Impacts of the proposal of the CNG2020 strategy on aircraft emissions of China-Foreign routes

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Abstract. Aviation emission accounting is the key to establishing market measures to control aviation pollutant emissions. Based on the Fuel Percentage Method (FPM), this paper applies the Improved BFFM2-FOA-FPM method to calculate the emissions of six pollutants (CO_2, CO, HC, NOx, SO_2, and PM2.5) between Chinese and foreign cities from 2014 to 2019, including CCD (Climb/Cruise/Descent) emissions and LTO (Landing and Take-off) emissions. The error rate between the

10 calculated results and the official data is about 2.75%. The results show that the emissions of six pollutants have changed before and after the proposal of the "Carbon Neutral Growth 2020 " strategy (CNG2020 strategy). Although the total amount has increased, the average emission per ton-kilometer of CO2, CO, HC, NOx, SO2, and PM2.5 has decreased by 17.77%, 17.26%, 25.15%, 14.32%, 17.77%, and 16.35%, respectively. The results of this paper can provide data basis and method reference for implementing the CNG2020 strategy and realizing global carbon emission reduction goals. The dataset is

15 available from https://doi.org/10.6084/m9.figshare.20071751.v1 (Cui., 2022).

1. Introduction

Air transportation between China and foreign countries has developed dramatically in recent years. The turnover of China's foreign routes totalled 46.37 billion ton-kilometres in 2019, an increase of 93.13% over 2014; China's foreign routes completed 74.25 million passengers, a rise of 135.34% over 2014; China's foreign routes met 2.42 million tons of goods and

20 mail, an increase of 44.0% over 2014 (CAAC, 2022). The significant growth of the air transport scale intensifies the environmental changes caused by air pollutant emissions, making the aviation industry one of the top ten greenhouse gas emission industries globally (Turgut et al., 2017). Aviation pollutants mainly come from NOx, CO2, HC, SO2, PM, and aviation-induced cloudiness (AIC) and other pollutants discharged after mixed combustion of aviation fuel and air during an eruption, which affects air quality and causes the temperature chamber effect. Therefore, aviation pollutant emission has

25 attracted more and more attention from the global community (Zheng et al., 2019; Shan et al., 2021).

To solve the problem of aviation pollutant emission, in October 2016, the 39th General Assembly of ICAO adopted two critical documents: ICAO comprehensive statement on sustainable policies and practices of environmental protection - climate change and ICAO complete statement on sustainable policies and procedures of environmental protection - global market measures mechanism, It aims to achieve the zero-carbon emission growth goal of the international aviation industry

from 2020 through the phased implementation of the "Carbon Neutral Growth 2020" strategy (CNG2020 strategy), from 2021 to 2035. The CNG2020 strategy had been implemented from 2021. The period from 2021 to 2023 is the pilot phase, and the period from 2024 to 2026 is the first phase. All countries can voluntarily participate in the plan in the pilot phase and the first phase. The second phase is from 2027 to 2035. In the second phase, most countries must participate in the plan, and the least developed countries, island and landlocked developing countries can participate voluntarily. ICAO has predicted the impacts of the CNG2020 strategy on the whole airline industry. The estimated quantity to be offset to achieve carbon-neutral growth from 2020 would be of the order of 142–174 million tons of CO_2 in 2025; and 443–596 million tons of CO_2 in 2035, with these ranges being determined by the definitions of 9 scenarios for CO2 trend assessment from the most optimistic scenario to the less optimistic one. Therefore, the CNG2020 strategy may have substantial impacts on the global airline industry (Cui and Li, 2018).

Aiming at the accounting method of gaseous pollutant emission of aero-engine during standard landing and take-off cycle, ICAO has successively developed simple method, advanced method, and complex method according to different calculation methods and data requirements since the 1970s (Kesgin, 20006; Altuntas, 2014; Winther et al., 2015; Cokorilo, 2016; Xu et al., 2020). The direct use of model reference value in ICAO's simple methods will uncertainty the accounting results. On the other hand, the advanced and complex methods further improved have highly accurate results. However, they have the limitations of high data requirements, complex implementation, high research cost, and are unsuitable for mass calculation. Therefore, the development of relevant research is relatively slow. Based on the ICAO calculation system, the U.S. Environmental Protection Agency (EPA) puts forward the EPA (Environmental Protection Agency) method combined with the actual situation (U.S. Environmental Protection Agency, 1999; Unal et al., 2005; Zhou et al., 2019; Baxter et al., 2020). Although the EPA method considers meteorological conditions and establishes the relationship between meteorological conditions and aero-engine emissions, it is helpful to understand the relationship between meteorological conditions and emissions. However, the calculation of emission inventory during the aircraft LTO cycle by the EPA method is the same as that of the ICAO simple B method. Therefore, there may be no significant difference between the two calculation results for a single aircraft. As a result, and because ICAO recommended method is applied by many organizations and projects, ICAO recommended method is the most effective method for LTO cycle pollutant emission evaluation (Kurniawan and Khard, 2011). Meanwhile, the European Environment Agency (EEA) has established the EMEP (European monitoring and evaluation program) cooperative action framework (Civil and military aviation., 2014; Park et al., 2014; Pereira et al., 2014). This method is an accounting method of pollutant emission of aircraft during the whole flight based on fuel statistical data. Still, it focuses on analysing the emission characteristics of aero-engine from the fuel perspective and ignores the differences between engine types. Furthermore, ICAO has further improved the calculation method and proposed the ICAO carbon emission calculator, which can estimate the aviation emission per unit passenger based on the data of various aircraft types (Wasiuk et al., 2016; Liu et al., 2019; ICAO, 2021).

However, there are some drawbacks to the approach provided by ICAO. First, the distance difference is not enough. For example, according to the VariFlight (VariFlight, 2022), A320-214 flew between 360 km and 3,649 km on domestic routes

in China in 2018, exceeding the methodology provided by the ICAO. Second, there is no distinction between specific aircraft. ICAO's calculation method only considers large sequences and does not consider differences between subsequence. For example, the A320 family has many families, such as the A320-100 and A320-200, with different engine types, which may lead to a significant difference in the carbon emissions of the two aircraft (Cui et al., 2022a). Third, various pollutants cannot be calculated at the same time. Aming at these problemes, Cui et al. built Improved BFFM2-FOA-FPM method and the ICAO method to calculate the emissions of CO_2, CO, HC, NOx, SO_2, and PM2.5. The Improved BFFM2-FOA-FPM method can divide each route into multiple distance segments every 500 kilometres, then calculate each aircraft's emission intensity in each distance segment (Cui et al., 2022b).

However, there is little study to analyse the emission difference before and after the construction of CNG2020 strategy. This study can make up for this deficiency. Generally, the entire flight process consists of seven steps: Engine Starting, Taxiing, Taking Off, Climbing, Cruising, Descending, and Landing (Cui, 2019). It is usually divided into the Landing and Take-Off (LTO) cycle and the Climbing/Cruising/Descending (CCD) stage. This paper will calculate the CCD emissions and LTO emissions of six pollutants (CO_2, CO, HC, NOx, SO_2, and PM2.5) from China-Foreign routes to analyse the impacts of the proposal of CNG2020 strategy. Since the COVID-19 began to rage around the world in 2020, the data of 2020 and 2021 cannot be used as a reference. Therefore, 2017-2019 is selected as the cycle after the proposal of the CNG2020 strategy. For comparison, 2014-2016 is selected as the cycle before the proposal, so the whole process of the selected data is 2014-2019. We analyse the impact of CNG20202020 strategy on China's external routes by comparing and analysing the emission difference and unit turnover emission difference in these two time periods.

2. Results

The original data were collected from VariFlight.com (VariFlight, 2022), and we compiled the data. The data on aircraft type, flying time, flying distance, transfer flight, and airlines are from VariFlight.com (VariFlight, 2022). The data on the engines are from ICAO Aircraft Engine Emissions Databank (EASA, 2022).

Statistical characteristics of China-Foreign routes from 2014 to 2019. This article collects the information of China-Foreign routes during 2014-2019, and the detailed statistical characteristics are shown in Figure 1. The China-Foreign routes from 2014 to 2019 involved 121 aircraft types (see details in Supplementary Tables). It can be seen from Figure 1(a) that China-Foreign exchanges have has developed very rapidly. The number of routes has increased from 367 in 2014 to 521 in 2019, and the number of airlines has increased from 76 in 2014 to 120 in 2019. These routes cover about 80 countries, showing that China has fixed flights globally with more than 80 countries. Figure 1(b) shows that the farthest route in 2014 is Guangzhou-New York, 12,583 km. However, in 2015 and 2016, the most distant route became Shanghai Pudong-Mexico City, and the distance is 12,908 km. In 2017, the farthest route was Beijing-Havana, and the distance was 13,091km, but in 2018, the most distant route became Guangzhou-Mexico City (14,115 km), and it returned to Beijing-Havana in 2019. From this result, the farthest place between China and foreign routes is in Central and North America. The nearest routes in 2014,

2015, 2017, and 2019 are Yanji-Vladivostok, whose distance is 224 km, while the nearest one is Nanning-Hanoi (289 km) in other years.

The accuracy of the method. According to the data released by The Civil Aviation Administration of China, in 2019, 46.374 billion tons-km of transportation was completed between China and foreign routes(CAAC, 2022). The fuel consumption per unit turnover of China's domestic and foreign routes is about 0.285 kg/ton-km, but the international routes can float 5% due to the large aircraft used. The fuel consumption of unit turnover is 0.2993 kg/ton-km. Using this standard, multiplied by the carbon emission coefficient per unit fuel consumption (3.157kg/kg), the carbon dioxide emissions of the China-Foreign routes in 2019 would be 43,818,333.50 tons. Therefore, the carbon dioxide emissions calculated in this paper are 45,021,648.34 tons, with a 2.75% error rate. Similarly, the error in 2018 is 5.21%. Considering the statistical data of various airlines may also appear errors, so the calculation results of this paper are very accurate.

The impact of CNG2020 on aircraft configuration. This article makes detailed statistics on the configuration of aircraft types involved in Chinese and foreign routes from 2014 to 2019. The top three aircraft types in each year are shown in Figure 2. Among them, the total frequency of 737-800 and 320-214 aircraft types has consistently ranked first and second respectively in these six years, and the full frequency of 737-800 aircraft types has increased from 1510 in 2014 to 2754 in 2019, The total frequency of 320-214 aircraft increased from 868 in 2014 to 1232 in 2019. As shown in Figure 2, before the CNG2020 strategy was put forward, the aircraft configuration changed little. For example, the third aircraft in 2014-2015 was 330-243E, and the third model in 2016 was 320-232. After the CNG2020 strategy is put forward, the aircraft configuration will be fine-tuned every year. For example, the third-ranked aircraft in 2017 was 321-231, the third-ranked aircraft in 2018 was 320-232, and the third-ranked aircraft in 2019 was 330-343E.

The emission intensity of the aircraft in the CCD stage. As mentioned earlier, different from the method of ICAO, we segment each route according to a section of 500 kilometres. Therefore, all routes are divided into 29 distance sections: 0-500km, 501-1000km, 1001-1500km, 1501-2000km, 2001-2500km, 2501-3000km, 3501-3500km, 3501-4000km, 4001-4500km, 4501-5000km, 5001-5500km, 5501-6000km, 6501-6500km, 7001-7500km, 7501-8000km, 8001-8500km, 8501-9000km, 9001-9500km, 9501-10000km, 10001-10500km, 10501-11000km 11001-11500km, 11501-12000km, 12001-12500km, 12501-13000km and 13001-13500km, 13501-14000km and 14001-14500km. In addition, we also considered the differences between subseries, such as 320-214 and 320-232. Then, we get the aircrafts' emission intensity of the six pollutions from 2014 to 2019 based on the Modified Fuel Percentage Method (MFPM) (see the Supplementary Tables). Since different models apply to different distances, we divide the total distance segment into 0-4000 km, 4001-9500 km, and 9501-14500 km. In the 0-4000km section, 320-214, 320-232, 737-700 and 737-800 cover almost all distance sections. 320-214 and 320-232 are subseries of the A320 series, and 737-700 and 737-800 are B737 series. Therefore, this comparison highlights the difference between this study and the ICAO method. In the 4001-9500km section, 330-243E and 330-343E cover most of the distance sections, so we will compare them. Sections 9501-14500 km, 777-300ER, and 787-8 cover the longest distance, so that we will compare them. We summarize the average carbon emission intensity and show the detailed

130 results in Figure 3. It should be noted that the data of 777-300ER and 787-8 are lost when the distance is greater than 13000km, so this part of the comparison cannot be made.

As shown in Figure 3(a), under the A320 series, 320-214 and 320-232 have similar carbon emission intensity in a 0-500km distance segment. However, in other distance segments, the carbon emission intensity of 320-214 is higher than that of 320-232. Therefore, 320-232 has a better performance in carbon emissions per kilometre, providing more references for airlines

135 in arranging aircraft types. Under the B737 series, 737-700 and 737-800 have similar carbon emission intensity in a distance segment of 0-1000km distance segment. In addition, the carbon emission intensity of 737-700 is higher than that of 737-800 at 3500-4000km, but 737-800 has a more significant intensity than 737-700 at other distances. Therefore, the 737-700 is better than that of the 737-800 at 0-4000km. Figure 3(b) shows that 330-343E has a lower carbon emission intensity than 330-243E in the 4000-9500km distance segment. These two aircraft have similar intensity at 4500-5000km, but the

140 intensities of 330-243E are larger than 330-343E in other distance segments. In the distance segment 9500-13000km, 787-8's carbon emission intensities are smaller than 777-300ER, so 787-8 has a better overall performance than 777-300ER in this distance segment.

The difference between the A320 series and the A330 series has little relationship with the engines, as the engines of the sub-series are the same. The engines of 320-214 and 320-232 are CFM56-5/V2500, and those of 330-343E and 330-243E are

145 PW4000/Trent 700/CF6-80E1. Therefore, their difference in carbon emission intensity may be related to airline route arrangement and actual flight operation. However, the engines of the other two pairs of aircraft are different. The engines of 737-700 are CFM56-7B20/CFM56-7B24, while those of 737-800 are CFM56-7B24/CFM56-7B27. The engines of 777-300ER are PW4090/Trent 895/GE90-94B, and those of 787-8 are Trent 1000/GEnx-1B. The engines of 737-800 and 777-300ER consume more fuel per kilometre so the engine difference may lead to the emission intensity of these aircraft.

150 **The impacts of CNG2020 on the overall emissions**. The primary emissions include CO_2, CO, HC, NOx, SO_2, and PM2.5. The CNG2020 strategy was proposed in the second half of 2016. Therefore, the data are divided into two groups for comparison: before the CNG2020 was proposed, i.e., 2014-2016, and after the CNG2020 strategy was proposed, i.e., 2017-2019. The overall emissions of six pollutants in the two groups were averaged and compared. After the proposal of the CNG2020 strategy, the overall emission of six pollutants is still increasing, but the increase rate is no more than 27%.

155 Taking carbon dioxide as an example, the average overall emission in 2017-2019 was 44787085.866 tons, increasing 21.50% (36,861,009.60) over the average overall emission in 2014-2016. The average overall emission of CO, HC, NOx, SO_2, and PM2.5 increased 18.36%, 9.62%, 26.73%, 21.50%, and 20.88%, respectively.

And we can also find that CO_2 accounts for the most significant proportion of various pollutants, which is much higher than other pollutants. Taking 2014-2016 as an example, the average overall emission of CO_2 was 36,861,009.60 tons, far

160 exceeding the second-ranked NOx (370,431.58tons). In addition to CO_2, the emissions of CO and NOx are also relatively large. Nitrogen oxides account for about 1.05% of the overall emissions, while carbon monoxide accounts for about 0.63%. In addition, Figures 4 and 5 more accurately show the changes in the overall emissions of six pollutants before and after the proposal of the CNG2020 strategy. We can find that the growth rate of the overall emissions of pollutants in 2017-2019 is

generally less than that in 2014-2016. Taking CO_2 as an example, the average annual growth percentage in 2014-2016 is 27.75%, and that in 2017-2019 is 1.19%. And the growth rate of CO and HC in 2017-2018 and the growth rate of CO_2 in 2018-2019 are negative, which shows that the growth rate of aviation pollutant emissions has slowed down and has a negative growth trend under the influence of the CNG2020 strategy. Since the CNG2020 strategy was put forward in 2016, airlines have made efforts to reduce the overall emissions of Chinese and foreign routes, which is a very positive signal that airlines are trying to achieve carbon-neutral growth in the aviation industry.

In addition, we compared the changes in unit turnover emissions of six pollutants before and after CNG2020. According to the relevant report data of the Civil Aviation Administration of China, the total transportation turnover in 2014-2019 was 24.011 billion tons-km, 29.111 billion tons-km, 34.06 billion tons-km, 38.848 billion tons-km, 45.319 billion tons-km, and 46.377 billion tons-km, respectively. The unit turnover emissions of six gases in 2014-2019 can be obtained. After averaging the first three years and the next three years, the results are shown in Figure 6 (1 represents 2014-2016 and 2 represents 2017-2019). Under the influence of the CNG2020 strategy, the unit turnover emissions of six gases have decreased. Taking CO2 with the largest emissions as an example, the carbon emissions per ton-kilometre in 2014-2019 were 1.17E-03 tons, 1.25E-03tons, 1.35E-03tons, 1.13E-03tons, 1.00E-03tons, and 9.71E-04 tons, respectively. The average carbon emission per ton-kilometre before the proposal of the CNG2020 strategy (2014-2016) was 1.26E-03 tons, and after the proposal of the CNG2020 strategy (2017-2019), the average carbon emission per ton-kilometre was1.03E-03 tons, a decrease of 17.77%. Similarly, CO, HC, NOx, SO_2, and PM2.5. The decline rates were 17.26%, 25.15%, 14.32%, 17.77% and 16.35%respectively.

The impacts of CNG2020 on the emissions of the routes. First, we will discuss the impact of the CNG2020 strategy on the average emissions of routes. The data are still divided into two groups for comparison. The average overall emission change of the two groups of routes can be calculated. During 2014-2016, The average overall emission of CO_2, CO, HC, NOx, SO_2, and PM2.5 of the routes is 84783.78 tons, 543.60 tons, 62.06 tons, 852.40 tons, 103.93 tons, and 7.45 tons, respectively. From 2017 to 2019, The average overall emission of CO_2, CO, HC, NOx, SO_2, and PM2.5 of the routes is 91185.39 tons, 587.56 tons, 60.17 tons, 955.15 tons, 111.78 tons, and 7.95 tons, respectively. Therefore, except for HC, the overall decline was 3.03%, and the average overall emissions of other pollutant routes in 2017-2019 increased relatively compared with 2014-2016. For example, CO2 increases by 7.55%, and NOx rises by 12.05%, but the increasing percentage is not more than 13%. Further, analyse the annual variation range of the average overall emission of each pollutant route in the two groups in detail. As shown in Figures 7 and 8, the growth range of the average overall emission of the routes in 2017-2019 is generally less than that in 2014-2016. Taking CO_2, which accounts for the most significant proportion of pollutants, as an example, the average annual growth percentage in 2014-2016 is 10.67%. In 2017-2019, it was negative growth (-2.46%). And during 2018-2019, the average overall emissions of all pollutants from all routes increased negatively, indicating that all routes have reduced aviation pollutants after the CNG2020 strategy was put forward.

Meanwhile, the average unit turnover emissions of each route before and after CNG2020 are shown in Figure 9 (1 represents 2014-2016 and 2 represents 2017-2019). The average emissions per ton-kilometre of CO_2, CO, HC, NOx, SO_2, and PM2.5

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routes before the CNG2020 strategy was proposed (2014-2016) are 1.92E-03 1.60E-05 tons, 1.70E-06 tons and 1.83E-05tons,2.40E-06 tons, and 2.00E-07 tons. After the CNG2020 strategy was put forward (2017-2019), The average emissions per ton-kilometre of CO_2, CO, HC, NOx, SO_2 and PM2.5 of the route were 1.56E-03 tons, 1.15E-05 tons, 1.10E-06 tons, 1.16E-05 tons, 1.90E-06 tons and 1.00E-07 tons, down 18.78%, 28.29%, 33.31%, 36.54%, 30.09% and 18.79% respectively.

Therefore, we further compare the average carbon emissions per unit turnover of the two groups of data hot routes, select the routes before and after the strategy, and study the impact of the CNG2020 strategy on them. Through analysis, 291 routes are available and can be defined as popular routes. As shown in Figure 10 (a), we summarize the five routes with the largest increase in average unit turnover carbon emissions and the five routes with the largest reduction before and after the strategy is proposed. The red is the route with the largest increase, and the green is the route with the largest reduction. The former can indicate that the carbon emission per unit turnover increased rapidly after the CNG2020 strategy was proposed. The latter can indicate that the carbon emission per unit turnover decreased rapidly after the CNG2020 strategy was proposed. Among the top five routes with the largest emission reduction, four are associated with Guangzhou, China. It shows that Guangzhou, as one of China's air transport centre cities, has achieved good results in carbon emission reduction. Figure 10 (b) shows the five routes with the smallest change in average carbon emissions per unit turnover before and after the proposal of the CNG2020 strategy. Three of them are routes between China and South Korea. In addition, two of the five routes are from Shanghai and Beijing. Compared with Guangzhou, Shanghai, and Beijing, as important aviation hubs in China, they are not sensitive to the CNG2020 strategy. It is worth noting that only 14 of the 291 routes have increased their average carbon emissions per unit turnover after the proposal of the CNG2020 strategy, and the emissions of the remaining 277 routes have decreased after the proposal of the CNG2020 strategy. Moreover, among the 14 routes with increased emissions, ten routes are less than 5000km, indicating that under the CNG2020 strategy, airlines do not control the carbon emissions of short-haul routes enough.

The impacts of CNG2020 on the emissions of the airlines. First, we will discuss the impact of the CNG2020 strategy on the average emissions of airlines. By comparing the changes of average overall emissions of airlines before and after the CNG2020 strategy, it can be concluded that airlines' average overall HC emissions decreased by 8.24% in 2017-2019 compared with 2014-2016. But the average overall emissions of CO_2, CO, NOx, SO_2, and PM2.5 increased by 1.93%, 2.24%, 6.17%, 1.93%, and 0.98%, respectively, and the growth rate was no more than 7%. Further analyse the annual variation range of the average overall emission of each pollutant airline in the two groups, as shown in Figures 11 and 12. Unlike the comparison results of routes, the growth rate of average overall emissions of airlines in 2017-2019 is generally more significant than that in 2014-2016. Taking CO_2, which accounts for an essential proportion of pollutants, as an example, the annual average growth percentage is 0.12% in 2014-2016 and 4.48% in 2017-2019. The main reason is that the average overall emissions of airlines increased significantly from 2017 to 2018. The number of airlines in 2018 was 107, 18 less than that in 2017, and the number of routes was also reduced by 9. However, the average overall carbon emissions of airlines increased by 20.45%. Therefore, the overall average total emissions in 2017-2019 were more significant than 2014-2016.

In addition, the average unit turnover emissions of airlines before and after CNG2020 are shown in Figure 13 (1 represents 2014-2016 and 2 represents 2017-2019). Airlines' average emissions per ton-kilometre of CO_2, CO, HC, NOx, SO_2, and PM2.5 before the CNG2020 strategy was proposed (2014-2016) are 1.69E-03 tons 1.20E-05 tons, 1.30E-06 tons, 1.41E-05 tons, 2.10E-06 tons, and 2.00E-07 tons, respectively. The average emissions per ton-kilometre of CO_2, CO, HC, NOx, SO_2 and PM2.5 of airlines after the proposal of CNG2020 strategy (2017-2019) were 1.48E-03 tons, 1.03E-05 tons, 1.00E-06 tons, 1.14E-05 tons, 1.80E-06 tons and 1.00E-07 tons, down 12.42%, 13.96%, 24.21%, 18.98%, 12.42% and 31.44% respectively.

Therefore, we further compare airlines' average carbon emissions per unit turnover in the two data hot spots. Through analysis, 63 airlines are available, which can be defined as popular airlines. We summarize the three airlines with the most significant increase in average unit turnover carbon emissions and the three airlines with the most significant decrease in carbon emissions of popular airlines before and after the strategy is proposed, as shown in Figure 14. The red ones are the airlines with the most significant increase, and the green ones are the airlines with the most significant decrease. It is worth noting that among the 63 airlines, only four airlines have increased their carbon emissions per unit turnover after the CNG2020 strategy was proposed, namely, Asian Air, Lucky Air, Eastar Jet, and Pakistan International Air. The rest have decreased, indicating that most airlines have better controlled their carbon emissions after the CNG2020 strategy was proposed.

3. Discussion

In this study, we discuss the impacts of the proposal of the CNG2020 strategy on the aircraft emissions of China-Foreign routes during 2014-2019. The emissions from 2014 to 2016 are considered the data before the CNG2020 strategy is put forward, and those from 2017 to 2019 are the data after the CNG2020 strategy is put forward. We collect the flight information (including aircraft types, flight frequency, airline, flying distance, flying time, etc.) of all the international routes between China and foreign countries. Then we calculate the overall emissions for each route and airline containing CO_2, CO, HC, NOx, SO_2, and PM2.5. The overall emissions include the CCD emissions and LTO emissions, the former is calculated through the Modified BFFM2-FOA-FPM method, and the latter is calculated based on the ICAO standard method. Accounting for emissions can better summarize the impact of the CNG2020 strategy on aircraft activities and provide data and method references to implement the CNG2020 strategy better.

We get some important results. First, after the proposal of the CNG2020 strategy, the overall emission of six pollutants is still increasing, but the increase rate is no more than 27%. The growth rate of the overall emissions of pollutants in 2017-2019 is generally less than that in 2014-2016. And CO_2 accounts for the most significant proportion of various pollutants, which is much higher than other pollutants. Second, under the influence of the CNG2020 strategy, the emissions of unit turnover of six gases have decreased. Third, except for HC, the overall decline was 3.03%, and the average overall emissions of other pollutant routes in 2017-2019 increased relatively compared with 2014-2016, but the increasing percentage is not

more than 13%. And the growth range of the average overall emission of the routes in 2017-2019 is generally less than that in 2014-2016. Fourth, each route's average unit turnover emissions have decreased after CNG2020 strategy. Fifth, as one of China's air transport centre cities, Guangzhou has achieved good results in carbon emission reduction, but Shanghai and Beijing, as important aviation hubs in China, are not sensitive to the CNG2020 strategy. And under the CNG2020 strategy, airlines do not control short-haul routes' carbon emissions. Sixth, airlines' average overall HC emissions decreased, but the average overall emissions of CO_2, CO, NOx, SO_2, and PM2.5 increased. And unlike the performance of the routes, the growth rate of average overall emissions of airlines in 2017-2019 is generally more significant than that in 2014-2016. Seventh, each airline's average unit turnover emissions have decreased after the CNG2020, too. Most airlines have better controlled their carbon emissions after the CNG2020 strategy was proposed. Only four airlines have increased their carbon emissions per unit turnover after the CNG2020 strategy was proposed.

The standard LTO method is adopted in the calculation of LTO phase emissions in this paper, without considering delays and flight turnover caused by weather. And this paper does not consider the emissions of freight transport. Future research could focus on emissions from delays, flight turnover and freight.

4. Methods

In this paper, the emissions in the CCD stage are calculated through the Modified BFFM2-FOA-FPM method. The LTO emissions are calculated based on the ICAO standard method.

4.1 Modified BFFM2-FOA-FPM method

In the Modified BFFM2-FOA-FPM method, the CCD emissions $E(Q)$ can be calculated by

$$E_j(Q) = I_j * F(Q) = I_j * M_{fuel} * weight(Q) = I_j * (1 - M_{ff}) * weight(Q)$$

$$= I_j * \left(1 - \prod_{i=1}^n \frac{W_i}{W_{i-1}}\right) * weight(Q) = I_j * \left[1 - e^{-\frac{dis*ratio_{cr}}{10*v}}\right] * weight(Q)$$

$$= I_j * \left[1 - e^{-\frac{dis*ratio_{cr}}{10*v}}\right] * (aircraft bare weight + 100 * (load\ factor * number\ of\ seats) + 50 * seat) \quad (1)$$

I_j is the emission coefficient of pollution j of aviation kerosene (EASA, 2022). $weight(Q)$ is the total weight of the aircraft. M_{fuel} is the fuel coefficient, $M_{ff} = \prod_{i=1}^n \frac{W_i}{W_{i-1}}$ is a fuel weight proportionality coefficient, which is usually calculated by Fuel Percentage Method (FPM). The total sections of a whole flight contain seven task sections: Engine Starting, Taxiing, Taking Off, Climbing, Cruising, Descending and Landing. $\frac{W_i}{W_{i-1}}$ as the fuel weight proportionality coefficient of task section i ($i = 1,2,...,7$). $number\ of\ seats$ is certified seat number, $seat$ is the actual passenger number.

As we only consider the CCD section in this study, so we define the $\frac{W_i}{W_{i-1}}$ of other sections is 1. The $\frac{W_i}{W_{i-1}}$ of Climbing and Descending are 0.980 and 0.990. The equation of the CCD section to calculate $\frac{W_i}{W_{i-1}}$ is $\frac{W_i}{W_{i-1}} = e^{-\frac{dis*c_{cr}}{10*v*LD_{cr}}}$. dis is

the cruising distance, v is the cruising speed, c_{cr} is the fuel consumption ratio when the aircraft is cruising, LD_{cr} is the lift-drag ratio when the aircraft is cruising. The value of c_{cr} and LD_{cr} has direct relationships with the aircraft type. We define $ratio_{cr} = \frac{c_{cr}}{LDc_{cr}}$, and then for the cruising task section, the W_i/W_{i-1} is $W_i/W_{i-1} = e^{-\frac{dis*ratio_{cr}}{10*v}}$.

295　The actual flying time of each flight is applied to check the results of $ratio_{cr}$, and get the emission intensity.

For CO_2, the emission coefficient is fixed, which is $I_{CO2} = 3.157\ kg/kg$;

For SO_2, the emission coefficient is fixed, which is $I_{SO2} = 3.870\ g/kg$;

　For CO and HC, $I_j = I_{j0}*\frac{\theta^{3.3}}{\delta^{1.02}}$. θ is the ratio of outside temperature to 288 K; δ is ratio of external pressure to sea level pressure. I_{j0} is the standard emission coefficient of a LTO stage OF CO or HC (g/kg).

300　For NOx, $I_{NOx} = I_{j0}*\frac{\delta^{0.51}}{\theta^{1.65}} * \exp(19.0 * (0.0063 - \frac{0.622*\varphi*Pv}{P-\varphi*Pv}))$. I_{j0} is the standard emission coefficient of a LTO stage of NOx (g/kg). θ is the ratio of outside temperature to 288 K; δ is ratio of external pressure to sea level pressure. φ is atmospheric relative humidity; P is external pressure; Pv is atmospheric saturation pressure, which is calculated by Goff-Gratch formula (Detwiler, 1983):

$$\lg Pv = 10.79574 * (1 - \frac{273.16}{T}) - 5.02800 * \lg\left(\frac{T}{273.16}\right) + 1.50475 * 10^{-4} * \left[1 - 10^{8.2969*\left(1-\frac{T}{273.16}\right)}\right] + 0.42873 * 10^{-3} *$$

305
$$\left[10^{4.76955*\left(1-\frac{T}{273.16}\right)}\right] + 0.78614.$$

According to relevant physical laws (Smith et al., 1970; Abu-Ghannam and Shaw, 1980; Detwiler, 1983), The external pressure P is $P = 101325 * (1 - \frac{H}{44300})^{5.256}$. H is height. The outside temperature T is $T = 291.15 - \frac{6*H}{1000}$. The atmospheric relative humidity φ is $\varphi = 100 * \frac{a*(1+T/273.16)}{0.8*Pv}$. a is absolute humidity, and it is a $= \frac{26}{233211} * T^3 - \frac{302}{3731} * T^2 + \frac{569}{29} * T - \frac{17461}{11}$.

310　For PM2.5, it can be divided into Nonvolatile Component Fine Particles (NCFP) and Volatile Component Fine Particles (VCFP). For NCFP, $I_{NCFP} = 0.054 * AFR * (SN)^{1.234} + 0.877$. The unit of I_{NCFP} is mg/kg. AFR is Air-Fuel Ratio, which is decided by height. SN is engine smoke, which can be found in the ICAO Aircraft Engine Emissions Databank 27. For VCFP, it contains Volatile Organic Components (VOC) and Volatile Sulfur Components (VSC). For VOC, $I_{VOC} = \sigma * I_{HC}$. σ is the ratio of VOC to the emission coefficient of HC, which can be found in ICAO Aircraft Engine Emissions Databank

315　(EASA, 2022). For VSC, $I_{VSC} = 3 * 10^6 * 0.2\% * 3.3\%$. 0.2% is fuel sulfur content and 3.3% is sulfur conversion coefficient. Therefore, for PM2.5, $I_{PM2.5} = I_{NCFP} + I_{VOC} + I_{VSC}$.

4.2 ICAO standard method to calculate LTO emissions

This paper uses the standard LTO cycle definition specified by ICAO to calculate the fuel consumption, including all activities at an altitude below 3000 feet (915m) near the airport. The calculation formula of the five non-CO_2 pollution

320　emissions in LTO stage is:

$$E_{LTO} = \sum_m P_a \times N_a \times C_m \times t_m \tag{2}$$

E_{LTO} is the emissions in the LTO stage; P_a is the standard emissions of the engine of aircraft type a (unit: kg); N_a is the number of engines of aircraft type a; C_m is the thrust setting of stage m; t_m is the working time of phase m. The value range of m is 1, 2, 3, and 4, respectively corresponding to the four stages of take-off and landing in the aircraft flight process: take off, climb, approach and taxiing. According to the standard LTO cycles defined by ICAO, when the aircraft is taking off, its engines are at 100% thrust and working time is 0.7 minutes; when the aircraft is climbing, its engines are at 85% thrust and working time is 2.2 minutes; when the aircraft is approaching, its engines are at 30% thrust and working time is 4 minutes; when the aircraft is taxiing, its engines are at 7% thrust and working time is 26 minutes. Therefore, in a standard LTO cycle, the total working time is 32.9 minutes.

The fuel consumption rate is calculated as:

$$F_{am} = \frac{1}{A} \sum_j K_j F_{jmi} \tag{3}$$

A is the total number of airlines with aircraft type a; j is the type of engine of the aircraft; K_j is the number of aircraft type a equipped with engine type j; F_{jmi} is the fuel consumption rate of engine type j under the m setting. The data is from the ICAO Aircraft Engine Emissions Databank (EASA, 2022). This formula is based on the weighted average of all possible engine types of the domestic routes in China.

5. Data sources

The original data are collected from the VariFlight.com (VariFlight, 2022) and we complied the data. The data on aircraft type, flying time, flying distance, transfer flight and airlines are from VariFlight.com (VariFlight, 2022). The data sources and specific steps for data collection in Supplementary Table S1. The data on the engines of each aircraft and the data on the engines are from ICAO Aircraft Engine Emissions Databank (EASA, 2022).

6. Data availability

The CCD and LTO emissions of each route and airline for the six pollutions from 2014 to 2019 can be found in Cui (2022) (https://doi.org/10.6084/m9.figshare.20071751.v1).

7. Acknowledgements

8. Author contributions

9. References

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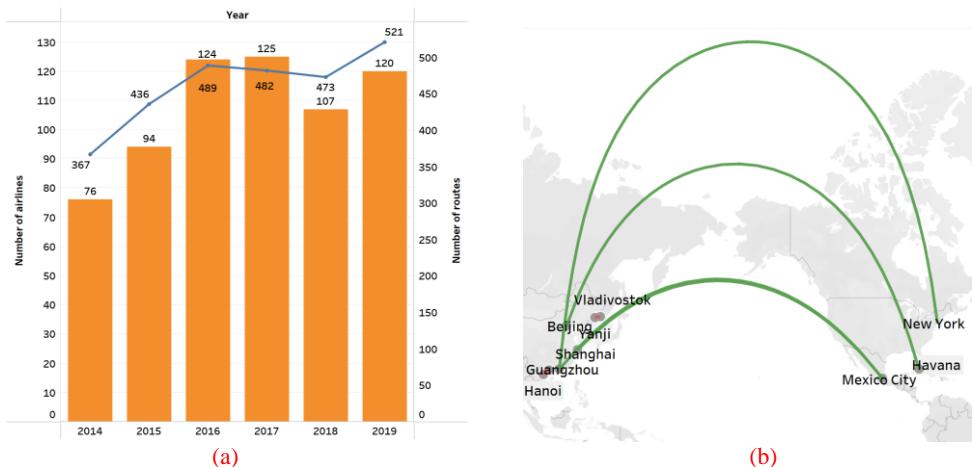
Figure 1 Statistical characters of the routes during 2014-2019. (a) Number of routes and number of airlines. (b) The farthest and nearest distance.

415 **Figure 2 Statistical characteristics of aircraft configuration during 2014-2019.**

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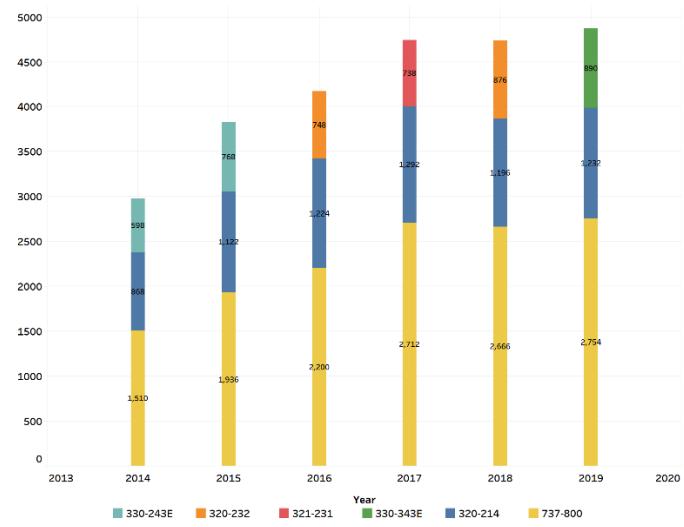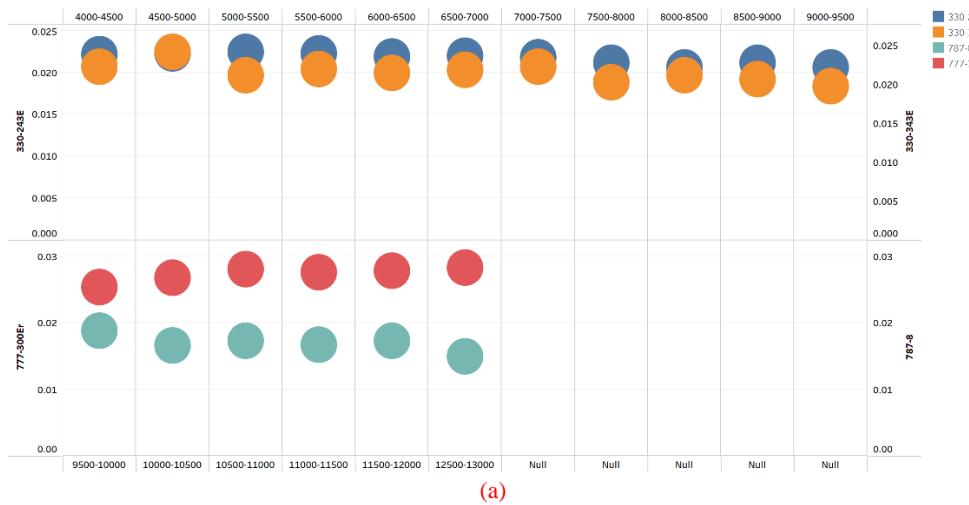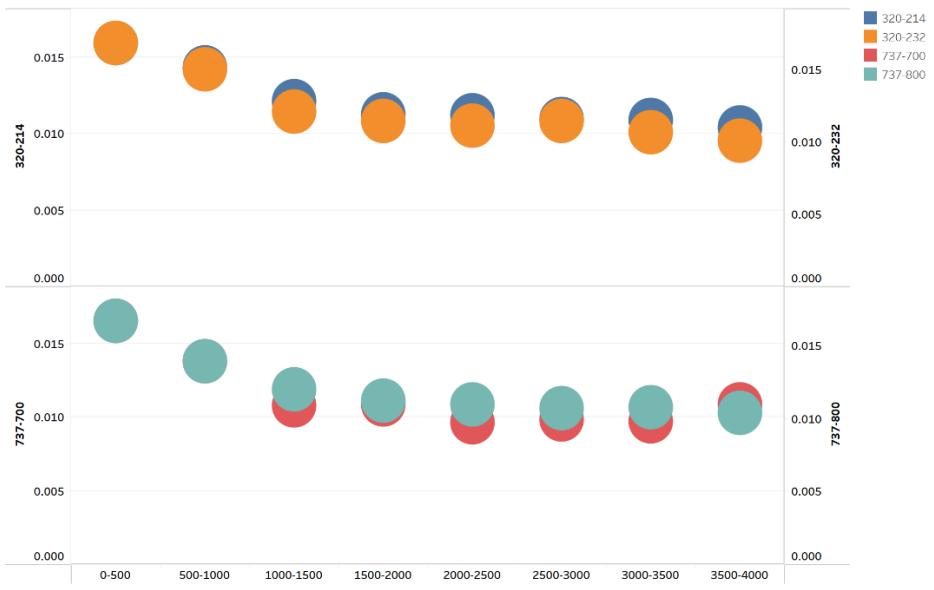
Figure 3 Carbon emission intensity of the aircrafts (ton/km). (a) Comparison of 320-214 and 320-232, and 737-700 and 737-800. (b) Comparison of 330-243E and 330-343E, and 777-300ER, and 787-8.

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(a)   (b)
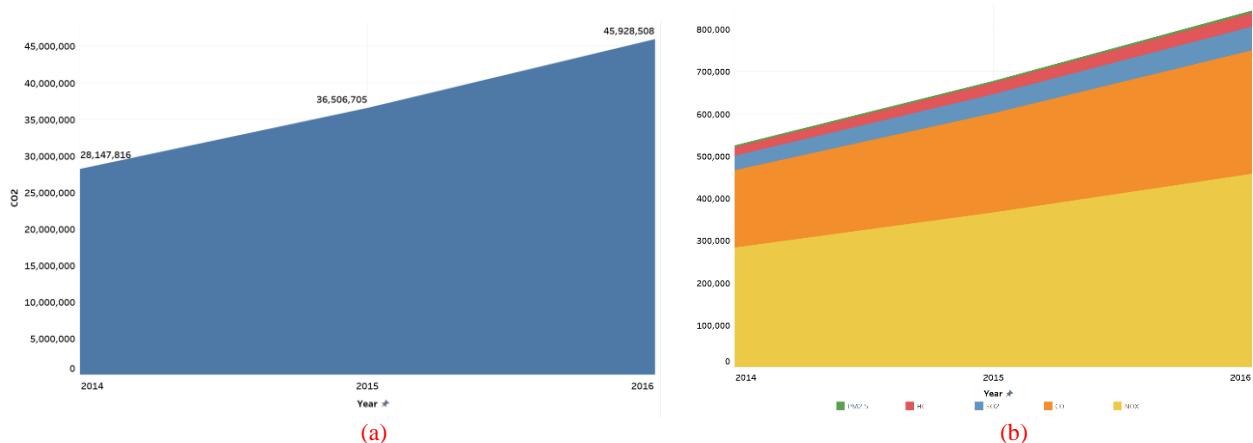
Figure 4 Annual overall emissions of six pollutants before CNG2020 strategy (tons). (a) Annual overall emissions of CO2 from 2014 to 2016. (b) Annual overall emissions of other five pollutants from 2014 to 2016.

(a)   (b)
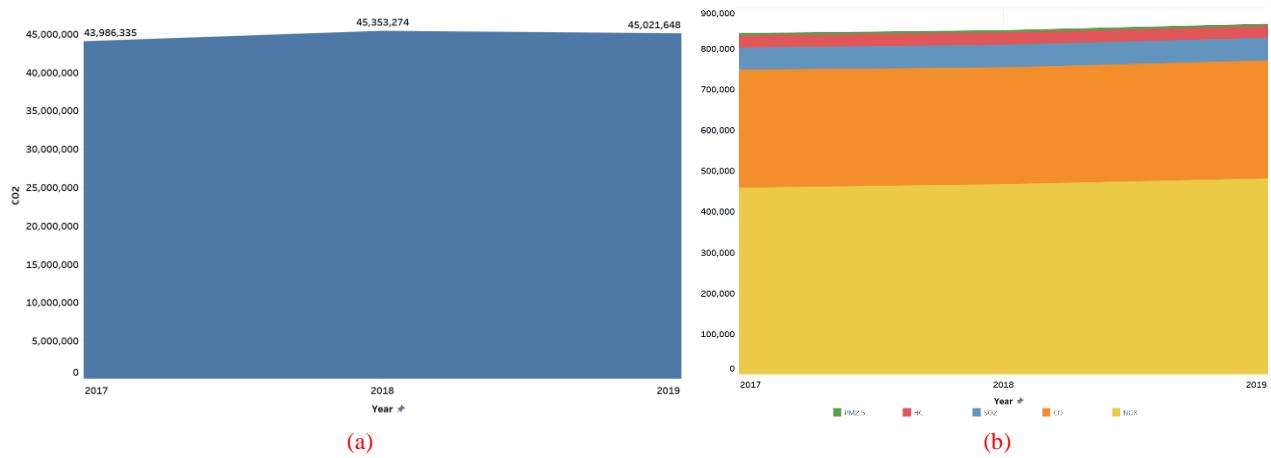
Figure 5 Annual overall emissions of six pollutants after CNG2020 strategy (tons). (a) Annual overall emissions of CO_2 from 2017 to 2019. (b) Annual overall emissions of other five pollutants from 2017 to 2019.

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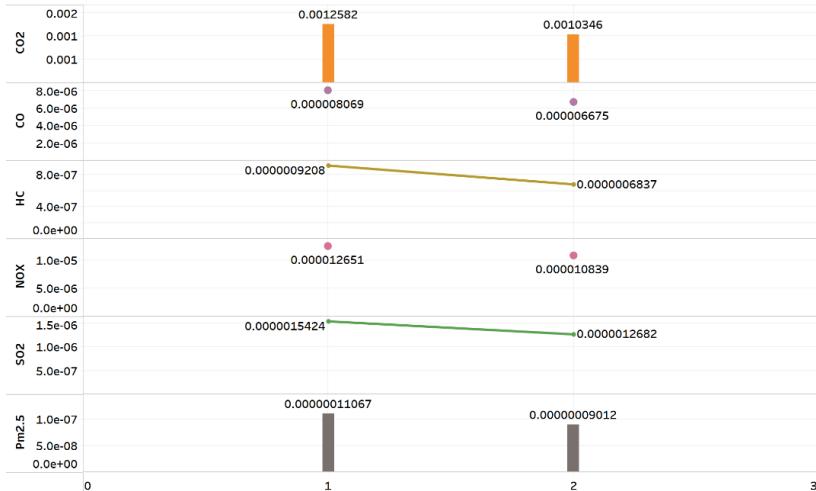
Figure 6 Changes in average unit turnover emissions of six pollutants before and after CNG2020 strategy(tons).

(a)

(b)
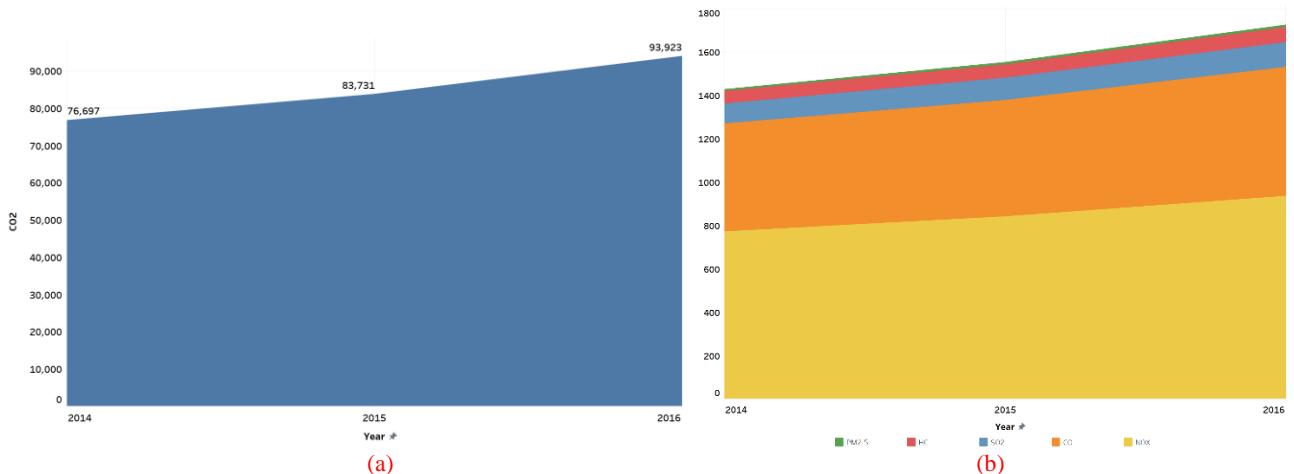
Figure 7 Annual average emissions of six pollutants from routes before CNG2020 strategy (tons). (a) Average overall route emissions of CO₂ from 2014 to 2016. (b) Average overall route emissions of other five pollutants from 2014 to 2016.


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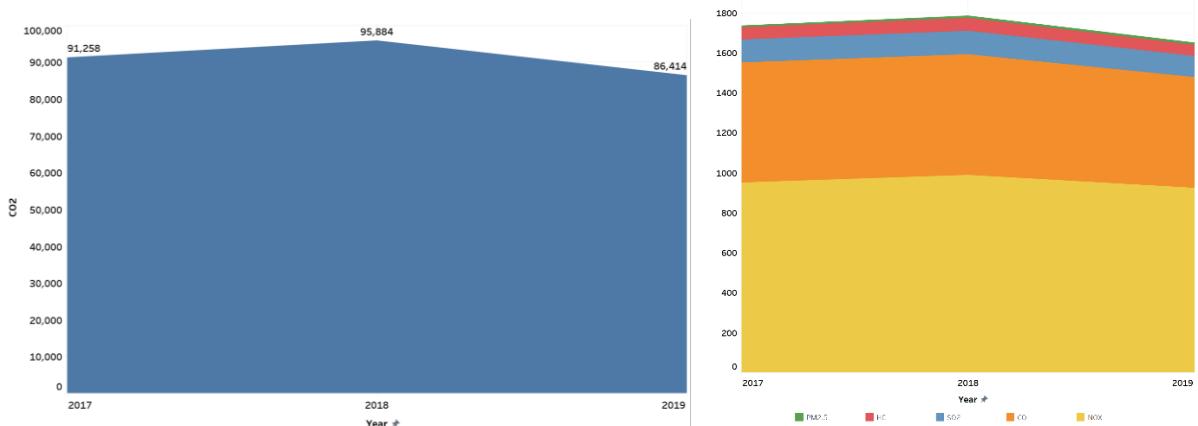
440   **Figure 8 Annual average emissions of six pollutants from routes after CNG2020 strategy (tons). (a) Average overall route emissions of CO_2 from 2017 to 2019. (b) Average overall route emissions of other five pollutants from 2017 to 2019.**
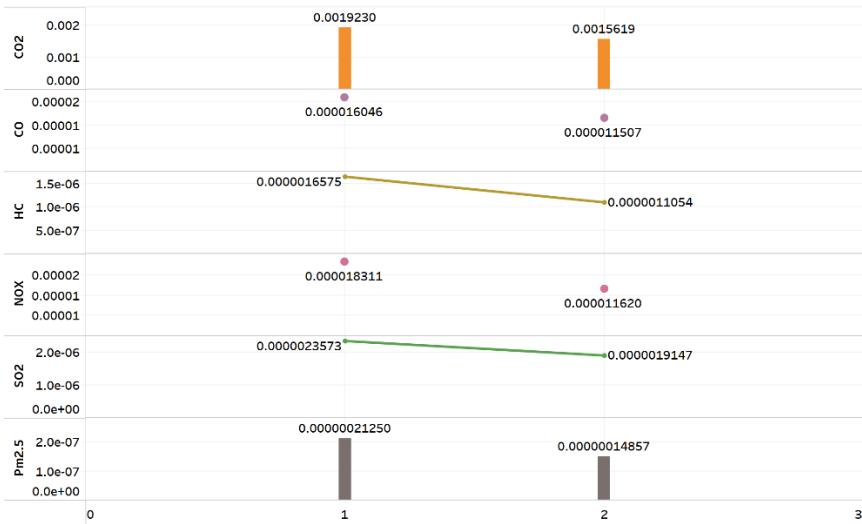
Figure 9 Average unit turnover emissions of each route before and after CNG2020.

445   (a)
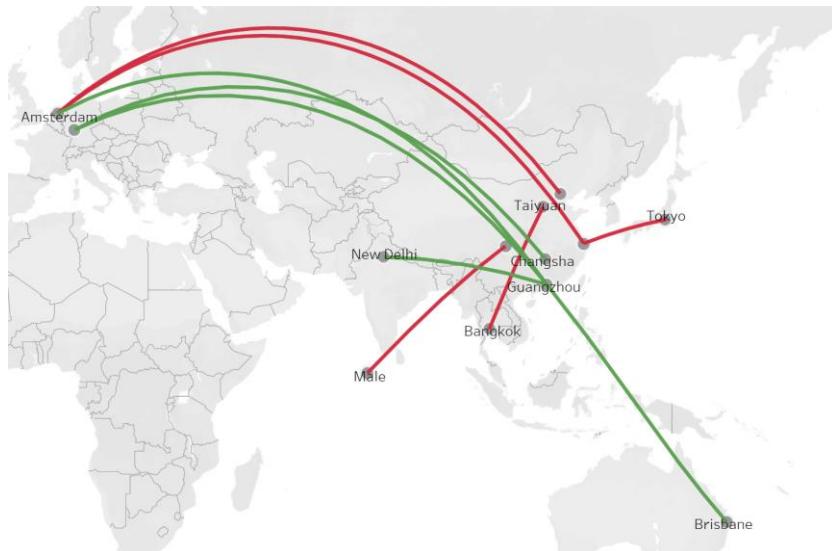
(b)

Figure 10 The impacts of CNG2020 on some popular routes. (a) The largest increase and the largest reduction routes in average unit turnover carbon emissions before and after CNG2020 strategy is proposed. (b) The strategy proposes the path of minimum change of average unit turnover carbon emissions before and after CNG2020 strategy is proposed.
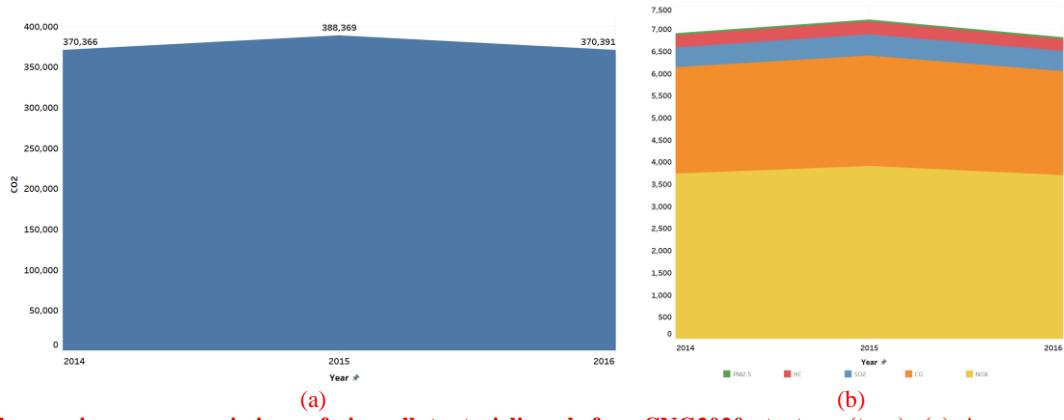

(a)

(b)

Figure 11 Changes in average emissions of six pollutant airlines before CNG2020 strategy (tons). (a) Average overall airline emissions of CO_2 from 2014 to 2016. (b) Average overall airline emissions of other five pollutants from 2014 to 2016.
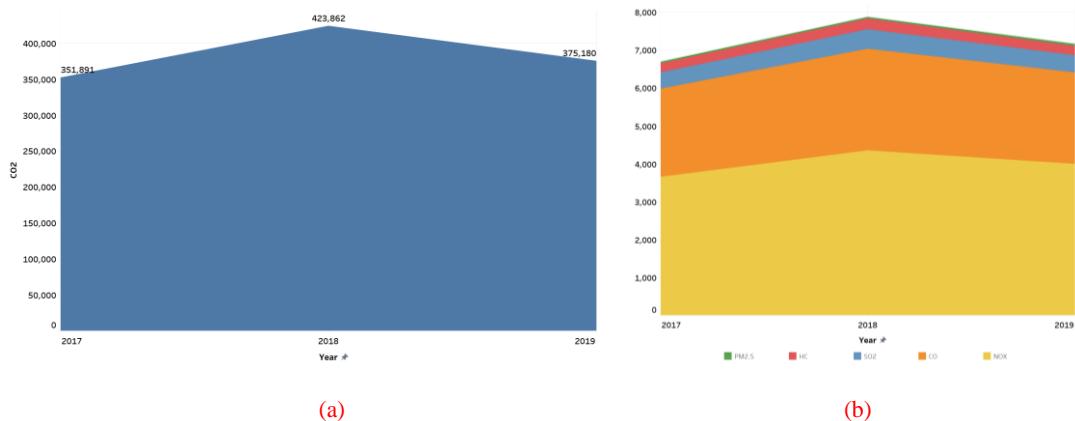

(a)

(b)

Figure 12 Changes in average emissions of six pollutant airlines after CNG2020 strategy (tons). (a) Average overall airline emissions of CO_2 from 2017 to 2019. (b) Average overall airline emissions of other five pollutants from 2017 to 2019.
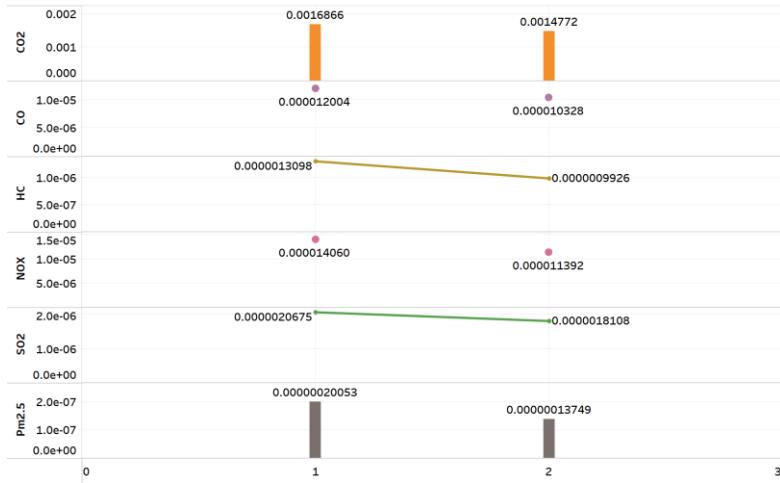
460   **Figure 13 Average unit turnover emissions of each airline before and after CNG2020.**
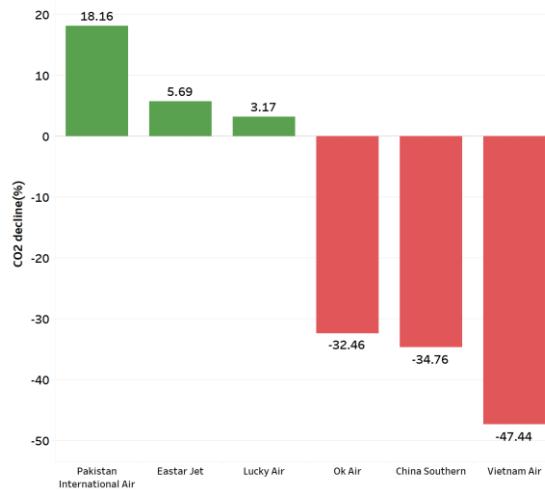
Figure 14 The impacts of CNG2020 on some popular airlines.