A 250m annual alpine grassland AGB dataset over the Qinghai Tibetan Plateau (2000-2019) in China based on in-situ measurements, UAV images, and MODIS Data

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16 Abstract. The alpine grassland ecosystem accounts for 53% of the Qinghai-Tibet Plateau (QTP) area , which and is an 17 important ecological protection barrier, but it is fragile and highly-vulnerable to climate change. Therefore, continuous 18 monitoring of grasslandof the aboveground biomass (AGB) of grassland is necessary. Although many studies have mapped 19 the spatial distribution of AGB over for the QTP, the results vary widely due to the limited ground samples and mismatches 20 with satellite pixel scales. This paper proposed a new algorithm using unmanned aerial vehicles (UAVs) as a bridge to re-21 estimate the grassland AGB over on the QTP from 2000 to 2019. The innovations were as follows: 1) In terms of In the 22 aspect of ground data collectionacquisition, the spatial scale matching among betweenamong the traditional ground-23 samplesquadrat sampling, UAV photos, and MODIS pixels was fully considered. From 2015 to 2019During 2015-2019, 24 906 pairs of quadrat-scale ground-UAV sample data at the quadrat scale and 2,602 sets of MODIS pixel-scale UAV data-25 matching the MODIS pixel scale were collected (over 37,000 UAV photos). Therefore, the ground validation samples was 26 were sufficient and scale-scale-matched. 2) In terms of model construction, the traditional quadrat scale $(0.25m^2)$ was 27 successfully upscaled to the MODIS pixel scale (6,2500 m²) based on the random forest method and stepwise upscaling 28 schememethods. Compared with previous studies, the scale matching of independent and dependent variables was 29 realized achieved, effectively reducing the impact of spatial scale mismatch. The results showed that the correlation between 30 the AGB values estimated by UAV and the MODIS vegetation indices was higher than that of the traditional sampling 31 method at the pixel scaleAt the pixel scale, the AGB value estimated by UAV had a more linear correlation with the MODIS-32 vegetation indices than the traditional sampling method. The multi-year independent cross-year -validation results showed that the constructed pixel scale AGB estimation had good robustness, with an average R^2 of 0.83 and RMSE of 34.13 g/m². 33

34 Our dataset provides an important input parameter for a comprehensive understanding of the role of the QTP in the process

of global climate change processes. The dataset is available from the National Tibetan Plateau/Third Pole Environment Data
 Center (https://doi.org/10.11888/Terre.tpdc.272587, Zhang et al., 2022).

37 1 Introduction

Grasslands, accounting for approximately 37% of the earth's surface, play an essential role in global carbon cycling and food supply (O'mara, 2012). However, most natural grasslands have been degraded to a certain extent due to overgrazing, farmland encroachment, soil erosion, and global climate change (Suttie et al., 2005; Ramankutty et al., 2008; O'mara, 2012). Therefore, timely monitoring of grassland health is crucial for sustainable development and understanding the global carbon cycling processing. Aboveground biomass (AGB) is a key indicator of grassland status and an important input parameter for the ecological model model and carbon storage estimation. Thus, accurate and rapid estimation of AGB is valuable for grassland monitoring.

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46 The advent of satellites has made it possible to map the spatiotemporal dynamics of large areas of grasslands over large 47 areas. The advent of satellites makes it possible to map the spatial distribution and temporal dynamics of grassland over large-48 areas. Spectral information from different satellites has been employed for biomass estimation, such as Sentinel-2, Landsat, 49 and MODIS (Wang et al., 2019; Zhang et al., 2016). Although there are differences in spatial and spectral resolution, the 50 core idea of building a biomass model is constructing the linear or nonlinear relationships between the field-field-measured 51 samples and various satellite spectral indices. Therefore, the estimation accuracy is closely related to the quality and quantity 52 of ground samples (Morais et al., 2021; Yu et al., 2021). However, tThere are still two deficiencies in ground data 53 eollectionacquisition: the large spatial scale gap between the traditional samples and satellite pixels, and the low efficiency.

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How to narrow the spatial gap between traditional samples and satellite pixels is an urgent problem to be solved. Since it is 55 impossible to harvest all the grasses within a pixel range, an the average of 3-5 quadrats size samples ($0.5 \text{ m} \times 0.5 \text{ m}$ or 1 m56 57 \times 1m) is usually used as the measurement (Dusseux et al., 2015; Yang et al., 2017), which results in a considerable spatial 58 gap. A lot of studies have been carried out to upscale ground measurements to satellite pixels (Crow et al., 2012; Bian and 59 Walsh, 1993), such as block Kriging geostatistical interpolation, different types of regression models, or machine learning 60 algorithms (Cheng et al., 2007; Wang et al., 2014; Cannavacciuolo et al., 1998; Dancy et al., 1986; Li et al., 2018). However, 61 the accuracy of these methods depends on the density of sampling points. In addition, fine-resolution satellites were used as 62 a bridge to reduce the impact of scale mismatch on AGB estimation (Yu et al., 2021; He et al., 2019). The reason is that the 63 finer the satellite resolution, the smaller the spatial gap with the ground samples The primary reason is that the spatial gap 64 between traditional ground data and fine-resolution satellites is much smaller than medium or coarse-resolution satellites 65 (Wang and Sun, 2014; Morais et al., 2021). -Therefore, obtaining a value ground samples that -matchhing the pixel scale is

the key to improving improving the accuracy of the satellite AGB inversion estimation using remote sensing data accuracy
 from remote sensing.

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69 Improving the efficiency of ground sampling is the other problem to be solved another issue that needs to be addressed. 70 Although the the traditional field sampling method can get yield high high-accuracy results, it is time--consuming and labor-71 intensive. Large-region grassland AGB inversion often requires years of accumulation to obtain ground observation samples-72 with sufficient spatial representation. For example, it took Yang et al. spent five years to complete completing the collection 73 of ground samples to investigate-invert the grassland AGB in-in China (Yang et al., 2010). Moreover, with limited original 74 ground data, some scholars had to useexpanded the sample size by using the data published by others- to expand the sample 75 size when the original ground data was limited (Xia et al., 2018; Jiao et al., 2017). However, datasets from different sources 76 may affect the overall accuracy due to the differences in sample plot size, sample size, and sampling methods. Considering-77 the differences in the plot area, guadrat size, and sampling method, datasets from different sources may affect the overall 78 inversion accuracy.

80 The development and popularization popularity of unmanned aerial vehicle (UAV) technology has provides provided new-81 ideas for to solving solve- solutions to the above problems. UAV images have been successfully used to estimate ecological 82 indicators-metrics such as FVC, biomass, and canopy height (Chen et al., 2016; Zhang et al., 2018; Bendig et al., 2015). The 83 use of UAVs has the following unparalleled advantages over traditional sampling methods. Compared with traditional sampling methods, the use of UAVs has the following incomparable advantages. First, UAVs can effectively obtain 2D-two-84 85 or 3D-three-dimensional vegetation information about vegetation structure in a non-destructive way without destroying 86 damaging it, which is helpful helpful for grassland the estimation of grassland biomass (Lussem et al., 2019; Zhang et al., 87 2022a; Zhang et al., 2018). Second, UAVs can easily-rapidly collect key parameters of grassland within satellite pixels (e.g., 88 FVC, Chen et al. 2016). Hence, UAV images can serve be used as a bridge to reduce the spatial gap between the field 89 samples and the satellite pixels. However, most current UAV-based grassland biomass estimations are small-scale, with with 90 few regional-scale studies. WIt is still unknown whether UAVs can be used to narrow-reduce the spatial gap between the 91 traditional ground samples sampling and satellite pixels remains an open question. In addition, due to the limited sample size, 92 previous studies regional-seale grassland AGB models-lacked independent years of cross-year validation-ross-validation to 93 test the robustness of the AGB estimation model over time due to the limited sample sizein different periods.

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This study proposed a new <u>method_approach that_eombining_combines_</u>traditional ground sampling, UAV photographingphotography, and satellite data to <u>generate_produce_a</u> new reliable AGB dataset of QTP grassland. The objectives of this study were: 1) to construct <u>a_the-</u>UAV-based grassland AGB estimation model_<u>s-</u>at <u>the_</u>quadrat/satellite pixel scales, respectively; 2) to investigate whether UAVs can be used as a bridge to <u>narrow_reduce_</u>the spatial gap between 99 traditional ground observation samples and satellite pixels, and to improve the estimation accuracy of grassland AGB; and

100 3) to map the AGB of alpine grasslands on the Qinghai-Tibetan Plateau (QTP) from 2000 to 2019.

101 2 Materials and Methods

102 2.1 Study Site

103 OTP is the highest and largest plateau on the earth $(26^{\circ}00'12'' \sim 39^{\circ}46'50''N, 73^{\circ}18'52'' \sim 104^{\circ}46'59''E)$, with an average 104 elevation of ~4000 m and an area of approximately 257.24×10^4 km² (Figure 1). It is located in western-western-China, and 105 thwith an average annual average temperature and precipitation of about are around 1.6°C and 413.6 mm, respectively. The 106 main grassland types are alpine meadows, alpine steppe, and sparse grassland, which play a critical role in climate regulation, 107 water conservation, and biodiversity protection (Ding et al., 2013). However, grassland ecosystems are fragile and 108 vulnerable to global climate change and human activities, and have high spatial heterogeneity. In this study, the boundary of 109 the OTP of China (Zhang et al., 2014) was downloaded from the National Earth System Science Data -Centerr, National 110 Science & Technology Infrastructure of China (http://www.geodata.cn(Zhang et al., 2014)). The-Ggrassland type data was 111 derived from the 1:1000000 Chinese digital grassland classification map provided by the China #Resource and 112 eEnvironmental science Science and data-Data Ceenter of China (https://www.resdc.cn/). This data set, generated through 113 field surveys in the 1980s and supplemented by satellite and aerial imagery, is the most detailed grassland-type map available. This data set was produced through field surveys and supplemented by satellite and aerial images in the 1980s and 114 115 is also the most detailed map of grassland types. For comparison with others studies, we combined the grassland types into 116 three categories: alpine meadow, alpine grassland, and sparse grassland, and resampled them to 250 meters (Table A1).



Figure 1. Distribution of field and UAV sampling sites <u>ofin</u>-2019 (a); <u>and</u>-UAV sampling sites <u>in of 2015-2018 in</u> alpine grasslands
 on the QTP from 2015-2018 -(b-e). Field_UAV_2019 represents the quadrat_-scale sampling sites for for the 2019 2019-UAV-Field
 synchronous grassland biomass experiment. UAV_year represents the UAV sampling point based on the GRID or RECTANGE
 mode of the corresponding year.

122 2.2 Overall technology roadmap

- 123 Figure 2 showswas the The overall flowchart of this studyof UAV-field investigation and the construction of grassland AGB-
- 124 estimation model at different spatial scales were shown in Figure 2, which It consisted s of mainly includes four main steps:
- 125 1) UAV and field investigation; 2) constructing constructing the the grassland AGB estimation model at the quadrat scale; 3)

- 126 upscaling the grassland the AGB to the MODIS pixel scales; 4) building the the final AGB estimation model at the MODIS
- 127 pixel scale and applying it to the QTP region. More detailed information about on each step was iswas described in the
- 128 following sections.



131 Figure 2. The overall flowchart of UAV_field investigation_survey and the construction of grassland AGB estimation models at different spatial scales.

133 2.3 Field investigation

134 2.3.1 UAV and route planning

DJI Phantom 3 professional Professional (DJI Company, Shenzhen, China), a popular consumer quadrotor UAV equipped with a high-resolution RGB camera, was used to collect UAV images of the QTP from 2015 to 2019. It has a 1/23-inch CMOS sensor and is capable of taking 12-megapixel photos. In addition, it uses a 3-axis stable gimbal to take photos downward vertically downward andto eliminate the distortion of UAV images. It has good environmental adaptability, with an operatingthe working temperature ranges from 0° to 40°, , and a maximum the highest take-off altitude can reach of 6000 meters. Therefore, it is well can adaptadapted well to the low temperature and high altitude of the QTP. More detailed information about the UAV system is was listed in Table A2.

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Fragmentation Monitoring and Analysis with aerial Photography (FragMap) system, which can realize capable of long-term collaborative observation, was used for UAV route planning (Yi, 2017). The repeatability of UAV observation is the basis for understanding the ecological process. Through During 2015-2019, we conducted UAV monitoring of the QTP grasslands using FragMap (Fig. 1). FragMap, we conducted UAV observations on the QTP _from 2015 to 2019 (Figure 1). Over 2,000 fixed flight routes were set-up during this period, during this period_,-and more than 37,000 UAV images were collected, providing a reliable UAV data-set for this study (Table 1).

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150 Table 1. UAV sampling information from 2015 to 2019

Year	Flight Mode	Number of routes	Photo number	Acquisition time
2015	RECTANGLE	214	2568	7.05~8.24
2016	RECTANGLE	334	4008	6.20~9.29
	GRID	150	2400	6.20~9.23
2017	RECTANGLE	315	3780	5.10~10.24
	GRID	322	5152	7.15~8.22
2018	RECTANGLE	79	948	7.22~8.03
	GRID	303	4848	7.04~8.29
2019	GRID	885	14160	7.12~9.21
	<u>Total</u>	<u>2602</u>	<u>37864</u>	

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- 152

153 GRID, RECTANGLE, and BELT are the most commonly used flight modes in the FragMap software. The-GRID and 154 RECTANGLE modes have 16 and 12 waypoints for capturing UAV images within a MODIS pixel range (Figure A1). Their 155 flying height and speed are set to 20 m and 3m/s, respectively. The spatial coverage of a 20-meter-high UAV photo is about 156 $26 \text{ m} \times 35 \text{ m}$. The BELT mode is similar to GRID, but is designed to obtain get-near-ground UAV image data with a-higher 157 resolution (Figure 3b). It can be combined with the traditional sampling method to ensure the that-consistency of UAV 158 images are consistent with the ground quadrats samples (Figure 3d). Generally Typically, the BELT size is set to $40 \text{ m} \times 40$ 159 m, and the flying height and speed are set to 2 m and 1 m/s to ensure that field workers crews have enough time to place a-160 sampling quadrat frames on-under the UAV shooting-waypoints. -As with the GRID mode, 16 UAV images can be captured-161 during in a single -one-flight. Compared with the MOSAIC flight flight-mode (which requires a guaranteed overlap rate 162 between photos to obtain a full view of an area), our design is more in line with the traditional ecological sampling concept. 163 It allows for a better balance of spatial representation and accessibility of samples, resulting in efficient sample collection.



Figure 3. Schematic diagram of the UAV-field synchronization experiment in 2019: a combination design of GRID (a) and BELT
(b) flight modes; a UAV image with a quadrat from the BELT mode at the height of 2 m (d); a 20-meter-high UAV image
including four sample quadrats (c); and the cropped UAV images at quadrat scale from 20 m (e) and 2 m (f) height, respectively.

169 2.3.2 Synchronization experiment of UAV and field sampling

170 A UAV-field biomass synchronization experiment was designed in 2019 to ensure spatial matching among betweenamong

171 satellites, UAVs, and ground sampling (Figure 3). The specific implementation steps were as follows. First, we set a GRID

172 flight mode with the a MODIS pixel size (250 m × 250 m) (Figure 3a). Then, three waypoints were randomly selected from

- 173 the GRID route-mode were randomly selected for to set setting the BELT flight routes modes (40 m \times 40 m). For each
- BELT, we placed a sampling quadrat frame (0.5 m × 0.5 m) was placed at its 6, 7, 10, and 11 waypoints to ensure that the

- 175 GRID image ean-could contain the four quadrats frames described mentioned above (Figure 3b-c). Then, at the end of all
- 176 <u>flights, the gGrassland AGB samples were then</u>-cut, bagged, and numbered-at the end of all flights. Finally, these samples 177 were oven-dried at 65°C to constant weight to obtain the field-measured AGB values.
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179 2.4 Data processing

180 **2.4.1 UAV photo pre_processing and indices calculation**

The_UAV photo prePre-processing of UAV photos_included image quality inspection, image-cropping, and calculation of different indices.-__First, we eliminated the overexposed or blurry-blurred_20-meter-high UAV images. Second, the pixels in the sampling quadrat-frames_were cropped and saved (Figure 3e). Third, we calculated the RGB indices for the cropped UAV images_were calculated. Similar to our previous study, indices included color space, histogram, and vegetation indices, the_details of which could can_be found in Zhang et al. (2022a). -reference (Zhang et al., 2022a). In addition, 30 other RGB vegetation indices were added as candidate independent variables. The names, formulas, and references of the above indiceswere were shown in Table A3.

188 2.4.2 MODIS vegetation index and other spatial data

189 The MOD13O1(v006) product was downloaded from the NASA earth explorer website (https://earthexplorer.usgs.gov/) for 190 the inversion of the alpine grassland AGB on the QTP. The data contained contained two commonly used vegetation indices, 191 the Nnormalized vegetation Vegetation index-Index (NDVI) and the enhanced enhanced vegetation vegetation index-Index 192 (EVI), with spatial and temporal resolutions of 250 m and 16 days, respectively. A total of 2,842 scenes from 2000 to 2019 193 were downloaded. Then, the MODIS images were reprojected and stitched using the MODIS projection Projection tool Tool 194 (MRT). After that, the corresponding vegetation indices closest to the time of the UAV sampling were extracted to 195 construct/validate a pixel-scale AGB estimation model to construct a pixel-scale AGB estimation model. After that, we used 196 the point extraction function in ArcGIS software to get the corresponding vegetation indices of the UAV samples to 197 construct the pixel-scale AGB estimation model. In addition, based on the NDVI index and the formula kNDVI= TANH-198 (NDVI²), the kNDVI index was calculated to overcome the NDVI saturation issue based on the equation kNDVI= TANH 199 (NDVI²) (Camps-Valls et al., 2021). The annual maximum vegetation indices were calculated by the maximum value 200 composition (MVC) algorithm of ENVI software to estimate the spatial AGB distribution of QTP from 2000 to 2019 201 (Holben, 1986; Wang et al., 2021; Gao et al., 2020).

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Furthermore, the meteorological, soil texture, and topographic data were also included as candidate independent variables for constructing the pixel-scale AGB estimation model. Meteorological factors, including the annual mean temperature (TA), annual mean precipitation (PREC), and annual total solar radiation (RAD), were calculated based on the daily meteorological dataset from the National Meteorological Information Center of China. The data processing steps mainly included interpolation, cumulative summation, and annual averaging processing to obtain the <u>a</u> meteorological raster dataset with a spatial resolution of 1000 meters (Li et al., 2021). Moreover, the spatial distribution data of soil texture <u>data</u> with <u>ata-1</u> 1 km spatial resolution, including the ratio of soil organic matter (SOM), clay, sand, and silt, were downloaded from the Resource and Science and Data Center of China (<u>https://www.resdc.cn/</u>). All the meteorological and soil datasets were resampled into 250 m by ArcGIS software to match the MODIS data.

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213 Terrain factors included the digital elevation model (DEM), slope, and aspect. The DEM was derived from shuttle-Shuttle

- 214 radar topography Topography mission Mission (SRTM) images imagery (version 004, 90 m) and resampled to 250 m.
- 215 Then use the terrain analysis tool of ArcGIS software to calculate the The slope and aspect data were derived -based onfrom
- 216 DEM data using the terrain analysis tool of ArcGIS software. Slope and aspect were then calculated from the DEM data
- 217 using the terrain analysis tools of ArcGIS software.

218 2.5 AGB modeling and computation at different scales

- 219 We estimated the grassland AGB at three scales: the quadrat scale, <u>the photo scale</u>, and <u>the satellite pixel scale</u> (Figure 4).
- 220 More detailed information wasiswas described as follows.



Quadrat scale

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Figure 4. Upscaling steps to estimate grassland AGB matching the MODIS pixel scale.

223 2.5.1 Modeling method

Random Forest (RF) (Breiman, 2001), (Breiman, 2001) is an ensemble-learning algorithm, was that has been widely used employed-to estimate AGB-at different scales due to its excellent performance in biomass estimation (Ghosh and Behera, 2018; Mutanga et al., 2012; Wang et al., 2016). The tTwo main parameters, namely the number of regression trees in the forest (*ntree*) and the number of feature variables required to create branches (*mtry*), were first optimized based on the root mean square error (RMSE) of training data-at first. Here, the value of *ntree* values werewas tested set from 100 to 5000 with an interval of 100, and-while the *mtry* was set as the square root of the number of training sample features. In addition, the importance of each predictor was ranked by calculating the percentage increase in mean square error (%IncMSE).

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The backward feature elimination method (BFE) was used to reduce the number of input variables to simply simplify the RF model (Vergara and Estévez, 2014). The main steps were as follows: 1) constructing an AGB RF model by including all predictor variables in the initial stages and calculating the %IncMSE index for each variable; 2) eliminating the least promising variable and then rerunning the RF model until only one independent variable was was left. Moreover, the corresponding coefficient of determination (R²) and the corresponding RMSE were calculated in each iteration ; 3) selecting the smallest subset of variables with the highest R² was selected as the final optimized indices.

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In addition, different training and validation strategies were used at different scales. At the quadrat scale, a 10-fold crossvalidation method was used due to the limited ground samples (Kohavi, 1995). At the pixel scale, 30% of the UAVestimated AGB samples in 2019 were randomly selected as an independent validation dataset due to the large sample size. Meanwhile, the UAV_AGB values from 2015 to 2018 were used for cross-year validation was performed using UAVestimated AGB values estimated by UAVs from 2015 to 2018 to test the robustness of the model over-over timedifferent periods. Statistical measuresmetrics, including the R² (Eq.1) and , the RMSE (Eq.2), and mean absolute percentage error (MAPE, Eq.3), were used to qualify evaluate the model-performance performance of the model.

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$$R^{2} = 1 - \frac{\sum_{i=1}^{n} (\hat{y_{i}} - y_{i})^{2}}{\sum_{i=1}^{n} (\hat{y_{i}} - y_{i})^{2}}$$
(1)

$$RMSE = \sqrt{\frac{\sum_{i=1}^{n} (\hat{y}_{i} - y_{i})^{2}}{n}}$$
(2)

where n is the number of samples, y_i and \hat{y}_i represent the measured and the predicted AGB value, respectively, \overline{y}_i is the mean value of measured AGB samples.

250 2.5.2 AGB RF estimation model at the quadrat scale (0.25 m²)

251 Since the spatial coverage of a 20m-high UAV photo (26 m×35 m) is much wider than a single 2m-high UAV photo, making

- 252 it easier to match to the MODIS pixel scale. Hence, the 20m-high UAV photos containing the sample frames were chosen
- 253 for constructing the quadrat-scale AGB estimation model. A total of 906 pairs of quadrat-scale UAV-field AGB observation

data were collected, with good spatial representativeness (Figure 1 a, red dots). The observed AGB values ranged from 0 g/m^2 -to 450 g/m², with mean and median values of 59.75 g/m² and 33.04 g/m², respectively, most of which were less than 100 g/m² (Figure 5a). Then, the The cropped 20-meter-high UAV image indices and the measured AGB values were used as the independent and dependent variables to build the RF model (Figure 2).

258 2.5.3 AGB calculation at the photo scale (~900 m²)

The steps for for AGB estimation of the entire whole 20-meter-high UAV photo were as follows: 1) First, each UAV photo was divided into ~2,000 quadrat-sized small patches. 2) Second, the AGB of each small patch was calculated based on the quadrat-scale AGB estimation model. 3) Finally, the average value of all the small patches was calculated as the AGB of the whole photo. Based on the above steps, the AGB values of 37,487 images in GRID or RECTANGLE mode were calculated using over-more than 74 million AGB values at the of the quadrat scale (Table 1).





Figure 5. Histograms of field-measured AGB values at quadrat scale (a) and UAV-estimated AGB values of different years at the photo scale (b).

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269 2.5.4 AGB RF model construction at MODIS pixel-scale (6,2500 m²)

270 The following steps were involved in constructing the AGB estimation model at the pixel scale. 1) Since the coverage of a

271 GRID or RECTANGLE <u>route</u> was similar to that of <u>the a</u> MODIS pixel, the average of its 16 or 12 photos was taken

- as the AGB value of the corresponding pixel. From 2015-2019, a total of ____2,602 UAV-estimated AGB samples were
- 273 obtained at the pixel scale from 2015 to 2019 (Table 1). 2) The MODIS vegetation indices and other spatial metrics

274 corresponding to each GRID or RECTANGLE route-mode were then extracted using the ArcGIS software. Here, the

275 MODIS NDVI, EVI, and kNDVI indices closest to the sampling time were chosen to minimize the time difference between

- 276 sampling and satellite overpass.3) Subsequently, the UAV-estimated AGB values and the extracted spatial indices were used
- 277 as dependent and independent variables to build the AGB estimated model at the pixel scale using the RF algorithm.

278 **2.6** Uncertainty analysis

279 Since the actual AGB values of MODIS pixels cannot be directly obtained, vegetation indices were used to quantify the 280 uncertainty of different AGB estimation methods. In other words, the higher correlation between the estimated AGB and 281 MODIS vegetation indices, the higher accuracy of the correlation between the estimated AGB and MODIS vegetation indices, 282 the more accurate the estimation model was-was. This stud The performance of the estimation model was evaluated through 283 the three aspects. In this study, wey firstly compared the correlation between the MODIS vegetation indices and AGB values 284 obtained by traditional sampling and UAV estimation methods. We also explored the uncertainties of UAV sampling 285 coverage by randomly combining the number of photos in a MODIS pixel, -and tested whether the estimated AGB was 286 closer to the true value as the number increased. Furthermore, the AGB validation results from GRID or RECTANGLE at 287 the pixel scale were compared to understand the uncertainties caused by different flight modes.

288 2.7 Trend analysis of grassland AGB

289 This study combined the Theil-Sen median trend analysis and Mann-Kendall test to analyze the temporal variation

- 290 characteristics of grassland AGB of -QTP (Jiang et al., 2015). Theil-Sen median trend analysis is a robust trend statistical
- 291 method with high computational efficiency, insensitive to outliers (Hoaglin et al., 1983). The Mann-Kendall test is a
- 292 nonparametric test for time series trends, which does not require the measurements to follow a normal distribution and is not
- 293 affected by missing values and outliers. The Theil-Sen Median trend analysis and Mann-Kendall trend test have been widely
- used to analyze vegetation index, cover, and biomass (Gao et al., 2020; Jiang et al., 2015; Fensholt et al., 2009). The
- 295 formulas for the Theil-Sen median trend analysis and the Mann-Kendall method are detailed in Jiang et al. (2015)(Jiang et al.,
- 296 2015).
- 297

298 3 Results

299 3.1 Independent variables selected for AGB modeling

300 The selected-independent variables for AGB estimation at the -quadrat and pixel scales were arewere listed presented -in

301 Table 2. A total of 36 independent variables were finally selected at the quadrat scale, including 26 vegetation RGB indices,

- 302 6 histogram indices, and 4 color space indices (Figure A2). At the pixel scale, five variables were selected, including NDVI,
- 303 kNDVI, EVI, PREC, and DEM (Figure A3).
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- 305

Scale	Model	Number	Independent variables
Quadrat	RF _Q	36	FVC, WI, GI, EXG, TGI, EXGR, VEG, GRATIO, COM, CIVE, RGBVI, EXR,
			GLA, GRRI, MVARI, MGRVI, GRVI, RGRI, GBRI, VARI, NDI, RRATIO,
			EXB, V, IPCA, INT,
			HOC_R_CORR, HOC_B_CHIS, HOC_R_CHIS, HOC_G_CHIS,
			HOC_G_CORR, HOC_B_CORR
			B, H, G, R,
Pixel	RF_P	5	NDVI, KNDVI, EVI, DEM, PREC

Table 2: Selected independent variables for the AGB modeling at quadrat and pixel scales. The full names of each variable at the quadrat scale were listed in Table A<u>3</u>2.

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309 3.2 Modeling and accuracy assessment

310 For the guadrat-scale the AGB estimation model at the guadrat scale, the results of 10-cross validations results showed that 311 there was a significant linear relationship between the estimated \overline{AGB} values of the model and the measured values (R²=0.73, 312 p < 0.001, Table 3, Table A4). The student's t-test was also used to evaluate assess whether there was a significant differences 313 existed between the predicted AGB values and the measured values at a coefficient confidence level of 95%. As shown in Table 4, -showed there was there was no significant difference between the predicted and measured average AGB values-314 (p=0.51>0.05) with an . The total RMSE of and MAPE of the prediction model were -32.94 g/m² and 48.94%, respectively. 315 316 The scatter plot showed that the model predicted well when the measured biomass was less than 150g/m², but showed some underestimation when it was greater more than 200g/m² (Figure -6a). The reason might be that It may be because the number 317 318 of samples the sample size larger more than 200g/m² was is relatively small, accounting for only 8.50% of allthe total 319 number of samples (Figure 5a). Although the UAV sample size of UAVs varied vearly from year to year, the estimated most 320 of the -AGB values at estimated from photosthe photo scale ranged from 0 to 300 g/m² (Figure 5b). ThMost of the-321 AGB averages estimated by UAVs were around 150 g/m² from 2016 to 2019, and slightly lower in 2015 (108 g/m²).e mean-

322 UAV AGB in 2016-2019 was around 150 g/m², while it was slightly lower in 2015 with 108 g/m².

323

For the pixel-scale AGB estimation model, <u>there was a strong linear relationship s existed between the predicted AGB and</u> UAV <u>estimates estimated values for 2015-2019</u> (Table A<u>4</u>-). In 2019, The fitting coefficient R² was 0.85 in for 2017-2019, an,d and slightly lower in for 2015-2016 at, at 0.63 and 0.77, respectively (Table 3, Figure 6b-f). The RMSE and MAPE of the pixel-scale model ranged from 23.36 g/m² to 34.07 g/m², 12.32% to 25.19%, respectively (Table 3). In addition, we found that there were no significant differences between the predicted and measured average AGB values, except in for 2017 and 2018 (Table 4). While While the average model projections the averages for or the 2017 and 2018 model estimates

330 are-were 14.72% and 13.78% lower than the UAV estimates, respectively, they are-were within an acceptable ranges.

331 Therefore, the cross-year validation results indicated that it the constructed pixel-scale AGB estimation model had good

332 good performance and robustness in different across in different years the cross year validation results indicated that it had
 333 good performance and robustness in different years (Figure 6b~f).

338 Table 3: Validation results of AGB models at quadrat and pixel scales

Scale	Year	Training s	et		Validation s	et
		R ²	RMSE(g/	[/] m ²)	R ²	RMSE(g/m ²)
Quadrat-scale	2019	0.94	20.18		0.73 ***	32.94
Pixel-scale	2019	0.96	10.68		0.85 ***	23.36
	2018				0.85 ***	24.83
	2017				0.85 ***	23.83
	2016				0.77 ***	31.28
	2015				0.63 ***	34.07
'***' significant at p	< 0.001					
Table 4: T-test results	between the predic	ted and measu	ired AGB values	for the mod	es of at the quadrat	and pixel scales
					_	
Validation model	Measured me	<u>ean Pr</u>	<u>edicted mean</u>	<u>t</u>	<u>df</u>	<u>p-value</u>
2019_Quadrat-scale	<u>51.57</u>	<u>54</u>	.35	<u>-0.66</u>	<u>939.35</u>	<u>0.51</u>
2019 Pixel scale	<u>136.68</u>	<u>13</u>	7.7461	<u>-0.15</u>	<u>340.78</u>	<u>0.88</u>
2018 Pixel scale	<u>152.49</u>	<u>13</u>	<u>1.48</u>	<u>4.01</u>	<u>723.81</u>	<u>6.63e-05</u>
2017 Pixel scale	<u>141.42</u>	<u>12</u>	0.60	<u>5.48</u>	<u>1225.2</u>	<u>5.26e-08</u>
2016 Divel geole	140 56	1.4	0.70	1 (0	0.61.00	

1.96

1225.2

0.05

344

2015 Pixel scale

345

346 3.3 Correlation analysis between AGB values and MODIS indices

108.65

347 The correlationss between the UAV-estimated AGB values and MODIS vegetation indices were much better than the 348 traditional ground sampling method (Figure 7a). For example, the correlation between NDVI and traditionally measured 349 AGB was only 0.53, much lower than that that obtained from a single UAV image (r=0.74). Moreover, the correlation 350 between NDVI and UAV-estimated AGB increased with the number of UAV photos. It increased rapidly as the number 351 increased from 1 to 4 (from 0.74 to 0.86), then slowed down and stabilized (from 0.87 to 0.88).

98.23

352

353 In addition, we compared the the scatter plots and fitting lines between NDVI and different AGB estimation methods (Figure 354 7b-f). The results showed a weak linear relationship between the traditionally measured AGB and NDVIa weak linear 355 relationship between the traditional measured AGB and NDVI, and with an the R² was only of 0.29. With the UAV sampling 356 method, Linearity the linear relationship was greatly improved ausing the UAV sampling method and increased with the number of photosphotographs. The fit coefficient R^2 increased from 0.54 to 0.78, much higher than the traditional sampling

357

358 method (Figure 7).







364 Figure 6Figure 6. Validation results of the AGB estimation models at the quadrat (a) and MODIS pixel scale for 2015-2019 (b~f).





Figure 7. The eCorrelations between the-MODIS vegetation indices and different AGB estimation methods (a); the scatter plots between of NDVI and with different AGB estimation methods (b-f). UAV_x, x represents the number of UAV photos used to 368 369 estimate the average AGB at the MODIS pixel- scale. Here, the value range of x is-ranges from 1 to 16.

372 3.4 Spatial distribution of grassland AGB

373 The spatial distribution of the average grassland AGB on the OTP from 2000 to 2019 was calculated (Figure 8). The 374 AGB gradually increased from west to east. As shown in Figure 8b, the average biomass of eastern OA1, IIAB, IB1, and 375 IIC2 eco-geographical regions ranged from 150 to 190 g/m^2 , and the average AGB of IC1 and IIC1 ranged from 80 to 110 376 g/m². The average AGB of IID2, IID3, IC2, and IID1 in the west was relatively low, ranging from 35 to 75 g/m². The ID1 377 region is was dominated by sparse grassland with the lowest average interannual AGB values, which fluctuated around 20 378 g/m^2 with interannual mean AGB values fluctuating around 20 g/m². (Figure 8b), From 2000 to 2019. The mean average 379 AGB on theof OTP showed an insignificant increasing trend between 2000 and 2019, with an average growth rate of 0.22 gm⁻²a⁻¹ (Figure 9a). The overall mean AGB of the QTP was 103.6 g/m², -and with 151.85 g/m², 60.85 g/m², and 28.91 g/m² 380 for the mean AGB of the alpine meadow, alpine steppe, and sparse grassland, r-were 151.85 g/m², 60.85 g/, and 28.91 g/m², 381 382 respectively (Figure 9b). In addition, the temporal trend of grassland AGB in each pixel was analyzed. As shown in Figure 383 10, the IID3, ID1, IID2, and IIC2 eco-geographical regions of the northern OTP showed an increasing trend from 2000 to 384 2019, while the IC2, IB1, and IIC1 regions showed some degradation. Therefore, although the overall AGB of the QTP-385 showed an increasing trend from 2000 to 2019, there was spatial heterogeneity in the temporal variation.

386

387 <u>-</u>





to 2019, the mean AGB on the QTP showed an insignificant increasing trend, with an average rate of 0.22 gm⁻²a⁻¹ (Figure 9a).
 The overall mean AGB of the QTP was 103.6 g/m², and the mean AGB of the alpine meadow, alpine steppe, and sparse grassland were 151.85 g/m², 60.85 g/, and 28.91 g/m², respectively (Figure 9b).

From 2000-



Figure 8.- (a) The spatial distribution of average grassland AGB -on the Qinghai-Tibet PlateauQTP during 2000-from 2000 to 2019. IID1, IID2, IID3, ID, IIC1, IIC2, IC1, IB1 IIAB1, and OA1 are the eco-geographical regions of the QTP(Zheng, 1996). The full names of each eco-geographical region were listed in Table A5. (b) AGB values of each eco-geographical region from 2000 to 2019. (c) Comparison of multi-year AGB averages in the different eco-geographical regions.









40.

404Figure 10. Spatial trends of grassland AGB on the
QTP from 2000 to 2019. IID1, IID2, IID3, ID, IIC1, IIC2, IC1, IB1 IIAB1, and
OA1 are the eco-geographical regions of the QTP (Zheng, 1996). The full names of each eco-geographical region were listed in
Table A5.

409 4. Discussion

410 4.1 Scale matching and its influence factor

411 In Unlike the previous studies, the AGB value of a satellite pixel was directly represented by the average value of 3-5 412 quadrat-scale samples, so there is a large spatial gap between the ground samples and the satellite pixels -(Yang et al., 2017; 413 Yang et al., 2009; Meng et al., 2020), which directly represented the AGB value of a satellite pixel with the average value of 414 3-5 quadrat-scale samples. The spatial gap between ground samples and satellite indices affects the accuracy of grassland 415 AGB models. The smaller the spatial gap between the two, the higher the accuracy of the model (Morais et al., 2021). To-416 address this issue, we used the UAV as a bridge to close the gap We addressed this issue using the UAVs as a bridge to 417 reduce the spatial gap, this study successfully upscaled the traditional quadrat scale to the MODIS pixel scale. We achieved 418 the Sspatial scale matching of dependent and independent variables was achieved in estimating when calculating the AGB 419 values at different scales. First, at the quadrat scale, the independent variables were all derived from cropped 20-meter-high 420 UAV images corresponding to the ground samples (Figure 3e). Then, the 20-meter-high UAV image was cropped into 421 ~ 2000 guadrat-sized small-patches to ensure consistency with the guadrat--scale model, and the average of these patches was 422 taken used as the final AGB at the photo- scale. Finally, by averaging the AGB of 16 or 12 UAV photos within the MODIS 423 pixel, the AGB AGB-value matching-that matched the MODIS pixel scale was calculated by the average value of averaging 424 the AGB of 16 or 12 UAV photos within the MODIS pixel (Figure A1). Through the above With these three steps, we 425 successfully upscaled the measured AGB from the traditional quadrat scale (0.5 m×0.5 m) to the photo scale (26 m×35 m) 426 and MODIS pixel scale (250 m×250 m). Our results showed that <u>at the pixel scale</u>, the correlations between the UAV-427 estimated UAV estimated AGB values estimated by UAV and the MODIS vegetation indices indexindices was were higher 428 than that of the traditional sampling method (Figure 7).

429

430 Furthermore, we found that the spatial coverage of the UAV sampling had an impact had a particular influence on the the-431 effect s of scale matching. Our results indicated showed that the closer the spatial coverage of the UAV sampling was to the 432 satellite pixel, the higher its correlation with MODIS spectral indices (Figure 7a). This It is also confirmed by 433 comparing The comparison of the validation results of of different flight modes also confirmed this. At the pixel scale, we 434 found that the R² between the model predictions and the AGB values estimated based on the GRID mode was better than that 435 of RECTANGLE (Figure 11). At the pixel scale, we found a higher correlation between the model predictions and the AGB-436 estimates obtained based on the GRID model than the RECTANGLE model (Figure 11). At the pixel scale, we found that 437 UAV AGB estimates from the GRID mode had a higher correlation with the mode predictions than the RECTANGLE flight 438 mode (Figure 10). The reason is that GIRD mode can take 16 pictures within a MODIS pixel, while RECTANGLE mode

- 439 only takes 12 pictures (Figure A1). The reason was that the GIRD mode could obtain 16 photos in the MODIS pixel at a time,
- 440 while the RECTANGLE mode could only take 12 photos.
- 441
- 442 The above results confirmed that UAVss could serve as a bridge to effectively narrow <u>closereduce</u> the spatial gap between
- 443 traditional samples and satellite data.







446 Figure 1011. Comparison of validation results for the GRID (a,c,e) and RECTANGLE (b,d,f) modes in 2016-2018.

447 4.2 Importance of the addition of non-vegetation samples

448 Compared with traditional sampling, UAV sampling has the advantage of wide spatial coverage (0.5 m×0.5 m vs. VS 35 449 m×26 m). Thus, vegetation and non-vegetation background information, such as roads, water, soil, gravel, riverbed, etc., 450 were captured on the UAV photos the UAV image could capture vegetation and non-vegetation background information. 451 such as roads, water, soil, gravel, riverbed, etc. (Figure A4Figure 1112). Adding The addition of non-vegetated vegetation 452 samples could improve the estimation accuracy of A-AGB estimation at the photo scale, especially for low-coveragecover 453 areas, -to avoid overestimation. The It same was also true for the pixel--scale AGB estimation model. However, the less-454 eonsideration was given to the non-vegetated areas in the traditional method traditional sampling method gave less 455 consideration to the non-vegetation vegetated areas. The sample plots were mainly set -in areas with the homogeneous 456 uniform spatial distribution, -and rarely but few in areas with spatial heterogeneity. This defect shortcoming might may limit 457 the accuracy of AGB estimation due to the high spatial heterogeneity of the QTP. Fortunately, the UAV sampling method 458 could can avoid this drawback. It can objectively record surface information and reduce the influence of manual plot 459 selection on AGB estimation. It could objectively record the ground surface information with both vegetated and non-460 vegetated areas, resulting in a more objective AGB estimation at the pixel scale.

461

9-124393.JPG	10-99489.JPG	10-108960.JPG	10-110458.JPG	10-110650.JPG
11-136340.JPG	11-143170.JPG	12-99734.JPG	12-108669.JPG	12-109042.JPG
13-124317.JPG	13-124406.JPG	14-98088.JPG	14-102399.JPG	14-124108.JPG
15-124210.JPG	15-124383.JPG	15-139325.JPG	15-146818.JPG	16-99687.JPG

463 Figure 1112. Examples of 20-meter-high UAV images with different non-vegetation background information.

464

462

466 4.3 Comparison of the estimated AGB with previous studies

467 We compared our resultsIn the following, the AGB estimation results of this study were compared with previous
 468 studiesthose of others at the quadrat, pixel, and regional scales.at the quadrat scale, pixel scale, and regional scale.

469

At the quadrat scale, consistent with our previous study, we further confirmed that the UAV RGB images could be used to estimate grassland AGB at the quadrat scale over a large region (Zhang et al., 2022a; Zhang et al., 2018). Similar to the 2meter-high UAV image, the indices from the 20-meter-high UAV images could also be used to estimate the grassland AGB at the quadrat_-scale_-(R²=0.73, RMSE=44.23_-g/m²_x). The quadrat-scale UAV model had an excellent grassland AGBestimation ability in the range of 0-150 g/m², and the verification points were mainly distributed near the 1:1 line_(Figure 6a). Compared with the 2-meter-high UAV image, the 20-meter-high UAV image is more suitable for matching the MODIS pixel due to its wider spatial coverage (26 m ×35 m).

477

478 At the pixel scale, compared with other studies, this paper achieved the spatial scale matching of independent variables and 479 dependent variables in the during the modeling modeling. -In previous studies process (Yang et al., 2009; Yang et al., 2017; 480 Meng et al., 2020), they directly constructed the models from the measured AGB values at the quadrat-scale and the 481 spectral indices of the satellites without considering the spatial scale difference. It partly explaineds why the R² of the AGB 482 linear model constructed by Yang et al. was only 0.4 (Yang et al., 2009). Our results also confirmed that after considering 483 the scale difference between measured AGB and NDVI, the R^2 of the linear model could be increased from 0.29 to 0.78 after 484 reducing the spatial gap between measured AGB and NDVI (Figure 7). In addition, thanks to the rapid sampling of UAV 485 AGB, a total of 2.602 samples matching the pixel scale were collected during 2015-2019. It allowed us to perform multi-year 486 cross-year validation to assess the robustness of the model at different timesover time of AGB models at pixel scale to verify 487 the model's robustness in different years, which washas rarely been performed in previous studies. Our results showed 488 similar validation results for 2017-2019 ($R^2=0.85$, p<0.001) des -despite -different sample sizes and spatial distributions 489 (Figure 1, Table 1).

However, previous studies only randomly selected 20-30% of samples and rarely considered the independence of samples
on the time scale.

In addition, we implemented large region and multi-year cross-validation in model verification. Despite differences in sample size and spatial distribution (Figure 1, Table 1), the validation results for 2017-2019 were similar (R²=0.85). But in 2015-2016, -,-R² was relatively low, at 0.63 and 0.77, respectively (Table 3, Figure 6). The reason was that during 2015-2016, due to the improper setting, many some photos with abnormal white balance were obtained due to improper settings, which reduced the accuracy of the estimation accuracy (Figure A5Figure 1213). The validation results showed that the pixel-scale AGB estimation model had good robustness in different regions and times when the photo quality was acceptable. The validation results indicated show that that the pixel-scale AGB estimation mode has 1 had good good

- 499 adaptability robustness in different regions and times when the photo quality is acceptable periods while obtaining high-
- 500 quality UAV images. Therefore, this method can be used to estimate the AGB values matching the satellite pixel scale in
- 501 large regions.



503 Figure 1213. An example of a set of GIRD photos with abnormal white balance in 2015.

504

505 Table 45: Comparison of AGB estimation results of different studies on the QTP

Mean	Alpine	Alpine	Study period	Approach	Input F	Data References
AGB	steppe	meadow			<u>parameter</u>	
(g/m^2)	(g/m^2)	(g/m ²)			source	
68.8	50.1	90.8	2001-2004	Linear regression	MODIS-EVI	(Yang et al., 2009)
	22.4	42.37	2000-2012	Linear regression	MODIS-NDVI	(Liu et al., 2017)
<u>120.73</u>	_	_	<u>1980–2014</u>	Exponential	<u>NDVI</u>	<u>(Jiao et al., 2017)</u>
				regression		
78.4			1982-2010	RF	GIMMS<u>NDVI,</u>	(Xia et al., 2018)
					<u>climate</u>	
77.12	76.43	154.72	2000-2014	RF	<u>NDVI, EVI,</u>	(Zeng et al., 2019)
					<u>climate,</u>	
					terrainMODIS	
59.63	42.75	77.56	2000-2017	RF	<u>MODISNDVI,</u>	(Gao et al., 2020)
					<u>climate</u>	
120.73	_	_	1980–2014	Regression	MODIS	(Jiao et al., 2017;
						Zhang et al.,
						2022b)
<u>102.4</u>			<u>2000-2020</u>	<u>RF</u>	climate, soil, and	<u>(Zhang et al.,</u>

<u>70.00</u>	_	_	<u>1960–2002</u>	Century	terrain celimate and soil	2022b) (Zhang et al.,
<u>119.78</u>	_	_	2002-2004	Orchidee	<u>data</u> <u>cClimate, soil and</u>	2007) (Tan et al., 2010)
103.6	60.85	151.85	2000-2019	RF	MODIS	this study

507 At the regional scale, consistent with previous results, we found an overall increase in AGB over the OTP from 2001 to 2019, 508 albeit with although there were fluctuations among years (Zeng et al., 2019; Gao et al., 2020). The annual average AGB of 509 grassland was 103.6 g/m², which was closest to Zhang et al. (Zhang et al., 2022b) and within the range of the previous 510 estimates (59.63-120.73 g/m²) (Table 5). The annual mean AGB of grassland was 103.6 g/m², within the previously estimated range (59.63-120.73 g/m²) (Table 4) .- The mean AGB varied among of different grassland types was different, 511 512 with 151.85 g/m² for among which the the alpine meadow was 151.85 g/m², and 60.85 g/m² for the alpine steppe was 60.85 513 g/m^2 . Our estimation results were similar to those of Zeng et al. (Zeng et al., 2019), but the overall average AGB was higher 514 than their estimated of 77.12 g/m². The spatial distribution of AGB was consistent with previous studies, showing a west-to-515 east increasing trend (Zhang et al., 2022b; Xia et al., 2018). Specifically, the average AGB of OA1, IIAB, IB1, and IIC2 eco-516 geographical regions in the east was significantly higher than that of IID2, IID3, IC2, IID1, and ID1 regions in the west 517 (Figure 8). In general, the average AGB estimates for each eco-geographical region in this paper were not much different 518 from those of Zhang et al. (2022b). Among them, our average AGB estimates for ID1, IID1, IID3, and IID2 regions were 519 slightly lower, but our values were closer to the measured values of these regions (Figure 8c). The reason may be that they 520 calculated the potential biomassAGB, while we calculated the actual biomassAGB, so our estimate was relatively low. The 521 reason may be that they calculate the potential biomass, while we calculate the actual biomass, so the estimate is low. In 522 terms of spatial and temprooral trends, the data results showed that the eco-geographical regions in the northern part of the 523 OTP demonstrated an increasing trend (IID3, ID1, IID2, and IIC2), while the IC2, IIC1, and IB1 regions exhibited 524 significant or non-significant decreases, which was consistent with the results of others (Gao et al., 2020; Liu et al., 2017).

525

526 The difference between our estimated grassland AGB and previous studies might be due to differences in data sources and 527 modeling methods. Firstly, the sample size and spatial distribution of ground samples were different. The number of ground 528 samples is the most important variable affecting the accuracy of the grassland AGB estimation model (Morais et al., 2021). -529 Unlike previous studies, we collected ground verification validation data by combining the traditional sampling method and 530 UAVs. The newly proposed method could overcome the shortcomings of traditional samplings (, such as the time-531 consuming and labor-intensive). It no longer took years of work to It no longer takes years to obtain spatially representative, 532 large-scale ground validation dataOobtaining sufficient sufficient spatially representative ground verification validation data 533 in over a large regionsarea no longer requires years of work (Yang et al., 2017). - With UAV sampling, -ground observations 534 matching the satellite pixel scale can be obtained in only 15-20 minutes, which is difficult to achieve in traditional surveys. 535 Our new sampling method not only accelerates the sampling speed and increases the sample size, but also improves the

536 spatial match between ground samples and satellite pixels. As a result, our ground validation data is superior to previous 537 studies in terms of quantity and spatial match to the satellite data. Secondly, the predictorinput variables parameters of AGB 538 estimation models were different. Some scholars used only a single vegetation index (NDVI or EVI), while others combined 539 the vegetation index with meteorological, soil, and terrain indices to construct the AGB estimation models (Table 5). In this 540 study, NDVI, kNDVI, EVI, DEM, and PREC were used as the final predictor variables to construct the AGB estimation 541 model at the pixel scale (Table 2). Thirdly, Through UAV sampling, only 15~20 minutes were needed to complete a ground-542 survey in a pixel range of 250 m × 250 m. In addition, it could effectively reduce the spatial gap between ground verification 543 samples and satellite pixels. 544

545 Meanwhile, different modeling approaches methods might also affect the simulation results. As shown in Table 5, the 546 overall AGB averages of the QTP estimated estimated varied considerably based on different methods (-such as linear or 547 nonlinear regression, machine learning, and ecological process model methods) varied considerably. Yang et al.(2017) found 548 that the model performance of ANN was much better than the linear regression model when using the same dataset to 549 estimate grassland AGB in the Three-River Headwaters Region of China (Yang et al., 2017). Jia -et al. (2016) reported that 550 the model forms could bring 13% uncertainty to the AGB estimation.n(Jia et al., 2016). Wang et al. -compared the RF with 551 the support vector regression (SVR) machine learning algorithm and found that the RF yielded the best performance in 552 grassland biomass estimation (Wang et al., 2017).-

553

554 4.4 Limitations and further work

We acknowledge that there are some shortcomings in this study. 1) The predicted values of the quadrat-scale model were 555 underestimated when the measured biomass values were greater than 250 g/m^2 (Figure 6). One reason may be that the 556 557 number of samples greater than 250 g/m² was relatively small, accounting for only 5.18 % of the total all samples. Another 558 reason may be that for high biomass grasslands, a single UAV RGB photo can only reflect information such as vegetation 559 cover and greenness, but not height information. This feature is bound to be very unfavorable for estimating AGB in 560 grassland areas with high vegetation coverage and height. Studies have shown that adding vegetation height information can 561 help improve the estimation accuracy of grassland AGB-1) The sample size greater than 200 g/m² was insufficient at the quadrat seale, leading to underestimation where AGB was high. We will enlarge the sample size to improve the simulation 562 563 accuracy in future research. 2) Although the grassland height information could help improve the estimation accuracy of 564 grassland AGB, it was still challenging to obtain grassland height information from UAV RGB images in a large area. (Zhang et al., 2022a; Lussem et al., 2019; Viljanen et al., 2018). In future work, aAn affordable DJI Zensil L1 Lidar UAV 565 566 will be introduced to invert the height of the grasslandt grass heights in future work. Thus, in the next step, we will consider using the affordable DJI Zensil L1 Lidar UAV to obtain grassland height information to improve the AGB estimation 567 568 capability.-2) At the pixel scale, limited by the estimation accuracy of AGB from UAV, there was also some underestimation

569 in the high biomass area. Although the MODIS index closest to the sampling time was chosen for the construction/validation 570 of the AGB estimation model, there iswas still a time difference between the measured samples and the MODIS indices, 571 which may might lead to estimation errors. In addition, the NDVI saturation problem was not considered in this study, which 572 might affect the AGB estimation accuracy of QTP (Tucker, 1979a; Gao et al., 2000; Mutanga and Skidmore, 2004; Tucker, 573 1979b). In the next step, we will continue to collect samples with high biomass and try to correct the NDVI saturation 574 problem to optimize the simulation accuracy of the data set. 3)3) During 2015-2016, we our study washed just 575 beginningstarted, - just started using UAVs to monitor the health of the grassland, and the appropriate suitable camera 576 parameters and methods were still under being exploration explored. Therefore As a result, some photos with abnormal white 577 balance were obtained, reducing the accuracy of AGB estimationreduced the accuracy of AGB estimation at the photo scale 578 (Figure 1213A5). 4) We only collected grassland AGB only in during the peak season of vegetation growth growing season, 579 and whether the applicability of the proposed method applies to other growing growing seasons- needs further studyremains-580 to be further investigated. 5) During the modeling process, due to the limited positioning accuracy, only the center points of the flight route path were used to find the matching corresponding MODIS pixels-due to the limited positioning accuracy. 581 582 Moreover, although the UAV images from in GRID or RECTANGLE mode could cover most areas of a MODIS pixel, 583 full pixel coverage was was still not achieved. Therefore, we will gradually seale upupscale to MODIS pixels by combining 584 UAVs with Sentinel-2 or Landsat images.

- 585
- 586

587 5. Data availability

588 The dataset is available from the National Tibetan Plateau/Third Pole Environment Data Center 589 (-athttps://doi.org/10.11888/Terre.tpdc.272587). The dataset contains 20 years of AGB spatial data of the QTP with a resolution of 250 m and is stored in TIFF format. The name of the file is "AGB yyyy.tif", where yyyy represents the year. 590 591 For example, AGB 2000.tif represents this TIFF file describing the alpine grassland AGB condition of QTP in 20052000. 592 The data can be readily imported into standard geographical information system software (e.g., ArcGIS) or accessed 593 programmatically (e.g., MATLAB, Python).

594 6. Conclusion

595 This study In this study, a new a new -AGB dataset for alpine grasslands on the QTP was calculated based on traditional

- 596 ground sampling, UAV photography, and MODIS imagery. presents a new gridded dataset of alpine grassland AGB over
 - 597 the OTP based on traditional ground sampling, UAV photographing, and MODIS images. The uniqueness of this dataset is
 - 598 the use of that when obtaining ground verification data, the UAVs is used as a spatial scale-matching bridge between

599 traditional local measurement samples and satellite pixels. The study confirmed that the UAV images could be used for AGB estimation at the quadrat /pixel scale, with R^2 of 0.73/0.83 and RMSE of 44.23/34.13 g/m², respectively. At the pixel scale, 600 the correlation between AGB estimated by UAV and MODIS vegetation index was higher than that of the traditional 601 602 sampling method (0.88 vs. 0.53-). At the pixel scale, the AGB estimated by UAV was more correlated with the MODIS-603 vegetation indices than the traditional ground sampling method (0.88 VS 0.53), and Moreover, the spatial scale matching of 604 the dependent and the independent variables was achieved during the model-modelingconstruction. In addition, we 605 performed a an independent cross-year validation of the pixel-scale AGB estimation model to confirm the robustness of the model and the accuracy of this dataset In addition, the constructed pixel scale model has been independently cross-validated 606 607 over many years (2015-2019), which confirmed the robustness of the model and ensured the accuracy of this dataset. Availability The availability of the new dataset is helpful in many applications. First, this dataset provides reliable regional 608 609 data for estimating grassland productivity, carbon storage, ecological environment-carrying capacity, and ecological service 610 functions (such as feed for grazing livestock) -onof the OTP. Second, the dataset can be used to understand the mechanisms 611 of environmental processes, such as hydrological cycle processes, soil erosion and degradation, and carbon cycle processes 612 in the QTP. In addition, this dataset can be used as input or validation parameters for various ecological models to 613 understand the response mechanism of the OTP to global climate change.

614 7. Author contributions

HZ contributed to the study conceptualization, methodology, funding acquisition, and the original draft of the manuscript. ZT, BW, and HK contributed to resources and formal analysis. QY and YS contributed to data collection and manuscript review. BM, ML, and JC contributed to the methodology and reviewed the manuscript. YL and JZ participated in reviewing and editing the manuscript. SN contributed to the data collection and review of the manuscript. SY contributed to the study conceptualization, funding acquisition, and manuscript review. All authorship have read and approved the manuscript.

620 8. Competing interests

621 The authors declare that they have no conflict of interest.

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628 Appendix









631

632 Figure A2. The importance values for each independent variable (a) and the R² results of the different number of input variables

633 at the quadrat scale.



Figure A3. The importance values for each independent variable (a) and the R² results of the different number of input variables

at the pixel scale.

9-124393.J	PG 10-99489.JPG	10-108960.JPG	10-110458.JPG	10-110650.JPG
11-136340.	JPG 11-143170.JPG	12-99734.JPG	12-108669.JPG	12-109042.JPG
13-124317.	JPG 13-124406.JPG	14-98088.JPG	14-102399.JPG	14-124108.JPG
15-124210.	JPG 15-124383.JPG	15-139325.JPG	15-146818.JPG	16-99687.JPG

Figure A4. Examples of 20-meter-high UAV images with different non-vegetation background information.



641 Figure A5. An example of a set of GIRD photos with abnormal white balance in 2015.

644 Table A1. Combined grassland types

New grassland type	Original grassland type
Alpine meadow	Alpine meadow, Lowland meadow, Montane meadow,
Alpine steppe	Temperate steppe, Alpine steppe, Alpine meadow steppe
Spare grassland	Temperate steppe desert, Alpine desert

646 Table A2. Features of DJI Phantom 3 Pro

	Features	Description
4	Sensor	1/23-inch; Effective-pixel: 12-megapixel
	Filed of view	FOV 94° 20 mm
	Aperture	f/2.8
	Shooting speed	Electronic shutter: 8-1/8000 s
	Photo size	4000×3000
	Flight time	~25 min
DJI Phantom 3 Pro	Image format	JPEG
	Hovering accuracy	± 0.5 m vertically; ± 1.5 m horizontally
	Weight	1280 g

Table A3: Details of the independent variables for quadrat-scale AGB estimation

Acronym	Index name	Formula	Reference
GRVI	Green Red Vegetation Index	(G-R)/(G+R)	(Tucker, 1979a)
EXG	Excess Green Vegetation Index	2G-R-B	(Woebbecke et al., 1995)
GLA	Green leaf area	(2G-R-B)/(2G+R+B)	(Louhaichi et al.)
MGRVI	Modified Green Blue Vegetation Index	(G2-R2)/(G2+R2)	(Bendig et al., 2015)
RGBVI	Red Green Blue Vegetation Index	(G2-B*R)/(G2+B*R)	(Bendig et al., 2015)
EXB	Excess Blue Vegetation Index	(1.4*B-G)/(G+R+B)	(Maimaitijiang et al., 2019)
NDI	Normalized difference index	(R-G)/(R+G)	(Woebbecke et al., 1993)
EXR	Excess Red Vegetation Index	1.4*R-B	(Meyer and Neto, 2008)
EXGR	Excess Green minus Excess Red index	ExG-ExR	(Meyer and Neto, 2008)
RRATIO	Red Ratio	R/(R+B+G)	(Woebbecke et al., 1995)
BRATIO	Blue Ratio	B/(R+B+G)	(Woebbecke et al., 1995)
GRATIO	Green Ratio	G/(R+B+G)	(Woebbecke et al., 1995)
VARI	Visible Atmospherically Resistance Index	(G - R)/(G + R - B)	(Gitelson et al., 2002)
NRBI	Normalized Red Blue Index	(R-B)/(R+B)	(Michez et al., 2016)
NGBI	Normalized Green Blue Index	(G-B)/(G+B)	(Michez et al., 2016)
VEG	Vegetative index	G/(RaB(1-a)),where a=0.667	(Hague et al., 2006)
WI	Woebbecke Index	(G-B)/(R-G)	(Woebbecke et al., 1995)
CIVE	Color Index of Vegetation	0.441R –	(Kataoka et al., 2003)
		0.881G+0.385B+18.78745	
COM	Combination Vegetative index	0.25ExG+0.3ExGR+0.33CIVE	(Guijarro et al., 2011)
		+0.12VEG	
TGI	Triangular Greenness Index	G-0.39R-0.61B	(Hunt et al., 2014; Michez et
			al., 2018)
RGBVI	Red Green Blue Vegetation Index	(G2-B*R)/(G2+B*R)	(Bendig et al., 2015)
GRRI	Green Red Ratio Index	G/R	(Maimaitijiang et al., 2019)
GBRI	Green Blue Ratio Index	G/B	(Maimaitijiang et al., 2019)
RBRI	Red Blue Ratio Index	R/B	(Maimaitijiang et al., 2019)
BRRI	Blue Red Ratio Index	B/R	(Jibo et al., 2018)
BGRI	Blue Green Ratio Index	B/G	(Jibo et al., 2018)
RGRI	Red Green Ratio Index	R/G	(Jibo et al., 2018)
INT	Color Intensity Index	(R+B+G)/3	(Ahmad and Reid, 1996)
MVARI	Modified VARI	(G-B)/(G+R-B)	(Cen et al., 2019)
IPCA	Principal Component Analysis Index	$0.994 \times R-B + 0.961 \times G-B +$	(Saberioon et al., 2014)
		0.914× G-R	

6	55	Table A3:	Details of the inde	pendent variables	for quadrat-scale	e AGB estimation	(continued)
							· · · · · · · · · · · · · · · · · · ·

Acronym	Index name	Formula	Reference
R	An average value of R channel of the quadrat-scale UAV image		
G	An average value of G channel of the quadrat-scale UAV image		
В	An average value of B channel of the quadrat-scale UAV image		
Н	An average value of H channel of the quadrat-scale image in HSV color space		
S	An average value of S channel of the quadrat-scale image in HSV color space		
V	An average value of V channel of the quadrat-scale image in HSV color space		
FVC	Fractional Vegetion Cover		
EGI	Extra Geen Index	EGI=2G-R-B	(71)
GI	Green Index	GI=9×(H×3.14159/180) +3×S+V	(Zhang et al., 2022a)
HOC_ <i>i</i> _C ORR	The histogram correlation coefficient between the i band and the black reference histogram, where the i represents the three bands of RGB	$corr \frac{\sum_{I} (H_{1}(I) - \overline{H}_{1})(H_{2}(I) - \overline{H}_{2})}{\sqrt{\sum_{I} (H_{1}(I) - \overline{H}_{1})^{2} \sum_{I} (H_{2}(I) - \overline{H}_{2})^{2}}}$	
HOC_i_ INTERSE C	The histogram intersection coefficient between the i band and the black reference histogram, where the i represents the three bands of RGB	$intersec = \sum_{I} \min \left(H_1(I), H_2(I) \right)$	
HOC_i_ BHATTA	The histogram Bhattacharyya distance coefficient between the i band and the black reference histogram, where the i represents the three bands of RGB	$bhatta = \sum_{I} \min \left(H_1(I), H_2(I) \right)$	
HOC_i_C HIS	The histogram correlation coefficient between the i band and the black reference histogram, where the i represents the three bands of RGB.	chis = $\sum_{I} \frac{(H_1(I) - H_2(I))^2}{H_1(I)}$	

657Table A4: Regression analysis for AGB estimation models at quadrat and pixel scales658for AGB estimation models at quadrat and pixel scales

Model name	Coefficient	Value	Standard Error	t-Value	p-value
2019_Quadrat-scale	Slope	0.67	0.016	42.58	9.05e-194
	Intercept	20.10	1.49	13.59	5.96e-37
2019_Pixel_scale	Slope	0.84	0.03	31.59	2.75e-73
	Intercept	23.20	4.04	5.74	4.24e-8
2018_Pixel_scale	Slope	0.73	0.02	45.81	8.28e-157
	Intercept	20.43	2.74	7.46	6.01e-13
2017_Pixel_scale	Slope	0.75	0.01	59.13	1.98e-260
	Intercept	13.89	2.04	6.82	2.19e-11
2016_Pixel_scale	Slope	0.94	0.02	40.45	4.69e-157
	Intercept	2.48	3.75	0.66	0.03
2015_Pixel_scale	Slope	0.82	0.04	18.88	2.59e-47
	Intercept	9.50	5.25	1.81	0.04

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660 Table A5: List of abbreviations of eco-geographical regions and the mean AGB of the QTP

Abbreviation	<u>Full name</u>
IB1	Golog-Nagqu high-cold shrub-meadow zone
IIAB1	Western Sichuan-eastern Tibet montane coniferous forest zone
<u>IC1</u>	Southern Qinghai high-cold meadow steppe zone
<u>IC2</u>	Qiangtang high-cold steppe zone
<u>ID1</u>	Kunlun high-cold desert zone
<u>IIC1</u>	Southern Tibet montane shrub-steppe zone
<u>IIC2</u>	Eastern Qinghai-Qilian montane steppe zone
<u>IID1</u>	Nagri montane desert-steppe and desert zone
<u>IID2</u>	Qaidam montane desert zone
<u>IID3</u>	Northern slopes of Kunlun montane desert zone
<u>OA1</u>	Southern slopes of Himalaya montane evergreen broad-leaved forest zone

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