



The HYPERMAQ dataset

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Abstract. Because of the large diversity of case 2 waters ranging from extremely absorbing to extremely scattering waters and the complexity of light transfer due to external terrestrial inputs, retrieving main biogeochemical parameters such as chlorophyll-a or suspended particulate matter concentration in these waters is still challenging. By providing optical and biogeochemical parameters for 180 sampling stations with turbidity and chlorophyll-a concentration ranging from 1 to 700 FNU and from 0.9 to 180 mg m⁻³ respectively, the HYPERMAQ dataset will contribute to a better description of marine optics in optically complex water bodies and can help the scientific community to develop algorithms. The HYPERMAQ dataset provides biogeochemical parameters (i.e. turbidity, pigment and chlorophyll-a concentration, suspended particulate matter), apparent optical properties (i.e. water reflectance from above water measurements) and inherent optical properties (i.e. absorption and attenuation coefficients) from six different study areas. These study areas include large estuaries (i.e. the Rio de la Plata in Argentina, the Yangtze Estuary in China and the Gironde Estuary in France), inland (i.e. the Spuikom in Belgium and Chascomús lake in Argentina) and coastal waters (Belgium). The dataset is available from Lavigne et al (2022), <https://doi.pangaea.de/10.1594/PANGAEA.944313>.

1 Introduction

In marine optics, certain water properties such as the concentration of chlorophyll-a ([Chl-a] hereafter) or suspended particulate matter (SPM hereafter) are inferred from water leaving reflectance allowing a powerful satellite based monitoring. However, although algorithms are well matured in clear case 1 waters (Morel and Prieur, 1975; Morel and Maritorena, 2001), it is not the case in optically complex case 2 waters where apparent optical properties (AOPs) and inherent optical properties (IOPs) are influenced not only by [Chl-a] but also by terrestrial optically-active substances such as suspended sediments and colored dissolved organic matter (CDOM) that do not covary



with [Chl-a]. Given the complexity of light transfer in these waters and the large diversity of case 2 waters, algorithm definition is much more challenging (Odermatt et al., 2012) and requires datasets covering the extreme variability of case 2 water conditions. Hence, any additional data obtained in optically complex waters are valuable to the scientific community as they will help to better understand marine optics in such waters and to design ocean color algorithms.

The present dataset (Lavigne et al., 2022; <https://doi.pangaea.de/10.1594/PANGAEA.944313>) has been collected as part of the HYPERMAQ project. During this project, different types of optically complex waters with turbidity ranging from moderate to extreme (1 to 700 FNU) and [Chl-a] ranging from low to very high (0.9 to 180 mg m⁻³) have been sampled in various locations around the world. Their main optical and biogeochemical parameters are shared in this dataset, including measurements of water-leaving reflectance, turbidity, non-water light absorption and attenuation coefficients as well as SPM, [Chl-a] and other pigment composition. In the next sections, study areas, sampling methodology and final HYPERMAQ datasets are described.

2 Sites

Different sites, generally characterized as optically complex case-2 waters with a turbidity and [Chl-a] ranging from moderate to extremely high, have been sampled in coastal and inland waters in Belgium, France, Argentina and China (Figure 1).

2.1 Belgian coastal waters

The Belgian coastal waters (latitudes: 51.27° to 51.59°N; longitudes: 2.50° to 3.15°E) have been sampled in April 2018 and July 2018 from the RV Simon Stevin (Table 1). Belgian coastal waters are dominated by Atlantic waters which enter from the English Channel (Lacroix et al., 2004) and experience very strong along shore tidal currents which cause sediment resuspension leading to high turbidity. SPM concentrations range from less than 1 g m⁻³ in offshore and deeper waters to more than 100 g m⁻³ in very shallow waters. Phytoplankton blooms, characterized by high [Chl-a] concentration (more than 10 mg m⁻³), develop in spring from March to May. During summer, the biomass remains rather high (5 to 10 mg m⁻³) compared to winter when phytoplankton growth is mostly limited by light (Lancelot et al., 2005). The blooming season is mostly dominated by two taxa: diatoms in early spring and summer and *Phaeocystis globosa* in April-May (Muylaert et al., 2006).

2.2 Spuikom lagoon

The Spuikom lagoon (latitude: 51.23°N, longitude: 2.95°E) is an artificial basin that is connected to Ostend harbor (Belgium) by a lock system. The Spuikom has a surface area of 0.82 km² and an average depth of 1.5 m. In the past, it has been used as a flushing basin to flush sediments from the harbor channel. Today it is used for leisure and commercial activities like sailing and shellfish farming. The Spuikom can experience events of phytoplankton blooms, of high turbidity (when strong winds cause the resuspension of bottom sediments), and of clear waters, which allow the development of microphytobenthic biofilms and macroalgae in the bottom (Castagna et al., 2022).



71 The system was sampled during the growth season of 2018 (April and July, Table 1). Measurements were
 72 performed from inflatable boats provided by Ghent University and VLIZ (Zeekat).

73 **2.3 Gironde Estuary**

74 The Gironde Estuary, southwest France, is a good example of sediment-dominated case 2 waters influenced by
 75 river inputs. The Gironde Estuary has been sampled between 17-20 September 2018 in two locations: Pauillac
 76 (latitude: 45.1975°N, longitude: -0.7422°E) close to the maximum turbidity zone and Le Verdon (latitude:
 77 45.5438°N, longitude: -1.042°E) close to the river mouth. In the Gironde Estuary, the origin of the particles is
 78 twofold: inputs from rivers Garonne and Dordogne and erosion of recently settled sediments by tidal currents
 79 (Castaing and Allen, 1981). The suspended matter is a mixture of organic and mineral composites, where the
 80 organic fraction represents less than 2% of the total material (Doxaran et al. 2002). The mineral fraction is com-
 81 posed of micas (63%) and quartz (25%), while clay phases contain four minerals: montmorillonite (30%), illite
 82 and interstratified minerals (40%), kaolinite (15%), chlorite and interstratified minerals (15%). [Chl-a] and CDOM
 83 concentrations are low, with [Chl-a] ranging from 1 to 3 mg m⁻³ (Irigoien & Castel, 1997), and dissolved organic
 84 carbon (DOC) ranging from 1 to 7 mgC L⁻¹ (Abril et al., 1999; Castaing and Allen, 1981). The Gironde Estuary
 85 has well-developed turbidity maximum zones, with both tidal asymmetry and density residual circulation involved
 86 in their formation (Castaing and Allen, 1981). It is characterized by SPM concentrations ranging from 10 to 1000
 87 g m⁻³ within surface waters (Doxaran et al. 2009a).

88 **2.4 Chascomús lake**

89 Chascomús lake, located in the Pampa Plain in the Buenos Aires Province in Argentina (latitude: -35.5828°N,
 90 longitude: -58.0202°E), with a surface area of ~ 30 km², is a highly turbid, shallow lagoon (average depth of ~ 1.9
 91 m), permanently mixed due to intense and persistent winds (Torremorell et al., 2007). Total suspended matter
 92 varies widely from 66.3 to 614 g m⁻³ with a mean value of 227.3 ± 133.7 g m⁻³ (Diovisalvi et al. 2014) and on
 93 average the inorganic content represented ~65%. Nephelometric turbidity also widely ranged from 76.46 to 509.74
 94 NTU, with a mean value of 209.18 ± 112.76 NTU. Turbidity was highly correlated to SPM while no significant
 95 correlation with [Chl-a] was found (Pérez et al. 2011). Total [Chl-a] concentration ranged from 50.6 to 856.3 mg
 96 m⁻³ (mean = 328.5 ± 173.4 mg m⁻³) during the 2005-2009 sampled period (Diovisalvi et al. 2014). The lake is
 97 characterized by high primary production (Torremorell et al., 2009) and a rich and diverse phytoplankton commu-
 98 nity, mostly composed of cyanobacteria. In terms of biovolume, cyanobacteria contribute 50% to total phytoplank-
 99 ton biovolume and 75% to total C in the water column (Diovisalvi et al. 2010). Despite the high CDOM absorption
 100 (a_{CDOM} , mean $a_{CDOM}(440)$ = 4.65 ± 0.91 m⁻¹), absorption by particulate fraction (a_p) has a prominent role in light
 101 absorption, for which both phytoplankton pigments (a_{phy}) and non-pigmented particles (a_{NAP}) contribute similarly
 102 to total particulate absorption (Pérez et al. 2011). Both SPM (especially the inorganic part) and [Chl-a] (less pro-
 103 nounced) show seasonal variation with increasing values in spring and summer (Mid-September to Mid-March),
 104 while the dissolved fraction did not show a significant seasonal difference (Pérez et al. 2011). The HYPERMAQ
 105 field campaign in Chascomús lake took place on 9-10 April 2018. Radiometric, in-water measurements as well as
 106 samples were collected at the end of a 164 m long pier.



107 2.5 Río de la Plata

108 The Río de la Plata is a large and shallow funnel shaped estuary with high values SPM, ranging from 100 to 300
 109 g m^{-3} (Framiñan and Brown, 1996) and reaching 940 g m^{-3} in the maximum turbidity zone (Dogliotti et al. 2014).
 110 Turbidity values widely vary between 2 and 680 FNU (Dogliotti et al. 2016). SPM, turbidity and [Chl-a] spatial
 111 distribution and temporal variability is highly variable. In the upper estuary, a freshwater with tidal regime area,
 112 turbidity increases from January to April/May (with higher values along the southern Argentinian coast of com-
 113 pared to the northern Uruguayan coast), and decreases from June to September (Dogliotti et al. 2016). In turn [Chl-
 114 a] also show high spatial variability, in the upper estuary higher values are generally found in the northern part
 115 (Uruguay) compared to the southern part (Argentina). In particular, high [Chl-a] have been recorded during spring-
 116 summer months related to cyanobacteria blooms both along the Uruguay (Aubriot et al. 2020) and Argentine
 117 (Dogliotti et al. 2021) coasts, when [Chl-a] values as high as 13.6 and 153 mg m^{-3} have been recorded, respectively.
 118 Measurements in the Río de la Plata were performed from a fixed 500 m long pontoon at the Palermo Pescadores
 119 Club in Buenos Aires (latitude: -34.5609°N , longitude: -58.3988°E) on 4 and 5 April 2018.

120 2.6 Yangtze Estuary

121 The Yangtze Estuary is located on the east coast of China and close to East China Sea (Figure 1). Influenced by
 122 the Yangtze River, the largest river in China and the third largest in the world, which discharges an annual average
 123 of $9 \times 10^{11} \text{ m}^3$ of freshwater and 4×10^8 tons of sediment into the estuary (Chen et al., 2003), the Yangtze Estuary is
 124 an extremely turbid area (Shen et al., 2010a). Taking 2009 as example, the annually averaged of SPM in surface
 125 waters varied from 58 g m^{-3} at the upstream limit of the estuary to about 600 g m^{-3} at the mouth area, and fell again
 126 to 57 g m^{-3} at the seaward limit of fresh water diffusion (Li et al., 2012). Due to the different river discharges, the
 127 SPM of the Yangtze Estuary exhibits seasonal variations (Shen et al., 2013), with SPM in the upper estuary (lower
 128 estuary) during flood season significantly higher (lower) than that during the dry season. Over the past 37 years,
 129 SPM in Yangtze Estuary demonstrated an overall declining pattern (Luo et al., 2022), with SPM in the inner
 130 estuary responding most promptly (40.3% reduction) after the operation of Three Gorges Dam. [Chl-a] also shows
 131 seasonal variations in Yangtze Estuary, ranging from 0.03 to 3.10 mg m^{-3} and from 0.88 to 31.5 mg m^{-3} during
 132 spring and summer seasons of 2008, respectively (Shen et al., 2010b). In addition, the Yangtze Estuary is an area
 133 with frequent outbreaks of algal blooms, with diatoms being the most frequently reported group (Shen et al., 2019;
 134 Zhu et al., 2019).

135 Two Hydrological Stations in Chongming Island, Shanghai, China, namely Chongxi (longitude: 121.193°E , lati-
 136 tude: 31.759°N) and Baozhen (longitude: 121.609°E , latitude: 31.520°N) have been sampled from 30 May to 8
 137 June in 2018 (Table 1).

139 3. Data collection

140 The dataset contains measurements of the turbidity and, if available, concomitant SPM, absorption and attenuation
 141 coefficients, [Chl-a] and reflectance measurements are also included (Lavigne et al., 2002; https://doi.pan-
 142 gaea.de/10.1594/PANGAEA.944313). An overview of the dataset, with the number of observations after quality



control for each site and parameter, is provided in Table 2. The measurement methodology for each parameter is described below.

3.1 Water-leaving reflectance

Above-water reflectance was determined using three TriOS/RAMSES hyperspectral spectroradiometers, two spectroradiometers measure radiance and one irradiance. The same TriOS instruments from RBINS institute were used for all campaigns except the ones which occurred in Argentina where only instruments from IAFE institute were available. The spectrometers measure in the 350-950 nm range with a sampling interval of 3.3 nm and effective spectral resolution of 10 nm. The instruments were mounted on a frame and placed in the bow of the vessels (Belgian coastal zone and Spuikom) or fixed to a rail when measurements were made from pontoons (Gironde, Chascomús and Rio de la Plata). Zenith angles of the sea- and sky-viewing radiance sensors were set to 40°. Prior to each measurement, the azimuth angle of the sensors was adjusted to obtain a relative azimuth angle with respect to the sun of 90°, either left or right to get the best unobstructed view of the water and minimize structure perturbation when measuring from pontoons. Simultaneous upwelling water radiance (L_u), downwelling sky radiance (L_{sky}) and downwelling irradiance (E_d) were collected every 10 s for 10 min. Data was acquired using MSDA-XE software and radiometrically calibrated using the latest calibration update from annual laboratory calibration. Water reflectance (ρ_w) was calculated following

$$\rho_w(\lambda) = \frac{L_u(\lambda) - \rho_{sky} L_{sky}(\lambda)}{E_d(\lambda)} \pi$$

Where ρ_{sky} is the air-sea interface reflection coefficient which is calculated based, when available, on wind speed as in Ruddick et al. (2006) or set to a fixed value of 0.0256 when measured in estuaries from fixed pontoon considering that surface waves are fetch-limited and not related to wind speed. The data processing, including quality control, are described in Ruddick et al. (2006).

3.2 Turbidity

Turbidity was measured with two handheld HACH 2100P/Q ISO turbidimeters from RBINS and IAFE institutions. In the HYPERMAQ dataset, turbidity data measured with the instrument from IAFE were provided by default as they cover the most of the campaigns. However, when turbidity data from IAFE instrument were not available (Belgian coastal waters, April 2018 and Spuikom April 2018), the values obtained with the instrument of the RBINS were used. Figure 2 shows the good consistency of both instruments ($r^2=0.99$). Water samples were collected from the surface with a bucket or from subsurface with a NISKIN bottle for measurements in coastal waters. A 10 mL vial was filled and turbidity was determined in Formazin Nephelometric Unit (FNU) with the ratio of light scattered at 90° compared to the transmitted light at 860 nm. Turbidity was recorded in triplicates and the median value was used. Turbidimeters were controlled with standards STABCAL Stabilized Formazin Turbidity of 0.1, 20, 100 and 800 FNU before and after each campaign. In water turbidity was also measured with an OBS501 (OBS hereafter) using a CR200 data logger. Turbidity measurements are derived from back-scattering with a field-of-view ranging from 125 to 170 degree and 90-degree side-scattering of a signal emitted at 850 nm and data are provided in Formazin Backscatter Unit (FBU) and in



178 Formazin Nephelometric Unit (FNU), respectively. When deployed from a pier, OBS was continuously recording
 179 data at subsurface throughout the whole day and values corresponding to specific stations were extracted from the
 180 time-series in a time-window of 10 minutes centered on the timing of the radiometric measurement and water
 181 sampling. When deployed from a boat, the OBS was maintained at subsurface (1 m depth) for at least 5 minutes.
 182 Then, from a visual check, leading and trailing data of each time-series were removed and the central values were
 183 averaged to obtain a final value.

184 **3.4 *In situ* absorption, beam attenuation and scattering coefficients**

185 The underwater absorption- and attenuation-meter (AC-9, WETLabs, Inc.) used was modified to cover the visible
 186 and near-infrared (NIR; 700 to 900 nm) spectral regions. It was designed with three visible (centered at 440 nm,
 187 555 nm and 630 nm) and six NIR (centered at 715 nm, 730 nm, 750 nm, 767nm, 820 nm and 870 nm) spectral
 188 channels, and a short pathlength (10 cm) appropriate for turbid coastal waters. At the sampling sites, the AC-9
 189 sensor was either deployed within the water column using an electrical water pump (SBE, SeaBird, Inc.) or used
 190 as a bench photometer passing the water samples right after collection through the tubes by gravimetry. The AC-
 191 9 data recorded just below the water surface were averaged over the last minute of acquisition to obtain the mean
 192 attenuation and absorption spectra for each station. Temperature and salinity corrections were applied as recom-
 193 mended by the manufacturer. As in Doxaran et al. (2007), the residual scattering effects on absorption measure-
 194 ments were corrected by applying the “proportional” method using 870 nm as the reference wavelength. The scat-
 195 tering coefficient was calculated as the difference between the measured beam attenuation coefficient, c_{nw} , cor-
 196 rected for temperature and salinity effects, and the absorption coefficient, a_{nw} , corrected for temperature, salinity
 197 and scattering effects. Those attenuation and absorption coefficients were referenced to pure water (non-water,
 198 subscript “nw”), so that the scattering coefficient obtained by difference corresponds to the scattering coefficient
 199 of marine particulates, b_p in m^{-1} . Small bubbles can contribute to the measured attenuation and scattering, but in
 200 turbid systems particles dominate the signal. One of the main issues encountered when sampling highly turbid
 201 waters was the saturation of the measured absorption and/or attenuation coefficients, which sometimes occurred
 202 at short visible wavelengths and even in near-infrared bands in the case of extremely turbid waters. This saturation
 203 was easily detected and the corresponding spectra were systematically removed from the dataset.
 204 When possible, after AC-9 data measurements, the water sample collected was directly filtered through disc filters
 205 (pore size 0.2 μm , Whatman). As in Doxaran et al. (2009b), the tube was rinsed twice with Milli-Q water and once
 206 with the filtrate, and then filled with the filtrate. The absorption signal of the filtrate was measured, providing
 207 a_{CDOM} in m^{-1} after applying corrections for temperature and salinity. The absorption coefficient of suspended par-
 208 ticles (a_p , in m^{-1}) was finally calculated by subtracting the signal from the non-water absorption coefficient.

209 **3.5 Concentration of Suspended Particulate Matter and Suspended Inorganic Particulate Matter**

210 SPM concentration was determined gravimetrically following the protocol of Tilstone et al. (2002) which is based
 211 on Van der Linde (1998). Water was sampled from the surface (maximum 2 m depth) with a NISKIN bottle on
 212 board the RV Simon Stevin or with a bucket in estuarine and inland waters. A sufficient volume of water was
 213 filtered on a pre-ashed GF/F filter and conserved at $-20^{\circ}C$ before analysis. The volume filtrated was determined



214 as a function of the turbidity following recommendations of Neukermans et al. (2012). Inorganic suspended par-
 215 ticulate matter (SPIM) was also calculated in all campaigns except in the Yangtze river. All the SPM measurements
 216 have been conducted with 3 replicates to assess variability except for the campaigns in the Yangtze Estuary where
 217 only one sample has been measured per each station. Filters were dried at 75°C for 24 hours and weighed in order
 218 to determine the suspended matter concentration (SPM, in g m^{-3}). For SPIM measurements, filters were then
 219 burned at 450°C for 4 hours to remove the organic part, and weighed again to estimate the suspended inorganic
 220 particulate concentration (SPIM, in g m^{-3}).

221 3.6. Chlorophyll-*a* and other pigment concentrations

222 [Chl] and phytoplankton pigments were determined using High Performance Liquid Chromatography (HPLC)
 223 following the protocol of Van Heukelem and Thomas (2001) in campaigns in the Belgian coastal waters, the Spu-
 224 ikom and the Gironde. In Belgian waters, measurements were provided by the LifeWatchBE sampling campaigns
 225 (Mortelmans et al., 2019, Flanders Marine Institute, 2021) of VLIZ. Pigment standards were acquired from the
 226 Danish Hydrographic Institute (DHI). In the Gironde, the analysis of pigments were performed by the SAPIGH
 227 analytical platform of the “Institut de la Mer de Villefranche” (CNRS-France). In the Argentinian campaigns
 228 [Chl-*a*] was determined spectrophotometrically using hot ethanol (60-70 °C) (Jespersen and Christoffersen 1987).
 229 As for turbidity and SPM, water samples have been collected from surface water with a bucket in inland waters or
 230 subsurface waters with a NISKIN in sea water.

231 4. Results and discussions

232 4.1 SPM and turbidity results

233 In the HYPERMAQ dataset SPM ranges between 1 g m^{-3} and 474 g m^{-3} (Table 3) and turbidity measured from
 234 HACH and OBS (side-scattering measurements) ranges between 0.9 and 771 FNU and between 0.2 and 632 FNU
 235 respectively. A very good relationship is observed between SPM and turbidity which almost follows the 1:1 line
 236 for both instruments (Figure 3). A linear model between both parameters gives very good coefficients of determi-
 237 nation ($R^2 = 0.98$ for HACH and $R^2 = 0.95$ for OBS) and slopes (0.92 for HACH and 0.86 for OBS). However, we
 238 can notice that for very high turbidity ($> 500 \text{ FNU}$), turbidity values measured by HACH tends to be slightly higher
 239 than SPM values (Figure 3A) but not OBS turbidity values. As expected from previous results, when comparing
 240 side scattering turbidity obtained from OBS and turbidity measured by HACH, a good relationship is retrieved
 241 (Figure 4A) with a R^2 of 0.96 and a slope of 0.84. Despite larger variability for very high turbidity, these results
 242 confirm that OBS is a good tool for continuous measurements of turbidity in turbid environments.
 243 The ratio of the side scattering versus the back scattering derived from OBS measurements has a particular interest
 244 as it can provide information on the size and properties of the particles, e.g higher ratio could be explained by
 245 larger particles (Nechad et al., 2016). In the HYPERMAQ dataset this ratio (Figure 4B) displayed a very high
 246 variability in low turbidity environments and an increasing slope for high turbidity environments (i.e. Pauillac) as
 247 also observed by Nechad et al. (2016). The very high variability when turbidity is low is explained by the strong
 248 impact of uncertainty on low back scattering values in the ratio calculation. In Figure 4B, it can be observed that



the side scattering versus back scattering ratio varies significantly between and within sampled sites. For instance, this ratio is higher in the Gironde Estuary at Le Verdon than in the Spuikom lagoon. It seems also to be higher in the Río de la Plata and in the Gironde at Pauillac than in the Chascomús lake, though the Río de la Plata showed high variability. Finally, the median ratio of the whole dataset is 1.77 which is close to the mean value of 1.72 found in Nechad et al. (2016) in turbid waters.

4.2 Chl-a and other pigments concentrations

[Chl-a] are extremely variable within HYPERMAQ test sites with values ranging between 0.91 mg m⁻³ in the Gironde Estuary at Le Verdon and 180.7 mg m⁻³ in the Chascomús lake, although most of the observations are within the range of 3 mg m⁻³ to 10 mg m⁻³ (Table 4). In addition, very high variability is observed within Belgian waters and Spuikom, with [Chl-a] values ranging by a factor 10. This variability is mainly due to the fact these study areas have been sampled at two different seasons (i.e. spring and summer).

Phytoplankton pigments derived from HPLC analysis were available in the Gironde Estuary, in the Belgian coastal waters and in the Spuikom. The relative contribution of some key pigments for phytoplankton groups identification (Uitz et al., 2006; Mackey et al., 1996) are represented on Figure 5. In the Gironde Estuary, at Le Verdon, significant concentration of fucoxanthin, peridinin and chlorophyll-b are observed suggesting that diatoms, dinoflagellates and chlorophytes are co-existing at similar levels. However, at Pauillac where phytoplankton biomass is higher (Table 4), the high concentration in fucoxanthin suggests that planktonic assemblage is dominated by diatoms. In Belgian waters, high value of fucoxanthin is also observed. This pattern was expected as fucoxanthin characterized the two phytoplankton groups which are dominant during spring and summer in the Southern North Sea: diatoms and the prymnesiophyte *Phaeocystis globosa* (Lancelot et al., 2005). The last one is also characterized by the presence of chlorophyll-c₃. In the Spuikom, fucoxanthin and chlorophyll-b show high concentrations indicating an important proportion of diatoms and chlorophytes.

4.3 Absorption and attenuation coefficients

Very wide ranges of light absorption and attenuation coefficients were measured as representative of low to extremely turbid waters. As expected in CDOM- and sediment-rich waters, the spectral variations of the non-water absorption coefficients were closely following an exponential function, with decreasing values from short visible to near-infrared wavelengths (Figure 6A). The respective contributions of CDOM and suspended particles to light absorption at 440 nm (Figure 7) were observed to vary from 20 % to 40% for CDOM and hence from 60 % to 80% for suspended particles, which could be expected in productive waters strongly influenced by sediment inputs from rivers and resuspension effects.

The spectral variations of the non-water attenuation coefficients (c_{nw} , Figure 6B) showed a smooth decrease with increasing wavelengths, closely following the power-law function with varying slopes. These variations of the spectral slope are expected to be representative of different particle size distributions due to the combined influences of wind-driven and tidal currents, and to the mixing between mineral-rich (sediments) and organic-rich (phytoplankton) particles.



284 4.4 Water reflectance

285 The large diversity of water-leaving reflectance spectra is displayed in Figure 8. Maximum reflectance in each
 286 spectrum varies from less than 0.02 on some spectra of the Belgian coastal waters to more than 0.15 in the Gironde
 287 Estuary at Pauillac. Shapes of the spectra are also very variable. The mark of strong chlorophyll-a absorption
 288 around 670 nm is well observed in the Chascomús and Spuikom lakes as well as in some spectra of the Belgian
 289 coastal waters. The two extremely turbid ~~sampling~~ stations, the Rio de la Plata and the Gironde at Pauillac, show
 290 some similarities in their spectral shapes although a large variability is observed at Pauillac due to a larger impact
 291 of tides.

292 The relationship of water reflectance at 645 and 860 nm and turbidity (Figure 9) shows expected patterns with a
 293 saturation of the reflectance at 645 nm when turbidity is higher than 200 FNU. Indeed, for these extreme turbidity
 294 values the band at 860 nm shows a more linear relationship. These results also show the limits of the standard
 295 algorithm of Nechad et al. (2009) for high turbidity or in case of a different environment like the Chascomús lake
 296 which is characterized by high [Chl-a] and turbidity.

297 5. Conclusion

298 Coastal and inland waters strongly interact with human activities. Some of these activities, like fisheries or tourism,
 299 rely on a good ecological status whereas the same activities but also others like farming, industry or urbanization
 300 tend to affect water quality. Hence, monitoring these waters is extremely important and for that optical remote
 301 sensing is a valuable tool as it allows a large spatial and temporal coverage. However, it is still challenging to
 302 retrieve biogeochemical parameters in complex case 2 waters (Odermatt et al., 2012) because the transfer of light
 303 in water is affected by temporally and spatially variable inputs of CDOM and terrestrial sediments. To help the
 304 scientific community to build comprehensive database for the development of algorithms, the HYPERMAQ da-
 305 taset provides data for seven different studies areas with SPM and [Chl-a] ranging from moderate to extremely
 306 turbid and productive, and located over three continents (i.e. Europe, South America and Asia). The HYPERMAQ
 307 dataset includes big river estuaries characterized by high turbidity, inland lagoons with productivity ranging from
 308 moderate to extreme and finally Belgian coastal waters in the North Sea characterized by the high spatio-temporal
 309 variability of optical properties (Vantrepotte et al., 2012). The parameters shared in the HYPERMAQ dataset
 310 include descriptors of biogeochemical conditions (i.e. [Chl-a], SPM, turbidity) as well as AOPs (i.e. water reflec-
 311 tance) and IOPs (a_{nw} and c_{nw}). Although this dataset does not aim to cover the whole variability of case 2 waters,
 312 it provides valuable information to describe turbid and even extremely turbid waters and has the potential to help
 313 the development of remote sensing algorithms. It can also contribute to the production of a larger optical database,
 314 based on in situ measurements for a comprehensive description of case 2 waters.

315



316 **Data availability**

317 Data is available from Lavigne et al. (2022), hosted at PANGAEA (<http://www.pangaea.de>) under the doi:
 318 <https://doi.pangaea.de/10.1594/PANGAEA.944313>

319 **Author contributions**

320 HL, AD, DD, FC, AC, XS, JIG, RP, MB, QV and KR participated to one or more field campaigns. Data processing
 321 has been made by HL, AD and JIG (turbidity), DD (absorption), KR, MB, QV, RP, AD (water reflectance), AC,
 322 AD, DD, FC, KS (chlorophyll-a, pigments and SPM). HL, DD, AD have compiled data and created the final
 323 dataset. All authors participated to manuscript redaction and revision.

324 **Competing interests.**

325 The authors declare that they have no conflict of interest.

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Table 1: Date, location and platform of the campaigns.

Campaign	Date	Platform	Latitude (deg)	Longitude (deg)
Spuikom	19 April 2018, 23,24 and 27 July 2018	Inflatable boat	51.23	2.95
Belgian Coastal waters	23-25 April 2018, 25- 26 July 2018	RV Simon Stevin	51.18-51.59	2.50-3.15
Gironde (Pauillac)	17 and 19 Sept. 2018	Harbor	45.1975	-0.7422
Gironde (Le Verdon)	18 and 20 Sept. 2018	Pier	45.5438	-1.042
Chascomús	9-10 April 2018	Pier	-35.5828	-58.0202
Rio de la Plata (Buenos Aires)	4-5 April 2018	Pier	-34.5609	-58.3988
Xangtze (Chongxi)	31 May, 1 and 3 June 2018	Pier	31.759	121.193
Xangtze (Baozhen)	4-8 June 2018	Pier	31.520	121.609

Table 2: Number of observations for each sampling site. *also include pigments concentrations from HPLC.

Campaign / Site	TriOS	TUR (HACH)	TUR (OBS)	$a_{nw} - C_{nw}$ AC-9	SPM	[Chl-a]
Spuikom	27	27	23	11	17	17*
Belgian coastal waters	18	19	17	10	19	18*
Gironde - Pauillac	25	26	26	23	13	13*
Gironde - Le Verdon	21	25	25	24	12	12*
Chascomús	5	5	5	5	5	3
Rio de la Plata- BA	16	22	22	18	10	10
Xangtze - Chongxi	-	19	-	-	17	-
Xangtze - Baozhen	-	37	-	29	37	-

Table 3: Distribution of SPM (g m^{-3}) in each sampling site.

Campaign / Site	SPM (g m^{-3})			
	min	median	mean	max
Spuikom	2.06	3.16	3.93	8.40
Belgian coastal waters	1.02	4.49	9.63	62.04
Gironde - Pauillac	22.5	181	177	474
Gironde - Le Verdon	5.85	7.80	10.2	23.5
Chascomús	81.0	175	141	189
Rio de la Plata	49.3	71.7	74.0	93.8
Xangtze - Chongxi	27.2	42.2	44.8	66.4
Xangtze - Baozhen	23.6	52.8	53.6	138.4



469 **Table 4: Distribution of Chl-a concentration (mg m^{-3}) in each sampling site.**

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Campaign / Site	Chl-a (mg m^{-3})			
	min	median	mean	max
Spuikom	2.40	9.16	10.64	22.70
Belgian coastal waters	1.99	6.33	7.49	17.36
Gironde - Pauillac	2.49	3.82	3.88	6.85
Gironde – Le Verdon	0.91	1.63	1.67	2.79
Chascomús	141.5	141.5	154.6	180.7
Rio de la Plata-BA	2.17	3.27	3.72	8.71
Xangtze - Chongxi	-	-	-	-
Xangtze - Baozhen	-	-	-	-

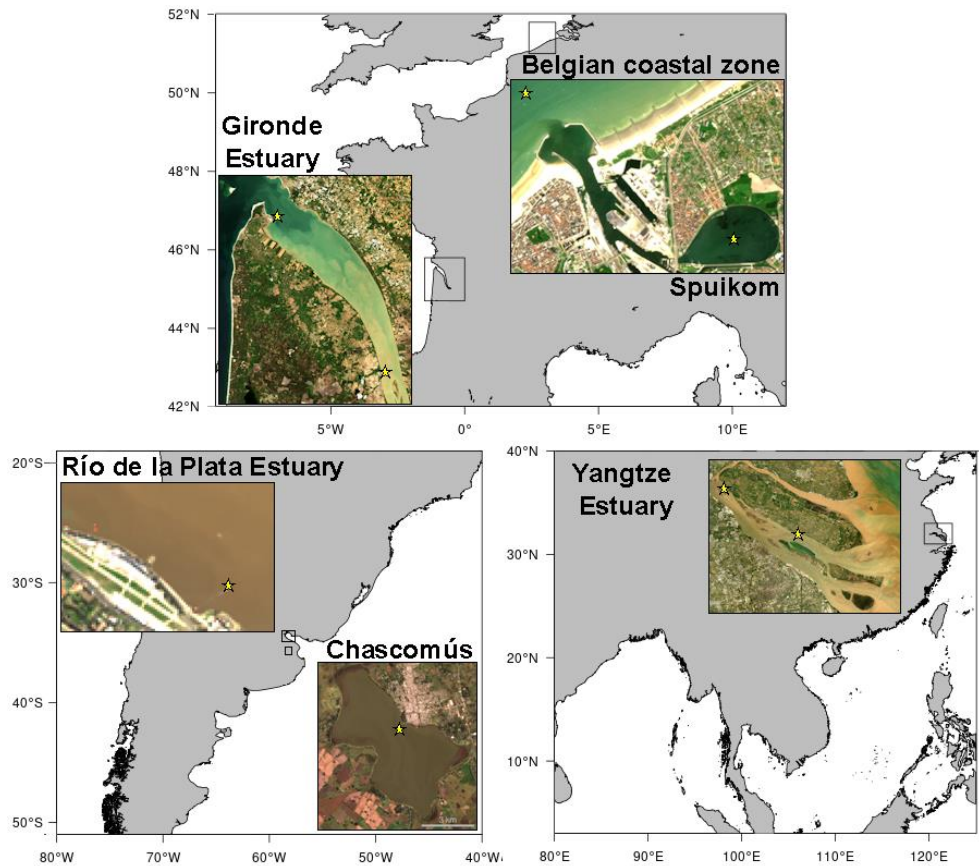
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Figure 1: Locations of the study areas. Satellite images are coming from Landsat 8 OLI (Yangtze: image taken on 2021-04-29, Chascomús: image taken on 10-05-2017, Río de la Plata: image taken on 2014-08-13) and Sentinel 2B MSI (Belgian Coastal Zone: image taken on 2021-05-30, Gironde: image taken on 2021-05-03)

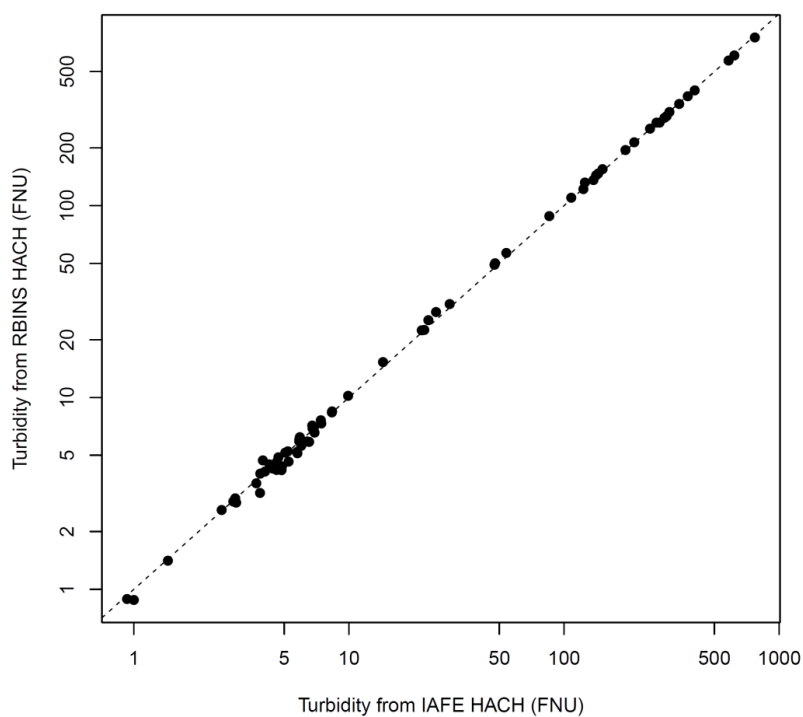
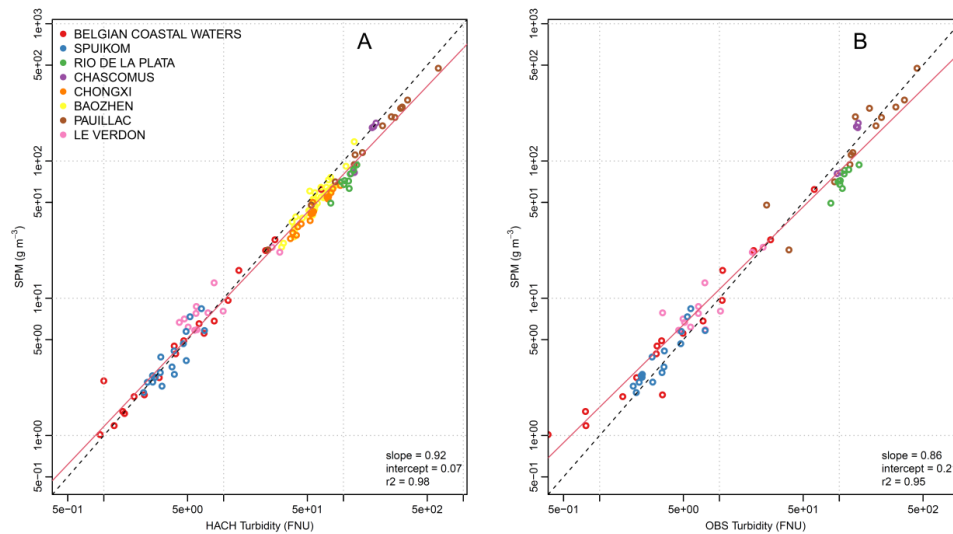


Figure 2: Comparison of simultaneous measurements of turbidity made from two different HACH instruments ($r^2=0.99$).



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489 **Figure 3: SPM as a function of turbidity. Turbidity from OBS is given by the side-scattering measurement. The dotted**
490 **line is the 1:1 line and the red line represents the linear regression between SPM and turbidity.**

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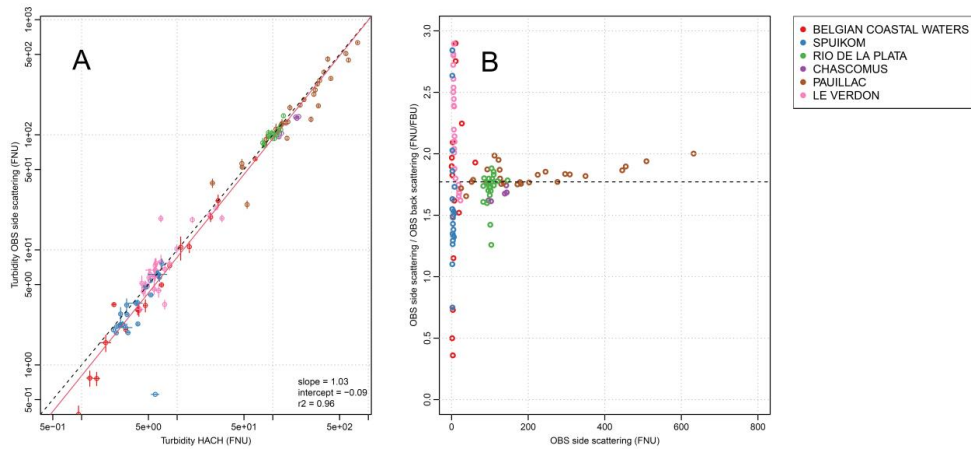


Figure 4: Turbidity measured by HACH instrument as a function of side-scattering turbidity measured by the OBS instrument (panel A). The red line shows the least squares regression between these variables. Panel B: ratio of the OBS side-scattering to backscattering ratio as a function of the OBS side-scattering. The horizontal dotted line represents the median value of the scattering ratio.

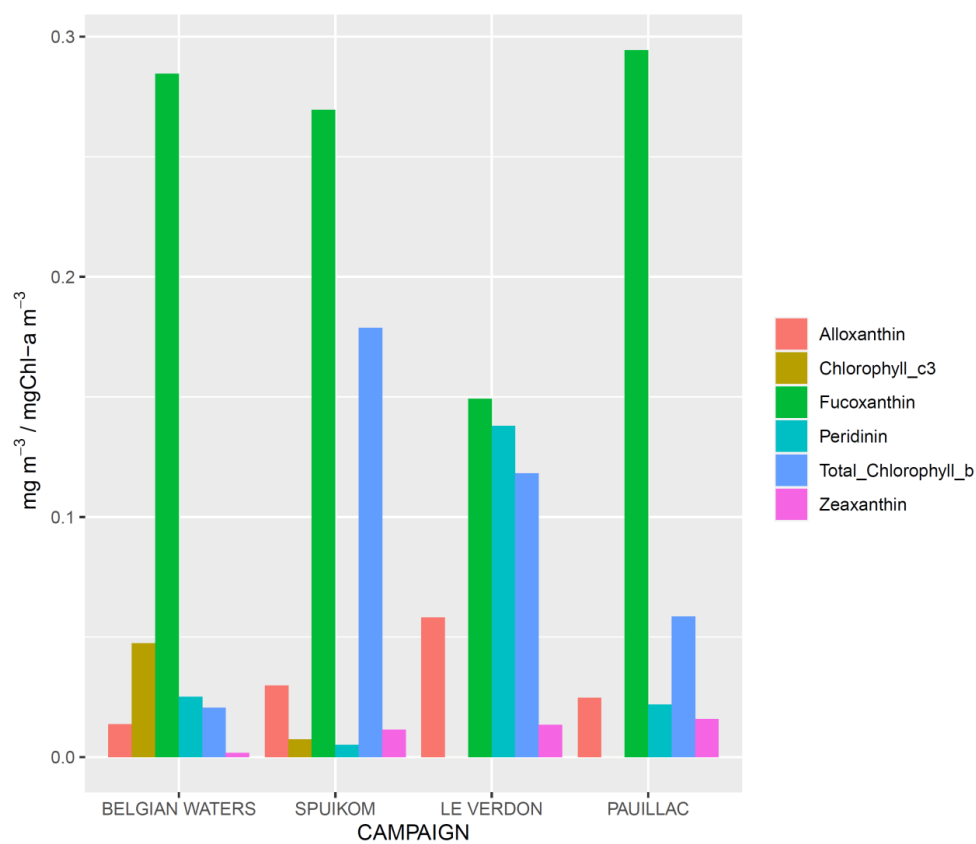


Figure 5: For each campaign, average concentration of alloxanthin, fucoxanthin, peridinin, chlorophyll c3, zeaxanthin and total chlorophyll b normalized by the concentration in chlorophyll-a

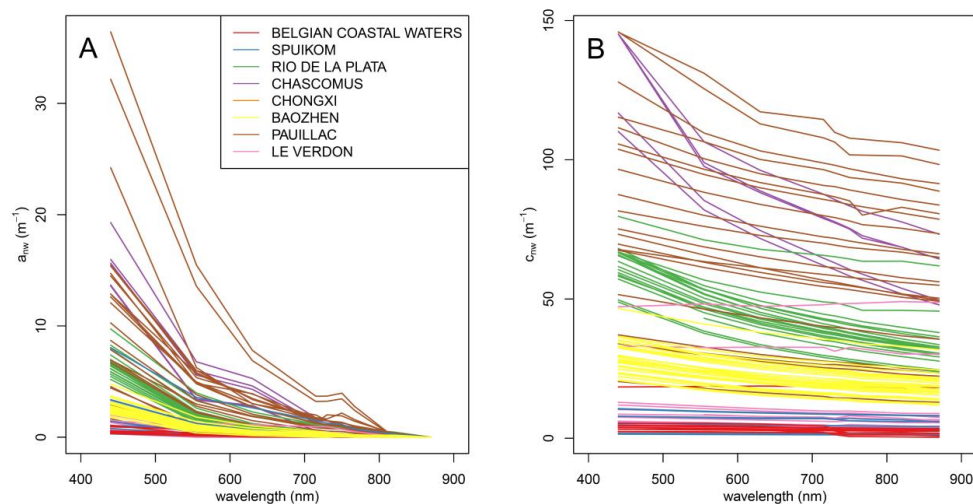


Figure 6: Non-water absorption (panel A) and attenuation coefficients (panel B) measured with the AC-9 instrument.

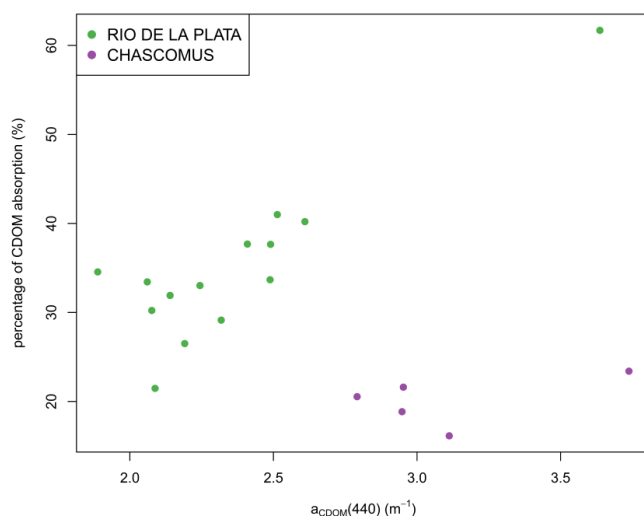


Figure 7: Percentage of CDOM absorption as a function of $a_{\text{CDOM}}(440)$. a_{CDOM} was only measured during two campaigns.

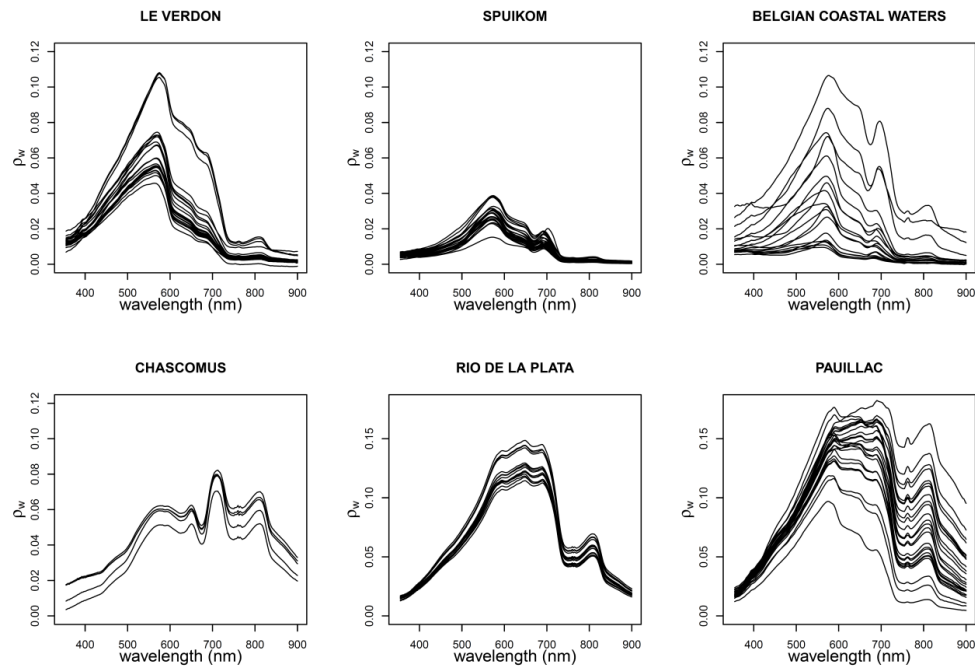


Figure 8: Water reflectance spectra from each sample site.

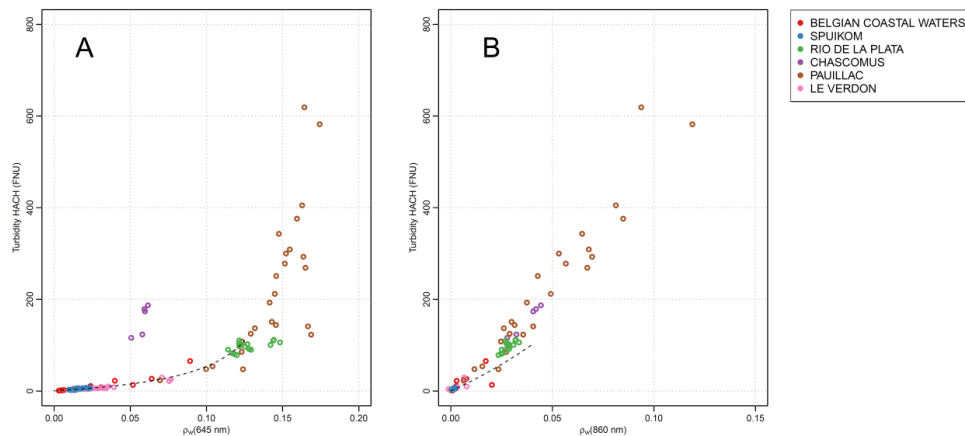


Figure 9: Turbidity as a function of water reflectance at 645 nm (panel A) and 850 nm (panel B). Black dotted line represents the model of Nechad et al. (2009) between 0 and 100 FNU.