1	A homogenized daily <i>in situ</i> PM _{2.5} concentration dataset from national air quality			
2	monitoring network in China			
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Abstract

20 In situ PM_{2.5} concentration observations have long been used as critical data sources in haze related 21 studies. Due to the frequently occurred haze pollution events, China started to regularly monitor $PM_{2.5}$ 22 concentration nationwide from the newly established air quality monitoring network since 2013. 23 Nevertheless, the acquisition of these invaluable air quality samples is challenging given the absence of public available data download interface. In this study, we provided a homogenized in situ PM2.5 24 25 concentration dataset that was created on the basis of hourly PM2.5 data retrieved from the China 26 National Environmental Monitoring Center (CNEMC) via a web crawler between 2015 and 2019. 27 Methods involving missing value imputation, change point detection, and bias adjustment were applied 28 sequentially to deal with data gaps and inhomogeneities in raw $PM_{2.5}$ observations. After excluding 29 records with limited samples, a homogenized PM_{2.5} concentration dataset comprising of 1,309 five-30 year long PM_{2.5} data series at a daily resolution was eventually compiled. This is the first thrust to 31 homogenize in situ PM2.5 observations in China. The trend estimations derived from the homogenized 32 dataset indicate a spatially homogeneous decreasing tendency of PM2.5 across China at a mean rate of 33 about -7.6% per year from 2015 to 2019. In contrast to raw PM_{2.5} observations, the homogenized data 34 record not only has a complete data integrity but is more consistent over space and time. This 35 homogenized daily in situ PM_{2.5} concentration dataset is publicly accessible at https://doi.pangaea.de/10.1594/PANGAEA.917557 (Bai et al., 2020a), which can be applied as a 36 37 promising dataset for PM_{2.5} related studies such as satellite-based PM_{2.5} mapping, human exposure 38 risk assessment, and air quality management.

39 Keywords: PM_{2.5}; Data homogenization; Bias correction; *In situ* observation; Air quality indicators

40 1 Introduction

41 A consistent PM_{2.5} concentration dataset is vital to the analysis of variations in PM_{2.5} loadings 42 over space and time as well as in support of its risk analysis for air quality management, meteorological forecasting, and health-related exposure assessment (Lelieveld et al., 2015; Yin et al., 2020). Ground-43 44 based monitoring network is commonly built to measure concentrations of air pollutants across the globe. Suffering from extensive and severe haze pollution events in the past few years (Guo et al., 45 2014; Ding et al., 2016; Wang et al., 2016; Cai et al., 2017; Huang et al., 2018; Luan et al., 2018; Ning 46 47 et al., 2018), China launched the operational ambient air quality sampling late in 2012 on the basis of 48 the sparsely distributed aerosol observation network. To date, this in situ network has been enlarged 49 to cover almost all major cities in China consisting of about 1500 monitoring stations. Concentrations 50 of six key air pollutants including PM_{2.5}, PM₁₀, NO₂, SO₂, CO, and O₃, are routinely measured on an 51 hourly basis while the sampled data are released publicly online by the China National Environmental 52 Monitoring Center (CNEMC) since 2013.

53 Although in situ PM_{2.5} concentration data have played critical roles in improving our 54 understanding of regional air quality variations and relevant influential factors (Yang D. et al., 2018; 55 Yang Q. et al., 2019; Zheng et al., 2017), little concern was raised to the quality of such dataset itself 56 (Bai et al., 2019a, 2019c; He and Huang, 2018; Zhang et al., 2019, 2018; Zou et al., 2016). Meanwhile, 57 few studies provided a detailed description of the accuracy or bias level (uncertainty) of the observed PM_{2.5} data in recent years (Xin et al., 2015; You et al., 2016; Guo et al., 2017; Shen et al., 2018). The 58 59 primary reason lies in the fact that neither quality assurance flag nor metadata information 60 documenting the uncertainty other than data samplings were provided, making such quality assessment infeasible. 61

The data quality, in particular the data homogeneity, is of critical importance to the exploration of the given dataset, especially for trend analysis (Bai et al., 2019c; C. Lin et al., 2018; Liu et al., 2018; Ma et al., 2015) and data integration (Bai et al., 2019b, 2020b; T. Li et al., 2017; Zhang et al., 2019) in which a homogeneous dataset is absolutely essential for downstream applications. Since two distinct kinds of instruments are used in the current air quality monitoring network to measure near surface PM_{2.5} concentration in China (Bai et al., 2020), imperfect instrumental calibration and intermittent replacement of instruments may thus introduce obvious issue of discontinuity in PM_{2.5} observations. Such inhomogeneity may result in large uncertainty and even biased results in the subsequent analysis, especially in context-based and data driven PM_{2.5} concentration mapping (Bai et al., 2019b, 2019a; He and Huang, 2018; Wei et al., 2020), in which *in situ* PM_{2.5} concentration observations are used as the ground truth to characterize complex statistical relationships with other possible contributing factors.

73 Given the absence of an open access and quality assured in situ PM_{2.5} concentration dataset in 74 China, in this study, we attempted to generate a long-term coherent in situ PM2.5 concentration dataset 75 for scientific community to use in future applications. A set of methods involving missing value 76 imputation, change point detection, and bias adjustment were geared up seamlessly in a big data 77 analytic manner toward the improvement of data integrity and the removal of possible discontinuities 78 in raw PM_{2.5} observations. Such an analytical process is also referred to as data homogenization in 79 data science or big data analytics (Cao and Yan, 2012; Wang et al., 2007). To our knowledge, this is 80 the first thrust to homogenize a large-scale dataset of *in situ* PM_{2.5} concentration observations in China. 81 In the following sections, we will introduce the data source as well as detailed big data analytics 82 methods used for the creation of a homogenized PM_{2.5} concentration dataset.

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2 In situ PM_{2.5} concentration observations

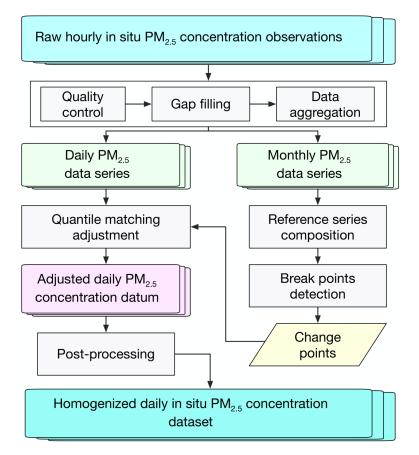
84 In this study, the hourly PM_{2.5} concentration data sampled from more than 1,600 state-controlled 85 air quality monitoring stations across China between January 1, 2015 and December 31, 2019 were utilized. These PM2.5 concentration data were measured on an hourly basis using either beta-86 87 attenuation monitors or Tapered Element Oscillating Microbalance (TEOM) analyzer. The ordinary 88 instrumental calibration and quality control were performed according to the national ambient air 89 quality standard of GB3095-2012 and HJ 618-2011 (Guo et al., 2009, 2017). Generally, TEOM can 90 measure PM_{2.5} concentration within the range of $0-5,000 \ \mu g \ m^{-3}$ at a resolution of 0.1 $\mu g \ m^{-3}$, with precisions of $\pm 0.5 \ \mu g \ m^{-3}$ for 24-h average and $\pm 1.5 \ \mu g \ m^{-3}$ for hourly average (Guo et al., 2017; Xin 91 92 et al., 2012; Xin et al., 2015). The $PM_{2.5}$ measurements were publicly released online by the China 93 National Environmental Monitoring Center (CNEMC) via the National Urban Air Quality Real-time
94 Publishing Platform (http://106.37.208.233:20035/) within one hour after the direct sampling.

95 Although the sampled data were publicly released, the acquisition of these valuable samplings is 96 always challenging because no data download interface is provided to the public by the CNEMC 97 website. Therefore, it is impossible for users to retrieve the historical observations from the given website. Rather, science community has to count on other measures such as an automatic web crawler 98 99 for the retrieval of these online updated data samples from the data publishing platform. Nevertheless, 100 the data records retrieved through such an approach suffered from significant data losses due to various 101 unexpected reasons like power outage and internet interruption. Consequently, the data integrity 102 becomes problematic and further treatments like gap filling are thus essential to accounting for such 103 defects at least.

104 Moreover, hourly PM_{2.5} concentration observations that were sampled at five embassies of United 105 States in China from January 2015 to June 2017 were used as an independent dataset to evaluate the 106 fidelity of the homogenized PM_{2.5} concentration dataset. Geographic locations of these five embassies 107 have been shown in Table S1. These PM_{2.5} data were measured independently under the U.S. 108 department of state air quality monitoring program and can be acquired from the 109 http://www.stateair.net/. To be in line with the homogenized dataset, the hourly PM_{2.5} concentration 110 data were aggregated to the daily level by averaging the 24-h observations sampled on each date while 111 daily averages were calculated only for days with more than 12 valid samples of a possible 24-h.

112 **3** Homogenization of *in situ* PM_{2.5} concentration data

For the creation of a long-term coherent *in situ* PM_{2.5} concentration dataset, it is necessary to create an analytical framework of the big data analytics which seamlessly gears up several methods as a whole for the purposes of missing value imputation, change point detection, and discontinuity adjustment, given the presence of data gaps and possible discontinuity in raw PM_{2.5} observations . Figure 1 shows a schematic illustration of the general workflow toward generating a homogenized PM_{2.5} concentration dataset and the whole process can be outlined as follows. (1) It is necessary to perform essential quality control and gap filling on raw PM_{2.5} observations so
 that the bias arising from large outliers and resampling errors due to incomplete observations can
 be reduced.



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Figure 1. A schematic flowchart for the creation of a homogenized daily *in situ* PM_{2.5} concentration
dataset.

125

126 (2) Short-term time series due to sites relocation were temporally merged to attain a long-term record. 127 Then, PM_{2.5} concentration time series with a temporal coverage of less than four-year during the study period were excluded. Subsequently, the quality-controlled observations of hourly in situ 128 129 PM_{2.5} concentrations were resampled to daily and monthly scales to initiate the homogeneity test. 130 (3) Reference time series were constructed for each long-term PM_{2.5} concentration record on the basis 131 of data measured from adjacent monitoring sites. For PM2.5 concentration records failing to 132 produce a reliable reference series, no homogeneity test was performed for such datum due to the absence of essential reference data series. 133

(4) The discontinuity identified in each daily long-term PM_{2.5} concentration time series were corrected
 using the quantile-matching (QM) adjustment method according to the change points detected in
 each monthly data record with the support of reference series.

(5) Post-processing measures such as nonpositive value correction and another round gap filling were
 further performed on the homogenized records to attain a quality-assured *in situ* PM_{2.5}
 concentration dataset. More details of each analytic method were described in the following
 subsections.

141 **3.1 Quality control**

Given the possibility of the presence of abnormal samplings, it is necessary to remove the outliers detected in raw $PM_{2.5}$ observations to reduce the false alarm rate in change point detection during the subsequent homogeneity test. Specifically, hourly $PM_{2.5}$ concentration data values meeting one of the following criteria were excluded: 1) out of the range between 1 and 1,000 µg m⁻³, and 2) more than three standard deviations from the median of observations within a 15-h time window. Both criteria aimed to remove large outliers which could result in biased daily averages. Overall, 3.46% of $PM_{2.5}$ samples were treated as outliers and were then excluded accordingly (treated as missing values).

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3.2 Gap filling and resampling

150 As indicated in our recent study (Bai et al., 2020b), missing value related data gaps become a 151 big obstacle in the exploitation of raw PM_{2.5} observations that were retrieved from the CNEMC website as PM_{2.5} observations on 40% of sampling days suffered from data losses due to unexpected reasons. 152 153 To reduce the impact of missing value related sampling (from hourly to daily) bias on the subsequent 154 homogeneity test, we filled those missing value related data gaps that were found in each 24-h PM_{2.5} 155 observations by applying the DCCEOF method developed very recently (Bai et al., 2020b). Such a 156 gap filling effort enabled us to improve the percentage of days without missingness during the study 157 time period from 58.8% to 97.3%.

In spite of the improvement of data integrity after gap filling, the resultant $PM_{2.5}$ time series remain temporally discontinuous due to the emergence of several long-lasting (e.g., more than 24 consecutive hours) data missing episodes. Also, the hourly time series are still too noisy to be handled 161 by the current homogeneity test software due to the significant variation in PM2.5 concentration over 162 space and time. In such context, the hourly PM2.5 concentration records were aggregated to daily and 163 monthly scales to initiate the homogeneity test. Moreover, the monthly series was primarily used to 164 detect the possible change points while the daily series was adjusted in reference to the corresponding 165 reference series based on the change points detected from the monthly series. To avoid large 166 resampling bias, monthly averages were calculated only for those with at least 20 valid daily means of 167 a possible month at each site. The frequency of missing values in each month was also calculated as a 168 possible metadata information to further examine the detected change points.

169 **3.3 Homogeneity test**

170 A commonly used homogeneity test software, the RHtestsV4 package, was hereby applied to 171 detect the possible discontinuities in raw PM_{2.5} data series that were retrieved from the CNEMC website. As suggested in Wang and Feng (2013), RHtestsV4 is capable of detecting and adjusting 172 173 change points in a data series with first-order autoregressive errors. Given the low false alarm rate via 174 change point detection and the capability to adjust discontinuity, the RHtests software packages have 175 been widely used to homogenize climate data records such as temperature (Cao et al., 2013; Xu et al., 176 2013; Zhao et al., 2014), precipitation (Wang et al., 2010a; Nie et al., 2019), and other datum like 177 boundary layer height (Wang and Wang, 2016). Two typical methods, namely the PMTred and 178 PMFred, were embedded in a recursive testing algorithm in RHtestsV4, with the former relying on the 179 penalized maximal t test (PMT) while the latter based on the penalized maximal F test (PMF) (Wang 180 et al., 2007; Wang, 2008a). With the incorporation of these empirical penalty functions (Wang, 2008a, 181 b), the problem of uneven distribution of false alarm rate is largely alleviated with the aid of RHtestsV4. 182 In contrast to the PMF which works without a reference series, the PMT uses a reference series to 183 detect change points and the results are thus far more reliable (Wang, 2008a, b). The way to generate 184 reference series will be described in the next subsection. Also, the RHtestsV4 is capable of making 185 essential adjustments to the detected discontinuities by taking advantage of the QM adjustment method 186 (Wang and Feng, 2013).

187 Here the PMT method rather than the PMF was used to detect change points given the higher 188 confidence of the former method in change point detection due to the involvement of reference series 189 (Wang and Feng, 2013). To ensure the reliability of detected discontinuities, change point was defined 190 and confirmed at a nominal 99% confidence level, and the data records were then declared to be 191 homogeneous once no change point was identified. Subsequently, the QM adjustment method was 192 applied to correct PM_{2.5} observations with evident drifts with the support of reference series, namely, 193 to homogenize PM_{2.5} concentration data series. To avoid large sampling uncertainty in the estimate of 194 QM adjustments, the Mq (i.e., the number of categories on which the empirical cumulative distribution 195 function is estimated) was automatically determined by the software to ensure adequate samples for 196 the estimation of mean difference and probability density function. Meanwhile, the number to 197 determine the base segment (i.e., *Iadj*) was set to zero so that datum in other segments were all adjusted 198 to the segment with the longest temporal coverage.

199

3.3.1 **Construction of reference series**

200 A good reference series is vital to the relative homogeneity test because it helps pinpoint possible 201 discontinuities in each base series (the data series to be tested) and determines the performance of the 202 subsequent data adjustment. In general, reference series can be organized by using one specific record 203 either measured from one adjacent station or aggregated from multiple observations (Cao and Yan, 204 2012; Peterson and Easterling, 1994; Xu et al., 2013; Wang et al., 2016). The most straightforward 205 way is to use the neighboring data series either measured at the nearest station or series that are highly 206 correlated with the base series (Peterson and Easterling, 1994; Cao and Yan, 2012; Wang and Feng, 207 2013). Such methods, however, fail to take the representativeness of the neighboring series into 208 account since the neighboring series may also suffer from discontinuities.

209 To avoid the misuse of inhomogeneous PM_{2.5} concentration records as reference series, a 210 complex yet robust data integration scheme was hereby developed to screen, organize, and construct 211 reference series for each in situ PM2.5 concentration data series. For each daily PM2.5 concentration 212 data series, all the neighboring series were firstly identified from its surroundings with a lag distance 213 as large as of 50 km. No reference series was constructed once there was no neighboring series 214 available within the given radius and in turn the homogeneity of the given record was not examined. Otherwise, both correlation coefficient (R) and coefficient of variation (CV) were calculated between the given base series and each selected neighboring series to assess their representativeness (Shi et al., 2018; Rodriguez et al., 2019). Then, neighboring series with R greater than 0.8 and CV smaller then 0.2 were selected as candidates to construct the reference series for a given base series.

219 The reference series was then constructed by averaging both the base and the candidate series at 220 each observation time if there was only one candidate series. For the situation with more than one 221 candidate series, the empirical orthogonal function (EOF) method was applied to these multiple 222 candidates and then the original fields were reconstructed with the leading principal components when 223 the accumulated variance explained by them exceeded 80%. This was expected to reduce the possible 224 impacts of abnormal observations and short-term discontinuities in the neighboring candidates on the 225 resultant reference series. Subsequently, the reference series were organized and constructed through 226 a spatial weighting scheme as each reconstructed record was assigned a spatially resolved weight 227 according to their relative distances to the base series over space. Here we applied a Gaussian kernel 228 function to estimate the weight of each neighboring observation that can influenced the base series in 229 space and such a scheme has been proven to be effective in assessing the spatial autocorrelation of PM_{2.5} concentration (Bai et al., 2019b). Mathematically, the reference series can be constructed from 230 231 the following equations:

$$PM_{ref} = \sum_{i=1}^{N} \frac{w_i * PM_{cand}^i}{\sum w_i}$$
(1)

233
$$w = \exp\left(\frac{-d^2}{2h^2}\right) \tag{2}$$

232

where PM_{ref} and PM_{cand} denote the reference and candidate series, respectively. *N* is the total number of candidate series while *w* is the spatially resolved weight assigned to each candidate series and *d* is the spatial lag distance between the base and the corresponding candidate series. *h* is a spatial correlation length that is used to modulate the relative influence of a distant observation on the data measured at the base site. In this study, an empirical value of 50 km was used according to the estimated semi-variogram results (Bai et al., 2019b).

For any record having neighboring series within 50 km but poorly correlated (R<0.8 or CV>0.2) to all its neighbors (meaning the base series differ from the neighbors), the reference series were 242 created by following the same procedures as those detailed above by taking the nearest neighbor as the 243 base series. For the situation with only one candidate series available, it is logical to compare both the 244 base and the candidate series against another data to check which one should be corrected. In this study, the PM_{2.5} time series estimated from the MERRA-2 aerosol reanalysis in the same way as described 245 246 in He et al. (2019) was used. The one with higher correlation to this external PM_{2.5} time series was then used as the reference (deemed as homogeneous) while the other was considered as the base series 247 248 (i.e., implies to be adjusted). Such an inclusive scheme empowered us to screen and construct reference 249 series for 1,262 long-term PM_{2.5} concentration records across the board. In contrast, no reference series 250 were constructed for 47 isolated records.

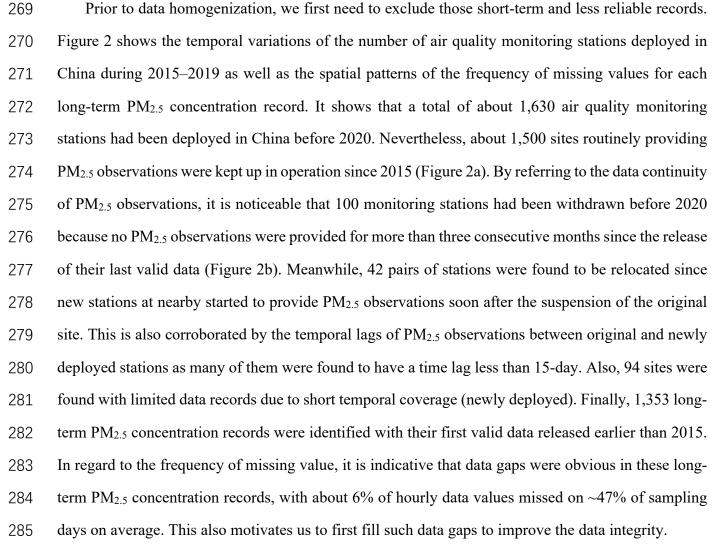
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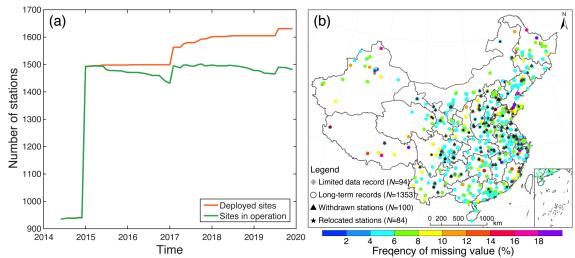
3.3.2 Post-processing measures

252 Several post-processing measures were applied to the adjusted data records to further improve 253 the quality of this dataset. Since nonpositive values may appear in the QM adjusted data series if the 254 original values are close to zero (Wang et al., 2010b), nonpositive values were replaced with the 255 smallest valid PM_{2.5} concentration amount measured at each monitoring site during the study period. 256 Subsequently, the data gaps in the adjusted datum due to long-lasting missingness were filled by first 257 calibrating the corresponding data values in the reference series measured on the same date (if available) 258 to the homogenized datum level. The modified quantile-quantile adjustment (MQQA) method 259 proposed in Bai et al. (2016) was hereby used given its adaptive data adjustment principle. For the 260 predicted values, such MQQA scheme rendered higher accuracy than those interpolated from data 261 values measured on adjacent dates because PM2.5 concentration is spatially more correlated than in the 262 temporal domain (Bai et al., 2019b). For the remaining data gaps, those missing values were 263 reconstructed in a similar procedure as the DCCEOF method (Bai et al., 2020b). Note that the matrix 264 used for EOF analysis in the context of DCCEOF was constructed using the neighboring data series 265 measured within a radius of 100 km with a temporal lag of 30 days at most. Finally, all data values 266 were rounded to integer to be in line with the original $PM_{2.5}$ concentration observations.

267 4 Results and discussion

268 4.1 Descriptive statistics





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Figure 2. Spatial and temporal patterns of air quality monitoring stations in China. (a) Temporal variations of the total number of air quality monitoring stations in China. (b) Spatial patterns of the frequency of missing value in each long-term hourly $PM_{2.5}$ concentration record measured from

January 1, 2015 to December 31, 2019. Stations were categorized into distinct groups according to their data length and temporal continuity. The frequency of missingness was calculated as the ratio of the number of missing values in each $PM_{2.5}$ concentration record to the total number of samplings from the time of the release of the first valid data to December 31, 2019.

294 4.2 Homogenization of *in situ* PM_{2.5} data

295 A total of 1,395 long-term (with five-year observations) PM_{2.5} concentration records were 296 acquired with the inclusion of 42 temporally merged data series at those relocated stations. After 297 removing those suffering from more than three consecutive months data losses, 1,309 long-term yet 298 consecutive PM_{2.5} concentration records were obtained. The homogeneity test was finally performed 299 on 1,262 records due to the availability of reference series. Figure 3 shows the spatial patterns of the 300 total number of change points detected in 1,262 monthly PM_{2.5} concentration records. The ubiquitous 301 change points imply that there is an obvious inhomogeneity in this in situ PM_{2.5} concentration dataset. About 57% (719 out of 1,262) of records failed to pass the homogeneity test due to the presence of 302 303 change points. Given the overall good agreement between the base and reference series (refer to Figure 304 S1 for the correlation coefficient and root mean square error between them), it indicted that these PM_{2.5} 305 concentration records did suffer from evident discontinuities. Meanwhile, the vast majority (~80%) of 306 the inhomogeneous PM_{2.5} records suffered from no more than two change points (Figure 3), suggesting 307 the mean shift could be the primary reason for the detected discontinuities. Moreover, 20 records were 308 even found suffering from no less than five significant change points, indicating phenomenal 309 discontinuities in these records.

310 Figure 4 shows the temporal variability of the number of change points detected in monthly PM_{2.5} 311 concentration records. As indicated, change points were detected in every specific month of the year 312 from May 2015 to July 2019, especially in late spring (e.g., May), in which change pointes were more 313 likely to be detected (Figure 4b). This is attributable to the seasonality of PM_{2.5} loading in China as high PM_{2.5} concentrations are always observed in the winter whereas low values in the summer. 314 315 Consequently, change points were more likely to be detected during the chronic transition periods (e.g., 316 spring to summer). In addition, it is noteworthy that a large volume of change points was detected in 317 early 2015, indicating the existence of phenomenal discontinuities during this period (Figure 4a). After 318 checking the temporal variations of $PM_{2.5}$ concentration, findings indicate that $PM_{2.5}$ observations 319 varied with large deviations among each other during this period. This could be linked to the imperfect 320 instrument calibration or irregular operation in the early stage.

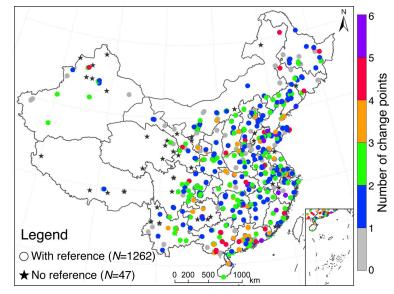




Figure 3. Spatial patterns of the total number of change points detected in each long-term yet consecutive PM_{2.5} concentration records. Gray dot indicates there was no change point detected in this PM_{2.5} concentration record.

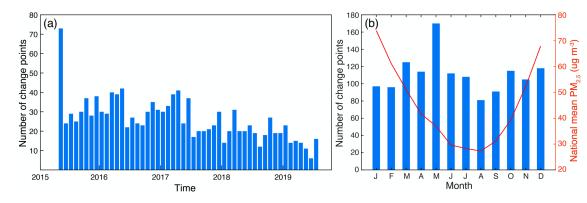
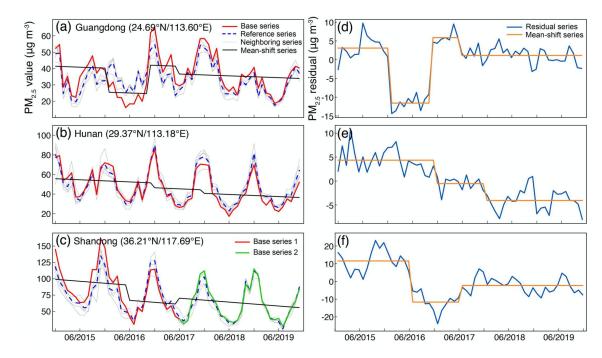


Figure 4. Temporal variations of the number of change points detected in (a) each specific month from 2015 to 2019 and (b) each month of the year. National mean PM_{2.5} concentration in each month of the year was calculated based on PM_{2.5} data measured at our selected 1309 sites during 2015–2019.

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330 Due to the lack of essential metadata information, it is a challenge for us to verify each detected 331 change point through a manual inspection. Rather, the variations in the base and reference series was 332 explored to identify the possible reasons for the detected discontinuities. Figure 5 presents three typical 333 inhomogeneous PM_{2.5} time series with different number of change points. The inter-comparisons 334 between the base and reference series indicate an overall good agreement among them in terms of the 335 long-term variation tendency. However, obvious drifts were still phenomenal in their residual series, 336 which were even more evident by referring to their mean-shift series. For example, both the residual 337 and mean-shift series shown in Figure 5d clearly illustrate a typical discontinuity as there was an 338 obvious departure of mean PM_{2.5} concentration level during the period of January to October 2016. In 339 contrast, the Figures. 5b and 5e present another typical inhomogeneity as statistically significant 340 decreasing trend was found in the residual series with monthly PM_{2.5} concentration deviations decreased from nearly 5 µg m⁻³ to -4 µg m⁻³ step wise. Such inhomogeneity would undoubtedly result 341 342 large bias in the trend estimations over that region. The bottom panel (Figures. 5c and 5f) shows the 343 change points detected in the merged PM_{2.5} time series at a pair of relocated sites. It is noteworthy that 344 the detected discontinuity should be largely ascribed to the inconsistency emerged in the first data 345 series rather than due to the site relocation.



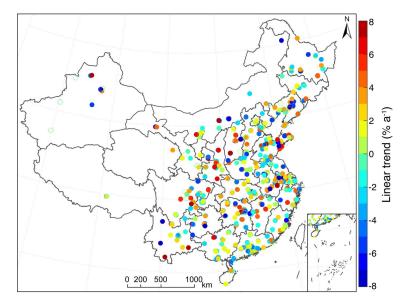
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Figure 5. Temporal variations of three typical inhomogeneous PM_{2.5} concentration records during 2015–2019. (Top) Significant deviations during a short time period, (middle) long-term chronic drifts with statistically significant varying trend detected in the residual series, (bottom) discontinuity due to site relocation. The left panel compares the base series with the reference and the neighboring series

351 used to compose the reference while the right panel shows the residual series between the base and 352 reference series as well as their mean-shift series.

353

354 Figure 6 shows the estimated linear trends for PM_{2.5} residual series that failed to pass the 355 homogeneity test. Approximately 89% of the residual series were found exhibiting statistically 356 significant linear trends, suggesting the vital importance to homogenize such PM_{2.5} concentration 357 records as the trend estimations at these stations could be prone to large bias if no essential adjustments 358 are performed. Further comparisons of the percentage of data gaps between homogeneous and 359 inhomogeneous records (Figure S2) as well as the spatial distance between the base and the reference 360 series (Figure S3) indicate that both the frequency of data gaps and spatial distance have no obvious 361 impact on the change point detection. In other words, the detected change points have no linkage with 362 neither missing value frequency nor spatial distance between the base and neighboring series, 363 suggesting a high confidence level of the identified discontinuities in these PM_{2.5} concentration records.

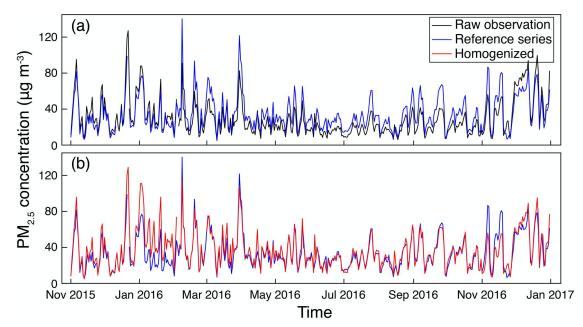


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Figure 6. Trend estimations for the residual $PM_{2.5}$ concentration data series that failed to pass the homogeneity test during 2015–2019. The solid circles indicate trends are statistically significant at the 95% confidence level.

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369 Given the emergence of obvious discontinuities in more than half of the selected long-term PM_{2.5} 370 concentration records, the QM adjustment method was applied to correct the discontinuities detected in each $PM_{2.5}$ concentration record. Figure 7 shows an example of homogenization on $PM_{2.5}$ concentration data series that suffered from evident drifts from its reference (large drifts shown in Figure 5d). The inter-comparisons of $PM_{2.5}$ concentration data between the base and reference series indicate that the $PM_{2.5}$ concentration level was obviously underestimated by the raw observations compared with the reference, especially during the middle of 2016 (Figure 7a). Such evident drifts were remarkably diminished after the homogenization (Figure 7b), which shows a good agreement of the mean $PM_{2.5}$ concentration level between the homogenized datum and the reference series.



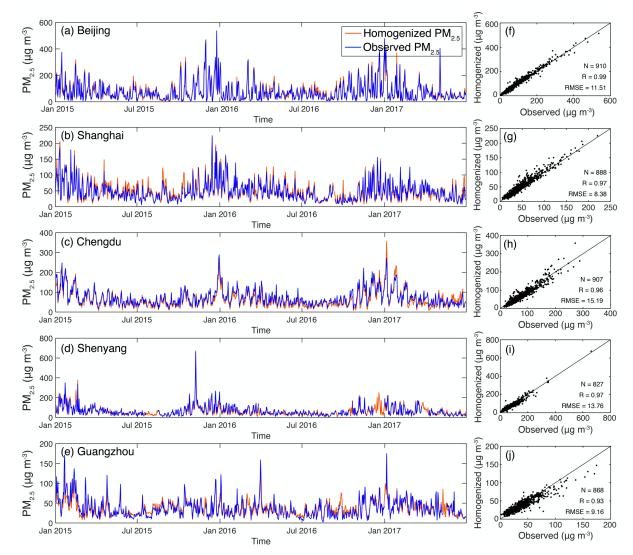
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Figure 7. Comparison of daily mean PM_{2.5} concentration before and after homogenization at one
monitoring site in Guangdong province (24.69°N/113.60°E) from November 2015 to December 2016
(large drifts shown in Figure 5d).

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383 4.3 Validation with independent dataset

In this study, $PM_{2.5}$ observations that were collected independently by five consulates of United States distributed in five major Chinese cities between 2015 and 2017 were used to evaluate the consistency of the derived $PM_{2.5}$ concentration records. Figure 8 shows site-specific comparisons of daily $PM_{2.5}$ concentration between homogenized and observed data in Beijing, Shanghai, Chengdu, Shenyang, and Guangzhou, respectively. It is indicative that the homogenized daily $PM_{2.5}$ concentration data were in good agreement with $PM_{2.5}$ observations sampled at US consulates, with a 390 correlation coefficient value of >0.95 and root mean square error of <15 μ g m⁻³. Given the 391 independent measurement of PM_{2.5} concentration data at US consulates, we argue that the 392 homogenized PM_{2.5} records are accurate enough in characterizing the variability of PM_{2.5} loadings in 393 China. It is also noteworthy that the homogenized PM_{2.5} records are temporally complete whereas 394 missing values are found in PM_{2.5} observations sampled at US consulates.



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Figure 8. Comparisons of the homogenized PM_{2.5} concentration (red) against PM_{2.5} observations (blue)
 measured at five consulates of United States in China from January 2015 to June 2017. (a–e) Temporal
 variations of daily PM_{2.5} concentration and (f–j) the associated scatter plots.

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400 **4.4 PM**_{2.5} trends estimated from the homogenized dataset

401 A homogenized data record is essential to trend analysis. Figure 9 presents the annual mean 402 concentration of $PM_{2.5}$ across China between 2015 and 2019. As shown, there is a phenomenal 403 reduction of $PM_{2.5}$ concentration in China in the past five years, especially over North China Plain (the 404 region outlined by a red rectangle shown in Figure 9f) where the annual mean $PM_{2.5}$ concentration 405 decreased from more than 100 µg m⁻³ in 2015 to about 60 µg m⁻³ in 2019. Such an evident decrease in 406 $PM_{2.5}$ concentration clearly demonstrates the effectiveness of clean air actions that were implemented 407 in recent years.

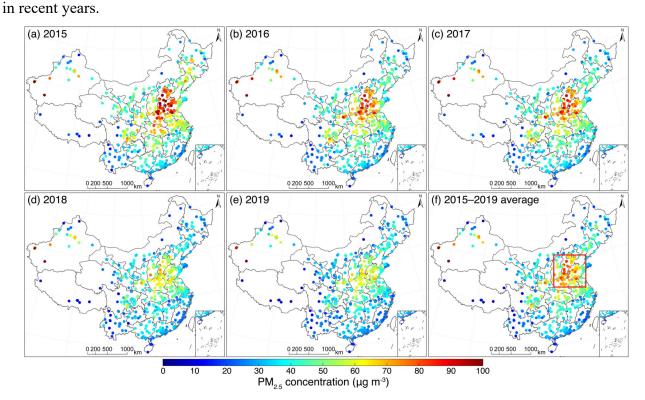
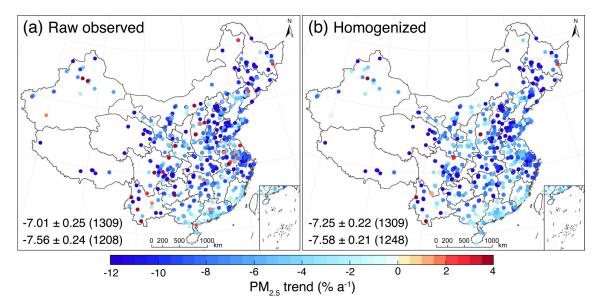


Figure 9. Annual mean PM_{2.5} concentration derived from the homogenized daily PM_{2.5} concentration
dataset at 1,309 monitoring stations in China between 2015 and 2019. The North China Plain was
outlined by the red rectangle in panel (f).

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To evaluate the benefits of data homogenization on PM_{2.5} trend estimations, PM_{2.5} trends estimated from both the raw observations and homogenized dataset were compared. Prior to trend analysis, each PM_{2.5} concentration record was standardized in reference to its mean annual cycle (i.e., PM_{2.5} concentration on the same date of the year between 2015 and 2019 was averaged) to reduce the impacts of seasonality and spatial variations. Figure 10 shows a site-specific comparison of PM_{2.5} trend estimations derived from raw observations and homogenized datasets during 2015–2019. In general, 19 trend estimations from both datasets showed an evident decreasing tendency of $PM_{2.5}$ concentration across China during the study period. Nevertheless, noteworthy is that trend estimations derived from raw $PM_{2.5}$ observations suffered from obvious inhomogeneity over space, being evidenced by antiphase (positive versus negative) trend estimations even at adjacent stations, especially for those with positive trends whereas all adjacent neighbors exhibited negative trends. These antiphase trend estimations over a small region also corroborate the existence of obvious inhomogeneity in raw observed *in situ* $PM_{2.5}$ concentration dataset.



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Figure 10. Linear trends for (a) raw observed and (b) homogenized daily $PM_{2.5}$ concentration data during 2015–2019. Solid circles indicate trends are statistically significant at the 95% confidence interval. Numbers shown in the lower left of each panel indicate the overall trend derived from (top) all available stations and (bottom) the stations with significant trends at the 95% confidence interval while the numbers shown in brackets are the corresponding number of data records. Each $PM_{2.5}$ time series were standardized by its mean annual cycle during the study period to account for spatial variations of $PM_{2.5}$.

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The dotted antiphase trend estimations were substantially diminished after data homogenization, resulting in a spatially much more homogeneous decreasing tendency of $PM_{2.5}$ concentration across China (Figure 10b). It is indicative that after data homogenization the national mean $PM_{2.5}$ trend was enlarged from -7.01% a⁻¹ to -7.25% a⁻¹ while the uncertainty was reduced from 0.25% a⁻¹ to 0.22% a⁻¹ 439 ¹. Also, the number of $PM_{2.5}$ records with statistically significant trends was increased from 1,208 to 440 1,248. These results collectively justify the effectiveness of the QM adjustment method in mitigating data inhomogeneity in PM_{2.5} observations, which also highlight the critical importance of data 441 442 homogenization in accounting for discontinuities in this in situ PM2.5 concentration dataset. Overall, 443 our results indicate an obvious decreasing trend of PM2.5 concentration in China in the past five years at a mean rate of $-7.25 \pm 0.22\%$ a⁻¹. Table 1 further compares the regional mean PM_{2.5} trend between 444 2015 and 2019. Compared with other regions of interest (ROIs) such as Pearl River Delta (PRD, refer 445 446 to Figure S4 for the location) and northern part of Xinjiang (XJ), PM_{2.5} loading over Beijing-Tianjin-447 Hebei (BTH), Heilongjiang-Jilin-Liaoning (HJL), and Central China (CC) decreased even more 448 prominently.

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Table 1. Regional mean trend for PM_{2.5} concentrations over eight major ROIs in China during 2015–
2019 before and after the data homogenization. Uncertainty in trend estimations were characterized at
the 95% confidence interval. Locations of these ROIs can be found in Figure S4.

ROI	Raw observation (% a ⁻¹)	Homogenized record (% a ⁻¹)
Beijing-Tianjin-Hebei (BTH)	-9.03 ± 0.78	$\textbf{-9.19}\pm0.69$
Yangtze River Delta (YRD)	-7.07 ± 0.54	-7.33 ± 0.40
Central China (CC)	-8.47 ± 0.51	$\textbf{-8.58} \pm \textbf{0.41}$
Sichuan Basin (SCB)	-7.39 ± 1.02	-7.84 ± 0.89
Pearl River Delta (PRD)	-4.30 ± 0.51	-4.60 ± 0.39
Heilongjiang-Jilin-Liaoning (HJL)	-8.89 ± 0.73	$\textbf{-9.15}\pm0.63$
Shaanxi-Gansu-Ningxia (SGN)	-4.85 ± 0.95	-5.30 ± 0.69
North Xinjiang (XJ)	-4.61 ± 1.96	-4.67 ± 1.60

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454 To further assess the improvement of the data quality after homogenization, the daily *in situ* 455 $PM_{2.5}$ concentration records at a 1° × 1° grid cell resolution were grouped across China. In each grid 456 cell, the regional mean correlation coefficient among $PM_{2.5}$ concentration time series and standard 457 deviation of $PM_{2.5}$ trends were estimated from the raw observed and homogenized daily $PM_{2.5}$ 458 concentration time series, respectively. Their relative differences were then calculated to show the 459 improvements of data homogeneity within each grid cell. As shown in Figure 11, the correlation among 460 PM_{2.5} concentration datum was enhanced ubiquitously after homogenization, especially in the 461 southwest of China (e.g., Yunnan) where obvious inhomogeneity was observed in the raw PM_{2.5} 462 observations (Figure 10a). Meanwhile, the standard deviation of PM2.5 trends within each grid cell was 463 also substantially reduced, even by more than two folds in the magnitude (Figure 11b). These results 464 also demonstrate the critical need to homogenize the observed PM2.5 concentration data from a large-465 scale monitoring network to reduce temporal inconsistency and spatial inhomogeneity that were not 466 even noticed before.

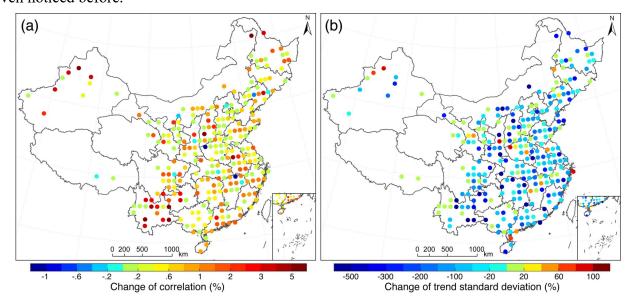


Figure 11. Spatial distributions of (a) the improvements of mean correlation coefficient among $PM_{2.5}$ concentration records before and after homogenization at a 1° × 1° grid cell resolution across China, and (b) their corresponding standard deviations of $PM_{2.5}$ trends.

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472 5 Data availability

The raw observations of *in situ* $PM_{2.5}$ concentration data in China used in this study were retrieved via a web crawler from the National Urban Air Quality Real-time Publishing Platform (http://106.37.208.233:20035) between 2014 and 2019. Given the deployment of many new monitoring sites in 2014, we decided to generate a coherent $PM_{2.5}$ concentration dataset starting from 2015 to include as many $PM_{2.5}$ data records as possible. The homogenized daily *in situ* $PM_{2.5}$ 478 concentration developed this study is publicly accessible dataset in at https://doi.pangaea.de/10.1594/PANGAEA.917557 (Bai et al., 2020a). To provide a long-term 479 coherent PM_{2.5} concentration dataset to the scientific community, the homogenized PM_{2.5} 480 concentration dataset will be regularly updated for each half a year by including new PM_{2.5} 481 482 observations that are retrieved during the past six months.

483 6 Conclusions

In this study, a homogenized yet temporally complete daily in situ PM2.5 concentration dataset 484 was generated based on the discrete hourly PM2.5 concentration records that were retrieved from the 485 China National Urban Air Quality Real-time Publishing Platform using a web crawler during the 486 487 period of 2015–2019. To create such a long-term coherent dataset, a set of analytic methods were geared up seamlessly and applied sequentially to the retrieved raw PM_{2.5} concentration records, 488 involving quality control, gap filling, data merging, change point detection, and bias correction. This 489 new dataset would help scientific community better elucidate the temporal and spatial variability of 490 491 haze pollution in China in the recent years, which is expected to improve the understanding of 492 underlying causes.

The raw PM_{2.5} concentration records were found to be suffering from phenomenal inhomogeneity caused by data inconsistency and temporal discontinuity as well as the relocation and repeal of a bunch of monitoring stations. More than half of the long-term PM_{2.5} concentration records were found failing to pass the homogeneity test due to the presence of considerable change points. Further investigation confirms that large yet short-term mean shifts and chronic drifts are two primary reasons for the detected discontinuities in raw PM_{2.5} concentration records.

Based on the homogenized dataset, the long-term trends of $PM_{2.5}$ concentration in China were estimated. In contrast to the inhomogeneous trend estimations that were derived from raw $PM_{2.5}$ concentration records, the homogenized dataset yielded a spatially much more homogeneous decreasing tendency of $PM_{2.5}$ concentration across China at a mean rate of about -7.3% per year. Such an improvement of homogeneity was also evidenced by the enhanced correlation and reduced standard deviation of trend estimations between homogenized $PM_{2.5}$ concentration time series in the surroundings. These results clearly demonstrate the benefits of data homogenization on the improvement of the quality of this $PM_{2.5}$ concentration dataset as evident discontinuities have been removed after homogenization. Overall, our results clearly indicate the presence of discontinuities in the raw *in situ* PM_{2.5} concentration observations that were measured in China, and the homogenization actions are essential to the acquisition of a long-term coherent PM_{2.5} concentration dataset that can be used to advance PM_{2.5} pollution related policy making and public health risk assessment.

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512 Author contributions

The study was completed with cooperation between all authors. JG and KB conceived of the idea behind generating homogenous PM_{2.5} dataset across China; KB and KL conducted the data analyses and KB wrote the manuscript; All authors discussed the experimental results and helped reviewing the manuscript.

517 **Competing interests**

518 The authors declare that they have no conflict of interest.

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