $1/1200^\circ$  (approximately 90 m),  
115 and averaged Huang's AGB map, canopy height map and tree cover map to the same resolution.

116 A reviewable, consistent ecosystem carbon stock inventory was conducted in China between  
117 2011 and 2015 (Tang et al., 2018). We requested the AGB carbon stock (AGBC) data at more  
118 than 5,000  $30\times 30$  m sized forest plots from the authors. Due to the scale mismatch between the  
119 maps of biomass, canopy height or tree cover and the field measurements, we dropped out the  
120 data within the  $1/1200^\circ$  resolution grids in which the standard deviation of tree cover was  
121 greater than 15%, according to (Chang et al., 2021), leaving 2444 homogeneous forest plots  
122 remaining (see Figure 2 for the spatial distribution of these forest plots and Figure S1a~b for

123 the cumulative frequency curve and histogram of the AGBC records). The AGBC records in  
124 these forest plots were further multiplied by the mean fraction of forestland over 2011–2015 in  
125 the corresponding grid, which was computed from the annual 30 m resolution China Land  
126 Cover Dataset (CLCD) (Yang and Huang, 2021). By comparison, GlobBiomass 2010 AGB  
127 matches the best with the grid-scale forest AGBC derived from plot measurements, with a  
128 correlation coefficient (CC) of 0.50, followed by tree cover (CC=0.42), the product of canopy  
129 height and tree cover (CC=0.38), and finally the canopy height (0.27) and Huang’s AGB (0.25).  
130 Therefore, to obtain an improved benchmark map of forest AGBC in China for the period of  
131 2011–2015, we chose the GlobBiomass 2010 dataset as our basis, and calibrated it against the  
132 in-situ observation-based grid-scale forest AGBC. To build an equation for the calibration, we  
133 divided the grid-scale AGBC values into 16 equidistant subranges (0~15, 15~30, ..., 225~240  
134 tC/ha), calculated the median of grid-scale AGBC values that are within each subrange, and  
135 then the median of GlobBiomass AGB values in the corresponding grids. According to previous  
136 studies, an exponential function would be suitable for calibrating the GlobBiomass map in a  
137 region such as China (Mialon et al., 2020). After the calibration, we averaged the benchmark  
138 AGBC map from 1/1200° to 1/120° (approximately 1 km) to further reduce the uncertainties.





139 74° E 78° E 82° E 86° E 90° E 94° E 98° E 104° E 110° E 116° E 122° E 128° E 134° E

140 **Figure 2.** The spatial distribution of 1) 2444 homogeneous forest plots with aboveground  
 141 ground biomass carbon stock measured between 2011 and 2015; and 2) 8182 forest plots with  
 142 both above- and belowground biomass records collated in this study.

143 **2.2 Temporally continuous forest AGBC mapping during 2002–2021**

144 Because the benchmark AGBC was mapped based on SAR data, the spatial pattern accuracy is  
 145 guaranteed, but the temporal coverage is limited to just a few years. Hence, to create a forest  
 146 AGBC time series over the past 20 years, we integrated the benchmark AGBC with long-term  
 147 continuous optical and passive microwave remote sensing data.

148 The spatial resolution of optical remote sensing is higher, and is thus preferred in this study. By  
 149 adopting the MODIS vegetation continuous fields (VCF) data (MOD44B v061) which includes  
 150 three ground cover components: percent tree cover, percent non-tree vegetation (i.e., short

151 vegetation) cover, and percent non-vegetated (Dimiceli et al., 2022), we first calculated the  
152 mean tree cover (hereinafter,  $TC_{\text{mean}}$ ) and short vegetation cover (hereinafter  $SVC_{\text{mean}}$ ) during  
153 2011–2015, and resampled them from 250 m to  $1/120^\circ$ , the same resolution as the benchmark  
154 AGBC map for 2011–2015. Because the canopy heights of trees are usually similar within a  
155 small area, the regional AGBC per  $TC_{\text{mean}}$  can be assumed as the same, which is referred to as  
156 the ‘homogeneous assumption’ hereinafter. Accordingly, for each grid, we searched the  $TC_{\text{mean}}$ ,  
157  $SVC_{\text{mean}}$  and AGBC within a  $3\times 3$  window ( $1/40^\circ\times 1/40^\circ$ ), and then regressed the AGBC values  
158 in 9 grids against both  $TC_{\text{mean}}$  (the primary, or key predictor of AGBC) and  $SVC_{\text{mean}}$  (assumed  
159 as a supplementary predictor) linearly. Specifically, when the regression coefficient of  $SVC_{\text{mean}}$   
160 was negative or the fitting efficiency was low ( $R^2<0.5$ ; significance  $p\text{-value}>0.05$ ), we excluded  
161 the supplementary predictor from the regression, only exploring the linear relationship between  
162  $TC_{\text{mean}}$  and AGBC. Afterwards, if the regression between  $TC_{\text{mean}}$  and AGBC was still invalid,  
163 we enlarged the searching window size to  $5\times 5$ , then  $7\times 7$ , and finally  $9\times 9$ , until the regression  
164 as well as the coefficients became valid. Then, the grid annual AGBC from 2002 to 2021 can  
165 be estimated from the TC or both TC and SVC in each year, following the regression results. If  
166 the regression failed even if the window size reached  $9\times 9$ , we stopped expanding the searching  
167 window to avoid the ‘homogeneous assumption’ being invalid. In those grids, following a  
168 previous study (Xu et al., 2021), we divided the estimated AGBC by the  $TC_{\text{mean}}$  during 2011–  
169 2015 and then multiplied the TC in each year to obtain the AGBC time series. The above  
170 method utilized spatial information to estimate the temporal variation, and can thus be referred  
171 to as the ‘space for time’ method.

172 Long-term continuous microwave VOD can also reflect forest biomass changes, although the  
173 relationship was nonlinear (Jackson and Schmugge, 1991; O'Neill et al., 2021; Liu et al., 2015;  
174 Wigneron et al., 1995). We selected the global land parameter data record (LPDR) v3 0.25°  
175 resolution VOD product, which was generated using similar calibrated, X-band brightness  
176 temperature retrieved from the Advanced Microwave Scanning Radiometer (AMSR-E) and the  
177 Advanced Microwave Scanning Radiometer 2 (AMSR2) (Du et al., 2017). As revealed by a  
178 recent evaluation study, LPDR VOD is better correlated with AGB than other long-term VOD  
179 products, especially in less-vegetated areas (Li et al., 2021). Because X-band VODs are still  
180 more sensitive to canopy cover than stem biomass and there is a data gap between October 2010  
181 and June 2011, while the plot investigations were all conducted in summers (Tang et al., 2018),  
182 we averaged the VOD data from mid-July (the 206<sup>th</sup> day) until the end of September (the 274<sup>th</sup>  
183 day) in each year to represent the annual AGB status. We also aggregated the benchmark AGBC  
184 map as well as the VCF data ( $TC_{\text{mean}}$  and  $SVC_{\text{mean}}$ ) to 0.25° resolution. After each round of  
185 searching, we applied the shape language modelling algorithm (D'errico, 2022) to fit the  
186 nonlinear but monotonous relationship between AGBC and VOD values within the searching  
187 window, and then fitted the bivariate linear regression between AGBC and VCF. If the nonlinear  
188 regression between AGBC and VOD is valid and the  $R^2$  is superior to the regression between  
189 AGBC and VCF data, LPDR VOD data is expected to outperform VCF in predicting the inter-  
190 annual AGBC changes in the corresponding 0.25° grid. Therefore, in these areas, we calibrated  
191 the VCF-derived high (1/120°) resolution annual AGBC by incorporating the ratio between the  
192 VOD-derived 0.25° AGBC and the aggregated VCF-derived AGBC in that year.

### 193 **2.3 Forest belowground biomass carbon (BGBC) mapping during 2002–2021**

194 This study mapped belowground forest biomass carbon (BGBC) following the random forest  
195 (RF) model approach (Huang et al., 2021). To reveal forests' above- and belowground biomass  
196 allocation rules in China, this study collated both AGB and BGB records at 8729 forest plots  
197 throughout China, which were obtained using allometric equations or clear-cutting methods  
198 from published papers, including (Luo, 1996), (Luo et al., 2014), (Guo and Ren, 2014), (Wang  
199 et al., 2014). Because forest stand age and tree species (forest type) information are also  
200 available at 8182 plots, while the climatic backgrounds are available from the WorldClim v2.1  
201 dataset (Fick and Hijmans, 2017), forest plots' AGB, forest type (hereinafter FOR\_T), stand  
202 age, mean annual temperature (MAT), temperature seasonality (standard deviation of monthly  
203 temperature $\times 100$ , abbreviated as Tsea), mean annual precipitation (MAP) and precipitation  
204 seasonality (coefficient of variation of monthly precipitation, Psea) were applied as predictors  
205 of forest plots' BGB. For simplicity, we distinguished all forests into 5 types: evergreen  
206 broadleaf forest (EBF), deciduous broadleaf forest (DBF), evergreen needleleaf forest (ENF),  
207 deciduous needleleaf forest (DNF), and mixed forest (MF). Using the data records at these 8182  
208 plots (see Figure 2 for the locations of these forest plots and Figure S1c~f for the cumulative  
209 frequency curves and histograms of the AGB and BGB data), we trained ten-fold RF models  
210 using MATLAB R2021a<sup>®</sup>. The number of regression trees was set to 500.

211 Because the 1/120° resolution grids where forest AGBC data were available are often mixed  
212 with forestland and some other land cover types, e.g., water bodies, bare ground, croplands, we

213 converted the annual grid-average AGBC into the AGBC per area forestland by incorporating  
214 the annual fraction of forestland computed from the CLCD at 30 m resolution. Considering the  
215 potential uncertainties in the forestland fraction as well as the inclusion of shrub or herbaceous  
216 plant AGB in the SAR-derived AGB, we only calculated the annual AGBC per area forestland  
217 in grids that were dominated by forestland (forestland fractions were consistently over 50%).  
218 In these forestland grids, we simulated the forest BGBC per area forestland during 2002–2021  
219 by inputting the estimated annual AGB (approximately 2 times of the AGBC) per forestland,  
220 annual forest type map derived from ESA CCI’s land cover classification dataset (Li et al.,  
221 2018), forest stand age (Besnard et al., 2021) and climatic background variables into the RF  
222 model. Afterwards, we multiplied the simulation results in every forestland grid with the annual  
223 forestland fractions to obtain the forests’ BGB and BGBC ( $0.5 \times \text{BGB}$ ) time series. Finally, for  
224 grids with forests but are not dominated by forestlands, we sequentially searched for at least  
225 five valid RSR values (the ratio of forests’ BGBC to AGBC) nearby (Chen et al., 2019b), and  
226 then multiplied the annual forest AGBC in the grid with the median of nearby RSR values in  
227 each year to estimate the annual forest BGBC.

## 228 **2.4 Evaluation and assessment**

229 We compared the inter-annual trend of forest biomass carbon calculated in this study against  
230 that of existing global/regional long-term woody biomass datasets, including the well-received  
231 global long-term terrestrial biomass data between 1993–2012, which was developed mainly  
232 based on a long-term integrated VOD dataset (Liu et al., 2015), as well as an updated woody

233 biomass dataset covering 2001–2019 whose long time series was derived from optical remote  
234 sensing data (i.e., MODIS VCF dataset) (Xu et al., 2021).

235 To justify the random forest models for BGBC predictions, we drew partial dependence plots  
236 (PDPs) in MATLAB R2021a<sup>®</sup> to show the marginal effect that one predictor has on the training  
237 target (e.g., BGB at forest plots) (Hastie et al., 2009). Here, for each predictor, we excluded the  
238 extreme values (the lowest 1% and the highest 1%) before calculating the corresponding PDP  
239 to avoid roughly extending the PDP lines to data-scarce areas. Ten-fold RF trainings were also  
240 performed to derive the mean PDP values as well as the standard deviations.

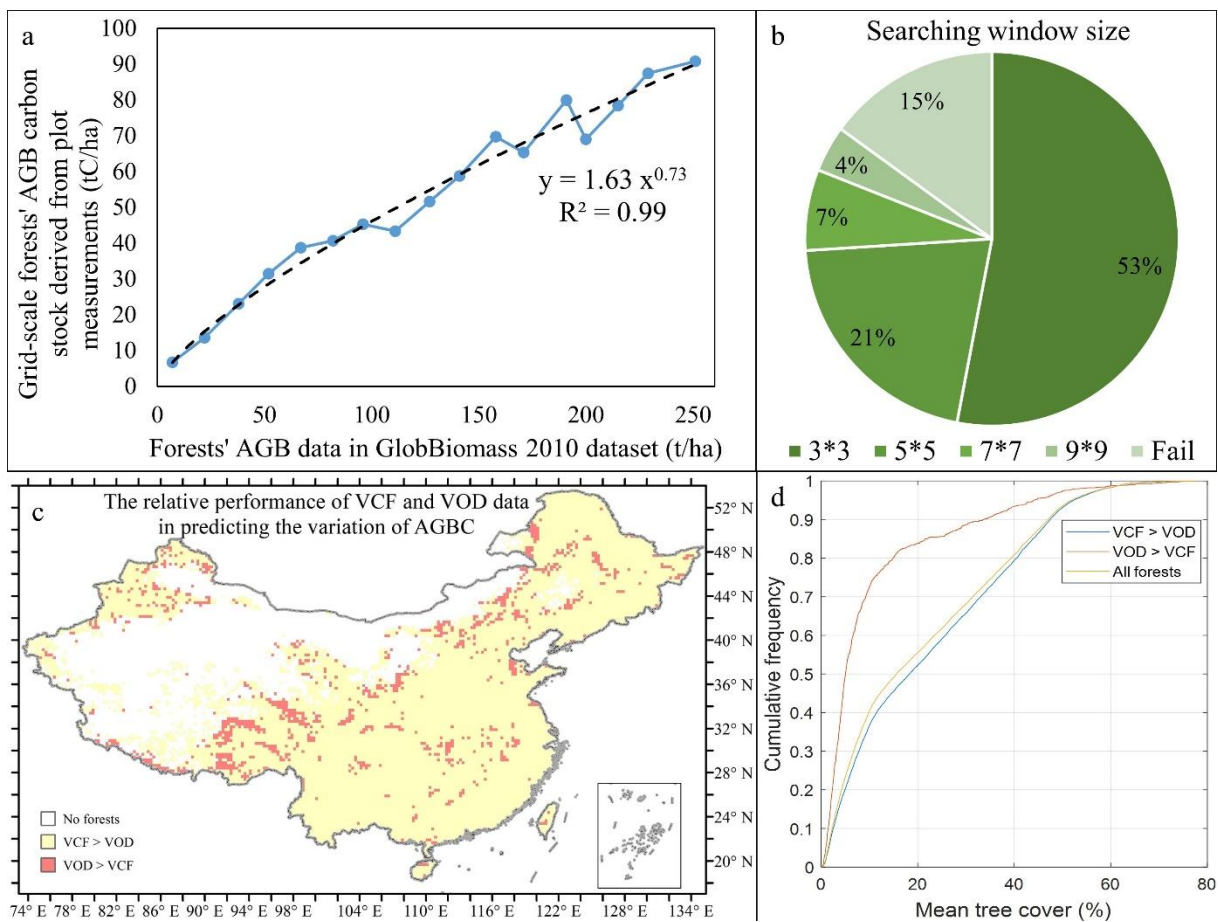
## 241 **3 Results and discussion**

### 242 **3.1 Evaluation of forests' AGBC and BGBC estimation**

243 First, according to Figure 3a, an exponential function:  $y=1.63 \times x^{0.73}$  can fit the relationship  
244 between the actual grid-scale forest AGBC over 2011–2015 (y) and the AGB values predicted  
245 by GlobBiomass 2010 (x). Hence, this function was applied to derive the benchmark map of  
246 forest AGBC across China.

247 Second, when using the spatial information of tree cover and short vegetation cover to estimate  
248 the temporal variation of AGBC in each grid, the spatial searching window was at its minimum  
249 of 3×3 in most (53%) grids with forests. Across China, the temporal extension of AGBC in only  
250 15% of all grids with forest cannot be achieved even when the searching window was enlarged  
251 to 9×9 (Figure 3b).

252 Next, as shown in Figure 3c and 3d, the grids where LPDR X-band VOD performed better than  
 253 MODIS VCF in predicting the temporal change in forest AGBC are usually located in regions  
 254 with low tree cover. These grids account for just 10.4% of all grids with forests, and may suffer  
 255 from high uncertainty within the optical-based variation in tree cover. Therefore, microwave-  
 256 based VOD is supposed to be more suitable for estimating the forests' AGBC changes in these  
 257 regions.



258  
 259 Figure 3. Evaluation of the forest AGBC and BGBC mapping in this study. (a) The regression  
 260 relationship between the grid-scale forest AGB carbon stock derived from plot measurements  
 261 during 2011–2015 and the GlobBiomass AGB dataset for 2010; (b) the minimum searching  
 262 window sizes of every 1/120° grid when the spatial variation in MODIS VCF was applied as

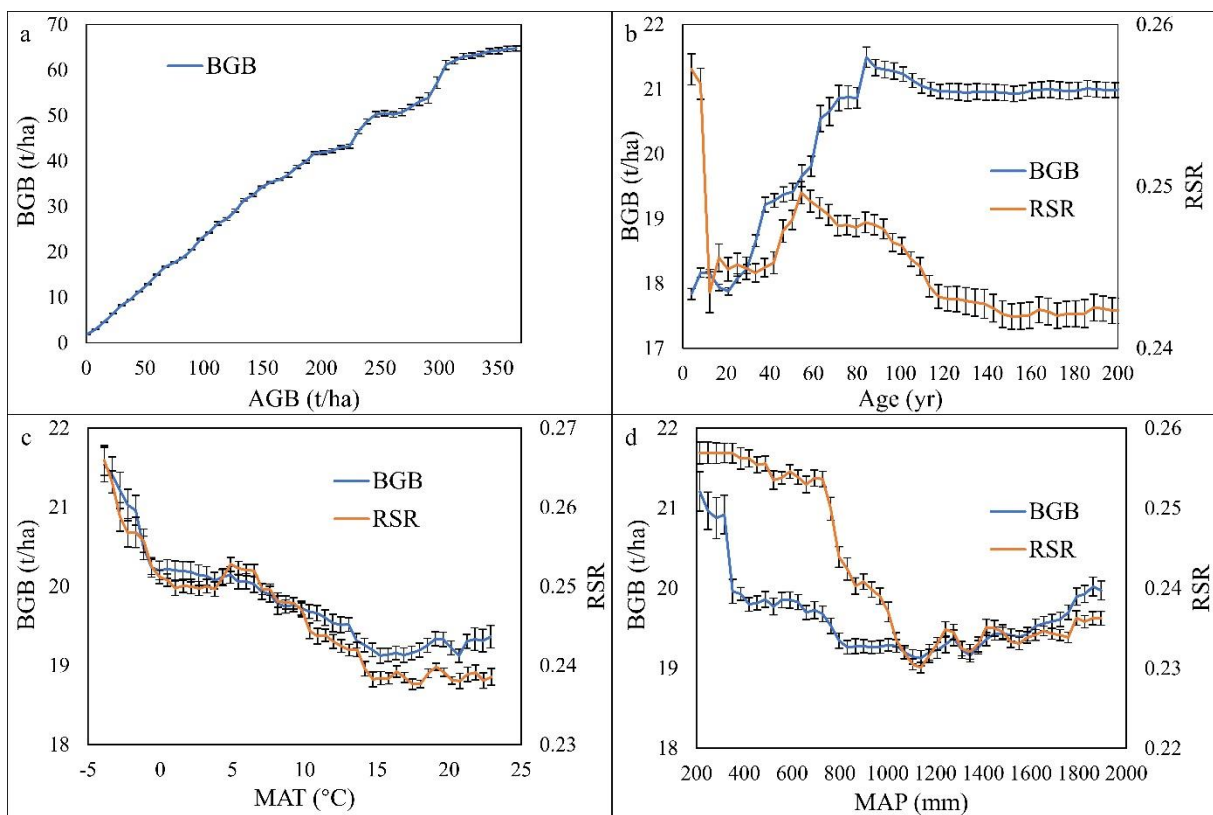
263 the predictor of AGBC changes; (c) the spatial pattern of the relative performances of MODIS  
264 VCF and LPDR VOD data in predicting the variation in AGBC; (d) comparison of the mean  
265 tree cover between the grids where VOD data were more suitable for predicting the variation  
266 of AGBC and the grids where VCF data were the better predictor.

267 The RF model designed for forest plot BGB estimation (see section 2.3) achieved a predictive  
268  $R^2$  of  $0.89 \pm 0.02$ , while the RMSE was  $6.3 \pm 0.5$  t/ha. AGB explained 53% of the variation in  
269 BGB among different plots. Long-term climate backgrounds, i.e., mean annual temperature,  
270 temperature seasonality, annual precipitation and precipitation seasonality accounted for 8%,  
271 6%, 8% and 7%, respectively. Forest type and stand age also contributed 12% and 8% to the  
272 training efficiency, indicating that the effects of these factors are nonnegligible. The selection  
273 of predictors of BGB basically followed the existing knowledge (Huang et al., 2021), and the  
274 seasonality of temperature and precipitation made sense in the prediction (see Text S1). On the  
275 other hand, although previous studies incorporated many edaphic factors as predictors of BGB  
276 (Huang et al., 2021), by comparing the training efficiencies when whether these edaphic factors  
277 are incorporated or not, we could justify the reasonability of our simplified set of predictors  
278 (Text S1).

279 According to the collected woody plots' data, AGB is a key driver of BGB (Figure 4). Yet, RSR  
280 changes among different forest growth stages, decreasing in general as reported (Mokany et al.,  
281 2006). The overall negative impact of mean temperature on BGB or RSR agrees with the  
282 mechanism that higher heat promotes nutrient accessibility (Luo et al., 2012; Ma et al., 2021),



283 and increases the turnover rates of roots at a higher magnitude than stems (Reich et al., 2014).  
 284 The ‘U-shaped’ relationship between precipitation and belowground biomass allocation follows  
 285 the ‘optimal biomass allocation’ theory, because arid climates promote root extension, yet too  
 286 heavy rainfall reduces nutrient availability through leaching and dilution effects (Luo et al.,  
 287 2012). Other factors, including temperature seasonality, precipitation seasonality and forest  
 288 type, have supplementary effects on biomass allocation (Figure S2).



289 **Figure 4.** Influence of key factors on forest belowground biomass (BGB) and root-shoot ratio  
 290 (RSR) in China. Subfigures (a~d) show partial influences of (a) AGB; (b) stand age; (c) MAT  
 291 and (d) MAP on BGB and RSR values of all forest plots. The error bars represent the standard  
 292 deviations of the ten-fold trainings. We did not draw the PDP for the impact of AGB on RSR,  
 293 since the dividend of RSR calculation is AGB.  
 294

### 295 3.2 Forest biomass carbon pool, allocation and change in China

296 Between 2002 to 2021, the forest above- and belowground biomass carbon (AGBC and BGBC)  
297 pools in China were  $8.6 \pm 0.6$  and  $2.2 \pm 0.1$  PgC, respectively (Table 1). The mean RSR for all  
298 forests was 0.25, basically equal to the global average (Huang et al., 2021). Separated by forest  
299 type, evergreen conifer forests (ENF) occupy the highest biomass carbon pool per unit area,  
300 mainly because ENF are mainly located in southwestern China and are more mature and natural  
301 (Yu et al., 2020; Zhang et al., 2017). Deciduous forests (DBF & DNF) in northern China (see  
302 Figure S3 for the distribution of different forest ecosystems) harbor less biomass carbon but  
303 higher BGBC (Figure 5a), which can be attributed to the higher RSR values (Table 1).

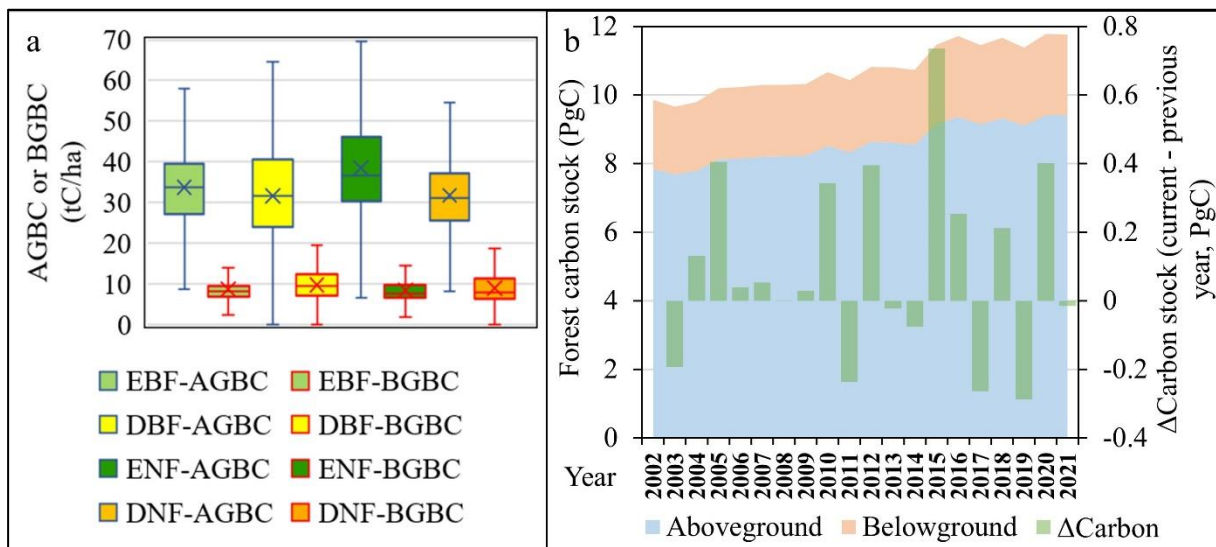
304 The forest biomass carbon stock in China increased at an average rate of  $114.5 \pm 16.3$  TgC/yr  
305 ( $p < 0.01$ ) during 2002–2021, and the annual biomass carbon gains were the greatest from 2014  
306 to 2015, reaching 736 TgC (Figure 5b). Changes in AGB and BGB accounted for 81.9% and  
307 18.1%, respectively, of the forest carbon stock gains over the past 20 years.

308 Our estimates of the forest biomass carbon pool, forest RSR and the recent inter-annual trend  
309 of forest biomass carbon are generally consistent with previous estimates based on massive  
310 field investigations (Table 1).

311 **Table 1.** Agreement of the estimated various forest RSR and the trend of forest biomass carbon  
312 in China with existing studies.

Variables	Our estimate	Previous estimates	Reference
Forests' AGBC	$8.6 \pm 0.6$ (2002–2021) $8.7 \pm 0.3$ (2011–2015)	$8.4 \pm 1.6$ (2011–2015)	(Tang et al., 2018)
Forests' BGBC	$2.2 \pm 0.1$ (2002–2021)	$2.1 \pm 0.4$ (2011–2015)	

	2.2 ± 0.1 (2011–2015)		
EBF's RSR	0.27±0.07	0.22±0.11	
DBF's RSR	0.31±0.05	0.28±0.15	(Tang et al., 2018)
ENF's RSR	0.22±0.04	0.24±0.11	
DNF's RSR	0.29±0.10	0.31±0.13	
Annual forest carbon stock increase	114.5 ± 16.3 TgC/yr (2002–2021)	116.7 TgC/yr (2000–2010)	(Fang et al., 2018)
	105.1 ± 42.2 TgC/yr (2002–2010)		



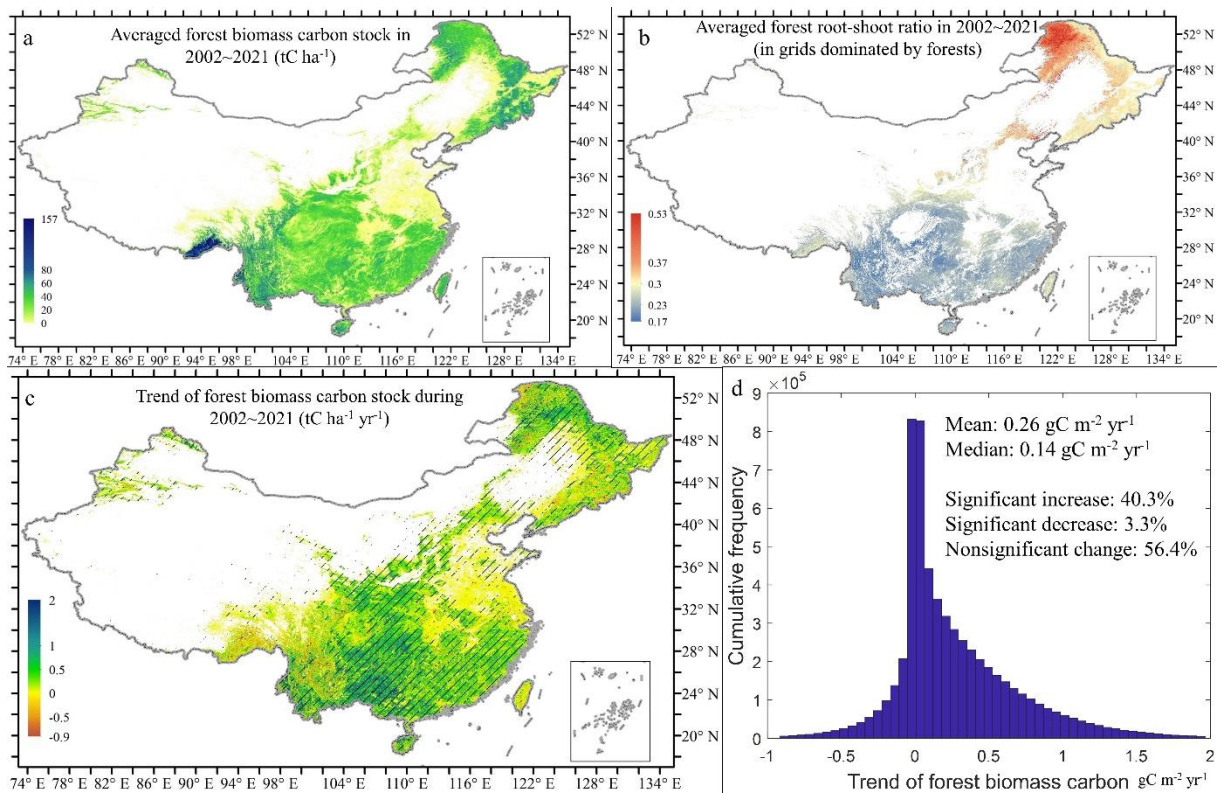
313

314 **Figure 5.** Forest biomass allocation and biomass change in China during 2002–2021: (a)  
 315 aboveground biomass carbon (AGBC) and belowground biomass carbon (BGBC) density of  
 316 different forest ecosystems in China; (b) the inter-annual changes of forest AGBC and BGBC  
 317 in China. Total forest biomass carbon stock changes from the previous to the current year are  
 318 represented by green columns.

### 319 3.3 Spatial pattern of the forest biomass carbon stock trend in China

320 The highest forest biomass carbon pools during 2002–2021 were observed in northeastern and  
 321 southwestern China, especially southern Tibet. Forest biomass carbon stocks were also high in  
 322 the natural or semi-natural forests in the Qinling Mountains, Hengduan Mountains, Hainan and  
 323 Taiwan (Figure 6a). Above- and belowground forest biomass allocation varies significantly

324 among regions. RSR is highest in northeastern deciduous conifer forests and northern China's  
 325 deciduous broadleaf forests but low in southern China (Figure 6b). The strongest forest biomass  
 326 carbon increases were found in central to southern China, including the Loess Plateau, Qinling  
 327 Mountains, southwest karst region and southeastern forests. Slight declines in forest biomass  
 328 carbon only occurred in some mature and natural forests, e.g., those in the Greater Khingan  
 329 Mountain, Hengduan Mountains and South Tibet (Figure 6c). A total of 40.3% of all forests in  
 330 China showed significant biomass carbon stock gains over the past 20 years, whereas only 3.3%  
 331 of forests experienced significant biomass carbon losses (Figure 6d).



332  
 333 **Figure 6.** Maps of forest biomass carbon pool, allocation and trend in China during 2002–2021.  
 334 (a) Spatial pattern of the forest biomass carbon pool in China; (b) all forestland pixels' RSR; (c)  
 335 map of the forest biomass carbon stock trend from 2002 to 2021, with shaded areas representing  
 336 statistically significant trends at the 95% confidence level; (d) histogram and basic statistics of

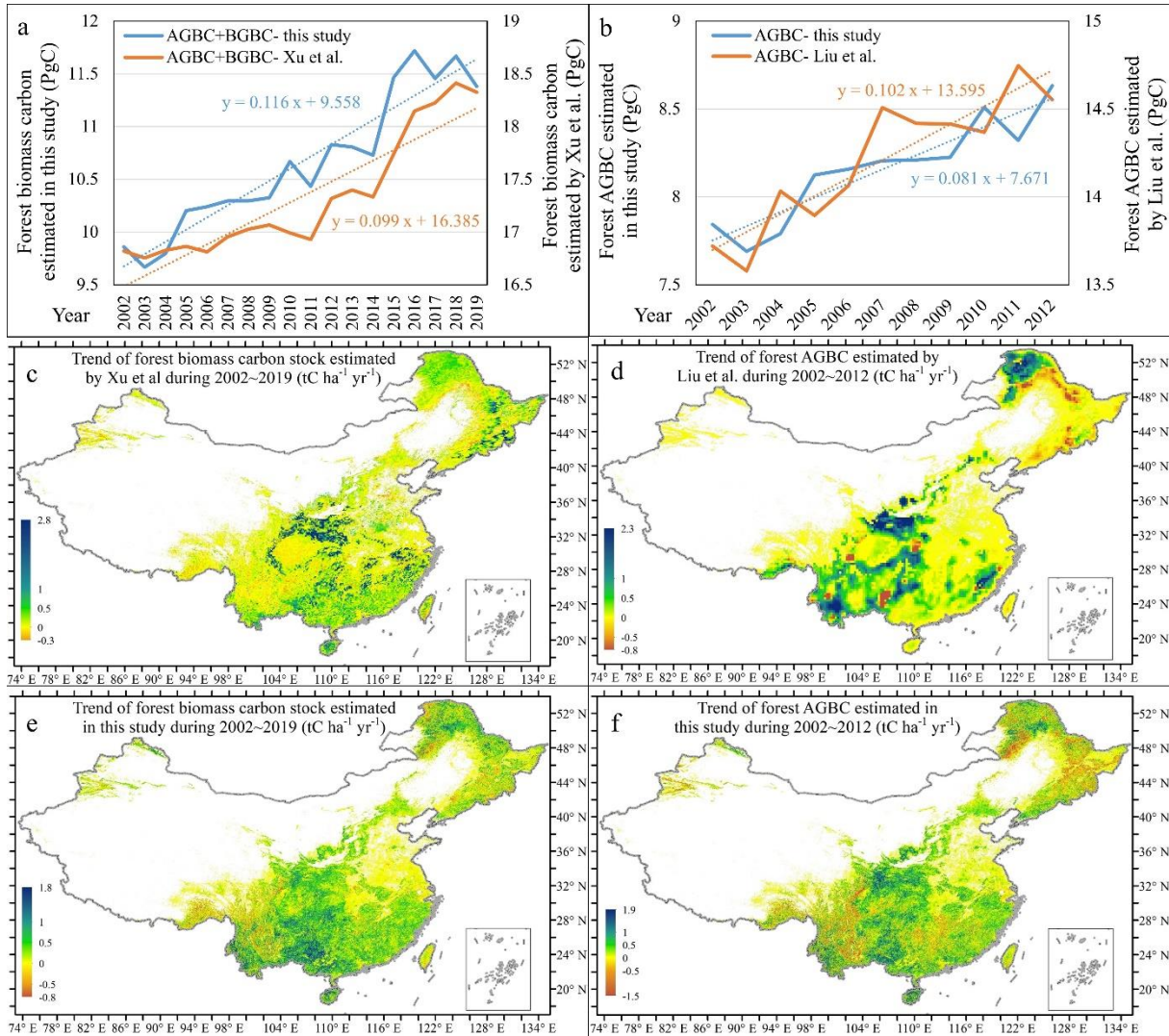
337 all forest biomass carbon stock trend.

## 338 **4 Discussion**

### 339 **4.1 Comparison of the estimated forest biomass carbon pool change in this study against** 340 **the existing datasets**

341 Although with potential overestimation, the inter-annual variation in forest AGBC in China  
342 according to Liu et al. (2015) and that of total biomass carbon according to Xu et al. (2021) are  
343 both highly correlated with our results ( $R^2= 0.65$  and  $0.88$ ). Liu et al. predicted a forest AGBC  
344 increase rate of  $102.2 \pm 35.8$  Tg/yr ( $p<0.01$ ), slightly higher than our estimate of  $80.8 \pm 25.1$   
345 Tg/yr during 2002–2012; while Xu et al. indicated a biomass carbon stock trend of  $99.4 \pm 23.2$   
346 Tg/yr ( $p<0.01$ ) from 2002 to 2019, slightly lower than the rate of  $115.6 \pm 20.2$  Tg/yr in this  
347 study (Figure 7a~b). The spatial maps of the forest biomass carbon trends estimated by Xu et  
348 al. and Liu et al. were slightly patchy (Figure 7c~d). Compared to this study, the two existing  
349 datasets (i.e., Liu et al. (2015) and Xu et al. (2021)'s datasets) predicted higher biomass carbon  
350 stock trends in the Qinling Mountains and the mature deciduous conifer forests in northeast  
351 China. Meanwhile, they predicted lower carbon sinks in southern China (Figure 7c~f), where  
352 reforestation and forest management-induced short term extensive carbon uptake (Tong et al.,  
353 2020) have been confirmed by atmospheric inversions (Wang et al., 2020; Yang et al., 2021).  
354 Finally, by comparing Figure 7e and 7f, we could also notice that the hotspot of forest biomass  
355 carbon gains has moved from the Loess Plateau over the first decade of our study period (2002–  
356 2012) to southern China (e.g., Guangxi Province) later. This change was probably due to the

357 large-scale implementation of the ‘Grain for Green’ project on the Loess Plateau (Liu et al.,  
 358 2020; Wu et al., 2019) before 2012, and the massive plantation of fast-growing trees in southern  
 359 China after 2010 (Tong et al., 2020).



360  
 361 Figure 7. Comparison of the estimated forest biomass carbon pool change in this study against  
 362 two existing datasets. (a) Comparison of the inter-annual variation of forest biomass carbon in  
 363 this study against the estimate by Xu et al. during 2002–2019; (b) comparison of the inter-  
 364 annual variation of forest AGBC calculated in this study against the estimate by Liu et al. over  
 365 2002–2012; (c) map of the inter-annual trend of forest biomass carbon stock in China during



366 2003–2019 according to Xu et al; (d) map of the forest AGBC trend in China during 2003–2012  
367 according to Liu et al; (e) map of the estimated trend of forest biomass carbon stock over 2002–  
368 2019 in this study; (f) map of the estimated forest AGBC trend over 2002–2012 in this study.

#### 369 **4.2 Some uncertainties of the forest biomass carbon dataset and future prospects**

370 During benchmark AGBC mapping, we converted the in-situ AGBC data at forest plots into the  
371 grid-scale average AGBC by multiplying by the fraction of forestland during the time period of  
372 field investigation. Considering the overall high-quality of the China's land-use/cover datasets  
373 developed via human–computer interactive interpretation of Landsat images (Liu et al., 2014;  
374 Yang and Huang, 2021), and that the producer's accuracy (PA) and user's accuracy (UA) for  
375 forestland classification in the CLCD dataset used in this study were 73% and 85% respectively,  
376 the errors within the benchmark AGBC mapping induced by the scale conversion based on the  
377 forestland area fraction were generally limited.

378 The variation in climatic conditions in the short term may have subtle influences on that in the  
379 BGB, but explicit knowledge on this effect is lacking. Instead, woody vegetation BGB is much  
380 more driven by AGB (vegetation density), as indicated by the very strong relationship between  
381 BGB and AGB ( $R^2 \geq 0.85$ ). Moreover, the long-term climatic background is expected to have a  
382 stronger influence on the RSR of perennial woody plants than the meteorological conditions in  
383 only a few years, since above- and belowground biomass allocation is the result of plants' long-  
384 term adjustment to the environment (Qi et al., 2019). Therefore, it is reasonable not to consider  
385 the influence of the specific climatic conditions in a year on the variation in BGB.

386 In the near future, P-band microwave sensors, which have higher penetrability into the canopy  
387 than L-band microwaves, will further improve AGB mapping. For example, BIOMASS, a fully  
388 polarimetric P-band SAR, is scheduled to be launched in 2022 (Le Toan et al., 2011). Therefore,  
389 in the future the relationship between P-band microwave retrievals and biomass should be  
390 addressed, as well as the calibration of historical AGB datasets (e.g., the long-term AGB dataset  
391 in this study) against the P-band SAR-based AGB benchmark map to extend the time series. In  
392 addition, an inter-calibration between the AMSR-E-based VOD and the AMSR2-based VOD  
393 will further reduce the potential bias within the long-term integrated VOD datasets (Wang et al.,  
394 2021a; Wang et al., 2021b). On the other hand, more in-situ AGB and BGB measurements in  
395 larger plots are needed to further improve the estimation of belowground biomass allocation.

## 396 **Data availability**

397 Annual forest above- and belowground biomass maps in China between 2002 and 2021 are now  
398 available at: <https://doi.org/10.6084/m9.figshare.21931161.v1>. This dataset will also be  
399 available on the National Tibetan Plateau/Third Pole Environment Data Center and PANGAEA  
400 soon (under checking now). Other open datasets that made this research possible and the related  
401 references are attached in Supplementary Information- Text S2.

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406 for sharing his forest aboveground biomass map.

## 407 **Conflict of interests**

408 The authors declare no conflict of interest.

## 409 **Credit author statement**

410 Y.C designed and conducted the research. B.F and X.F funded the research. Y.Z wrote the draft  
411 of the manuscript; X.F and all other authors read and revised the manuscript.

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