



# 1 **Spatially resolved hourly traffic emission over megacity Delhi** 2 **using advanced traffic flow data**

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14 **Abstract.** This paper presents a bottom-up methodology to estimate multi-pollutant hourly  
15 gridded on-road traffic emission using advanced traffic flow and speed data for Delhi. We have  
16 used the globally adopted COPERT (Computer Programme to Calculate Emissions from Road  
17 Transport) emission functions to calculate the emission as a function of speed for 127 vehicle  
18 categories. At first the traffic volume and congestion (travel time delay) relation is applied to  
19 model the 24-hour traffic speed and flow for all the major road links of Delhi. The modelled  
20 traffic flow and speed shows an anti-correlation behaviour having peak traffic and emissions  
21 in morning-evening rush hours. We estimated an annual emission of 1.82 Gg for PME  
22 (Exhaust particulate matter), 0.94 Gg for BC (Black Carbon), 0.75 Gg for OM (Organic  
23 matter), 221 Gg for CO (Carbon monoxide), 56 Gg for NO<sub>x</sub> (Oxide of Nitrogen), 64 Gg for  
24 VOC (Volatile Organic Carbon), 0.28 Gg for NH<sub>3</sub> (Ammonia), 0.26 Gg for N<sub>2</sub>O (Nitrous  
25 Oxide) and 11.38 Gg for CH<sub>4</sub> (Methane) for 2018. The hourly emission variation shows  
26 bimodal peaks corresponding to morning and evening rush hours and congestion. The  
27 minimum emission rates are estimated in the early morning hours whereas the maximum  
28 emissions occurred during the evening hours. Inner Delhi is found to have higher emission flux  
29 because of higher road density and relatively lower average speed. Petrol vehicles dominate  
30 emission share (> 50%) across all pollutants except PME, BC and NO<sub>x</sub>, and within them the  
31 2W (Two-wheeler motorcycles) are the major contributors. Diesel fuelled vehicles contribute  
32 most of the PME emission. Diesel and CNG vehicles have a substantial contribution in NO<sub>x</sub>  
33 emission. This study provides very detailed spatio-temporal emission maps for megacity Delhi,



34 which can be used in air quality models for developing suitable strategies to reduce the traffic  
35 related pollution. Moreover, the developed methodology is a step forward in developing real-  
36 time emission with the growing availability of real-time traffic data. The complete dataset is  
37 publicly available on Zenodo at <https://doi.org/10.5281/zenodo.6553770> (Singh et al., 2022).

38

39 **Key words:** COPERT, Multi-pollutant emission inventory, Diurnal Emission, Road transport,  
40 Exhaust emissions, Air quality.

41

## 42 **1 Introduction**

43 Exposure to vehicular emissions poses a greater risk to the air quality and human health (Lipfert  
44 et al., 2008; Salo et al., 2021, GBD 2021). On-road transport is the major contributor to the  
45 ambient air pollution and greenhouse gas emissions in urban areas, mainly near roads (Singh  
46 et al., 2014), therefore they are an important component of the local air quality management  
47 plans and policies (Gulia et al., 2015; DEFRA, 2016; NCAP, 2019; Sun et al., 2022). The actual  
48 traffic emission depends on several dynamic factors, such as emission factors, traffic volume,  
49 speed, vehicle age, road network and infrastructure, road type, fuel, driving behaviour,  
50 congestion etc. (Pinto et al, 2020; Jiang et al., 2021; Deng et al., 2020). Traffic emission  
51 modelling has evolved and improved over recent years, however gaps still exist because of the  
52 complexity and data involved in the emission inventory development. Moreover, the reliability  
53 of the emission decreases further when the emissions are spatially and temporally segregated  
54 (Super et al., 2020, Osses et al., 2021). There are differences in the reliability of emission  
55 inventories of developed and developing countries because of lack of space-time input data in  
56 developing countries (Pinto et al, 2020). The uncertainty associated with emission inventory is  
57 further propagated in air quality models making mitigation studies more challenging, mainly  
58 for developing countries such as India which is already facing air pollution issues (Pandey et  
59 al., 2021).

60 India is among the top 10 economies (6th GDP rank) in the world in 2020 (GDP, 2020) and is  
61 recognized as a developing country. The population and economic growth have led to dense  
62 urbanisation with poor air quality in cities (Ravindra et al., 2019; Liang et al., 2020; Singh et  
63 al., 2021). India hosts 22 cities among the top 30 polluted cities in the world (IQAIR, 2020).  
64 The national capital of India, Delhi, has pollution levels exceeding NAAQS and WHO  
65 guideline values (Singh et al., 2021). Earlier studies have estimated on-road traffic as the major  
66 local contributor to Delhi pollution (CPCB 2010; Sharma et al., 2016) along with long range



67 transport sources associated with stubble burning and dust leading to severe pollution episodes  
68 (Liu et al., 2018; Bikina et al., 2019; Khaiwal et al., 2019; Beig et al., 2020; Singh et al.,  
69 2020).

70 Delhi traffic exhaust (tailpipe) emissions have been studied extensively using different  
71 methodology for years. The emissions estimated by various studies show large variations (see  
72 comparison tables in Guttikunda and Calori, 2013; Goyal et al., 2013; Sharma et al., 2016;  
73 Singh et al., 2018, and in Table 5) suggesting that the emissions have large uncertainties  
74 associated with the method and data used. Most of the studies adopted a bottom-up  
75 methodology to calculate the total emission over Delhi based on the registered vehicles and  
76 average vehicle kilometre travelled (VKT) multiplying with emission factors. A few studies  
77 (e.g., Sharma et al., 2016; Singh et al., 2018, 2020) use an on-road traffic flow approach where  
78 emission is estimated for each line source (road link) then spatially segregated (Tsagatakis et  
79 al., 2020, Spatial of emissions methodology). CPCB (2010), Goyal et al. (2013) further  
80 spatially desegregated the total emissions to 2 km × 2km resolution but the method of gridding  
81 is not discussed in detail. Sharma et al. (2016) and TERI (2018) also estimated 2km × 2km and  
82 4km × 4km gridded emission respectively, by adopting a per grid traffic flow method.  
83 Guttikunda and Calori (2013) estimated the 1km × 1km gridded emission by disaggregating  
84 the net emission using various spatial proxies like gridded road density. Though these studies  
85 with coarser resolution are helpful for identifying the emission hotspots but they lack actual  
86 traffic flow information disaggregated by road type and vehicle type within the grids. Moreover  
87 their emission estimate show large variations. For e.g., Das and Parikh (2004) and Nagpure et  
88 al. (2013) estimated traffic emission using VKT methodology for the same base year 2004,  
89 however their estimates varied by a factor of two or more. The annual emission estimate around  
90 year 2010 by CPCB (2010), Sahu et al. (2011, 2015), Goyal et al. (2013), Guttikunda and Calori  
91 (2013) and Singh et al. (2018) varied considerably from 3.5 Gg to ~15Gg for PM emission and  
92 30 Gg to 200 Gg for NO<sub>x</sub> emissions. The VKT based estimation approaches (Nagpure et al.,  
93 2013; Goel et al., 2015a; TERI 2018) tend to estimate higher emission compared to the traffic  
94 flow methodology (Sharma et al., 2016; Singh et al., 2018). A 40% increase in PM<sub>2.5</sub> emission  
95 in 2018 as compared to 2010, is reported by SAFAR (2018) attributed to the increase in  
96 vehicular growth.

97

98 Most of the studies for Delhi use EFs developed by ARAI (Automotive research association of  
99 India, ARAI; 2008) and a few studies have used EFs from IVE (International Vehicular  
100 Emission Model by USEPA, Davis et al., 2005) and COPERT (Ntziachristos et al., 2019).



101 ARAI EFs are measured in laboratory conditions, operating the vehicles in variable speed  
102 known as the Indian driving cycle (IDC, ARAI., 2008). The IVE emission factors are a function  
103 of the power bins of the vehicle engine, whereas in COPERT emission factors are a function  
104 of average vehicle speed, vehicle technologies, estimated pollutants, correction methods, and  
105 adjustments to local conditions. (Cifuentes 2021). Goyal et al. (2013) used the IVE model to  
106 estimate the traffic emission over Delhi for the year 2008 and also studied the diurnal emission  
107 at a specific location. However the study is limited to a fixed major traffic intersection only.  
108 Kumari et al. (2013) used the COPERT-3 emission factor to estimate emission for Indian cities,  
109 focusing on the multi-year (19991-2006) evolution of vehicular emission. However, this study  
110 estimates the total emissions based on registered vehicles and does not provide spatial  
111 segregation. COPERT Tier-3 emissions have been used for comparison with real-world  
112 measured emission factors (Jaikumar et al., 2017; Choudharyu et al., 2019). Jaikumar et al.  
113 (2017) identified vehicle idling is the major factor in the deviation between model-based  
114 estimation and measured emission as the vehicles spend 20% of their time in idling mode.

115

116

117 The traffic volume and speed information over each road are vital for accurate emission  
118 estimation. The data over Delhi has been very limited, therefore studies have used the VKT  
119 approach which uses the number of registered vehicles to estimate the emission.

120 To the best of our knowledge, despite several studies for Delhi, none of the studies have studied  
121 Delhi emissions using advanced and detailed traffic data and speed based EFs to estimate the  
122 hourly gridded emissions at high resolution. Moreover, most of the studies are limited to the  
123 estimation of PM, NO<sub>x</sub>, CO and HC only. The availability of recent detailed traffic data and  
124 speed volume relation (Malik et al., 2018) as a part of the Transportation research and injury  
125 prevention programme (TRIPP) of IIT Delhi provides an opportunity to estimate and improve  
126 the emissions over Delhi. To the best of our knowledge, this is the first study of its kind which  
127 considers advanced traffic flow data and estimates the hourly multi-pollutant emissions as a  
128 function of speed.

129

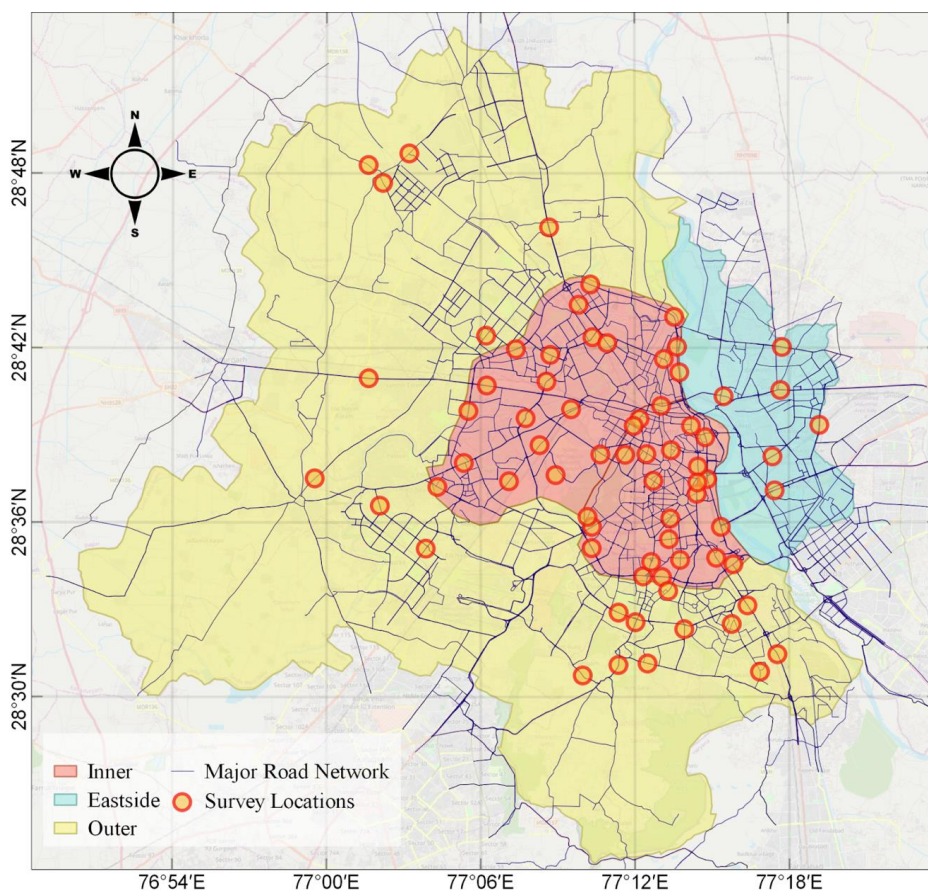
130 In this study, we have adopted a globally accepted methodology based on COPERT-5 Tier3 to  
131 estimate the hourly gridded emission for Delhi at high resolution for 2018. COPERT EFs have  
132 been used in many studies Alamos et al. (2021) for Chile, Mangones et al. (2019) for Bogota  
133 Cifuentes et al. (2021) for Manizalesto, Wang et al. (2010) for Chinese cities, Vanhulsel et al.  
134 (2014) for Belgium, Tsagatakis et al., (2019) for the national emission inventory over the UK



135 and also has been used by many around the globe (<https://www.emisia.com/utilities/copert/>).  
136 We combine advanced traffic volume and speed data (TRIPP, Malik et al., 2018) with speed  
137 based emission factors to calculate the emissions. The methodology considers different vehicle  
138 types, fuel type, engine capacity, emission standard and other key parameters such as  
139 congestion to estimate the emission for each road. We estimate the emission of particulate and  
140 gaseous pollutants, PME (Particulate matter exhaust) BC (Black carbon), OM (Organic  
141 matter), CO (Carbon monoxide), NO<sub>x</sub> (Oxides of Nitrogen), VOC (Volatile Organic  
142 Compound), NH<sub>3</sub> (Ammonia) and greenhouse gases, N<sub>2</sub>O (Nitrous Oxide) and CH<sub>4</sub> (Methane).  
143 We study the diurnal and spatial variability in the emission and identify the most polluting  
144 vehicle category, hotspots and the time when traffic emissions are highest. This study provides  
145 very detailed spatio-temporal emission maps for megacity Delhi that can be used in air quality  
146 models for developing suitable strategies to reduce the traffic related pollution. Moreover, the  
147 developed methodology is also a step forward in developing real-time emission models in the  
148 future with growing availability of real-time traffic data.  
149

## 150 **2 Methodology:**

151 We estimated the emissions for 2018 over the National Capital Territory (NCT) of Delhi having  
152 an area of 1483 sq. km (Fig. 1) and a population of 16.8 million (Census, 2011). The domain  
153 has been further divided into three regions (viz. Inner, Outer and Eastside), as shown in Fig. 1,  
154 to study the spatial variation in the emissions. Inner Delhi constitutes the major business hubs  
155 and workplaces within the ring road and the Outer is the area away from the ring road whereas  
156 the Eastside is the east part beyond the Yamuna river.



157

158 Figure 1. Map showing the study domain with TRIPP survey locations and the major road links  
159 over Delhi. The domain is segregated to three regions (Inner, Eastside and Outer) shown in  
160 different colours. The background map is from ©OpenStreetMap contributors. Distributed  
under ODbL v1.0 (<https://www.openstreetmap.org/>).

161 A bottom-up emission methodology has been adopted and a python-based model has been  
162 developed to estimate gridded hourly emissions of major pollutants over an urban area. The  
163 model estimates emission of PME, BC, OM, CO, NO<sub>x</sub>, VOC, NH<sub>3</sub>, N<sub>2</sub>O and CH<sub>4</sub>. The model  
164 uses hourly traffic activity and COPERT based emission factors as a function of hourly speed  
165 for each road link across Delhi. The major vehicle categories include 2W (Two wheeler motor  
166 bikes), 3W (Auto rickshaws), CAR (Passenger cars), BUS (Buses), LCV (Light Commercial  
167 Vehicles) and HCV (Heavy Commercial Vehicles).  
168



## 169 **2.1 Traffic Activity**

170 Classified traffic volume and speed study of Delhi (Malik et al., 2018) provides traffic count  
171 and speed for the roads of Delhi based on the Traffic volume and speed measurements  
172 conducted at 72 locations (Fig. 1) over Delhi in the year 2018 as a part of Transportation  
173 research and injury prevention programme (TRIPP) of IIT Delhi. We will refer to this dataset  
174 as TRIPP data from now on. TRIPP provides hourly traffic from 08:00-14:00 hours for eight  
175 fleet types (2W, 3W, Cars, Buses, Minibuses, HCV, LCV and NMV: Non-motorized vehicle)  
176 on over twelve thousand major road links over Delhi (Malik et al., 2018). These road links are  
177 further classified into five road classes (RClass1 to RClass5) based on the width of the road  
178 (Table S2). More detail of TRIPP traffic flow and its methodology is available elsewhere  
179 (Malik et al., 2018; Malik et al., 2021). As the TRIPP data is only available for 0800-1400  
180 hours, we use speed-flow-density relationship by Malik et al. (2021) to estimate the hourly  
181 traffic for each road link in Delhi.

### 182 **2.1.1 Generating traffic flow from congestion**

183 The relation between traffic volume and congested speed has been studied extensively using  
184 Greenshield model, the Greenberg model and the Underwood model (Wang et al., 2014;  
185 Hooper et al., 2014) and used by many studies (Jing et al., 2016; Yang et al., 2019) to estimate  
186 the traffic from the congestion for emission development. For Delhi, this relation is  
187 mathematically represented in Eq. (3) of Malik et al. (2021). By rearranging, the same can be  
188 written as Eq. (1) of this paper.

$$x_i = c_i \left( \frac{1}{\alpha} \left( \frac{V_{o,i}}{V_{Congested,i}} - 1 \right) \right)^{\frac{1}{\beta}} \quad (1)$$

189

190 Where,

191  $x_i$  = Traffic flow for road link  $i$

192  $c_i$  = Traffic capacity for road link  $i$

193  $V_{Congested,i}$  = Speed during congestion (km/h) for link  $i$

194  $V_{o,i}$  = Free flow velocity (FFV) of traffic for road link  $i$

195  $\alpha$  and  $\beta$  = constants (Table 1, Malik et al., 2021)

196



197 Traffic volume and road capacity determines the traffic speed. Increasing traffic volume leads  
198 to travel time delay (congestion) which further results in road traffic congestion resulting in  
199 increased traffic volume and decreased speed leading to traffic delays. Congested traffic speed  
200 ( $V_{congested}$ ) is inversely proportional to the *congestion* (Afrin and Yodo., 2020). Here we define  
201 *congestion* as percentage increase in travel time, i.e. 50% congestion level in a city means that  
202 a trip will take 50% more time than it would during baseline uncongested conditions. In this  
203 study, we have used hourly *congestion* data for Delhi obtained from TomTom  
204 ([https://www.tomtom.com/en\\_gb/traffic-index/about/](https://www.tomtom.com/en_gb/traffic-index/about/)). TomTom is one of the leading  
205 mapping and navigation services providing urban congestion worldwide. Congestion data has  
206 been taken for different days of the week then combined to create weekdays (Monday to  
207 Friday) and weekend (Saturday and Sunday) profiles. Because FFV ( $V_o$ ) and *congestion* are  
208 known for a road link,  $V_{congested}$  for weekdays and weekend has been calculated for each road  
209 link using the Eq. (2).

$$V_{congested} = \frac{V_o}{1 + congestion} \quad (2)$$

210 Further, substituting the value of  $V_{congested}$  in Eq. (1), we get a relation between congestion and  
211 traffic flow (Eq. 3) that has been used to estimate the weekdays and weekend traffic flow for  
212 all the road links in personal car units (PCU).

213

$$x_i = c_i \left( \frac{congestion}{\alpha} \right)^{\frac{1}{\beta}} \quad (3)$$

214

215 PCU values for Delhi are taken from Malik et al. (2021) and are as follows (a) 1.0 for CAR,  
216 (b) 0.5 for 2W, (c) 1.0 for 3W, (d) 3.0 for BUS, (e) 1.5 for LCV and (f) 3.0 for HCV. We use  
217  $C_i$  from TRIPP and *congestion* from TomTom. The values  $\alpha$ ,  $\beta$  and  $C_i$  used in this study are  
218 taken from Malik et al., (2021), and are shown in Table S2. Further, the speed and traffic  
219 volume has been corrected for each road link to match the observed PCU in TRIPP dataset for  
220 a better agreement. The hourly estimated traffic for each road link is further decomposed from  
221 PCU to different fleet categories using the percentage share provided by Malik et al., 2018.  
222 The hourly estimated traffic has been further corrected for the LCV and HCV using the  
223 percentage share provided by CRRI (Central Road Research Institute; Errampalli et al., 2020)  
224 to account for the travel restrictions of good vehicles peak traffic hours. To validate our results,





225 the annual VKT estimated for each fleet category is found to be comparable with that available  
226 in literature (Goel et al., 2015b; Malik et al., 2019). For simplicity, minibus has been combined  
227 with the bus category and NMVs are not used in this study.

## 228 **2.2 Vehicular Classification:**

229 The six types of primary vehicle categories (2W, 3W, CAR, BUS, LCV and HCV) have been  
230 further classified into 127 categories (Table S1) according to fuel, engine capacity and emission  
231 standards to match the COPERT-5 vehicular classification. The fuel share of petrol/gasoline,  
232 diesel and CNG/LPG vehicles in Delhi for passenger and freight vehicles has been obtained  
233 from Dhyani and Sharma. (2017) and Malik et al. (2019) respectively. The engine share for  
234 primary vehicle categories has been taken from working papers (Sharpe and Sathiamoorthy.,  
235 2019; Anup and Yang., 2020; Deo and Yang., 2020) of the International Council on Clean  
236 Transportation (ICCT). In India, the emission norms/standards, known as Bharat Stage (BS),  
237 can be considered equivalent to the European Emission Standards - Euro, have been introduced  
238 in a phased manner. These norms were introduced for passenger cars then later extended to  
239 other vehicle categories. For example, the BS-I (India-2000) for passenger cars was  
240 implemented in 2000 followed by BS-II, BS-III and BS-IV in 2005, 2010 and 2017  
241 respectively. The BS-VI for passenger cars are introduced recently in 2020 therefore has not  
242 been considered in our study. For Delhi, the timeline of BS implementation for passenger cars  
243 and other vehicles are shown in Table S3. The vehicles prior to the implementation of BS  
244 norms have been considered as Conventional (or BS-0 for simplicity). The BS share of the  
245 vehicles has been derived using the survival function method described in (Goel et al., 2015b;  
246 Malik et al., 2019). The vehicle survival was calculated for the past twenty year by considering  
247 2018 as the base year and then the BS share was calculated based on the age of the vehicle with  
248 respect to 2018 (Table S4). The final share of the primary vehicle category as per fuel, engine  
249 and BS norms has been calculated by multiplying the fuel share, engine share and BS norms  
250 share and shown in Table S1. In this study, BS and EURO/Euro have been used  
251 interchangeably, and BS-I to BS-IV or BS1 to BS4 or EURO1 to EURO4 represent the same  
252 emission standard.



253

### 254 **2.3 Emission Factors**

255 Emission factor (EF) is a crucial parameter required for emission estimation. Road traffic  
256 vehicular emission depends on a variety of factors such as vehicle type, fuel used, engine types,  
257 driving pattern, road type, emission legislation type (BS/EURO) and speed of the vehicle. We  
258 have adopted the recent COPERT-5 tier-3 methodology and used the speed based emission  
259 factor (<https://www.emisia.com/utilities/copert/>) for 127 vehicle types (Table S1) and  
260 according to the emission legislation up to BS/EURO-4 (As in 2018 BS-VI is not  
261 implemented). The EF as a function of vehicle speed ( $v$ ) is calculated using Eq. (4).

$$EF(v) = \frac{(\alpha \times v^2) + (\beta \times v) + \gamma + \left(\frac{\delta}{v}\right)}{(\varepsilon \times v^2) + (\zeta \times v) + \eta} \quad (4)$$

262

263

264 Where,

265  $v$  is the speed,

266  $\alpha, \beta, \gamma, \delta, \varepsilon, \zeta$  and  $\eta$  are coefficients that varies with vehicle type

267

268 The coefficients for each pollutant and vehicle category are taken from the COPERT-5  
269 database (COPERT-5 Guide book, 2020). The emission factors are further corrected for the  
270 emission degradation occurring in older vehicles considering the mileage as discussed in  
271 (COPERT-5 Guide book, 2020). The emissions are further adjusted with a factor of 1.2 to  
272 account for real-time driving behaviour (frequent braking, acceleration, deceleration) as per the  
273 study by Lejri et al., (2018). The non-exhaust emissions (Singh et al., 2020) have not been  
274 calculated in this study. As COPERT does not provide the EFs for the 3W CNG category, we  
275 have used EFs of CNG mini CAR for this. BC and OM emission are computed using the  
276 fraction (by COPERT-5 Guide Book, 2020) from PM exhaust.

277

### 278 **2.4 Emission calculation**

279 The model calculates hourly emissions for each road link of finite length and uses hourly traffic  
280 volume and emission factors as a function of speed for 127 vehicle categories (Table S1). The



281 hourly emission rate ( $Q$ ) for each road link is calculated using Eq. (5). The total emission for a  
282 given hour is calculated by taking the sum of emission across all vehicle categories.

283

284

$$Q_{i,h}^p = \sum_j V_{i,j,h} \times EF_j^p(v_{i,h}) \times L_i \quad (5)$$

285

286

287 Where

288  $Q_{i,h}^p$  is emission rate of a pollutant  $p$  for road link  $i$  and at hour  $h$ , where  $h=0$  to  $23$

289  $V_{i,j,h}$  is the traffic volume of vehicle category  $j$  for road link  $i$  at hour  $h$ , where  $j=1$  to  $127$

290  $L_i$  is the length of road link  $i$

291  $EF_j^p(v_{i,h})$  is the emission factor of pollutant  $p$  for vehicle category  $j$  as a function speed  $v_{i,h}$

292 for road link  $i$  at hour  $h$ .

293 The hourly emissions have been calculated for each pollutant over each road link then gridded

294 at  $100 \text{ m} \times 100 \text{ m}$  resolution using the methodology described in Singh et al., (2018, 2020) to

295 produce the hourly gridded emission inventory for Delhi.

## 296 **3 Results**

### 297 **3.1 Diurnal variation of traffic volume and speed**

298 The estimated hourly traffic volume (in PCU) and speed profiles for Delhi are shown in Fig. 2.

299 An anticorrelated diurnal variation is seen in the traffic volume and speed. The weekdays traffic

300 volume tends to have a bimodal profile with a morning peak (09:00-11:00) and an evening

301 peak (18:00-20:00). A similar traffic volume profile has also been observed by other studies

302 over Delhi (Dhyani and Sharma., 2017; Sharma et al., 2019). Similar bimodal traffic profile is

303 also observed over the cities around the world subject to the city specific travel demand (Järvi

304 et al., 2008 for Helsinki; Jing et al., 2016 for Beijing) The evening peak traffic volume tends

305 to be 40% higher than the morning peak. The night-time goods vehicle share is more in

306 comparison to the passenger and personal vehicles (Fig S1). However, the vehicular

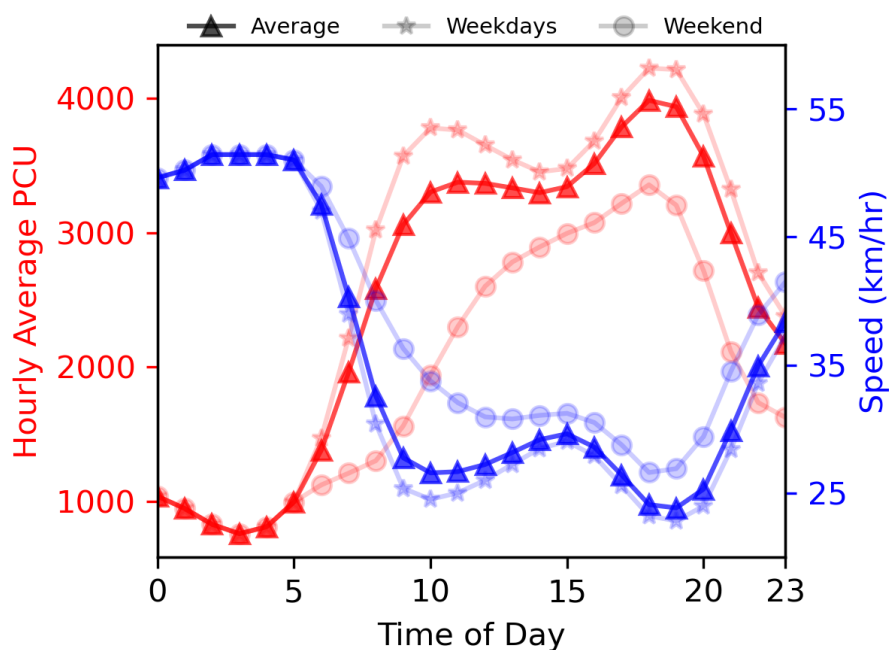
307 composition varies with respect to the road classes (Table S5). The weekend traffic volume

308 does not show a morning peak due to closure of the offices/workplaces and shows evening

309 peaks due to shopping and other weekend activities. As usual the minimum traffic volume is



310 observed at night (00:00-04:00 hours) because of the reduced human and commercial activities.  
311 Due to the minimum traffic at night, the traffic moves with an average speed of  $51 \pm 6$  km/h  
312 with almost no congestion. As traffic volume increases, it starts to build congestion, leading to  
313 reduced speed. The average speed during the weekdays morning peak hours is estimated to be  
314  $30 \pm 14$  km/h whereas the evening speed is estimated to be  $28 \pm 15$  km/h. The evening congestion  
315 leads to an average 46% reduction in the average speed increasing the travel time by a factor  
316 of two. We calculated the average profiles for each road link by combining weekdays and  
317 weekends and used them in the emission calculations. The estimated profiles averaged across  
318 all road links are shown in Fig. 2. We estimate 27, 31, 0.95 and 3.1 billion-VKT driven by  
319 CAR, 2W, HCV and LCV categories respectively, which compares well with earlier estimates  
320 by (Malik et al., 2019) and (Goel et al. 2015b).  
321



322

323 Figure 2. Weekdays, weekend and average diurnal profile for traffic volume in average PCU  
324 (red) and average speed (blue) over Delhi. The legend reflects the different markers used for  
325 weekdays, weekend and average profile.

### 326 3.2 Emission inventory

327 A multi-pollutant hourly and high spatial resolution (100m × 100m) emission inventory has  
328 been prepared for Delhi. As an example, the spatial distribution of  $\text{NO}_x$  emission at 03:00-  
329 04:00, 09:00-10:00, 15:00-16:00 and 18:00-19:00 hours, representing early morning, morning



330 peak, afternoon and evening peak respectively, has been shown in Fig. 2. The emission rate  
331 during the evening peak hours is the highest during the day followed by morning peak hours.  
332 The high traffic volume along with traffic congestions lead to more emissions during the peak  
333 traffic hours (Jing et al., 2016). The emission during the afternoon hours is comparable or less  
334 than that of the morning hours whereas the early morning emissions are lowest because of low  
335 traffic volume moving with free flow speed. The diurnal profile of emissions has been  
336 discussed in detail in Section 3.5.

337 The annual emissions have been calculated by summing the hourly emissions to get daily  
338 emissions and then multiplying with 365 (number of days in a year) to get annual emissions.  
339 We estimated an annual emission of 1.82 Gg for PME, 0.94 Gg for BC, 0.75 Gg for OM, 221  
340 Gg for CO, 56 Gg for NO<sub>x</sub>, 64 Gg for VOC, 0.28 Gg for NH<sub>3</sub>, 0.26 Gg for N<sub>2</sub>O and 11.38 Gg  
341 for CH<sub>4</sub> in 2018.

342

### 343 **3.3 Spatial variation**

344 The hourly emissions over Delhi have been summed together to calculate the daily emissions  
345 for all the pollutants. The spatial variation of daily mean emission rate has been analysed over  
346 three selected regions, viz. inner, outer and eastside Delhi (as shown in Fig. 1). The total  
347 emission for each pollutant and for each region has been tabulated in Table S6. Outer Delhi  
348 region has the highest emission (51-53%) for all the pollutants because of its largest area of  
349 1106 km<sup>2</sup> which is 4.5 times of inner Delhi. To avoid the influence of area on the emissions,  
350 we have calculated the emission flux (i.e. emission per unit area) and shown in Table S7. The  
351 emissions flux is highest for inner Delhi followed by eastside and outer Delhi region. For all  
352 pollutants, the emissions flux in inner Delhi is 40 - 50 % higher than the average emission of  
353 Delhi whereas the emissions flux in outer Delhi is ~46% lower. The emission flux is  
354 consistently high along the grids containing major roads (Fig. 3), intersections and major  
355 business hubs. Inner Delhi consists of major business hubs, workplaces and government  
356 offices, which entertain more vehicular activity in this region resulting in congestion leading  
357 to reduced speed and enhanced emissions. The daytime average speed across all roads in Inner  
358 Delhi is 29 km/h which is lower than the daytime average speed of 32 km/h in outer Delhi. The  
359 lower speed and higher traffic density influences the economic driving behaviour resulting in  
360 frequent braking, idling, acceleration and deceleration that enhances the vehicular emission.  
361 Moreover, the morning and evening peak hours with higher traffic and lower speed have the  
362 highest emission as compared to the rest of the day. In these heavy congested hours the vehicle  
363 is forced to run in lower speed which boosts the emission.



364

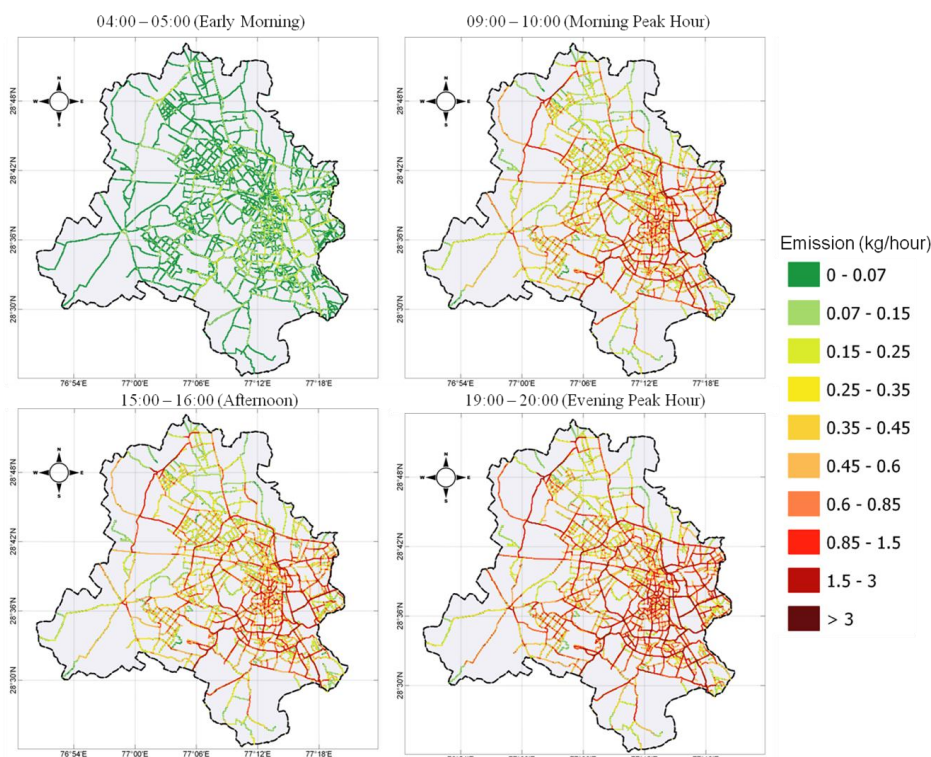
365 **3.4 Emissions along the Road class**

366 The emissions along the five road classes used in this study have been calculated and shown in  
 367 Table 1 and the hourly variation of emission has been shown in Fig. 4. RClass3 has a  
 368 substantial emission share (~35%) across all pollutants followed by RClass5 and RClass2,  
 369 whereas RClass1 holds the minimum emissions share (~2-3%). The dominant emission share  
 370 of RClass3 is due to the optimum vehicular activities over the longer road length. RClass2,  
 371 which are the feeder roads to the RClass3, RClass4 and RClass5, contribute ~23% to the  
 372 emission. The multi-lane wider roads, RClass4 and RClass5 contribute ~13-15 % and ~21-25  
 373 % respectively to the total emission. To remove the dependency of the road length, we  
 374 calculated the emission per km segment of a road. The emissions (per km) over multi-lane  
 375 wider roads (RClass4 and RClass5) are almost two times of the RClass3 (Table S8 and Fig.  
 376 S2) due to more traffic flow irrespective of the congested conditions. However, the emission  
 377 per lane per kilometre (Table S9) for RClass1 is found to be the highest because of lower speed  
 378 and congestion and major share of 2W. This shows that effective management of traffic in  
 379 narrow roads to reduce the congestion will be beneficial in reducing the pollution without  
 380 impacting the traffic volume. The multi-lane wider roads (RClass4 and RClass5) help the  
 381 vehicle to maintain an economic speed resulting in minimum congestion and lower emission,  
 382 however they are the emission hotspots in Delhi.

383 Table 1. Emission in Mega gram (Mg) per day (% share) across different road types.

RClass	PME	BC	OM	CO	NO <sub>x</sub>	VOC	NH <sub>3</sub>	N <sub>2</sub> O	CH <sub>4</sub>
RClass1	0.16 (3%)	0.09 (3%)	0.07 (3%)	19 (3%)	4 (2%)	5 (2%)	0.02 (2%)	0.02 (2%)	1.0 (3%)
RClass2	1.17 (23%)	0.61 (23%)	0.49 (23%)	139 (23%)	35 (23%)	41 (23%)	0.16 (21%)	0.16 (22%)	7.3 (23%)
RClass3	1.77 (35%)	0.9 (34%)	0.75 (36%)	228 (37%)	52 (34%)	67 (38%)	0.27 (35%)	0.25 (35%)	11.29 (36%)
RClass4	0.72 (14%)	0.38 (14%)	0.29 (14%)	84 (13%)	22 (14%)	23 (13%)	0.12 (15%)	0.11 (15%)	4.43 (14%)
RClass5	1.16 (23%)	0.62 (23%)	0.46 (22%)	132 (21%)	38 (25%)	37 (21%)	0.19 (25%)	0.17 (23%)	7.19 (23%)

384  
 385



386

387 Figure 3. Estimated gridded NO<sub>x</sub> emission at 100m × 100m spatial resolution at different  
388 time of the day representative of different congestion levels.

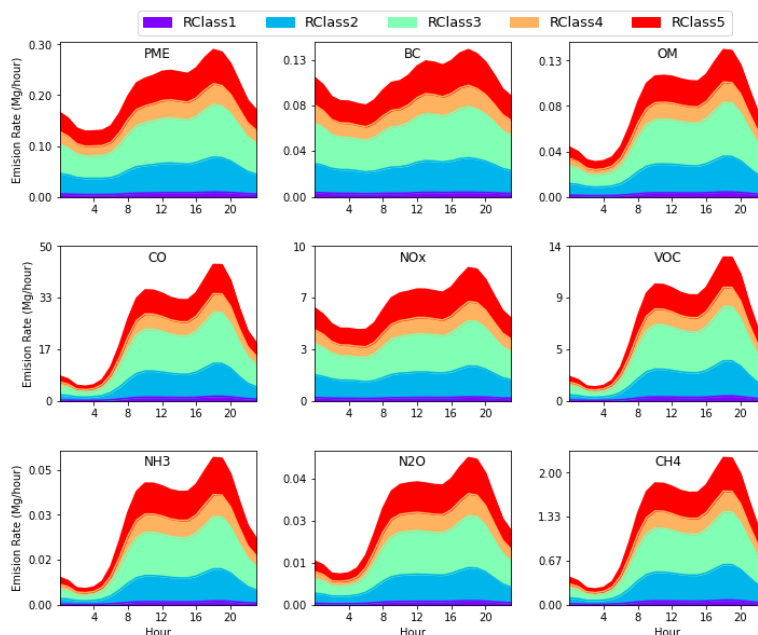
389

### 390 3.5 Diurnal variation of emission

391 Dynamic traffic volume and speed, as discussed in section 3.1, results in diurnal variation in  
392 the emissions during a day. Fig. 4 shows the hourly emissions (Mg/h) and contribution of each  
393 road class at each hour in Delhi. The temporal evolution of emission is linear with the traffic  
394 variation in a day with the minimum variation during the night-time and remarkable variation  
395 during the human active hours (08:00-20:00). Among different road types and for all the  
396 pollutants RClass1 has the lowest and RClass3 has the highest emission proportional to the  
397 traffic volume. A similar temporal variation of NO<sub>x</sub> emission rate is observed in a study, for  
398 different road types of Beijing (Jing et al., 2016). For most of the pollutants (except PME, BC  
399 and NO<sub>x</sub>), daytime (08:00 to 20:00) contributes ~70% to the daily emissions whereas the  
400 morning (09:00 to 11:00) and evening (18:00 to 20:00) rush hours alone altogether add 30-40%  
401 to the total emissions. The increasing activity of goods vehicle (HCV + LCV) during afternoon  
402 and night-time (Fig. S1) elevates the emission of PME, BC and NO<sub>x</sub> from these vehicles (Fig.  
403 5) resulting a different diurnal profile compare to other pollutants. The NO<sub>x</sub> and particulate



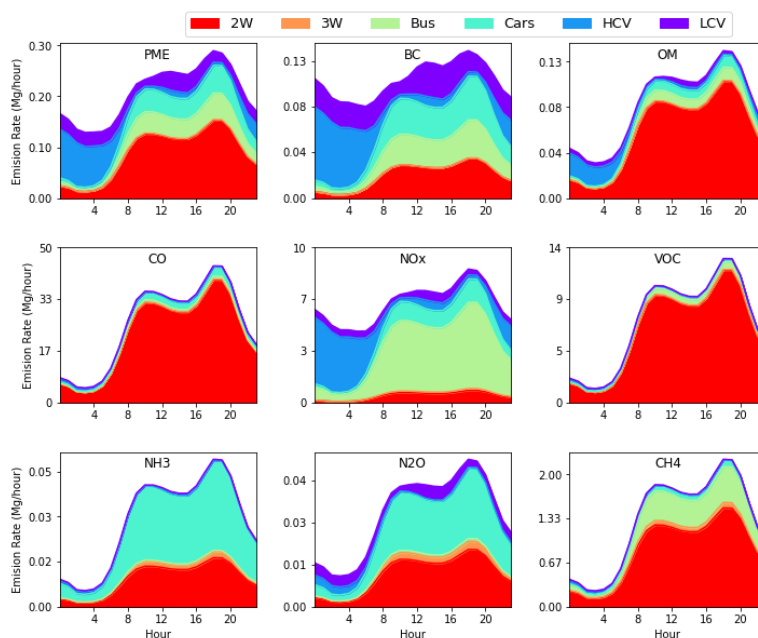
404 pollutants (PME and BC) emissions during late night hours (11:00-05:00) is relatively higher,  
405 adding up to 60% and 75% of total particulate and  $\text{NO}_x$  night-time emissions respectively as  
406 shown in Fig. 5. The contribution of vehicle type has been discussed in detail in section 3.6.  
407 The diurnal evolution of emission is also visible in the hourly spatial map shown in Fig. 3.  
408 Early morning with minimum traffic volume has lower emission whereas the evening rush hour  
409 with increasing congestion has higher emission. The density of higher emission grids (Fig. 3)  
410 in the inner Delhi region is higher compare to other regions throughout the day.  
411



412

413 Figure 4. Variation of hourly emission (in mega gram/hour) of the nine pollutants averaged  
414 across Delhi according to the five road classes (RClass1 to RClass5). Different colors  
415 indicate the hourly contribution of each RClass to the total emission.





416

417 Figure 5. Variation of hourly emission (mega gram/hour) of the nine pollutants averaged  
 418 across Delhi according to the major vehicle type. Different colors indicate the hourly  
 419 contribution of each vehicle type to the total emission.

420

### 421 3.6 Vehicular emission share

422 The percentage share of major vehicle types to the total emission of nine pollutants has been  
 423 calculated and shown in Table 2 and its hourly contribution is shown in Fig. 5. The 2W  
 424 vehicles, having major vehicular share (Table S5), are the major contributors to the total  
 425 emissions for all the pollutants except for BC, NO<sub>x</sub> and N<sub>2</sub>O. The goods vehicles (HCV and  
 426 LCV) contribute substantially, mainly during night-time, to the PME, BC and NO<sub>x</sub> emissions.  
 427 Buses have highest contribution to NO<sub>x</sub> emissions and substantial contribution to PME, BC  
 428 and CH<sub>4</sub>. Cars are the dominant source for NH<sub>3</sub> and N<sub>2</sub>O and contribute substantially to PME,  
 429 BC and NO<sub>x</sub> emissions. However, most of the emissions are from diesel cars.

430 Table 2. Emission in kg/day (% share) according to the vehicle types.

Vehicle	PME	BC	OM	CO	NO <sub>x</sub>	VOC	NH <sub>3</sub>	N <sub>2</sub> O	CH <sub>4</sub>
2W	2102 (41.6%)	500 (19.0%)	1475 (71.5%)	532316 (88.0%)	10600 (6.8%)	159582 (90.5%)	249 (32.6%)	249 (35.4%)	20588 (66.0%)
Cars	740 (14.6%)	537 (20.4%)	146 (7.1%)	42276 (7.0%)	20185 (12.9%)	3546 (2.0%)	458 (60.0%)	308 (43.8%)	1425 (4.6%)
3w	25 (0.5%)	3 (0.1%)	11 (0.5%)	3305 (0.5%)	1593 (1.0%)	952 (0.5%)	32 (4.2%)	35 (5.0%)	1151 (3.7%)



Buses	691 (13.7%)	459 (17.4%)	160 (7.8%)	12739 (2.1%)	75536 (48.4%)	9249 (5.2%)	4 (0.5%)	12 (1.7%)	7456 (23.9%)
HCV	787 (15.8%)	546 (21.2%)	171 (8.3%)	8645 (1.4%)	35404 (23.0%)	2057 (1.2%)	9 (1.2%)	24 (3.4%)	452 (1.4%)
LCV	636 (12.8%)	534 (20.7%)	87 (4.2%)	4803 (0.8%)	10547 (6.9%)	884 (0.5%)	11 (1.4%)	75 (10.7%)	126 (0.4%)

431

432 Table 3. Emission in kg/day (% share) according to fuel type.

Fuel	PME	BC	OM	CO	NO <sub>x</sub>	VOC	NH <sub>3</sub>	N <sub>2</sub> O	CH <sub>4</sub>
CNG	95 (1.9%)	14 (0.5%)	43 (2.1%)	12703 (2.1%)	45832 (29.8%)	9335 (5.3%)	68 (8.9%)	73 (10.4%)	9547 (30.6%)
Diesel	2698 (54.1%)	2052 (79.5%)	491 (23.9%)	25583 (4.2%)	91144 (59.2%)	5308 (3.0%)	36 (4.7%)	225 (32.0%)	805 (2.6%)
Petrol	2191 (44.0%)	514 (19.9%)	1517 (74.0%)	565799 (93.7%)	16890 (11.0%)	161628 (91.7%)	662 (86.4%)	406 (57.7%)	20848 (66.8%)

433

434 The vehicular fuel share to the total emission for each pollutant is shown in Table 3. Petrol  
 435 vehicles are the largest contributors to the CO (~94%), VOC (91%), NH<sub>3</sub> (86%), OM (74%),  
 436 CH<sub>4</sub> (67%) and N<sub>2</sub>O (58%) whereas diesel vehicles are the largest contributor to the BC  
 437 (~80%), NO<sub>x</sub> (59%) and PME (54%) emissions. The contribution of the CNG vehicles is  
 438 relatively smaller except for the NO<sub>x</sub> and CH<sub>4</sub> where they contribute to ~30 %, almost one  
 439 third, to the total emissions.

440

441 The larger contribution of petrol to the VOC, CO, OM and CH<sub>4</sub> emissions are dominated by  
 442 2W where we estimated that 2W in Delhi alone contribute 90%, 88%, 71%, and 66%  
 443 respectively as shown in Table 2. The contribution of 2W is also highest to PME (42%). The  
 444 larger share of 2W towards the CO emissions has also been reported earlier, 61% in Goyal et  
 445 al., (2013); 43% in Sharma et al., (2016) and 37% in Singh et al., (2018). Higher emission  
 446 share of 2W is due the higher emission factor of VOC in petrol fuelled 2W (Hakkim et al.,  
 447 2021) that has been also reported in a multi-year emission study over Delhi by Goel et al.  
 448 (2015a).

449

450 The PME emissions are dominated by diesel fuelled HCVs (16 %), LCVs (13%), Buses (14  
 451 %) and Cars (~13 %), whereas 2W are the main source in petrol fuelled vehicles contributing  
 452 ~42% to the total PME emissions. Earlier, Sharma et al. (2016) reported 33% share of 2W  
 453 emission in 2014. The share of petrol cars and CNG buses towards the PME, BC and OM  
 454 emissions is less than 2%. While it is clear that diesel powered vehicles are the major source  
 455 of PME emission, earlier studies have reported similar results but with large variations of HCVs  
 456 in emission share. The largest share of diesel fuelled HCV is reported as 92% by Goyal et al.



457 (2013), 46% by Sharma et al. (2016) and 33% by Singh et al. (2018). All these studies reported  
458 minimal emission share (less than 10% combining both diesel and petrol cars). The largest  
459 share of HCV, LCV and diesel Cars to BC emission is because of higher emission factors  
460 (Zavala et al., 2017) contributing to total urban BC emission as shown by Bond et al., (2013).

461

462 The petrol cars contribute more than half of the total  $\text{NH}_3$  emissions and among them the Euro  
463 2 with higher emission factor has the largest share of 39%. The diesel vehicles (HCVs, LCVs,  
464 diesel Buses and Cars) altogether contribute significantly to the PME, BC and  $\text{NO}_x$  emissions.  
465 The higher emission factor of diesel fuelled vehicles (Wu et al., 2012) clearly reflects in the  
466 emission share.

467

468 CNG buses have the highest share (27%) in  $\text{NO}_x$  emission and around 23% in  $\text{CH}_4$  emissions.  
469 The highest share of CNG is due to higher  $\text{NO}_x$  emission factor for CNG vehicles compared to  
470 petrol vehicles (Dimaratos et al., 2019). The larger share of ~15% from CNG buses to the total  
471 traffic  $\text{NO}_x$  emission is also reported in a study of CPCB (2010). In terms of Euro or BS  
472 standard, Euro 3 vehicles have the highest share (Table S10) in the total emission except for  
473  $\text{N}_2\text{O}$  and  $\text{NH}_3$ . This is mainly because of the highest share of Euro 3 vehicles in 2W, Buses,  
474 HCV and LCV (Table S4 in the Supplement). In the case of  $\text{N}_2\text{O}$ , the emissions are dominated  
475 by Euro 4 cars which have around 84% share to the total cars. For  $\text{CH}_4$ , the highest share of  
476 Euro 3 vehicles is due to the higher emissions from Euro 3 2W as the emission factor of petrol  
477 vehicles is higher (Clairotte et al., 2020).

478

479 In order to have a clear picture of the dominant polluting vehicle categories, we grouped  
480 different vehicle types into 35 categories and calculated the percentage share to the total  
481 emission of nine pollutants as shown in Fig. 6. We further identified the top five polluting  
482 vehicle categories for each pollutant and tabulated in Table 4. For PME, the top five polluting  
483 vehicles account for 55% of the total emissions which is dominated by petrol Euro 3 petrol  
484 2W and Euro 3 diesel HCVs. The BC emission is mainly driven by Euro 3 diesel HCVs, LCVs,  
485 Buses and the top five polluting vehicles account for 66% of the total emissions. The OM, CO,  
486 VOC emissions are dominated by 2W and the top five accounts for 71%, 89% and 91% of total  
487 emissions respectively.

488

489 Petrol fuelled cars and 2W hold the dominant share of  $\text{NH}_3$  emissions because of the larger EF  
490 compared to other categories (COPERT-5 Guide Book, 2020). For  $\text{N}_2\text{O}$ , 2W Euro 3 holds the

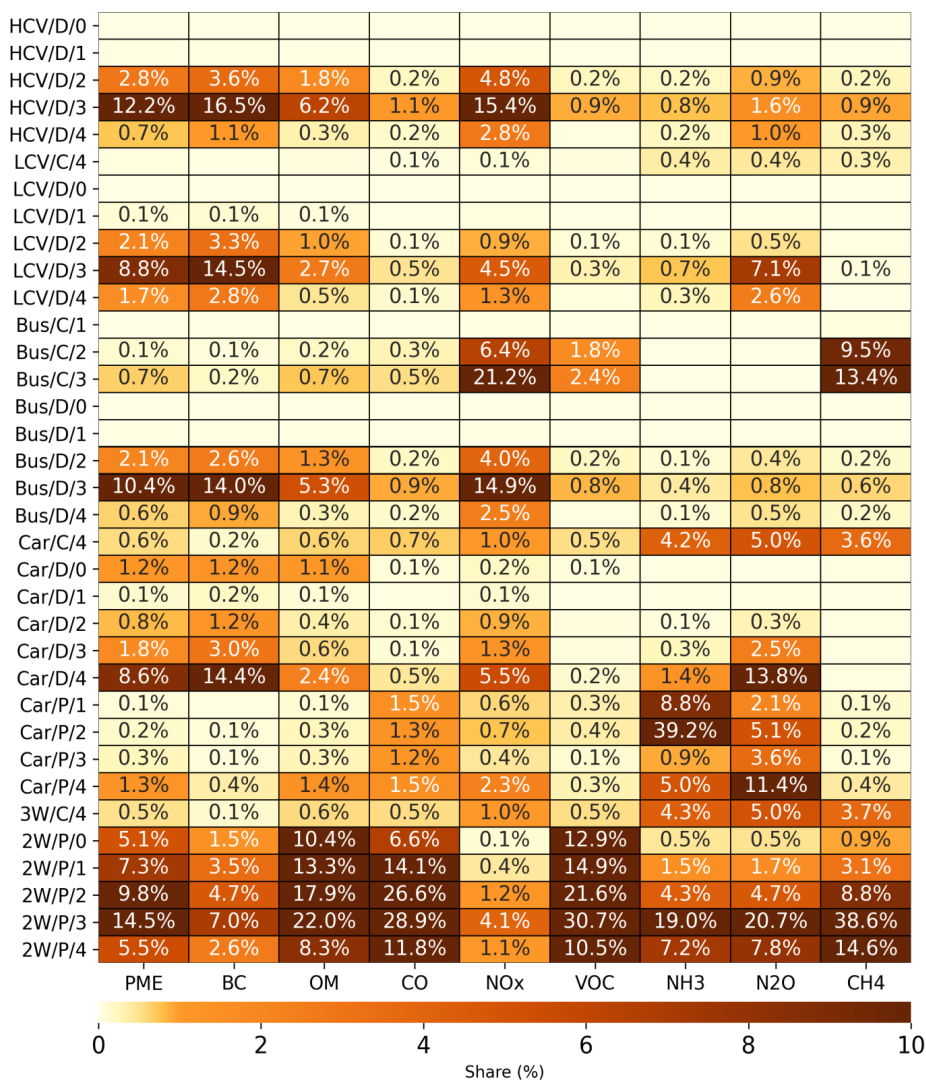


491 highest share of 21%, followed by EURO IV diesel and petrol cars. The top five contributors  
 492 to CH<sub>4</sub> emissions account for 86% of the total emissions which are dominated by 2W and CNG  
 493 buses. These two categories of vehicles altogether contribute to ~97% of the emissions.

494 Table 4. Top five polluting vehicle categories for each pollutant.

PME	BC	OM
Top 5 accounts for <b>55%</b> emissions 1. 14% from 2W (Petrol, Euro 3) 2. 12% from HCV (Diesel, Euro 3) 3. 10% from Bus (Diesel, Euro 3) 4. 10% from 2W (Petrol Euro 2) 5. 9% from LCV (Diesel Euro 3)	Top 5 accounts for <b>66%</b> emissions 1. 17% from HCV (Diesel Euro 3) 2. 14% from LCV (Diesel Euro 3) 3. 14% from Car (Diesel Euro 4) 4. 14% from Bus (Diesel Euro 3) 5. 7% from 2W (Petrol Euro 3)	Top 5 accounts for <b>71%</b> emissions 1. 22% from 2W (Petrol, Euro 3) 2. 18% from 2W (Petrol, Euro 2) 3. 13% from 2W (Petrol, Euro 1) 4. 10% from 2W (Petrol, Euro 0) 5. 8% from 2W (Petrol, Euro 4)
CO	NO <sub>x</sub>	VOC
Top 5 accounts for <b>89%</b> emissions 1. 29% from 2W (Petrol, Euro 3) 2. 27% from 2W (Petrol, Euro 2) 3. 14% from 2W (Petrol, Euro 1) 4. 12% from 2W (Petrol, Euro 4) 5. 7% from 2W (Petrol, Euro 0)	Top 5 accounts for <b>63%</b> emissions 1. 21% from Bus (CNG, Euro 3) 2. 15% from HCV (Diesel, Euro 3) 3. 15% from Bus (Diesel, Euro 3) 4. 6% from Bus (CNG, Euro 2) 5. 6% from Car (Diesel Euro 4)	Top 5 accounts for <b>91%</b> emissions 1. 31% from 2W (Petrol, Euro 3) 2. 22% from 2W (Petrol, Euro 2) 3. 15% from 2W (Petrol, Euro 1) 4. 13% from 2W (Petrol, Euro 0) 5. 10% from 2W (Petrol, Euro 4)
NH <sub>3</sub>	N <sub>2</sub> O	CH <sub>4</sub>
Top 5 accounts for <b>79%</b> emissions 1. 39% from Car (Petrol, Euro2) 2. 19% from 2W (Petrol, Euro3) 3. 9% from Car (Petrol, Euro1) 4. 7% from 2W (Petrol, Euro4) 5. 5% from Car (Petrol, Euro4)	Top 5 accounts for <b>61%</b> emissions 1. 21% from 2W (Petrol, Euro 3) 2. 14% from Car (Diesel, Euro 4) 3. 11% from Car (Petrol, Euro 4) 4. 8% from 2W (Petrol, Euro 4) 5. 7% from LCV (Diesel, Euro 3)	Top 5 accounts for <b>86%</b> emissions 1. 39% from 2W (Petrol, Euro 3) 2. 15% from 2W (Petrol, Euro 4) 3. 13% from Bus (CNG, Euro 3) 4. 10% from Bus (CNG, Euro 2) 5. 9% from 2W (Petrol, Euro 2)

495  
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497

498

499 Figure 6. Heat map showing the emission share of vehicles of different class, fuel and  
 500 BS/EURO standards. Contributions less than 0.1% are not shown here. Contributions more  
 501 than 10% are shown in the same colour. (D: Diesel, P: Petrol, C: CNG and number 0-4  
 502 represents the Euro type starting from 0 being conventional to 4 as Euro 4).

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511 **4 Limitations:**

512 Geotagged dynamic traffic information and emission factors are the backbone of the emission  
513 inventory model. The traffic volume information is very crucial and traditionally obtained by  
514 manual counting or automated counters or through video surveillance at a few locations.  
515 However, in a real-world scenario, the traffic volume and speed can have large variations  
516 within a segment of a road. In this study we have adopted the congestion based approach (Jing  
517 et al., 2016; Yang et al., 2019) to model the traffic volume for each hour of the day. We use  
518 the same diurnal congestion profiles for all roads that could lead to emission uncertainty (Malik  
519 et al., 2021). In reality, some of the roads can be more congested than other roads based on the  
520 local population and traffic management.

521

522 The fleet composition can be different for different locations and at a given time of the day  
523 (Sharma et al., 2019). We have used the fleet composition based on surveyed composition at  
524 72 locations during the daytime (08:00-14:00) (TRIPP). To account for the peak hour and day-  
525 time entry restrictions of goods vehicles, we have used the share of goods vehicle (HCV and  
526 LCV) from the study by Errampalli et al. (2020). We use a constant share of fuel type, engine  
527 type and Euro type across all road links. The availability of such data, though challenging, can  
528 improve the emission estimates.

529

530 Although the COPERT emission functions provide the speed dependent emission factors for  
531 various classes of vehicles, they have been developed for European conditions. This adds to  
532 uncertainties while applying for Indian vehicles. The COPERT speed dependent EFs are  
533 available only for the criteria pollutants such as PM<sub>10</sub>, CO, NO<sub>x</sub> and VOC. The emission factors  
534 used here are functions of average speed for each hour. This does not account for the emission  
535 due to acceleration, deceleration and idling of the vehicles (Lyu et al., 2021). We have tried to  
536 address these by adding another 20% emission across all roads based on the earlier study (Lejri  
537 et al., 2018), however these could be uncertain.

538

539 This study only focuses on the hot emissions and does not include cold start, evaporative  
540 emission. We don't consider change in the emissions due to the change in the ambient  
541 temperature and humidity (Franco et al., 2013). Additionally, we don't consider emissions  
542 associated with road slope, vehicle degradation and maintenance in detail. But, we have



543 considered the vehicle degradation effect occurring in older vehicles considering the mileage  
544 as discussed in the COPERT-5 guide book.

545

546 Non-exhaust particulate matter emissions, such as dust resuspension, BW (Brake wear), TW  
547 (Tire wear), RW (Road wear) have not been considered in this study because of larger  
548 uncertainty. However, the non-exhaust emission of PM will be the dominant source of PM  
549 pollution in Delhi (Sharma et al., 2016; TERI, 2018; Singh et al., 2020).

550

551 Residential roads, the small roads in residential areas, account for 80% of the total length of  
552 Delhi, however their emission share has been reported to be only ~3% (Singh et al., 2018). We  
553 did not use these roads in our study, firstly, because of small share, secondly, we did not have  
554 a good quality data and thirdly, we wanted to optimise the computational cost.

555

556 The emissions estimated in this study for Delhi are comparable to the emission estimated for  
557 other megacities. For e.g. road transport emission of NO<sub>x</sub> and PM<sub>2.5</sub> for London was 20.8 Gg  
558 and 1.12 Gg respectively in 2016 (LAEI, 2016). The megacity Beijing, which has three times  
559 larger road network, had 4.1 Gg of traffic PM emission in 2013 (Jing et al., 2016). While our  
560 estimates are comparable to other megacities, these are lower as compared to the one reported  
561 by earlier studies for Delhi (Table 5). The lower emissions for Delhi can be expected because  
562 India has implemented the recent emission standards in a phased manner (Table S3) which  
563 should reflect in the traffic emission calculations. In many parts of the world, the road transport  
564 emission has decreased, despite an increase in transport vehicles, because of the improvements  
565 in engine technology (Winkler et al., 2018, Sun et al., 2019). One of the reasons for higher  
566 emission estimation by earlier studies for Delhi is the use of old EFs developed by ARAI way  
567 back in 2008. Therefore these ARAI EFs tend to overestimate the emissions as it does not  
568 represent the recent emission standard technologies (i.e. Euro 3 and Euro 4). It is important to  
569 use recent emission factors such as COPERT-5 which can account for technology related  
570 emissions. Although we have considered advanced traffic flow data and estimated the hourly  
571 emission as a function of speed, the accuracy of the emissions are subject to quality of the input  
572 data and emission factors. In this study we have shown a data driven approach where the quality  
573 of input data is likely to improve the emission estimates.



574 Table 5. Traffic emission studies over Delhi.

Studies	Area	Year	Method	EF	Diurnal	Resolution	PME (Gg)	BC (Gg)	OM (Gg)	CO (Gg)	NOx (Gg)	VOC (Gg)	NH <sub>3</sub> (Gg)	N <sub>2</sub> O (Gg)	CH <sub>4</sub> (Gg)
<i>Das and Parikh (2004)</i>	Delhi	2005	VKT	ARAI	NO	-	5.4			203	39				
<i>Nagpure et al. (2012)</i>	Delhi	2005	VKT	Variety of emission factor	NO	-	10			350	104	221			
<i>Goyal et al. (2012)</i>	Delhi	2008	VKT	IVE	Yes	2 km	5.3			186	71				
<i>CPCB (2010)</i>	Delhi	2010	VKT	ARAI	NO	2 km	3.5				30.73				
<i>Sahu et al. (2010, 2015)</i>	NCR Delhi	2010	VKT	ARAI	NO	1.67 km	30.3			427	162				
<i>Guttikunda and Calori (2013)</i>	NCT Delhi	2010	VKT	ARAI and Other	NO	1 km	14			256	199	132			
<i>Singh et al. (2018)</i>	NCT Delhi	2010	Non-VKT	ARAI	NO	100 m	4.5			114	51.5				
<i>Goel et al. (2015a)</i>	NCT Delhi	2012	VKT	COPERT-3 and ARAI	NO	-	12.7			300	184	71.6			
<i>Sharma et al. (2016)</i>	NCT Delhi	2014	Non-VKT	ARAI	NO	2 km	4.7			117	41.5				
<i>TERI (2018)</i>	NCT Delhi	2016		ARAI	NO	4 km	12.4			501	126	342			
<i>SAFAR (2018)</i>	NCR Delhi	2018	VKT	ARAI	NO	400 m	43.2	15.5		483.1	257.7	614.5			
<i>This Study</i>	NCT Delhi	2018	Non-VKT	COPERT-5	YES	100 m	1.82	0.94	0.75	221	56	64	0.28	0.26	11.38

575 \* NCT area is around 1483 sq. km; NCR area is around 4550 sq. km.





576

## 577 **5 Conclusion**

578 Here we present a methodology to estimate high-resolution spatially resolved hourly traffic  
579 emission over Delhi using advanced traffic flow and speed. We estimated the emissions of  
580 major pollutants, viz. PME, BC, OM, CO, NO<sub>x</sub>, VOC, NH<sub>3</sub>, N<sub>2</sub>O and CH<sub>4</sub>.

581 We have used traffic volume and speed measurements conducted at 72 locations over Delhi in  
582 the year 2018 as a part of TRIPP of IIT Delhi. Additionally, we have used the hourly congestion  
583 data from TomTom to account for hourly changes in the speed. The studies relation between  
584 traffic volume and speed has been utilised to generate the hourly traffic volume and speed  
585 profile for each road link. The vehicles have been classified into 127 categories according to  
586 vehicle types, fuel type, engine capacity, emission standard. The COPERT-5 emission  
587 functions of speed are applied at a micro level for each hour along each road link to calculate  
588 the emissions that accounts for congestion and spatial variation in emission. To the best of our  
589 knowledge, this is the first study of its kind which considers advanced traffic flow data and  
590 estimates the hourly multi-pollutant emissions as a function of speed. We make the following  
591 conclusions:

- 592 1. We estimated an annual emission of 1.82 Gg for PME, 0.94 Gg for BC, 0.75 Gg for OM,  
593 221 Gg for CO, 56 Gg for NO<sub>x</sub>, 64 Gg for VOC, 0.28 Gg for NH<sub>3</sub>, 0.26 Gg for N<sub>2</sub>O and  
594 11.38 Gg for CH<sub>4</sub> in 2018.
- 595 2. The modelled traffic volume (in PCU) and speed profiles show bimodal distribution  
596 exhibiting an anti-correlation behaviour. The traffic volume peaks during morning and  
597 evening rush hours resulting in lower speed. There is a mild enhancement in speed during  
598 the afternoon due to the less traffic. During the early morning hours, the vehicles almost  
599 achieve the free flow speed.
- 600 3. The diurnal variation of emission of pollutants are like traffic variations and show distinct  
601 bimodal distribution with morning and dominant evening peaks for almost all pollutants.  
602 However, the difference in night-time and day-time emissions are less for PME, BC and  
603 NO<sub>x</sub> due to the enhanced share of goods vehicles during the night-time. The good vehicles  
604 significantly contribute to the night-time emission in Delhi. These emissions along with  
605 unfavourable meteorology (e.g. lower PBL and wind speed) might help in sustained PM  
606 levels during the night-time in Delhi.
- 607 4. In terms of the spatial distribution of the emissions, the emissions are higher along the  
608 major roads and the emission hotspots are near the traffic junctions. The emission flux in



609 inner Delhi is highest due the higher road and traffic density, and lower average speed. This  
610 is 40-50% higher than the mean emission flux of Delhi. However, the total emission is  
611 higher for outer Delhi due to its larger area having a total road length more than inner Delhi.

612 5. According to the road classes (RClass1 to RClass5, from single lane to multi-lane roads),  
613 we find that RClass3 has the highest emission share due to highest total road length.  
614 However, the emission per km is highest over multi-lane wider roads (RClass4 and  
615 RClass5) that is almost two times RClass3 because of high traffic volume. Moreover, the  
616 emission per lane per kilometre is highest for RClass1 because of lower speed and  
617 congestion. While the effective management of traffic in narrow roads could be beneficial,  
618 the multi-lane roads act as emission hotspots. An analysis of the choice of road width should  
619 be performed to achieve the optimum emission without increasing the pollution exposure  
620 near the roads.

621 6. Petrol vehicles contribute to over 50% emission of OM, CO, VOC, NH<sub>3</sub>, N<sub>2</sub>O and CH<sub>4</sub>  
622 emissions. For OM, CO, VOC, N<sub>2</sub>O and CH<sub>4</sub> the petrol share is dominated by 2W whereas  
623 for NH<sub>3</sub>, share is dominated by petrol cars. The diesel vehicles are the dominant contributor  
624 to PME, BC and NO<sub>x</sub> emission.

625 7. In terms of emission standards, Euro3 vehicles contribute the highest to all pollutants  
626 followed by Euro4 with an exception to NH<sub>3</sub> where Euro2, mainly petrol cars, are the  
627 dominant source.

628 8. Among vehicle classes, the 2Ws contribute the most to the total emissions for all the  
629 pollutants except for BC, NO<sub>x</sub> and N<sub>2</sub>O. The diesel vehicles including goods vehicles (HCV  
630 and LCV) contribute substantially to the PME, BC and NO<sub>x</sub> emissions. The goods vehicles  
631 have a dominant share in the night-time emissions. CNG Buses have the highest  
632 contribution to NO<sub>x</sub> and CH<sub>4</sub> emissions whereas diesel Buses have substantial contributions  
633 to PME emissions. Petrol cars are the dominant source for NH<sub>3</sub> whereas diesel cars  
634 contribute substantially to PME, BC and NO<sub>x</sub> emissions. The contribution of petrol cars to  
635 the PME emission is less than 2%.

636 9. For all the pollutants, the top 5 polluting vehicle categories account for more than half (55%  
637 - 91%) of the emissions. The pollutants such as CO, VOC, CH<sub>4</sub> and OM have a distinct  
638 source such as 2W. However, the PME and BC have mixed sources including 2W and  
639 diesel vehicles. NO<sub>x</sub> emissions are mainly due to CNG and diesel vehicles. NH<sub>3</sub> is mainly  
640 emitted from petrol and diesel cars and N<sub>2</sub>O has mixed sources including 2W and cars.

641 This spatio-temporal emissions can be used in air quality models for developing suitable  
642 strategies to reduce the traffic related pollution in Megacity Delhi. Moreover, the developed



643 methodology is a step forward in developing real-time emission prediction in the future with  
644 growing availability of real-time traffic data.

#### 645 **Data availability**

646 The emission dataset can be accessed through the open-access data repository  
647 <https://doi.org/10.5281/zenodo.6553770> (Singh et al., 2022), under a CC BY-NC-ND 4.0  
648 license. This dataset is presented as a netDCF covering the rectangular domain around National  
649 Capital Territory (NCT) of Delhi. The data and analysis presented in the paper is only over the  
650 NCT area as shown in Figure 3. TOMTOM averaged congestion data is available online  
651 ([https://www.tomt.com/en\\_gb/traffic-index/new-delhi-traffic/](https://www.tomt.com/en_gb/traffic-index/new-delhi-traffic/)). COPERT-5 emission  
652 factors are obtained from the EMISIA online platform  
653 (<https://www.emisia.com/utilities/copert/>) of Aristotle University, Thessaloniki.

#### 654 **Author contribution**

655 **Vikas Singh** and **Akash Biswal**: Conceptualization, investigation, visualization, formal  
656 analysis, writing original draft, writing, reviewing and editing; **Leeza Malik** and **Geetam**  
657 **Tiwari**: Traffic data validation, investigation, discussion, reviewing and editing; **Ravindra**  
658 **Khaiwal** and **Suman Mor**: Investigation, discussion, reviewing and editing.

#### 659 **Declaration of competing interest**

660 The authors declare that they have no conflict of interest.

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