Four-century history of land transformation by humans in the United States: 1630-2020

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Abstract. The land of the conterminous United States (CONUS) has been transformed dramatically by humans over the last four centuries through land clearing, agricultural land expansion and intensification, and urban sprawl. Spatial-temporal data on long-term historical changes in land use and land cover (LULC) across the CONUS is essential for understanding and predicting the dynamics of coupled natural-human systems. A few efforts have focused on reconstructing historical databases to characterize changes in cropland and urban extent in the CONUS. However, the high-resolution and long-term trajectories of multiple LULC types remain unclear. By integrating multi-source data, such as high-resolution remote sensing image-based LULC data, model-based LULC products, and historical census data, we reconstructed the LULC history of land use and land cover for the United States (HISLAND-US) at an annual time scale and 1 km x 1 km spatial resolution for the CONUS in the past 390 years (1630–2020). The results show widespread expansion of cropland and urban land associated with rapid loss of natural vegetation. Newly reclaimed croplands are mainly converted from forest, shrubland, and grassland, especially in the Great Plains and North Central. Forest planting and regeneration accelerated the forest recovery in the Northeast and Southeast since the 1920s. The geospatial and long-term historical LULC data from this study can be applied to assess the LULC impacts on regional climate, hydrology, carbon and nitrogen cycles, and greenhouse gas emissions. The datasets are available at https://doi.org/10.5281/zenodo.6469247, https://doi.org/10.5281/zenodo.7055086 (Li et al., 2022).

1 Introduction

Land use and land cover (LULC) change is an essential component of global change, and humans have altered over one-third of the Earth’s land surface (Foley et al., 2005; Winkler et al., 2021). The human-induced LULC changes, such as cropland expansion, deforestation, wood harvest, and tree planting, have profound impacts on climate change, carbon and nitrogen...
cycles, and biodiversity (Houghton et al., 1999; Dangal et al., 2014; Domke et al., 2020; Lark et al., 2020; Tian et al., 2020). In particular, managing agriculture and forest-related land use activities have been recognized as a critical pathway to achieve climate mitigation targets (Grassi et al., 2017; Griscom et al., 2017). Thus, a better understanding of historical LULC and its spatial-temporal dynamics is critical for quantifying the effects of LULC change on the ecosystem and climate.

In the past four centuries, the conterminous United States (CONUS) has experienced dramatic land use and land cover (LULC) changes associated with land clearing, cropland, and urban land expansion (Steyaert and Knox, 2008; Drummond and Loveland, 2010; Oswalt et al., 2014; Sohl et al., 2016). Before the arrival of Europeans, indigenous agriculture and crop planting existed in the eastern woodlands, the Great Plains, and the southwestern US South (Hurt, 2002). Since the establishment of the first colony in Virginia was established in 1607, cropland and pasture—land began to expand by land clearing, which mainly initially occurred in the eastern United States. During the Colonia Era, most people lived in the east of the Appalachian Mountains, and agriculture was the primary livelihood for 90% of the population during the colonial era (Steyaert and Knox, 2008). Driving by the westward movement in the 19th century, territorial expansion (e.g., Louisiana Purchases) opened up new areas for agriculture. Driven by the western movement, land clearing, agriculture expansion, and deforestation expanded across the Appalachian Mountains into Ohio, the upper-Mississippi River basins, and the Great Lakes region (Cole et al., 1998; Billington et al., 2001; Steyaert and Knox, 2008; Yu and Lu, 2018). In the Mississippi River Valley and Alabama, hardwood forests were cleared for cotton and grain production (Hanberry et al., 2012). The center of lumber production was shifted from the Northeast to the Great Lakes in the 1850s (Fickle et al., 2001). In California, agriculture and ranching expanded throughout the state and soon became an exporter of wheat as the gold mining waned (Olmssted and Rhode, 2017). Entered the 20th century, cropland and pasture land in New England, and the Atlantic coast, and the Southeast were abandoned, and the forest grew again in the late 19th century (Foster, 1992; Hall et al., 2002; Jeon et al., 2014). Though the environmental protection movement originated in the 1880s. Both tree planting and forest regeneration from abandoned agricultural land accelerated forest restoration began to increase until the 1930s (Stanturf et al., 2014). In the following 90 years, the national total plantation forest area increased to 27 Mha (Oswalt et al., 2014, 2019; Chen et al., 2017). While general trends in historical US landscape change are known, however, it we still lacks a long-term and spatial-explicit land-use LULC dataset to characterize historical LULC trajectories for the CONUS.

Several efforts have produced LULC data for the CONUS in the past several decades. For example, multiple contemporary and spatially explicit LULC products with a resolution from 30 m to 1 km are available, including Global Land Cover (GLC) 2000 (Bartholome and Belward, 2005), MODIS land cover (Friedl et al., 2010), GlobeLand30 (Chen et al., 2015), National Land Cover Database (NLCD) (Yang et al., 2018; Homer et al., 2020), and Cropland Data Layer (CDL) (Boryan et al., 2011; Lark et al., 2017, 2021). However, these datasets were generated using remote sensing images and cannot be used to characterize the century-long land use dynamics. Global-scale and long-term coverage land use datasets (e.g., Land and Use Harmonization (LUH2), the History Database of Global Environment (HYDE)) are widely used in global climate simulations and carbon budget projects (Goldewijk et al., 2017a, 2017b; Hurt et al., 2006, 2020). However, these datasets have a coarse resolution (from 5 arcmins to 0.25 degrees) and substantial uncertainties, which cannot present regional-scale details well (Li
et al., 2016; Yu and Lu, 2018). Moreover, the data uncertainties will significantly impact the quantification of LULC effects on the ecosystem (Peng et al., 2017; Yu et al., 2019). Some studies focused on reconstructing historical single-type land use datasets (e.g., settlement built-up area and cropland) for the US (Zumkehr and Campbell, 2013; Yu and Lu, 2018; Lerk et al., 2020). Nevertheless, the dynamics of pasture, forest, shrub land, and grassland also profoundly impact the ecosystem carbon dynamics (Chen et al., 2006; Tian et al., 2012). Therefore, developing a long-term and high-resolution land use LULC dataset with multiple land use classes types for the CONUS is essential for understanding the LULC change history and LULC impact on ecosystem dynamics, regional climate, hydrology, carbon and nitrogen cycles, and greenhouse gas emissions.

In this study, we aim to reconstruct the HISTORY of LAND use and land cover for the United States (HISLAND-US) and analyze the spatial and temporal pattern of LULC changes in the CONUS during 1630–2020 by integrating high-resolution satellite data, reliable inventory data, and model-based LULC data. This study consists of three parts: a description of input data and methods, an analysis of spatiotemporal characteristics of LULC in the past four centuries, and a comparison between our results and other studies. We also discussed the driving forces of LULC changes and the uncertainties of the newly developed dataset.

2 Materials and Method

This study reconstructed the land use LULC history (1630–2020) at annual time step and 1 km x 1 km spatial resolution for the CONUS (48 states) using remote sensing-based LULC data, model-based land use data, and historical census data. In addition, we aggregated the state-level data to eight subregions to analyze the regional divergence of land use LULC changes. These subregions include Northeast, Northeast, North Central, Southeast, South Central, Great Plains, Intermountain, Pacific Northwest, and Pacific Southwest (Oswalt et al., 2014, 2019) (Figure 1).

The reconstruction process of historical land use and land cover LULC data mainly included two parts: (1) reconstructing the historical urban land, cropland, pasture, and forest area at the state level (Section 2.2), (2) generating 1 km x 1 km spatial resolution gridded land use and land cover LULC data (Section 2.3). Figure 2 shows the general workflow for generating historical land use and land cover LULC data. The following sections provide a detailed description of the input data and how we process the data.
Figure 1: The division of the conterminous United States into eight subregions for data synthesis and analysis in this study.
Figure 2: Workflow for generating fractional and Boolean historical land use and land cover data for the conterminous United States. NLCD: National Land Cover Database; HISDAC: Historical Settlement Data Compilation; CAHA: Census of Agriculture Historical Archive; ERS: Economic Research Service; CPHR: the Crop Production Historical Report; HYDE: History Database of the Global Environment; NRI: National Resource Inventory; FIA: Forest Inventory and Analysis; USDA-FR: USDA Forest Resources of the United States, 2017; ANN: Artificial Neural Network; BPS: Biophysical Settings.
2.1 Input datasets for land use and land cover reconstruction

The input datasets included satellite-based land use and land cover (LULC) data (National Land Cover Database, NLCD), model-based land use datasets (i.e., HYDE3.2 baseline), land-use-census and inventory data, and other auxiliary data (Table 1, Table S1-S4). The spatial data were resampled or aggregated to 1 km x 1 km resolution for further processing. We also collected some other land use LULC datasets to validate our results, the newly developed dataset, including Historical Settlement Data Compilation (HISDAC) (Leyk and Uhl, 2018; Uhl et al., 2021), Yu and Lu (2018)’s cropland density (Yu and Lu, 2017), Zumkehr and Campbell (2013)’s historical fractional cropland fraction areas (Zumkehr and Campbell, 2013), Economic Research Service (ERS) Major Land Uses data (Bigelow and Borchers, 2017), Land Use Harmonization (LUH2) (Hurtt et al., 2020), CONUS historical land use and land cover (Sohl et al., 2016), county-level crops area (Crossley, 2020) and hay area (Haines et al., 2018)’s hay area (Table A2). All the spatial data were resampled to 1 km x 1 km resolution for further processing. Table 1 and Table A1 show a detailed description of the data used in this study.
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<td>2000-2020</td>
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2.2 Historical land use and land cover area reconstruction

2.2.1 Urban land

This study regarded the developed land in the NLCD dataset as urban land. The developed land area during 2001–2016 was set as the baseline to reconstruct the historical time-series total urban land area. We first calculated the urban land per capita at the state level over the period using NLCD developed land area and population. Following previous studies (Liu and Tian, 2010; Goldewijk et al., 2017a), we estimated the total urban land area during 1630–2020 by multiplying the urban land per capita and total population at the state level. In this study, we used the same definition for the developed land as NLCD for urban land. The developed land in NLCD includes four components: open space, low intensity developed land, medium intensity developed land, and high intensity developed land (Table 2). We used the NLCD developed land area during 2001–2019 as the urban land area baseline. Before 2001, we applied Historical Settlement Data Compilation for the United States (HISDAC-US) (Leyk et al., 2020; Uhl et al., 2021) as input to reconstruct the historical urban land area. The HISDAC-US built-up areas describes the built environment for most of the CONUS from 1810 to 2015 at 5-year temporal and 250 m spatial resolution using built-up property records, locations, and intensity data (Leyk and Uhl, 2018; Uhl et al., 2021). Here, we assumed that the HISDAC built-up areas data could capture the trend of urban land development. Then, the historical urban land can be estimated as follows:

\[ \text{HistUrban}_{s,t} = \text{HistUrban}_{s,t+1} \times \frac{\text{HISDAC}_{s,t}}{\text{HISDAC}_{s,t+1}} \]  

where \( \text{HistUrban}_{s,t} \) and \( \text{HistUrban}_{s,t+1} \) are the reconstructed urban land area of state \( s \) in year \( t \) and \( t+1 \); \( \text{HISDAC}_{s,t} \) and \( \text{HISDAC}_{s,t+1} \) are the HISDAC built-up area of state \( s \) in year \( t \) and \( t+1 \).

There is no census data on urban land area before 1810. Following Liu et al. (2010), we used population to estimate the urban land area by assuming that urban land expanded at the same rate as total population during 1630–1810. The urban land area of each state can be calculated as follows:

\[ \text{HistUrban}_{s,t} = \text{HistUrban}_{s,t+1} \times \frac{\text{Pop}_{s,t}}{\text{Pop}_{s,t+1}} \]  

where \( \text{HistUrban}_{s,t} \) and \( \text{HistUrban}_{s,t+1} \) are the reconstructed urban land area of state \( s \) in year \( t \) and \( t+1 \); \( \text{Pop}_{s,t} \) and \( \text{Pop}_{s,t+1} \) are the total population of state \( s \) in year \( t \) and \( t+1 \).
2.2.2 Cropland

The definition of cropland varies in the existing literature and datasets (Zumkehr and Campbell, 2013; Bigelow and Borchers, 2017; Goldewijk et al., 2017; Homer et al., 2020, Table S5). Cropland, defined by the U.S. Department of Agriculture (USDA) Economic Research Service (ERS), includes five components: cropland harvested, crop failure, cultivated summer fallow, cropland pasture, and idle cropland (Table 2). In this study, we only count the cropland harvested area, which includes row crops and closely sown crops, hay and silage crops, tree fruits, small fruits, berries, and tree nuts, vegetables and melons, and miscellaneous other minor crops (https://www.ers.usda.gov/data-products/major-land-uses/glossary/#cropland). USDA Census of Agriculture Historical Archive (CAHA) recorded state-level cropland harvested areas at 4 to 10 years intervals (Table 1 and Table S5), which was used for historical cropland area reconstruction between 1879 and 2017. The CAHA cropland was interpolated into annual using the linear method first. To subtract the double-cropped area, we applied the annual national cropland harvested area without double-cropped area from ERS Major Land Uses data to adjust the interpolated cropland harvested area. The adjustment can be expressed as follows:

\[ \text{HistCrop}_{s,t} = \frac{\text{Cropland Harvested}^{\text{linear}}_{s,t}}{\text{Cropland Harvested}^{\text{ERS}}_{\text{conus},t}} \times \text{Cropland Harvested}^{\text{ERS}}_{\text{conus},t} \]  

(3)

where \( \text{HistCrop}_{s,t} \) is the reconstructed cropland area of state \( s \) in year \( t \); \( \text{Cropland Harvested}^{\text{linear}}_{s,t} \) is the linearly interpolated cropland harvested area of state \( s \) in year \( t \) based on CAHA cropland harvested area; \( \text{Cropland Harvested}^{\text{ERS}}_{\text{conus},t} \) is the national total cropland harvested area without double-cropped area in year \( t \). For 2018–2020, the state-level cropland area was calculated based on the state-level area weight in 2017.

For 1879–1910, there was no national-level cropland harvested area without double-cropped area. Therefore, we applied the trend of the CAHA cropland harvested area to reconstruct the historical cropland:

\[ \text{HistCrop}_{s,t} = \text{HistCrop}_{s,t+1} \times \frac{\text{CAHA}_{\text{CHA}_{s,t}}}{\text{CAHA}_{\text{CHA}_{s,t+1}}} \]  

(4)

where \( \text{HistCrop}_{s,t} \) and \( \text{HistCrop}_{s,t+1} \) are the reconstructed cropland area of state \( s \) in year \( t \) and \( t+1 \); \( \text{CAHA}_{\text{CHA}_{s,t}} \) and \( \text{CAHA}_{\text{CHA}_{s,t+1}} \) are the cropland harvested area of state \( s \) in year \( t \) and \( t+1 \).

Because there was no available cropland census data at the state level before 1879, the HYDE cropland was used. We first estimated the cropland per capita by applying the trend of HYDE cropland per capita. Then, the total cropland area can be calculated by multiplying cropland per capita and total population. The data harmonization process can be expressed as follows:

\[ \text{HistCrop}_{s,t} = (\text{HistCrop}_{p,s,t+1} \times \frac{\text{HYDE}_{\text{Crop}_{p,s,t}}}{\text{HYDE}_{\text{Crop}_{p,s,t+1}}}) \times \text{Pop}_{s,t} \]  

(5)

where \( \text{HistCrop}_{s,t} \) is the reconstructed cropland area of state \( s \) in year \( t \); \( \text{HistCrop}_{p,s,t+1} \) is the reconstructed cropland per capita of state \( s \) in year \( t+1 \); \( \text{HYDE}_{\text{Crop}_{p,s,t}} \) and \( \text{HYDE}_{\text{Crop}_{p,s,t+1}} \) are HYDE cropland per capita of state \( s \) in year \( t \) and
Cropland is defined as the areas used for to produce crops, such as corn, soybeans, and cotton (Homer et al., 2020). In this study, we counted cropland area as the area of land on which crops are planted within a year, excluding crop failure, summer fallow, idle crop, and cropland pasture (Bigelow and Borchers, 2017; Yu and Lu, 2018). The United States Department of Agriculture (USDA) provided many agricultural census data, the most important reference to reconstruct the historical cropland area. However, the USDA did not provide the physical cropland area. Therefore, we used the planted area to estimate the physical cropland area at the state level by subtracting the double-cropped area.

The Crop Production Historical Report (CPHR) provided state-level and annual cropland planted crop area data from 1975 to 2020 (Table 1). Meanwhile, the USDA Census of Agriculture Historical Archive (CAHA) recorded state-level cropland harvested areas during 1879–2020 at 5 to 10 years intervals (Table 1). Because the annual state-level crop planted area data were only available in the past 45 years, we used the state-level crop harvested area for further processing to keep the consistency of historical data. First, we linearly interpolated the crop harvested area during 1879–2020 between the time points. Then, we used the CPHR annual national-level total crop harvested area during 1909–2020 to adjust the interpolated crop harvested area, making the inter-annual variations more reasonable (Figure A2). The adjustment was based on the ratio of the state crop harvested area accounting for the national total.

After getting the state-level cropland harvested area, we used a conversion factor to estimate planted area. The conversion factor was calculated using a linear fit method and state-level crop harvested area and planted area during 1978–2017 ($y = 1.0665x$, $R^2 = 0.99$, $p < 0.05$; Figure A3). We calculated the physical cropland area by subtracting the double-cropped area.

The regional double-cropped portion was derived from Borchers et al. (2014) and then disaggregated to the state level based on the cropland planted area as the weight value (Figure A4). The state-level double-cropped portion and the ratio of planted area and harvested area were assumed to be consistent. The reconstructed cropland area in 1879 showed a significant difference from that in 1889, so we only used the state-level cropland area since 1889 for further analysis.

There was no available agricultural census data at the state level between 1630 and 1889, so we used the HYDE3.2 data to reconstruct for this period. We first summarized the national total cropland area and population to calculate the national cropland per capita. Then, we estimated the state-level cropland per capita during 1630–1889 based on the trend of HYDE (Figure A5) and cropland per capita in 1889. In this step, we assumed that the changes in cropland per capita at the state level were consistent with that at the national level. Then, we calculated the national cropland area during 1630–1889 by multiplying state-level total population and cropland per capita. Finally, we combined the results in these two periods and got the cropland area at the state level during 1630–2020. Figure 3b shows all the cropland data used in different periods.

### 2.2.3 Pasture

The definition of pasture also varies among multiple datasets (Goldewijk et al., 2017; U.S. Department of Agriculture, 2020; Table S6). In this study, we use the definition from the National Resource Inventory (NRI), in which pasture is the land that has a vegetation cover of grasses, legumes, and forbs, regardless of whether it is being grazed by livestock, planted for livestock grazing, or the production of seed or hay crops (U.S. Department of Agriculture, 2020; Homer et al., 2020Table 2). In this
study, we set the National Resource Inventory (NRI) pasture area from 1982 to 2017 as the baseline for its historical reconstruction. For the early period, though USDA provided the information on plowable pasture (1940 and before) or cropland used for pasture (1945 and after), farm woodland pasture, and other pasture (Wasianen and Bliss, 2002), the pasture definitions changed several times, and the pasture and rangeland are not separated. Therefore, we used the HYDE pasture before 1982. First, the NRI data were interpolated linearly between 1982 and 2017. Then, we used the pasture per capita at the state level (NRI data-based) in 1982 and the national level (HYDE data-based) to estimate the pasture per capita during 1630–1982 (Figure A5). In this step, we assumed the trend of pasture per capita at the state level was consistent with that at the national level. Then, each state’s pasture land area was calculated by multiplying the pasture per capita and population. Additionally, we assumed that the area of pasture land during 2018–2020 was the same as that in 2017. Figure 3c shows all the pasture data used in different periods. The NRI provides state-level pasture area with 5-year interval between 1982 and 2017, and we set the pasture area as the baseline for historical reconstruction. Because there was no available pasture census data at the state level before 1982, the HYDE pasture was applied. We first estimated the pasture per capita by applying the trend of HYDE pasture per capita. Then, the total cropland area can be calculated by multiplying pasture per capita and total population. The data harmonization process can be expressed as follows:

\[
\text{HistPasture}_{s,t} = (\text{HistPasture}_{p,s,t+1} \times \frac{\text{HYDE Pasture}_{p,s,t}}{\text{HYDE Pasture}_{p,s,t+1}}) \times \text{Pop}_{s,t} \]

where \(\text{HistPasture}_{s,t}\) is the reconstructed pasture area of state \(s\) in year \(t\); \(\text{HistPasture}_{p,s,t+1}\) is pasture area per capita of state \(s\) in year \(t+1\); \(\text{HYDE Pasture}_{p,s,t}\) and \(\text{HYDE Pasture}_{p,s,t+1}\) are the HYDE pasture per capita of state \(s\) in year \(t\) and \(t+1\).

2.2.4 Forest

In this study, we use the forest definition from Forest Inventory and Analysis (FIA), identical to the Forest Inventory and Analysis (FIA), in which the forest is defined as land at least 10 percent stocked by forest trees of any size, or formerly having such tree cover, with a minimum area classification of 1 acre. Two datasets were used for the historical forest area reconstruction. The first is The USDA Forest Resources (USDA-FR) of the United States 2017 (Oswalt et al., 2019). It provides state-level forest areas from 1630 to 2017, with, including twelve snapshots (i.e., 1907, 1920, 1938, 1953, 1963, 1977, 1987, 1997, 2007, 2012, 2017) and a shot in 1630. Another is FIA’s Forest area trend data (FATD), which includes provided state-level forest area from 1760 to 2000 at 10-year intervals and a snapshot in 1630. The data was rebuilt by integrating FIA field data and reports (1950–2000), field inventories (1910–1940), Bureau of the Census land clearing statistics (1850–1900), and clearing estimates proportional to population growth (1760–1840), and USDA forest report. The USDA Forest Resources (USDA-FR) of the United States 2017 (Oswalt et al., 2019) provided state level forest areas from 1630 to 2017, including twelve snapshots (i.e., 1907, 1920, 1938, 1953, 1963, 1977, 1987, 1997, 2007, 2012, 2017) and a shot in 1630. For 1907–2017, the USDA-FR data was used without adjustments. We combined the two data sets and reconstructed a new historical forest inventory dataset. For the period 1907–2017 and 1630, USDA FR data was used. Before 1907, we calculated
the ratio of forest area in 1760–1900 with 1630 (i.e., $\text{FATD}_t/\text{FATD}_{1630}$, $1760 \leq t \leq 1900$) and then multiplied the forest area from USDA-FR to generate the forest area in 1760–1900. For 2018, 2019, and 2020, we first collected the latest forest area of each state. If one state did not publish the forest area of the latest year, we assumed that the area during these three years was the same as that in 2017. The latest forest area data can be accessed at https://www.fia.fs.fed.us/tools/data/ (last accessed: April 18, 2022). The data used in different periods is shown in Figure 3d. Before 1907, to keep the raw data consistent, we adopted USDA-FR in 1630 as the initial point and gap-fill the missing years by using the changes reflected by FATD data to reconstruct the forest area between 1630 and 1907. The following harmonization method was conducted to combine the two datasets:

$$\text{HistForest}_{s,t} = \frac{\text{USDA}_s \text{FR}_{s,1630} \times \text{FATD}_{s,t}}{\text{FATD}_{s,1630}}$$

where $\text{HistForest}_{s,t}$ is the reconstructed forest area of state $s$ in year $t$; $\text{USDA}_s \text{FR}_{s,1630}$ is the USDA–FR forest area of state $s$ in 1630; $\text{FATD}_{s,t}$ and $\text{FATD}_{s,1630}$ are the FATD forest area of state $s$ in year $t$ and 1630, respectively.

For 2018, 2019, and 2020, we first collected the latest forest area of each state. If one state did not publish the forest area of the latest year, we assumed that the area during these three years was the same as that in 2017. The latest forest area data can be accessed at https://fia-usfs.hub.arcgis.com/ (last accessed: Aug 30, 2022).
Figure 3: Workflow for reconstructing historical land use and land cover area at the state-level. NLCD: National Land Cover Database; HISDAC: Historical Settlement Data Compilation; ERS: Economic Research Service; CAHA: Census of Agriculture Historical Archive; NRI: National Resource Inventory; HYDE: History Database of the Global Environment; FIA: Forest Inventory and Analysis; USDA-FR: USDA Forest Resources of the United States, 2017.; FATD: Forest Area Trend Data; HistUrban, HistCrop, HistPasture, HistForest, and HistLULC refer historical urban land, historical cropland, historical pasture, historical forest, and historical land use and land cover.

2.2.5 Post-processing of historical urban, cropland, pasture and forest land area

Due to the difference in data sources in the reconstruction step, the total area of urban land, cropland, pasture, and forest may exceed the state’s total land area (TLA). Therefore, we calibrated the reconstructed historical land use and land cover area using the following equations:

\[
\begin{align*}
A_{i,rc}^t(s) &= A_{i,r}^t(s) \quad \text{if } TA_{i,r}^t(s) \leq TLA(s) \\
A_{i,rc}^t(s) &= \frac{A_{i,r}^t(s)}{TA_{i,r}^t(s)} \times TLA(s) \quad \text{if } TA_{i,r}^t(s) > TLA(s)
\end{align*}
\]  

(84)

\[TA_{rc}^t = \sum_{i=1}^{n} A_{i,r}^t(s)\]  

(92)

where \(t\) is the current year; \(A_{i,rc}^t(s)\) and \(A_{i,r}^t(s)\) are re-calibrated area and reconstructed area for the land use class \(i\) in the state \(s\), respectively; \(TA_{rc}^t\) is the total area of urban, cropland, pasture and forest; \(n\) is total number of land use types; \(s\) is the state index in the range from 1 to 48.

Table 2: Definitions of urban, cropland, pasture, and forest in this study

<table>
<thead>
<tr>
<th>LULC</th>
<th>Definition</th>
</tr>
</thead>
<tbody>
<tr>
<td>Urban land</td>
<td>Same as the definition of developed land in National Land Cover Database (NLCD). Developed land in NLCD include four components: open space, low intensity developed land; medium intensity developed land, and high</td>
</tr>
</tbody>
</table>

Cropland  Same as the definition of cropland in U.S. Department of Agriculture (USDA) Economic Research Service (ERS) major land use. Cropland defined by USDA ERS includes five components: cropland harvested, crop failure, cultivated summer fallow, cropland pasture, and idle cropland (https://www.ers.usda.gov/data-products/major-land-uses/glossary/#cropland). In this study, we only count the cropland harvested area subtracting the double-cropped area.

Pasture  Same as the definition of pasture in National Resource Inventory (NRI). Pasture is a land cover/use category of land managed primarily for the production of introduced forage plants for livestock grazing.

Forest  Same as the definition of forest from Forest Inventory Analysis (FIA). Forest is the land at least 10 percent stocked by forest trees of any size, or formerly having such tree cover, with a minimum area classification of 1 acre (https://www.fia.fs.fed.us/tools-data/maps/2007/descr/yfor_land.php).

2.3 Approach for generating gridded land use and land cover data

2.3.1 Calculating the land use and land cover probability

We reconstructed the historical land use and land cover data with 1 km x 1 km spatial resolution based on the state-level land use area and land use probability (Figure 2). Following previous studies, we applied the “Top-down” strategy to allocate the state-level LULC area to the grid level based on probability or suitability surfaces (Fuchs et al., 2013; West et al., 2014; He et al., 2015; Sohl et al., 2016). Previous spatially explicit land use and land cover simulation models, such as Conversion of Land Use and its Effects (CLUE) model and Forecasting Scenarios of Land use Change (FORE-SCE) model, used the logistic regression (LR) model to develop land use and land cover probability-of-occurrence (Verburg et al., 2009; Sohl et al., 2014, 2016; Li et al., 2016; Yang et al., 2020; Li et al., 2016). However, it needs to train the LR model for the different units (e.g., county, grid) to calculate a good probability map due to the spatial heterogeneity of land conversion. In comparison, artificial neural networks (ANNs) can learn and fit complex relationships between input data and training targets and can be used to solve various non-linear geographical problems (Hagenauer and Helbich, 2022). Moreover, ANN performs better performance than LR in land use and land cover change simulation (Liu et al., 2016). Therefore, we estimated the land use probability for urban, cropland, pasture, and forest using ANN and NLCD data we used the ANN-based Probability of Occurrence Estimation tool in Future Land Use Simulation (FLUS) software to generate the LULC probability (Liu et al., 2017). The independent variables for the ANN model training and prediction include terrain (elevation, and slope), climate (annual mean temperature, annual precipitation, annual maximum temperature (July), and annual minimum temperature (January)), crop productivity index, population density, distance to the city, distance to the road, distance to the railway, distance to the river, soil (soil organic carbon, soil sand, and soil clay); (Table A1) shows the detailed information on the independent variables. The Boolean type NLCD data in 2001 was used for ANN modeling training.
Over the past four centuries, the rules of LULC probability change a lot due to the interaction between natural environment and socioeconomic factors. The contemporary pattern of LULC probability is not representative for the early period (Sohl et al., 2016). Following Goldewijk et al. (2017), we improved the LULC probability by combining the biophysical probability and contemporary probability, as well as population density, human settlement extent, and satellite data. The total probability for each grid cell can be expressed as follows:

\[
TP_t = \begin{cases} 
S_{hist} \times w_1 + S_{satellite} \times w_2 \times (1.0 + r) & t \leq 2001 \\
S_{satellite} + Frac_{dt_{satellite}} \times (1.0 + r) & t > 2001 
\end{cases}
\]

where \(S_{hist}\) and \(S_{satellite}\) is the LULC fraction generated by using the historical \((Prob_{hist})\) and satellite \((Prob_{satellite})\) probability; \(w_1\) and \(w_2\) is probability weight; \(w_1\) is set to zero in 2001 and 100% in 1850 (and the pre-1850 period as well), while \(w_2\) is set to 0 in 1850 (and the pre-1850 period as well) and 100% in 2001; \(Frac_{dt_{satellite}}\) is the NLCD LULC fraction dynamics between year \(t\) and 2001; \(SE_{weight_t}\) is settlement weight in year \(t\), which is calculated based on the settlement in \(t0\) year and \(t1\) year; \(r\) is a random item with a range of \([0, 0.5]\), \(Prob_{bio}\) is the LULC probability that only use biophysical variables (terrain, climate, and soil variables), \(Prob_{2001}\) is the LULC probability that use all the variables; \(popd_t\) is population density (Figure S5), the extent of the settled area in the CONUS expanded from the Northeast to the West coast, making the impacts of the natural environment and socioeconomic factors on LULC change gradually. We divided the study period into four sub-periods (p1: 1630–1790; p2: 1790–1850; p3: 1850–1920; p4: 1920–2020) to improve the ANN modeled probability. In the early period, the population density was an essential factor because the total area and spatial distribution of the human-dominated land use types (urban, crop, pasture) were always related to the population density. Thus, we use the extent of the settled area and population density to restrict the land use change boundary (Figure A6). In the p4 period, the natural environment was not the decisive factor for human–dominated land use types due to the technology development. We further used the remote sensing–based land use map in the 2000s to constrain the land use probability (Goldewijk et al., 2017a). As a result, we calculate the final probability as follows:

\[
Prob_{k_t} = (1 - w_t) \times Prob_{k_{pop\_set}} + w_t \times Frac_{k, 2000} 
\]

\[
Prob_{k_{pop\_set}} = Prob_{k_{ann}} \times Pop\_weight_t \times ES\_weight_t 
\]

\[
Pop\_weight_t = \frac{pop_{dt_t}}{pop_{mean}} 
\]

\[
ES\_weight_t = SE_{t0} + w_t \times SE_{t1} 
\]

\[
w_t = \frac{t - t0}{t1 - t0} 
\]
where $\text{Prob}_{k,t}$ is the probability of land use type $k$ in $t$ year; $\text{Frac}_{k,2000}$ is the fraction of land use type $k$; $\text{Prob}_{k,\text{ANN}}$ is the probability of land use type $k$ determined by natural environmental conditions; $\text{Pop}_{\text{weight},t}$ is the population adjustment factor in $t$ year, $\text{Pop}_{d,t}$ is population density at $t$ year, $\text{Popd}_{\text{mean},t}$ is the mean population density at state level in $t$ year; $\text{ES}_{\text{weight},t}$ is settlement weight in $t$ year, which is calculated based on the settlement in $t_0$ year and $t_1$ year. For the p1 sub-period, we used the population weight in 1790 due to the lack of population density data. For the p1, p2 and p3 sub-period, we assumed that the land use dynamics was mainly constrained by natural environmental conditions and population density, and the weight of $\text{Frac}_{k,2000}$ was set as 0.

2.3.2 Strategies to generate fractional and Boolean land use and land cover data

In order to generate the fractional grid data, we assumed that the fraction of each land use type at the grid level was determined by the probability (Fuchs et al., 2013; Tian et al., 2014; West et al., 2014; He et al., 2015). A grid (land use type $k$) with a high probability will have a high fraction. Based on this principle and the state-level land use area, we generated the fractional land use data at 1 km x 1 km resolution and annual time scale.

We further generated the Boolean type land use and land cover data at 1 km x 1 km resolution using the fractional data for four land use types (urban, crop, pasture, and forest). First, the total number of potential pixels or the land use demand was determined based on the reconstruction results in Section 2.2. Then, the area difference of land use type $k$ between the target and current map was calculated. If the difference is negative, land use type $k$ will lose. In that case, the pixels of type $k$ with the high fraction will keep the condition of the current LULC map, and the rest pixels with a low fraction will be converted to other types. If the difference is positive, land use type $k$ will expand, the pixels of type $k$ in the current map will be assigned the value of $k$, then the pixels (non-$k$ type) ranging on top will be assigned as $k$ to meet the rest demand. Once a pixel has been assigned to more than one land type, we will compare their probability among different categories and assign the type with the highest probability to the conflicted pixel. Only urban land, cropland, pasture, and forest can be allocated spatially. The pixels not assigned value will be updated using the NLCD and LANDFIRE Biophysical Settings data. Finally, each state was iterated annually from 1630 to 2020.

Two types of gridded LULC data with 1 km x 1 km spatial resolution were generated. The first is fractional type, in which the dataset includes four fractional components: urban, cropland, pasture, and forest. Another is Boolean type with nine LULC types: urban, cropland, pasture, forest, shrub, grassland, wetland, water, and barren.

To generate the fractional gridded LULC data, we assumed that the fraction of each LULC type at the grid level was determined by the total probability (Fuchs et al., 2013; Tian et al., 2014; West et al., 2014; He et al., 2015). It means that a grid cell (LULC type $k$) with a high probability will have a high fraction. Based on this principle and the state-level LULC area, we generated the fractional LULC data at 1 km x 1 km resolution and annual time scale. The detailed information for generating fractional LULC data is shown in the following steps (Figure 4): (1) prepare the input data: state-level historical LULC area and probability; (2) calculate the state target LULC fraction for type $k$ and initialize an empty LULC fraction of surface; (3)
calculate a temporal fraction layer; (4) modify the temporal fraction, we assume that the fraction of water and barren is stable, and the sum of urban, crop, pasture, and forest fraction is lower than the maximum fraction in each grid cell; (4) add the temporal fraction data to the empty LULC fraction; (5) judge whether the unallocated LULC area is smaller than 0.01 km², if yes, the iteration will stop and begin to allocate another LULC type, else the unallocated area will be assigned to target fraction and return to step (3). The allocation was processed until the unallocated area was less than the threshold (0.01 km²). The above steps will be conducted for each state, and urban, cropland, pasture, and forest fractional map in the CONUS will be output.

Based on the LULC fraction map, we generated the Boolean type LULC data at 1 km x 1 km resolution. The detailed information is shown in the following steps (Figure 4): (1) prepare the input data: state-level historical LULC area and LULC fraction data; (2) generate a temporal LULC map (HistB) through identifying the dominate LULC type in each grid cell and initialize an empty LULC map (HisBE); (3) calculate the area difference for LULC type k between the HistB map and target area; (4) if the area difference is negative, we first sort the LULC fraction data where HistB equals to k, the top m (equals to target area) grid cells where HisBE not be assigned a value will be assigned as k, then if the available number of grid cell (type k) is less than the target area, we will sort the LULC fraction data where HistB map not equal to k, and the top n (equals to unallocated area) grid cells where HisBE not be assigned a value will be assigned as k; (5) if the area difference is positive, the grid cells where HistB data equals to k and the will be assigned k to HisBE not be assigned a value; then we will sort the LULC fraction data where HistB data not equals to k, and the top n (equals to unallocated area) grid cells where HisBE not be assigned a value will be assigned as k. If step (4) and (5) finish, the next LULC type will begin to allocate. After the four LULC types of allocation finish, the grid cell that is not assigned a type will be updated using the HistB data and LANDFIRE Biophysical Settings data (Figure A1; Rollins et al., 2009).
The lack of actual spatial explicit reference data made a complete formal validation impractical. Though the LULC definitions in this study are different from other LULC datasets, data comparison is a way to assess the accuracy of the reconstructed LULC area and spatial pattern. Thus, we conducted three data comparisons to increase the confidence of the newly developed LULC datasets. First, the state-level LULC area derived from the multisource datasets was used for comparison. Considering the differences in the cover period of multiple LULC datasets, we derived the average state-level statistics area for urban, cropland, pasture, and forest from 2000 to 2020 for comparison. Second, we collected the USDA county-level cropland area between 1840 and 2012 and compared the cropland proportion with that derived from our data in four selected years (1850, 1920, 1960, and 2002). Third, we compared urban, cropland, pasture, and forest from the newly developed LULC dataset with the NLCD during 2001–2019 at the grid level. To validated the newly developed land use dataset, we compared it with other land use datasets. Considering the time cover period of land use datasets, we derived the average state-level statistics area for urban, cropland, pasture, and forest from 2000–2020 for comparison. However, the land use datasets, except for NLCD, only include parts of four land use types. Therefore, the comparison is as follows: urban (This study, NLCD, and HISDAC), cropland, pasture, and forest.
cropland (This study, NLCD, HYDE, ERS, and YLmap), pasture (This study, NLCD, and HYDE), and forest (This study, NLCD, and LUH2). In the discussion section, we compared the national-level statistics area with NLCD, HYDE, LUH2, ZCmap, YLmap, Sohl et al. (2016), Haines et al. (2018), ERS, and HISDAC data. We also analyzed the spatial consistency and differences between our data and other land use datasets. The spatial comparison is as follows: urban (This study, HISDAC, and Sohl et al. (2016)), cropland (This study, HYDE, YLmap, and ZCmap), pasture (This study, HYDE, and LUH2), and forest (This study and LUH2).

3 Results

3.1 Comparison with other datasets Validation of the newly developed land use dataset

3.1.1 State-level land use and land cover area comparison

We compared the state-level urban, cropland, pasture, and forest areas using data derived from ERS, HISDAC, HYDE, NLCD, HYDE-LUH2, ERS, YLmap, and HISDAC with our data the newly developed LULC dataset (Figure 4). Generally, the newly developed land use dataset matches well with the data used for comparison (Figure 5). The urban land acreages from this study are close to NLCD data (Figure 5a; R² = 0.9993, Slope = 0.9361) because ERS urban land only includes the densely-populated areas with at least 50000 people (urbanized areas) and densely-populated areas with 2500 to 50000 people (urban clusters). In contrast, whereas lower than HISDAC urban land data (Figure 5b; R² = 0.8688, Slope = 1.3334), especially in Georgia, New York, North Carolina, Ohio, and Tennessee. It is because the HISDAC data is rebuilt using the detailed property records and have a relatively coarse resolution (Leyk et al., 2020). Our The cropland acreages area derived from this study are is consistent with NLCD (Figure 5b; R² = 0.99, Slope = 0.99102) and YLmap (Figure 5b; R² = 0.99, Slope = 0.934). However, ERS and HYDE data tend to overestimate the cropland (Figure 4b, SlopeERS = 1.26, SlopeHYDE = 1.14). Nevertheless, the ERS cropland is higher than our data (Figure 5b; R² = 0.96; Slope = 1.26) because the ERS cropland here includes the area the cropland harvested area, crop failure, cultivated summer fallow, cropland used for pasture, and idle cropland. The coefficients of determination between our pasture acreages and NLCD (Figure 4e5c; R² = 0.93, Slope = 1.02) and HYDE (Figure 4e5c; R² = 0.87, Slope = 0.99) are higher than 0.87. For the forest, both NLCD and LUH2 data are lower than our data, especially in the Rocky Mountain states (Figure 5d; SlopeNLCD = 0.72, SlopeLUH2 = 0.66). The differences in definition and data development method could result in LULC area differences for both pasture and forest (Table S1-S4), making it hard to compare. For example, the LUH2 forest area in Rocky Mountain states is lower than our data and NLCD because they applied biomass density data to determine the forest extent. Though there still are some uncertainties, the comparison results show that the newly developed dataset can provide a relatively accurate LULC area at the state level.
3.1.2 Comparison with cropland census data at county-level

An accurate cropland map is quite critical for historical LULC reconstruction. We compared our data with county-level census data to assess the accuracy. This study’s spatial pattern of cropland proportion (i.e., cropland area/county area) is close to the census data in 1850, 1920, 1959, and 2002 (Figure 6). In 1850, both the newly developed cropland and census data showed high cropland density in the Black Belt, New England, and the North Central. In contrast, our data was higher in North Central, the east of Virginia and North Carolina, and the south of Georgia (Figure 6). Cropland derived from this study was higher than the census data in the Atlantic coast, the Mississippi Alluvial Plain, the northwest of Texas, the west of Oklahoma, and California in 1920, 1959, and 2002. However, the cropland proportion in the Appalachian Mountains and the south of the Great Plains was lower than the census data (Figure 6). This underestimation may result from the low cropland fraction in satellite data because it is difficult for satellite data to identify the small area cropland patch in the mountain region and classify...
the pasture or grassland with cropland in the south of the Great Plains. Moreover, both datasets showed the cropland expansion in the North Central, the Great Plains, the Mississippi Alluvial Plain, and California between 1850 and 2002. The cropland abandonment can also be found in the Appalachian Mountains between 1920 and 2002. The statistical comparison also shows that our data fits well with the census data in 1920 ($R^2 = 0.68$), 1960 ($R^2 = 0.89$), and ($R^2 = 0.91$) (Figure A2). Overall, the newly developed cropland has a relatively accurate spatial pattern and proportion.

Figure 6: Spatial comparison of county-level cropland proportion between our reconstruction and census data in 1850, 1920, 1959 and 2002. First column: cropland proportion from census data; Second column: cropland proportion derived from this study; Third column: cropland proportion between this study and census data.

### 3.1.3 Comparison with NLCD at grid-level

The spatial patterns of urban, cropland, pasture, and forest in this study are close to the satellite-based data from NLCD, and most grid cells have a relatively small difference between 2001 and 2019 (Figure 7). Our results have a higher urban land fraction in the NLCD low urban density area, but the difference in 87% of urban grids is smaller than 10%. Cropland with a positive difference is mainly distributed in the Northeast, Alabama, and Missouri, in which 65.95% of grids have slight
3.2 Land use and land cover change during 1630–2020 in CONUS

The results showed that the land use and land cover (LULC) change from 1630 to 2020 was characterized by the expansion of cropland and urban land and the shrinking of natural land cover (e.g., forest, grassland, and shrubland) (Figure 58, Figure A37-A49A6). In 1630, the primary landscape was the forest in the eastern USCONUS and Pacific Coastal region, grassland in the Great Plains, and shrubland in the Rocky Mountains (Figure 58). Urban land, cropland,
and pasture were mainly distributed in the east of CONUS-US before 1850. Rapid cropland and pasture expansion occurred in the North Central region—(e.g., Iowa, Illinois, Minnesota), the Great Plains, and the Mississippi River Valley during 1850–1920 (Figure 58 and Figure A8A4). After 1920, the distribution of major land use classes LULC types became relatively stable (Figure 69). In the 2000s, the cropland in the Corn Belt regions, Central California, and Mississippi Alluvial Plain had the highest cropland density (Figure A8A4), and the highest pasture density was found in the east of Texas, Oklahoma, Missouri, and Kentucky (Figure A59).
Figure 58: Spatiotemporal patterns of land use and land cover in the conterminous United States during 1630-2020
Figure 69: Changes in areas of land use and land cover changes in the conterminous United States from 1630 to 2020

The United States experienced the colonial era, the war of independence, and territorial expansion during the colonial era between 1630 and 1850. However, in this period, the total increase of urban land was only increased by 3.23 ± 0.80 Mha with a total population growth of 23 million (Figure 69), and was mainly distributed in the Northeast (Figure 5 and Figure A10). In the middle of the 1800s, the cheap land and the industrial revolution growth prospect attracted many European and Mexican immigrants, which accelerated urban development. In the second half of the 19th century, the population tripled and the total urban land increased to 4.3 Mha in 1900, the national total urban land area reached 7.64 Mha (Figure 69). Entered the 20th century, both the rapid growth of population and urban land per capita accelerated the urban land expansion. Our result show that the urban land per capita increased from 0.02 ha/person in 1900 to 0.14 ha/person in 2020 (Figure S7). Though the government limited the total amount of immigrants, the urban land area in the CONUS still increased by 21.85 Mha during...
1880–1965. After that, the new immigration policy promoted the increase in population and urban sprawl. As a result, the national total urban land area increased to 47.81 Mha in 2020.

Cropland expanded slowly by 21.62 Mha from 1630 to 1850, and it increased substantially to 145.37 Mha in the following 70 years (Figure 96). Agriculture turned to be intensified after 1920, and but the national total cropland area in the CONUS did not change significantly was relatively constant, with a peak area of 455.37 Mha in 1930–1932 (Figure 69).

Due to the competition of newly reclaimed cropland with the high production land in the Midwest, cropland abandonment occurred in the Northeast, South, and Southeast (Bigelow and Borchers, 2017; Yu and Lu, 2018). During 1950–1975, the rise of the manufacturing and service industry resulted in agricultural labor and cropland area reduction. As the demand for biofuel and bulk grain grew in the 2000s, cropland began to extend again, and the national-total cropland area in 2020 was 427.32 Mha (Figure 69). Pasture showed an increasing trend with a slowly increasing rate during 1630–1850. It expanded more than 20 times from 1850 to 1950 and reached the maximum historical area (59.67 Mha) in 1950–1959. The national total pasture area in the CONUS kept relatively stable and decreased slowly in the following 70 years (Figure 69).

Forest was the dominant land use type in the CONUS before the colonial era, which accounted for about 47% of the total land area. The trends in forest area were contrary to that of agricultural land in the past four centuries before 1920. During 1630–1850, the national total forest loss was 33.91 Mha (Figure 96). Over the second period (1850–1920), forest area decreased by 83.02 Mha because of agricultural land occupation, lumber cut, and fuelwood consumption (Steyaert and Knox, 2008). In the third period (1920–2020), forest area has been relatively stable through forest management and planting (Figure 96).

3.3 Land use and land cover transitions during 1630–2020

The changes in the land use-LULC area only reflected its quantitative changes. However, the land use-LULC transition map takes a further step to illustrate the spatial distribution of conversion between two land use-LULC types (Figure 710). Over the past 390 years, cropland expansion by occupying forest, shrubland, and grassland was the primary land use-LULC change characteristic (Figure 710). The natural land loss was mainly distributed in the North Central region (e.g., Ohio, Indiana) and Southern states such as Tennessee, Texas, Alabama, and Georgia (Figure 710d). New-reclaimed cropland reclamation encroached 363.75 Mha (10.03% of total forest in 1630) of forest and 66.2 Mha (18.12% of total shrub and grassland in 1630) of grassland and shrubland. Meanwhile, 28.98 Mha of forest and 11.15 Mha of shrubland and grassland were converted to pasture. Moreover, urban land occupied more than 27.33 Mha of forest and 47.11 Mha of grassland and shrubland (Figure 74 Table 3). During the early period (1630–1850), forest converted to cropland was the dominant land use-LULC transition type, especially which was mainly distributed in the Eastern U.S. (Figure 710a). The U.S. experienced the most significant dramatic land use-LULC changes with large forest and grassland loss in North Central and Great Plains during 1850–1920 (Figure 10b), characterized by grassland converted to cropland. Cropland expansion encroached 56.21 Mha of forest and 59.01 Mha of grassland, and pasture development also occupied more
than 27.61 Mha of forest in the Midwest and North Central and forest converted to pasture in Southern (Figure 7b Table 3). Furthermore, urban land expansion and abandoned cropland converted to the forest (47.3622.35 Mha) distributed in the Northeast and Southern states was the essential feature of land-use LULC changes from between 1920– and 2020 (Figure 710c).

**Table 3. Net land use and land cover change during 1630-1850, 1850-1920, 1920-2020, and 1630-2020.**

<table>
<thead>
<tr>
<th>LULC transition type</th>
<th>1630-1850</th>
<th>1850-1920</th>
<th>1920-2020</th>
<th>1630-2020</th>
</tr>
</thead>
<tbody>
<tr>
<td>Cropland to Urban</td>
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<td>0.46</td>
<td>10.92</td>
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<td>0.00</td>
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<td>3.67</td>
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<td>Sub-total</td>
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<td>0.97</td>
<td>0.00</td>
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<td>56.21</td>
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<td>54.38</td>
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<td>59.01</td>
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<td>1.56</td>
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<td>9.74</td>
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Figure 7.10: Land transition (1 km x 1km spatial resolution) between 1630 and 1850 (a), 1850 and 1920 (b), 1920–2020 (c), 1630 and 2020 (d) in the conterminous United States.
3.4 Land use and land cover changes during 1630–2020 at regional level

Given the differences in natural environmental conditions and social-economic development, land use and land cover changes showed significant spatial heterogeneity in the CONUS during 1630–2020. Since 1630, the South Central region experienced the most intensive urban land expansion (10.62–11.21 Mha), followed by the North Central (10.25–28 Mha), Southeast (7.38 Mha), and Northeast (6.00 Mha), respectively (Figure 81a). Rapid cropland expansion first occurred in the North Central, Northeast, South Central, and Southeast in the 1830s. Cropland in the Intermountain and the Great Plains began to develop after 1860. The trends of cropland in eight regions except South Central and Southeast were consistent with the national total. Over the past four centuries, the North Central region had the largest cropland expansion area (46.01–47.89 Mha), followed by the Great Plains (31.41–33.03 Mha) and the South Central (20.10–20.38 Mha) (Figure 81b). The trends of cropland in eight regions were consistent with the national total. Cropland in the South Central and Southeast had decreased by 3.55–4.91 Mha and 44.50–12.44 Mha since the 1930s due to the increasing urbanization pressures and low cropland profitability.

Similar to cropland, the Northeast region was the first to develop pasture. The pasture experienced a rapid expansion during 1790–1950 and finally reached the maximum historical area (4.56–Mha) in the 1950s, and then gradually decreased (Figure 81c). For a long period (1865–1980), the South Central region had the largest pasture area. The maximum historical area was 22.42 Mha in 1950 and accounted for 37% of the national total. However, the pasture area in the North Central region began to decrease since 1960, and only 11.17 Mha of pasture was left by 2020 (Figure 81c).

Agricultural land encroachment, land clearing, and wood harvest/deforestation resulted in forest loss in eight regions (Oswalt et al., 2014, 2019). In the past four centuries, the North Central region lost the most forest area (24.85–12 Mha), followed by the South Central region (36.12–24.85 Mha). During 1850–1920, the forest area decreased rapidly in the North Central (24.96–24.97 Mha), South Central (30.29–29.39 Mha), Southeast (14.03–01 Mha), and Northeast regions (6.59 Mha). Most of the lost forest converted to cropland and pasture (Figure 81d). Since the 1920s, the regional forest area has been relatively stable with small fluctuations. Notably, the forest land recovered gradually, especially in the Northeast, South Central, and Southeast. Compared with the 1920s, the total forest area in the Northeast increased by 6.86–87 Mha (Figure 81d).
Figure 8. Changes in areas of urban land (a), cropland (b), pasture (c), forest (d) in different geographic regions during 1630–2020.
4 Discussion

4.1 Comparison with the previous datasets

This study reconstructed a gridded time series of LULC data for the CONUS from 1630 to 2020. Compared with the ERS and HYDE data, the reconstructed urban land was higher (Figure 9.12a), attributed to the definition differences in urban land with NLCD. ERS urban areas include densely populated areas with at least 50,000 people and densely populated areas with 2,500 to 50,000 people, which have changed several times over the past 70 years (Bigelow and Borchers, 2017). The ERS urban area includes the densely populated areas with at least 50,000 people (urbanized areas) and densely populated areas with 2,500 to 50,000 people (urban clusters). The total urban land area from HISDAC data was higher than the newly developed data in the recent four decades (Figure 12a). Because the HISDAC built-up area dataset was developed by using the detailed property records data at a relatively coarse resolution. In addition, the HISDAC settlement data was developed using the detailed address points data (Lerk et al., 2020), and some small-scale built-up land cannot be remote-sensing images cannot identify. Small-scale built-up land using satellite images and NLCD may underestimate the total urban land area. Moreover, as a result, the HISDAC settlement was higher than the reconstructed urban land in the recent four decades (Figure 9a). Moreover, the HISDAC settlement dataset built-up areas may underestimate the total urban area in the early years due to the high missing rate of property records due to the lack of address points or approximate location data in the early period (Lerk et al., 2020). Therefore, our data may also underestimate the total urban land area because we applied the trend of HISDAC between 1810 and 2001. We assumed that the developed land per capita was unchanged, and our results may overestimate the total urban land area in the early period. The spatial pattern of Boolean-type urban land was consistent with the Sohl et al. (2016) data and was mainly distributed in the area near the city, road, and railway (Figure 4.13). The spatial allocation rule determined that the grid with a high probability of occurrence would be allocated first, which may underestimate the developed land in the rural area (Verburg et al., 2009; Yang et al., 2020). Though the urban data had some uncertainties in the urban data exist, we provided a long-term description of urban land with higher resolution and consistency for the CONUS.
Figure 9.12: Comparison with other datasets for the conterminous United States: urban land (a); cropland (b); pasture (c); forest (d). NLCD: National Land Cover Database; HYDE: History Database of the Global Environment; HISDAC: Historical Settlement Data Compilation; ERS: Economic Research Service; YLmap: Yu and Lu (2018) cropland density; ZCmap: Zumkehr and Campbell (2013) historical fractional cropland area; LUH2: Land Use Harmonization; FATD: Forest Area Trend Data; USDA-FR: USDA Forest Resources of the United States of 2017.
This study

HISDAC

Sohl et al.
For the cropland, our result—the reconstructed cropland area—was close to NLCD, YLmap, and ZCmap during 2001–2016 in the 2000s (Figure 9). Our data and YLmap applied the cropland harvested area to estimate the historical cropland area and showed the same trend during 1850–2016 (Figure 12b). Because we used the crop planted area to estimate the physical cropland area, our data were consistent with YLmap during 1850–2016 (Figure 9b). The cropland area derived from ZCmap and ERS was higher than our data over the research period (Figure 12b) because cropland harvested, crop failure, cultivated summer fallow, cropland use for pasture, idle cropland all counted (Zumkehr and Campbell, 2013; Bigelow and Borchers, 2017). However, the ZCmap and HYDE cropland were higher than our data over the research period (Figure 9), which could be explained by the fall, idle, and pasture land area counted in ZCmap (Zumkehr and Campbell, 2013). The area trend between 1630 and 1879 was close to HYDE because we use it’s cropland per capita trend (Figure 12b). Applying the HYDE cropland historical trend made our result close to it during 1630–1889 (Figure 9b). Spatially, four fractional cropland maps show the similar state and expansion patterns. The highest cropland density can be found in the Corn Belt, Central California, and
Mississippi Alluvial Plain in the 2000s. Meanwhile, cropland expansion initially occurred in the east of Mississippi River, then moved to the Midwest and the Great Plains between 1850 and 1920 (Figure 14). Four fractional cropland maps all showed rapid cropland expansion in the Midwest and the Great Plain during 1850–1920 and cropland abandonment in the Northeast and Southeast during 1920–2010 (Figure 11, Figure A8). But our results can reflect the cropland abandonment in New England, the South, and the Southeast since the 1920s, which is consistent with previous studies (Reuss et al., 1948; Land, 1974; Foster, 1992). Moreover, the newly developed cropland improved the spatial resolution compared with HYDE and ZCmap, making it possible to catch more detailed information (Figure 15). In YLmap, there are some coarse grids in the early years (Figure 15) because they applied HYDE data to reconstruct the cropland expansion and abandonment (Yu and Lu, 2018). Our data was processed at 1km resolution and fixed this problem (Figure 15). Compared with the above cropland data, our product has higher spatial resolution and more extended temporal coverage, making it capable of depicting the cropland dynamics better in CONUS over the past four centuries (Reuss et al., 1948; Land, 1974; Foster, 1992). Compared with other LULC datasets, our product has higher spatial resolution and more extended temporal coverage, a better understanding of cropland dynamics in CONUS over the past four centuries.
Figure 14: Comparison of cropland maps among four datasets for the conterminous United States: this study (first column), the History Database of Global Environment (HYDE), Yu and Lu (2017) cropland density (YLmap), and Zumkehr and Campbell (2013) historical fractional cropland areas (ZCmap).
Figure 15: Visual comparison between our cropland data and the History Database of Global Environment (HYDE), Yu and Lu (2017) cropland density (YLmap), and Zumkehr and Campbell (2013) historical fractional cropland areas (ZCmap) in four different sites (a-d). The locations of image center points are as follows: a. Ohio (83.05 °W, 40.17 °N), b. Georgia (83.58 °W, 32.77 °N), c. Arkansas (90.56 °W, 34.76 °N), d. Texas (100.92 °W, 32.81°N).

To the best of our knowledge, accurate temporal and spatially explicit data are still lacking to describe the pasture dynamics for the CONUS. This study set the state-level pasture area from the National Resource Inventory (NRI) as the baseline data for historical pasture reconstruction, which made our data more reliable than HYDE. During 2001–2020, the total national area of pasture located in non-federal land ranged from 48 to 53 Mha, which was close to the NLCD (53 Mha) and HYDE (52 Mha) (Figure 912c). We also found that NLCD pasture/hay decreased during 2001–2016, while NRI pasture land kept relatively stable. The likely reasons for NLCD pasture/hay loss include normal crop cycling and more permanent conversion (Homer et al., 2020). The difference in pasture trends between NRI and NLCD may result from the definitional difference (Table S6). Moreover, Haines et al. (2018) pasture only includes hay, making it significantly lower than our result (Figure 912c). The ERS land use data also provided grazing land area, but the rangeland and pasture were not separated (Bigelow and Borchers, 2017). The application of the HYDE pasture per capita historical trend made our result close to it and reached the maximum historical value in the 1950s (Figure 129c). The three maps all show the highest pasture density in
eastern Texas, Oklahoma, and Missouri on three maps (Figure 1). At the regional scale, the spatial patterns of pasture land from this study are close to the HYDE and LUH2 data, but our data can characterize the historical changes of pasture with higher spatial resolution than current LULC products (Figure 17). Our results characterized the historical changes of pasture with higher spatial resolution than current LULC products.
Figure 162: Comparison of pasture patterns and changes among three data products for the conterminous United States: this study (upper panel), the History Database of Global Environment (HYDE) (middle panel), and Land Use Harmonization (LUH2) (lower panel).

Figure 17: Visual comparison of our pasture data with History Database of Global Environment (HYDE), and Land Use Harmonization (LUH2) in four different sites (a-d). The locations of image center points are as follows: a. Iowa (93.64 °W, 42.03 °N), Virginia (78.72 °W, 37.96 °N), c. Illinois (90.07 °W, 38.68 °N), d. Arkansas (92.56 °W, 34.97 °N).

We used the inventory-based data sets (FATD and USDA-FR) to reconstruct historical forest areas, which was more reliable than the satellited-based forest (NLCD) and biomass density-based forest (LUH2). Because Compared with satellited-based forest (NLCD), and Sohl et al. (2016), and LUH2 data, the total forest area in our data is higher. This area difference mainly resulted from the differences in forest definition. For example, NLCD and Sohl et al. (2016) define forest as the areas dominated by trees generally greater than 5 meters tall and greater than 20% of total vegetation cover, higher than that in our forest definition (forest cover greater than 10%) (Sohl et al., 2016; Homer et al., 2020; Table S7). Thus, the forest area in this study was higher than the NLCD, Sohl et al. (2016) data (Figure 9d). Moreover, the forest in LUH2 is determined by the vegetation biomass density and country-level forest area (Hurtt et al., 2020), underestimating the forest land in the western US forest distribution in the area with low biomass density. Spatially, our data and LUH2 can describe the high density in the
eastern US and Pacific Coast area, but LUH2 underestimates the forest fraction in Rock Mountain and Texas (Figure 18 and Figure 19). Our data fixed the above problem and improved the spatial resolution from 0.25 degree to 1 km. Meanwhile, the newly developed forest data has good performance in capturing forest dynamics. For example, previous studies reported deforestation in southern Michigan and forest cutting for agriculture and fuel in Virginia during the early settlement period (Carl, 2012; Mergener et al., 2014), also shown in our maps during the 19th century (Figure A649). Forest loss during the westward expansion period can be captured in the Northeast, Midwest, and Great Plains (Figure 183, Figure A649). The LULC conversion map can reveal the forest regrowth on much cutover and abandoned agricultural land in the Northeast and Southeast since the 20th century (Foster et al., 1998; MacCleery, 2011) (Figure 10, Figure A6). We also found the forest regrowth on much cutover and abandoned land in the late 20th century (Foster et al., 1998; MacCleery, 2011) (Figure A10, Figure 8). In addition, our data overcome the underestimation in the Rock Mountain region and Texas in LUH2 forest (Figure 13).

![Figure 183: Comparison of forest distribution between this study (upper panel) and Land Use Harmonization (LUH2) (lower panel) for the conterminous United States.](image-url)
Figure 19: Visual comparison between our forest data and Land Use Harmonization (LUH2) in four different sites (a-d). The locations of image center points are as follows: a. Colorado (106.47° W, 38.97° N), Wisconsin (89.85° W, 44.54° N), c. Alabama (86.72° W, 33.33° N), d. New York (75.14° W, 42.21° N).

4.2 Drivers of land use and land cover changes

Agricultural land expansion and natural vegetation loss (forest, grassland, and shrub) area is the primary characteristic of LULC change in the CONUS over the past four centuries. The complex interactions among land suitability, climate, population, transportation, agricultural technologies, and policy shaped the contemporary LULC pattern. In the Colonial Era, the migration of Europeans into the Northeast and Mid-Atlantic converted the Eastern forests to cropland and pasture (Waisanen and Bliss, 2002). More than 90% of people lived in the east of the Appalachian Mountains, and most farms were subsistence in this period. The forced migration of slaves contributed to the plantation agriculture expansion in Virginia, Maryland, South Carolina, and the Black Belt. After the new nation was established, numerous lands like Louisiana, Florida, Texas, Oregon, and New Mexico were acquired during 1800–1860 (Dahl and Allord, 1996; Fretwell, 1996). The westward movement opened new areas for agricultural development. With the building of canals and inland waterways, agricultural products from the cropland developed west of the Appalachians could be brought to the market (Meinig, 1993). In the second half of the 19th century, the rapid population growth and food demand resulted in the cropland expansion because farmers needed to reclaim another three to four acres to feed one person (MacCleey, 2011). After the 1920s, cropland, pasture, and forest area became relatively constant despite the growing population. The applications of hybrid crops and fertilizers and the increasing number of motor vehicles and farmer tractors improved agricultural productivity, which played an essential role in stabilizing cropland area (Waisanen and Bliss, 2002; MacCleey, 2011). Cropland abandonment in the east was affected by the fluctuations in crop prices, changes in labor markets, and competition from the high productivity in the Midwest (Hart, 1968; Williams, 1989; Bigelow and Borchers, 2017). Reversion of marginal cropland in the east and large-scale tree planting in the South contributed to the forest recovery (Clawson, 1979; Smith et al., 2001; Thompson et al., 2013). For example, many croplands in the South were abandoned following the disintegration of the post-bellum sharecropping system and later converted to plantations forest (Hart, 1968), and the plantation forest area increased from near zeros in the 1930s to 27 Mha.
Climate change also impacts the LULC change. For example, the Dust Bowl in the 1930s led to widespread crop failure in the Great Plains (Heimlich and Daugherty, 1991). Land marked by crop failure due to severe drought, extensive flooding, or wet weather has ranged between 5 and 22 million acres since 1949 (Bigelow and Borchers, 2017). Human activities are the main drivers of land use and land cover changes. In the CONUS, agricultural land changes were influenced by climate, soil productivity, population size, economy, and technological improvement (Waisanen and Bliss, 2002), and agricultural land expansion resulted in forest, shrubland, and grassland loss (Oswalt et al., 2014).

The population increased to 3 million from 1630–1775, but more than 97% of people lived in the east of the Appalachian Mountains. Though agriculture had developed in Virginia and Maryland, colonists in the Northeast or Mid-Atlantic region worked in small-scale farming. Therefore, urban land, cropland, and pasture showed a lower increase rate, and forest was the dominant land use type in the Eastern US. Numerous lands like Louisiana, Florida, Texas, Oregon, and New Mexico were acquired during 1800–1860 (Dahl and Allord, 1996; Fretwell, 1996). The westward expansion or movement opened up new agricultural areas and significantly affected the US land use pattern. For example, the rapid inland movement resulted in the conversion of wetland to cropland in the Ohio and Mississippi River Valleys (Dahl and Allord, 1996). In the Mississippi Valley and Midwest, forest land and grassland were cleared and put into growing crops (Steyaert and Knox, 2008; Hanberry et al., 2012). The national policy also had a substantial influence on LULC changes. For example, the Homestead Act issued in 1862 aimed to attract immigrants to develop the Western US (Shannon, 1977). As a result, cropland expanded with a rate of the population increase to produce more food during 1860–1920 (Fred, 1945; MacCleery, 2011). In addition, the development of fertile cropland in the Midwest resulted in the cropland abandonment in the Northeast and South, and the rapid development of railroads drove the growing cities and urban land increase. Though the population has more than tripled since the 1910s, cropland and forest land area maintained stable. The development of hybrid crops and the use of chemical fertilizers improve crop intensive level and productivity, reducing cropland reclamation (MacCleery, 2011). The conservation policy framework issued in the 1930s emphasized the importance of forest protection. Tree planting and stabilizing timber consumption also played an essential role in keeping forests stable (Chen et al., 2017). Though forest clearing for cropland reclamation continued in some states, offset by cropland abandonment and forest regeneration in other areas, like New England (Foster et al., 1998; Hall et al., 2002) and Wisconsin (Rhemtulla et al., 2009). However, urban land increased at a higher rate, mainly driven by the rapid urbanization and population increase (Leyk et al., 2020).

4.3 Uncertainties and future perspectives

This study provides a four-century LULC dataset at annual time step and 1 km x 1 km spatial resolution for the CONUS. However, some uncertainties may affect the accuracy of this dataset. This study aims to reconstruct the spatial and temporal dynamics of LULC changes in the CONUS and how agricultural activities and urbanization induce natural vegetation degradation or loss. For instance, both the reliability of input data and the harmonization method are critical for the historical LULC area reconstruction. Most census data used in this study was recorded at 4 to 10-year interval, making interannual fluctuations impossible to capture. The rebuilt state-level LULC area is also coarse if there are significant spatial shifts (e.g.,
cropland abandonment in some counties but reclamation in others) for a LUCC type. Moreover, the definitional differences among datasets increased the difficulties and uncertainties in the harmonization process. Though we tried to gather the most reliable LULC datasets, the definitions of LULC vary (Table S5-S7). For example, we applied three datasets (i.e., ERS cropland harvested area, CAHA cropland harvested area, HYDE cropland) to generate the cropland area for the study period, but the definitions of cropland harvested area and cropland are different among three datasets (Table S1). In addition, the definitions of four LULC types do not belong to a universal classification system, making it hard to process the total area, and a post-processing step needs to be conducted. Though we have gathered the most reliable land use census or inventory data, the historical census data only records net changes in the area. Moreover, the original census data was recorded at 5–10 years intervals except in recent decades, which made some insignificant fluctuations cannot be captured. The assumptions made in the reconstruction section also increased the data uncertainties. In the generating spatial data section, we assumed that environmental factors based land use probability was unchanged following previous land use simulation models (Verburg et al., 2006, 2009; Sohl et al., 2014; Liu et al., 2016). Though we improved the land use probability by integrating population density, human settlement extent, and current land use pattern, there were still uncertainties in describing the historical land use trajectory.

More efforts are needed to generate accurate historical LULC maps for understanding the history of regional LULC changes. An accurate LULC probability or suitability surface is the key to generate spatial data. In this study, we assumed that the ANN-based LULC probability was unchanged following previous LULC simulation models (Verburg et al., 2006, 2009; Sohl et al., 2014; Liu et al., 2016). However, we found that the contemporary probability surfaces could not represent the historical LULC pattern, especially for agricultural land (Sohl et al., 2016). To solve the problem, we modified the LULC probability by using population density, human settlement extent, and satellite observed LULC fraction, making it match the historical LULC pattern. The LULC pattern is highly related to that in the previous year, and the grid value is also affected by the fraction and type in the neighbor grid cells. In our spatial allocation strategy, we generate LULC map for each year based on the LULC probability or LULC fractional data, which ignores the LULC pattern interactions between the adjacent years. Some studies generated a LULC map by allocating the LULC net change area to a base map (West et al., 2014; Liu et al., 2016; Cao et al., 2021), but such an algorithm would underestimate the LULC gross change (Winkler et al., 2021). Therefore, an improved spatial allocation strategy should be developed to simulate LULC conversion better.

More efforts are needed to generate accurate historical LULC maps to understand regional LULC history. On the one hand, more detailed data are required, such as the historical cropland map and survey data, the wood harvest, and tree planting information at the county or site level (Zumkehr and Campbell, 2013; Yu et al., 2019). The subclass of LULC (e.g., tree species, crop types) also needs to reconstruct more accurately (Thompson et al., 2013; Chen et al., 2017; Crossley et al., 2021). On the other hand, land use simulation models should be improved to depict land use dynamics accurately by integrating machine learning methods. In the future, the impact of extreme climate events, war, and policies could be taken into account to better simulate LULC changes.
The newly developed LULC dataset reconstructed the LULC history with more LULC types than ZCmap and YLmap and has higher spatial resolution than HYDE and LUH2. Our LULC data emphasizes the accuracy of area change resulting from LULC conversion rather than the changes in LULC structure or attributes. For example, forest management (e.g., wood harvest and thinning) results in the forest cover decreases and ecosystem function change, but the LULC type is unchanged. HYDE and LUH2 not only have a more extended cover period, but also provide more sub-types and LULC attributes. HYDE classified cropland into rain-fed rice, irrigated rice, rain-fed other crops, and irrigated other crops (Goldewijk et al., 2017). LUH2 divides cropland into C3 crops and C4 crops and includes the wood harvest (traditional fuelwood, commercial biofuels, and industrial roundwood) and primary/secondary forest age (Hurtt et al., 2020). In the future, the LULC sub-types (e.g., tree species, crop types) and attributes (e.g., forest age, management intensity) through collecting from agricultural census data and forest inventory data can be incorporated into our dataset (Thompson et al., 2013; Chen et al., 2017; Crossley et al., 2021).

5 Data availability

The land use and land cover datasets for the conterminous United States are available at https://doi.org/10.5281/zenodo.6469247https://doi.org/10.5281/zenodo.7055086 (Li et al., 2022). The annual grided datasets (1km x 1km spatial resolution) with GeoTiff format include fractional and Boolean types. An Excel table is used to organize the annual urban, cropland, pasture, and forest area at the state level. A detailed data description is also provided.

6 Conclusions

This study developed spatially explicit LULC data at a spatial resolution of 1 km x 1 km and an annual time scale in the CONUS during 1630–2020 by integrating multi-source data sets and the machine learning method. The results showed that extensive cropland and pasture expansion and natural vegetation loss occurred from 1630 to 2020 in the CONUS. New reclaimed cropland was primarily converted from forest, shrubland, and grassland. Tree planting and forest regeneration increased the forest cover in the Northeast and the Southern in the recent century. Compared to other LULC datasets, our data provided more accurate information with higher spatial and temporal resolution and better captured the characteristics of LULC changes. The LULC data can be used for regional studies on various topics, including LULC impacts on the regional climate, ecosystems, biodiversity, water resource, carbon and nitrogen cycles, and greenhouse gas emissions.

Appendices

Table A1: Spatially explicit variables adopted for artificial neural network (ANN) modelling

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<tr>
<td>Variable</td>
<td>Description</td>
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<td>------------------------------</td>
<td>------------------------------------------------------------------------</td>
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Table A2: Land use and land cover datasets used for comparison.

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<th>Data variables</th>
<th>Time period</th>
<th>Resolution</th>
<th>Data sources</th>
</tr>
</thead>
<tbody>
<tr>
<td>Land Use Harmonization (LUH2) Cropland density (YUmap)</td>
<td>1600-2020</td>
<td>0.25 degree Annual</td>
<td><a href="http://luh.umd.edu/">http://luh.umd.edu/</a></td>
</tr>
<tr>
<td>Historical fractional cropland areas (ZCmap)</td>
<td>1850-2000</td>
<td>1 km Annual</td>
<td>Yu and Lu (2017). <a href="https://doi.pangaea.de/10.1594/PANGAEA.881801">https://doi.pangaea.de/10.1594/PANGAEA.881801</a></td>
</tr>
<tr>
<td>Hay area</td>
<td>1840-2012</td>
<td>County-level 10-year interval</td>
<td>Zumkehr and Campbell (2013). <a href="https://portal.nersc.gov/project/m2319/">https://portal.nersc.gov/project/m2319/</a></td>
</tr>
<tr>
<td>Crop area</td>
<td>1840-2017</td>
<td>County-level 10-year interval</td>
<td>Sohl et al. (2018) <a href="https://www.sciencebase.gov/catalog/item/59d3c73de4b05fe04c3d1d1">https://www.sciencebase.gov/catalog/item/59d3c73de4b05fe04c3d1d1</a></td>
</tr>
</tbody>
</table>

Note: ERS: Economic Research Service, U.S. Department of Agriculture; YUmap: cropland density (Yu and Lu, 2017); ZCmap: historical fractional cropland areas (Zumkehr and Campbell, 2013).

Figure A1: Vegetation type pre-Euro-American settlement.
Figure A2: National crop harvested and planted area during 1889–2020.

Figure A3: Scatter plot of crop harvested area and planted area during 1974–2017.
Figure A4: Portion of double-cropped area at state-level.

Figure A5: Changes of the History Database of Global Environment cropland and pasture per capita during 1600–2017.
Figure A6: Extent of settled area in 1700, 1800, 1830, 1850, 1890, and present.
Figure A2: Statistical comparison between cropland area from this study and census data in 1850, 1920, 1959, and 2002.
Figure A37: Fractional urban land in the conterminous United States during 1630–2020.
Figure A48: Fractional cropland in the conterminous United States during 1630–2020.
Figure A59: Fractional pasture in the conterminous United States during 1630–2020.
Figure A610: Fractional forest in the conterminous United States during 1630–2020.
Author contributions

HT designed the research; XL implemented the research and analyzed the results; XL, HT, SP, and CL wrote and revised the manuscript.

Competing interests

The authors declare that they have no conflict of interest.

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