

Response to Reviewer #1 comments

The PM_{2.5} data has been widely used for human exposure risk assessment and air quality management. However, as the author said, given the absence of an open access and quality assured in situ PM_{2.5} concentration dataset in China, it is urgent need to open a stable and reliable PM_{2.5} data access method. This paper attempted to generate a long-term coherent in situ PM_{2.5} concentration dataset for scientific community to use in future applications. Methods involving missing value reconstruction, change point detection, and bias adjustment were applied sequentially to deal with data gaps and inhomogeneities in raw PM_{2.5} observations. It is a nice and well-organized paper with a clear focus. In my opinion, there are some minor problems need to be solved before publishing. My biggest concern is whether the data set will continue to be updated. I suggest that the author add a statement in the conclusion, stating the update frequency and download link of the homogenized PM_{2.5} datasets. In the change points detection, how long is the breakpoint interval?

Reply: We are grateful to the anonymous referee for his or her valuable comments on our manuscript. All of these comments and concerns raised by the referee have been explicitly considered and incorporated into this revision. For clarity, we have listed the referee's comments in black plain font, followed by our point-by-point replies in green plain font.

1) "My biggest concern is whether the data set will continue to be updated"

Reply: The homogenized *in situ* PM_{2.5} concentration dataset will be regularly updated for every six-month based on our newly retrieved data records, and the extended dataset is also freely accessible per the user's request. A full dataset will be then published online on PANGAEA once we have one-year's new measurements.

2) "I suggest that the author add a statement in the conclusion, stating the update frequency and download link of the homogenized PM_{2.5} datasets"

27 Reply: Per your suggestion, we will clearly state the updating frequency of the dataset in our revised
28 manuscript.

29

30 3) “In the change points detection, how long is the breakpoint interval?”

31 Reply: The PMT method was hereby applied to detect possible break points in each $PM_{2.5}$
32 concentration time series in reference to the generated reference series. As the default configuration in
33 the RHtests v4 software package, a length scale of 5 was defined as the minimum interval between
34 two possible change points, which means that no change point would be detected from the 5 adjacent
35 observations. More technique details of PMT method can be found in the following reference, which
36 has been also cited in section 3.

37

38 References:

39 Wang, X.L. Accounting for Autocorrelation in Detecting Mean Shifts in Climate Data Series Using
40 the Penalized Maximal t or F Test. *J. Appl. Meteorol. Climatol.* **2008**, *47*, 2423–2444,
41 doi:10.1175/2008JAMC1741.1.

42 **Response to Reviewer #2 comments**

43

44 This paper developed a homogenized daily in situ PM_{2.5} concentration dataset from national air
45 quality monitoring network in China. The topic has important climate implications in evaluating air
46 quality variations at an interannual scale. The paper is well organized and written. The findings of this
47 study are worth of publication in the journal after minor revision as following:

48 **Reply:** We are grateful to the anonymous referee for his or her valuable comments on our manuscript.
49 All of these comments and concerns raised by the referee have been explicitly considered and
50 incorporated into this revision. For clarity, we have listed the referee's comments in black plain font,
51 followed by our point-by-point replies in green plain font.

52

53 1. The reference station is very import for the adjustment. So the regional representativeness for the
54 selected stations should be clarified.

55 **Reply:** Thanks for your insightful comment. Yes, the reference series is vital to the detection and
56 adjustment of possible inhomogeneities in each data series. In this study, we have developed a complex
57 data integration scheme to derive reference series rather than using one data series sampled at an
58 adjacent station. The representativeness of each selected data series was also taken into account which
59 was even used as the first screening criteria ($R>0.8$ and $CV<0.2$). More details related to the
60 construction of reference series can be found in section 3.3.1 in the revised manuscript.

61

62 2. The scales of most maps are missing.

63 **Reply:** Thanks for pointing it out. We have added the scale bar in each map in the revised
64 manuscript.

65

66 3. Why you only chose these three stations for analysis in figure 5?

67 **Reply:** Figure 5 illustrated three typical inhomogeneities that frequently emerged in PM_{2.5} time series,
68 including abrupt changes during a short time period, a long-term chronic drift, and site relocation
69 related drifts. So, the reason to choose these three stations is mainly due to the variation pattern of

70 inhomogeneities detected in these PM_{2.5} time series is informative and thus can be used as a good
71 illustration. We have clarified this in the revised manuscript to ease the readership.

72

73 4. Suggest that regional trend in the Northwest and Northeast China should be added in table 1.

74 **Reply:** Per your suggestion, we have added the regional trend of PM_{2.5} concentration in these two
75 regions in Table 1 in the revised manuscript.

76

77 5. What is your standard on the daily average from hourly data? Similar with China National
78 Environmental Monitoring Center?

79 **Reply:** Actually, there is no data gap in our derived PM_{2.5} dataset since we had filled the missing
80 values in raw PM_{2.5} time series using the gap filling method that we developed recently. In other words,
81 PM_{2.5} daily averages were calculated based on 24-h observations rather than only using available
82 observations within each 24-h. Such a treatment significantly reduced the bias level in PM_{2.5} daily
83 averages given no missing values. More details related to the gap filling method can be found in section
84 3.2. Missing value only presented for days with less than four observations during each 24-h of the
85 day.

86

87 **A homogenized daily *in situ* PM_{2.5} concentration dataset from national air quality**
88 **monitoring network in China**

89
90 Kaixu Bai^{1,2,3}, Ke Li³, Chengbo Wu³, Ni-Bin Chang⁴, Jianping Guo^{5*}

91
92 ¹Key Laboratory of Geographic Information Science (Ministry of Education), East China Normal University,
93 Shanghai, China

94 ²Institute of Eco-Chongming, 20 Cuiniao Rd., Chongming, Shanghai, China

95 ³School of Geographic Sciences, East China Normal University, Shanghai, China

96 ⁴Department of Civil, Environmental, and Construction Engineering, University of Central Florida, Orlando,
97 FL, USA

98 ⁵State Key Laboratory of Severe Weather, Chinese Academy of Meteorological Sciences, Beijing, China

99
100
101
102
103
104

*Correspondence to: Dr./Prof. Jianping Guo (jpguocams@gmail.com)

105 **Abstract**

106 *In situ* PM_{2.5} concentration observations have long been used as critical data sources in haze related
107 studies. Due to the frequently occurred haze pollution events, China started to [regularly](#) monitor PM_{2.5}
108 concentration nationwide [from the newly established air quality monitoring network since 2013](#).
109 Nevertheless, the acquisition of these invaluable air quality samples is challenging given the absence
110 of public available data download interface. In this study, we provided a homogenized *in situ* PM_{2.5}
111 concentration dataset that was created [on the basis of](#) hourly PM_{2.5} data retrieved from the China
112 National Environmental Monitoring Center (CNEMC) via a web crawler between 2015 and 2019.
113 Methods involving missing value [imputation](#), change point detection, and bias adjustment were applied
114 sequentially to deal with data gaps and inhomogeneities in raw PM_{2.5} observations. After excluding
115 records with limited [samples](#), a homogenized PM_{2.5} concentration dataset comprising of 1,309 five-
116 year long [PM_{2.5} data series at a daily resolution](#) was eventually compiled. This is the first thrust to
117 homogenize *in situ* PM_{2.5} observations in China. The trend estimations derived from the homogenized
118 dataset indicate a spatially homogeneous decreasing tendency of PM_{2.5} across China at a mean rate of
119 about -7.6% per year from 2015 to 2019. In contrast to raw PM_{2.5} observations, the homogenized data
120 record not only has a complete data integrity but is more consistent over space and time. This
121 homogenized daily *in situ* PM_{2.5} concentration dataset is publicly accessible at
122 <https://doi.pangaea.de/10.1594/PANGAEA.917557> (Bai et al., 2020a), which can be applied as a
123 promising dataset for PM_{2.5} related studies such as [satellite-based](#) PM_{2.5} mapping, human exposure
124 risk assessment, and air quality management.

125 **Keywords:** PM_{2.5}; Data homogenization; Bias correction; *In situ* observation; Air quality indicators

Deleted: routinely

Deleted: using

Deleted: reconstruction

Deleted: temporal coverage

Deleted: daily

Deleted: ¶

¶
¶
¶

136 **1 Introduction**

137 A consistent PM_{2.5} concentration dataset is vital to the analysis of variations in PM_{2.5} loadings
138 over space and time as well as in support of its risk analysis for air quality management, meteorological
139 forecasting, and health-related exposure assessment (Lelieveld et al., 2015; Yin et al., 2020). Ground-
140 based monitoring network is commonly built to measure concentrations of air pollutants across the
141 globe. Suffering from extensive and severe haze pollution events in the past few years (Guo et al.,
142 2014; Ding et al., 2016; Wang et al., 2016; Cai et al., 2017; Huang et al., 2018; Luan et al., 2018; Ning
143 et al., 2018), China launched the operational ambient air quality sampling late in 2012 on the basis of
144 the sparsely distributed aerosol observation network. To date, this *in situ* network has been enlarged
145 to cover almost all major cities in China consisting of about 1500 monitoring stations. Concentrations
146 of six key air pollutants including PM_{2.5}, PM₁₀, NO₂, SO₂, CO, and O₃, are routinely measured on an
147 hourly basis while the sampled data are released publicly online by the China National Environmental
148 Monitoring Center (CNEMC) since 2013.

149 Although *in situ* PM_{2.5} concentration data have played critical roles in improving our
150 understanding of regional air quality variations and relevant influential factors (Yang D. et al., 2018;
151 Yang Q. et al., 2019; Zheng et al., 2017), little concern was raised to the quality of such dataset itself
152 (Bai et al., 2019a, 2019c; He and Huang, 2018; Zhang et al., 2019, 2018; Zou et al., 2016). Meanwhile,
153 few studies provided a detailed description of the accuracy or bias level (uncertainty) of the observed
154 PM_{2.5} data in recent years (Xin et al., 2015; You et al., 2016; Guo et al., 2017; Shen et al., 2018). The
155 primary reason lies in the fact that neither quality assurance flag nor metadata information
156 documenting the uncertainty other than data samplings were provided, making such quality assessment
157 infeasible.

158 The data quality, in particular the data homogeneity, is of critical importance to the exploration
159 of the given dataset, especially for trend analysis (Bai et al., 2019c; C. Lin et al., 2018; Liu et al., 2018;
160 Ma et al., 2015) and data integration (Bai et al., 2019b, 2020b; T. Li et al., 2017; Zhang et al., 2019)
161 in which a homogeneous dataset is absolutely essential for downstream applications. Since two distinct
162 kinds of instruments are used in the current air quality monitoring network to measure near surface

Deleted: in due course

164 PM_{2.5} concentration ~~in~~ China (Bai et al., 2020), imperfect instrumental calibration and intermittent
165 replacement of instruments may thus introduce obvious issue of discontinuity in PM_{2.5} observations.
166 Such inhomogeneity may result in large uncertainty and even biased results in the subsequent analysis,
167 especially in context-based and data driven PM_{2.5} concentration mapping (Bai et al., 2019b, 2019a; He
168 and Huang, 2018; Wei et al., 2020), in which *in situ* PM_{2.5} concentration observations are used as the
169 ground truth to characterize complex ~~statistical~~ relationships with other possible contributing factors.

170 Given the absence of an open access and quality assured *in situ* PM_{2.5} concentration dataset in
171 China, in this study, we attempted to generate a long-term coherent *in situ* PM_{2.5} concentration dataset
172 for scientific community to use in future applications. A set of methods involving missing value
173 ~~imputation~~, change point detection, and bias adjustment were geared up seamlessly in a big data
174 analytic manner to ~~ward the~~ improvement of data integrity and ~~the removal of possible~~ discontinuities
175 ~~in~~ raw PM_{2.5} observations. Such an analytical process is also referred to as data homogenization in
176 data science or big data analytics (Cao and Yan, 2012; Wang et al., 2007). To our knowledge, this is
177 the first thrust to homogenize a large-scale dataset of *in situ* PM_{2.5} concentration observations in China.
178 In the following sections, we will introduce the data source as well as detailed big data analytics
179 methods used for the creation of a homogenized PM_{2.5} concentration dataset.

180 2 *In situ* PM_{2.5} concentration observations

181 In this study, the hourly PM_{2.5} concentration data sampled ~~from~~ more than 1,600 state-controlled
182 air quality monitoring stations across China ~~between~~ January 1, 2015 ~~and~~ December 31, 2019 were
183 utilized. ~~These~~ PM_{2.5} concentration data ~~were~~ measured on an hourly basis using ~~either~~ beta-
184 attenuation monitors ~~or~~ Tapered Element Oscillating Microbalance (TEOM) analyzer. The ordinary
185 instrumental calibration and quality control ~~were~~ performed according to the national ambient air
186 quality standard of GB3095-2012 and HJ 618-2011 (Guo et al., 2009, 2017). Generally, TEOM can
187 measure PM_{2.5} concentration within the range of 0–5,000 µg m⁻³ at a resolution of 0.1 µg m⁻³, with
188 precisions of ±0.5 µg m⁻³ for 24-h average and ±1.5 µg m⁻³ for hourly average (Guo et al., 2017; Xin
189 et al., 2012; Xin et al., 2015). ~~The~~ PM_{2.5} measurements ~~were~~ publicly released online by the ~~China~~

Deleted: around

Deleted: reconstruction

Deleted: the

Deleted: to

Deleted: e

Deleted: present

Deleted: at

Deleted: from

Deleted: to

Deleted: a

Deleted: routinely

Deleted: instruments such as

Deleted: and

Deleted: are

Field Code Changed

Deleted: All

Deleted: are

206 [National Environmental Monitoring Center \(CNEMC\)](#) via the National Urban Air Quality Real-time
207 Publishing Platform (<http://106.37.208.233:20035/>) within one hour after the direct sampling.

208 Although the [sampled data were](#) publicly released, the acquisition of these valuable samplings is
209 always [challenging](#) because no data download interface is provided to the public by the CNEMC
210 website. Therefore, it is impossible for users to retrieve the historical observations from the given
211 website. Rather, science community has to count on other measures such as an automatic web crawler
212 for the retrieval of these online updated data samples from the data publishing platform. Nevertheless,
213 the [data records retrieved](#) through such an approach suffered from significant data losses due to various
214 unexpected reasons like power outage and internet interruption. Consequently, the data integrity
215 becomes problematic and further treatments like gap filling are thus essential to accounting for such
216 defects at least.

217 Moreover, hourly PM_{2.5} concentration observations that were sampled at five embassies of United
218 States in China from January 2015 to June 2017 were used as an independent dataset to evaluate the
219 fidelity of the homogenized PM_{2.5} concentration dataset. Geographic locations of these five embassies
220 have been shown in Table S1. These PM_{2.5} data were measured independently under the U.S.
221 department of state air quality monitoring program and can be acquired from the
222 <http://www.stateair.net/>. To be in line with the homogenized dataset, the hourly PM_{2.5} concentration
223 data were aggregated to the daily level by averaging the 24-h observations sampled on each date while
224 daily averages were calculated only for days with more than 12 valid samples of a possible 24-h.

225 3 Homogenization of *in situ* PM_{2.5} concentration data

226 For the creation of a long-term coherent *in situ* PM_{2.5} concentration dataset, it is necessary to
227 create an analytical framework of the big data analytics which seamlessly gears up several methods as
228 a whole for the purposes of [missing value imputation](#), change point detection, and discontinuity
229 adjustment, [given the presence of data gaps and possible discontinuity in raw PM_{2.5} observations](#).
230 Figure 1 shows a schematic illustration of the general workflow toward generating a homogenized
231 [PM_{2.5} concentration](#) dataset and the whole process can be outlined as follows.

Deleted: under the China National Environmental Monitoring Center (CNEMC)

Deleted: se data sampling

Deleted: a

Deleted: exhibits a big

Deleted: c

Deleted: archived

Deleted: gap filling

Formatted: Subscript

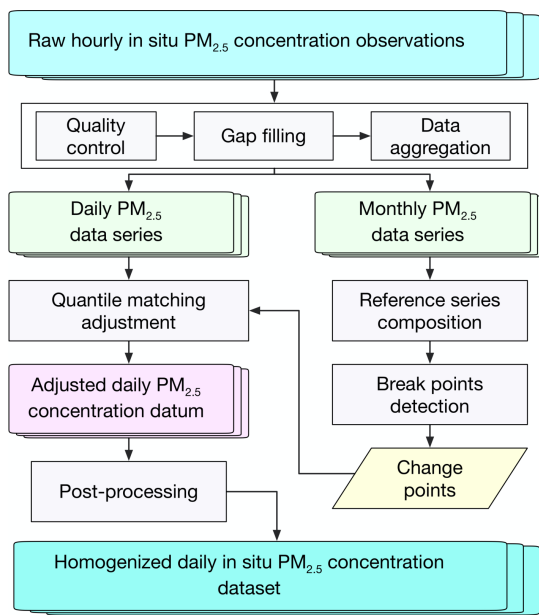
Formatted: Subscript

Deleted: .

Deleted: T

242 (1) It is necessary to perform essential quality control and gap filling on raw PM_{2.5} observations so
 243 that the bias arising from large outliers and resampling errors due to incomplete observations can
 244 be reduced.

Deleted: the
 Deleted: the



245
 246 **Figure 1.** A schematic flowchart for the creation of a homogenized daily *in situ* PM_{2.5} concentration
 247 dataset.

248
 249 (2) Short-term time series due to sites relocation were temporally merged to attain a long-term record.
 250 Then, PM_{2.5} concentration time series with a temporal coverage of less than four-year during the
 251 study period were excluded. Subsequently, the quality-controlled observations of hourly *in situ*
 252 PM_{2.5} concentrations were resampled to daily and monthly scales to initiate the homogeneity test.
 253 (3) Reference time series were constructed for each long-term PM_{2.5} concentration record on the basis
 254 of data measured from adjacent monitoring sites. For PM_{2.5} concentration records failing to
 255 produce a reliable reference series, no homogeneity test was performed for such datum due to the
 256 absence of essential reference data series.

Deleted: and
 Deleted: using
 Deleted: at
 Deleted: in the surroundings

263 (4) The discontinuity identified in each daily long-term PM_{2.5} concentration time series were corrected
264 using the quantile-matching (QM) adjustment method according to the change points detected in
265 each monthly data record with the support of reference series.

Deleted: adjusted

Deleted: detected

Deleted: the

266 (5) Post-processing measures such as nonpositive value correction and another round gap filling were
267 further performed on the homogenized records to attain a quality-assured *in situ* PM_{2.5}
268 concentration dataset. More details of each analytic method were described in the following
269 subsections.

Deleted: to improve the quality

Deleted: are

270 3.1 Quality control

271 Given the possibility of the presence of abnormal samplings, it is necessary to remove the outliers
272 detected in raw PM_{2.5} observations to reduce the false alarm rate in change point detection during the
273 subsequent homogeneity test. Specifically, hourly PM_{2.5} concentration data values meeting one of the
274 following criteria were excluded: 1) out of the range between 1 and 1,000 µg m⁻³, and 2) more than
275 three standard deviations from the median of observations within a 15-h time window. Both criteria
276 aimed to remove large outliers which could result in biased daily averages. Overall, 3.46% of PM_{2.5}
277 samples were treated as outliers and were then excluded accordingly (treated as missing values).

Deleted: essential

Deleted: removing

Deleted: the original

Deleted: which

Deleted: filled with Nan to indicate

278 3.2 Gap filling and resampling

279 As indicated in our recent study (Bai et al., 2020b), missing value related data gaps become a
280 big obstacle in the exploitation of raw PM_{2.5} observations that were retrieved from the CNEMC website
281 as PM_{2.5} observations on 40% of sampling days suffered from data losses due to unexpected reasons.
282 To reduce the impact of missing value related sampling (from hourly to daily) bias on the subsequent
283 homogeneity test, we filled those missing value related data gaps that were found in each 24-h PM_{2.5}
284 observations by applying the DCCEOF method developed very recently (Bai et al., 2020b). Such a
285 gap filling effort enabled us to improve the percentage of days without missingness during the study
286 time period from 58.8% to 97.3%.

Deleted: a

Field Code Changed

Deleted: voids

Deleted: re

287 In spite of the improvement of data integrity after gap filling, the resultant PM_{2.5} time series
288 remain temporally discontinuous due to the emergence of several long-lasting (e.g., more than 24
289 consecutive hours) data missing episodes. Also, the hourly time series are still too noisy to be handled

Deleted: discrete

304 by the current homogeneity test software due to the significant variation in PM_{2.5} concentration over
305 space and time. In such context, the hourly PM_{2.5} concentration records were aggregated to daily and
306 monthly scales to initiate the homogeneity test. Moreover, the monthly series was primarily used to
307 detect the possible change points while the daily series was adjusted in reference to the corresponding
308 reference series based on the change points detected from the monthly series. To avoid large
309 resampling bias, monthly averages were calculated only for those with at least 20 valid daily means of
310 a possible month at each site. The frequency of missing values in each month was also calculated as a
311 possible metadata information to further examine the detected change points.

312 3.3 Homogeneity test

313 A commonly used homogeneity test software, the RHtestsV4 package, was hereby applied to
314 detect the possible discontinuities in raw PM_{2.5} data series that were retrieved from the CNEMC
315 website. As suggested in Wang and Feng (2013), RHtestsV4 is capable of detecting and adjusting
316 change points in a data series with first-order autoregressive errors. Given the low false alarm rate via
317 change point detection and the capability to adjust discontinuity, the RHtests software packages have
318 been widely used to homogenize climate data records such as temperature (Cao et al., 2013; Xu et al.,
319 2013; Zhao et al., 2014), precipitation (Wang et al., 2010a; Nie et al., 2019), and other datum like
320 boundary layer height (Wang and Wang, 2016). Two typical methods, namely the PMTred and
321 PMFred, were embedded in a recursive testing algorithm in RHtestsV4, with the former relying on the
322 penalized maximal *t* test (PMT) while the latter based on the penalized maximal *F* test (PMF) (Wang
323 et al., 2007; Wang, 2008a). With the incorporation of these empirical penalty functions (Wang, 2008a,
324 b), the problem of uneven distribution of false alarm rate is largely alleviated with the aid of RHtestsV4.
325 In contrast to the PMF which works without a reference series, the PMT uses a reference series to
326 detect change points and the results are thus far more reliable (Wang, 2008a, b). The way to generate
327 reference series will be described in the next subsection. Also, the RHtestsV4 is capable of making
328 essential adjustments to the detected discontinuities by taking advantage of the QM adjustment method
329 (Wang and Feng, 2013).

Deleted: available

Deleted: variability

Deleted: of

Deleted: resampled

334 Here the PMT method rather than the PMF was used to detect change points given the higher
335 confidence of the former method in change point detection due to the involvement of reference series
336 (Wang and Feng, 2013). To ensure the reliability of detected discontinuities, change point was defined
337 and confirmed at a nominal 99% confidence level, and the data records were then declared to be
338 homogeneous once no change point was identified. Subsequently, the QM adjustment method was
339 applied to correct PM_{2.5} observations with evident drifts with the support of reference series, namely,
340 to homogenize PM_{2.5} concentration data series. To avoid large sampling uncertainty in the estimate of
341 QM adjustments, the Mq (i.e., the number of categories on which the empirical cumulative distribution
342 function is estimated) was automatically determined by the software to ensure adequate samples for
343 the estimation of mean difference and probability density function. Meanwhile, the number to
344 determine the base segment (i.e., $ladj$) was set to zero so that datum in other segments were all adjusted
345 to the segment with the longest temporal coverage.

Deleted: 0

346 3.3.1 Construction of reference series

347 A good reference series is vital to the relative homogeneity test because it helps pinpoint possible
348 discontinuities in each base series (the data series to be tested) and determines the performance of the
349 subsequent data adjustment. In general, reference series can be organized by using one specific record
350 either measured from one adjacent station or aggregated from multiple observations (Cao and Yan,
351 2012; Peterson and Easterling, 1994; Xu et al., 2013; Wang et al., 2016). The most straightforward
352 way is to use the neighboring data series either measured at the nearest station or series that are highly
353 correlated with the base series (Peterson and Easterling, 1994; Cao and Yan, 2012; Wang and Feng,
354 2013). Such methods, however, fail to take the representativeness of the neighboring series into
355 account since the neighboring series may also suffer from discontinuities.

Deleted: as well as

Deleted: at

Deleted: the

Deleted: adjacent

Deleted: adjacent

Deleted: method

Deleted: repetitiveness

356 To avoid the misuse of inhomogeneous PM_{2.5} concentration records as reference series, a
357 complex yet robust data integration scheme was hereby developed to screen, organize, and construct
358 reference series for each *in situ* PM_{2.5} concentration data series. For each daily PM_{2.5} concentration
359 data series, all the neighboring series were firstly identified from its surroundings with a lag distance
360 as large as of 50 km. No reference series was constructed once there was no neighboring series
361 available within the given radius and in turn the homogeneity of the given record was not examined.

Deleted: in constructing

371 Otherwise, both correlation coefficient (R) and coefficient of variation (CV) were calculated between
372 the given base series and each selected neighboring series to assess their representativeness (Shi et al.,
373 2018; Rodriguez et al., 2019). Then, neighboring series with R greater than 0.8 and CV smaller than
374 0.2 were selected as candidates to construct the reference series for a given base series.

375 The reference series was then constructed by averaging both the base and the candidate series at
376 each observation time if there was only one candidate series. For the situation with more than one
377 candidate series, the empirical orthogonal function (EOF) method was applied to these multiple
378 candidates and then the original fields were reconstructed with the leading principal components when
379 the accumulated variance explained by them exceeded 80%. This was expected to reduce the possible
380 impacts of abnormal observations and short-term discontinuities in the neighboring candidates on the
381 resultant reference series. Subsequently, the reference series were organized and constructed through
382 a spatial weighting scheme as each reconstructed record was assigned a spatially resolved weight
383 according to their relative distances to the base series over space. Here we applied a Gaussian kernel
384 function to estimate the weight of each neighboring observation that can influenced the base series in
385 space and such a scheme has been proven to be effective in assessing the spatial autocorrelation of
386 PM_{2.5} concentration (Bai et al., 2019b). Mathematically, the reference series can be constructed from
387 the following equations:

$$388 \quad PM_{ref} = \sum_{i=1}^N \frac{w_i * PM_{cand}^i}{\sum w_i} \quad (1)$$

$$389 \quad w = \exp\left(\frac{-d^2}{2h^2}\right) \quad (2)$$

390 where PM_{ref} and PM_{cand} denote the reference and candidate series, respectively. N is the total
391 number of candidate series while w is the spatially resolved weight assigned to each candidate series
392 and d is the spatial lag distance between the base and the corresponding candidate series. h is a spatial
393 correlation length that is used to modulate the relative influence of a distant observation on the data
394 measured at the base site. In this study, an empirical value of 50 km was used according to the estimated
395 semi-variogram results (Bai et al., 2019b).

396 For any record having neighboring series within 50 km but poorly correlated ($R < 0.8$ or $CV > 0.2$)
397 to all its neighbors (meaning the base series differ from the neighbors), the reference series were

Deleted: >

Deleted: <

Deleted: analysis

Deleted: one

Deleted: on

Deleted: other

Deleted: assigned

405 created by following the same procedures as those detailed above by taking the nearest neighbor as the
406 base series. For the situation with only one candidate series available, it is logical to compare both the
407 base and the candidate series against another data to check which one should be corrected. [In this study,](#)
408 [the PM_{2.5} time series estimated from the MERRA-2 aerosol reanalysis in the same way as described](#)
409 [in He et al. \(2019\) was used. The one \[with higher correlation\]\(#\) to this external PM_{2.5} time series was](#)
410 then used as the reference (deemed as homogeneous) while the other was considered as the base series
411 (i.e., implies to be adjusted). Such an inclusive scheme empowered us to screen and construct reference
412 series for 1,262 long-term PM_{2.5} concentration records across the board. In contrast, no reference series
413 were constructed for 47 isolated records.

414 3.3.2 Post-processing measures

415 Several post-processing measures were applied to the adjusted data records to further improve
416 the quality of this dataset. Since nonpositive values may appear in the QM adjusted data series if the
417 original values are close to zero (Wang et al., 2010b), nonpositive values were replaced with the
418 smallest valid PM_{2.5} concentration amount measured at each monitoring site during the study period.
419 Subsequently, the data gaps in the adjusted datum due to long-lasting missingness were filled by first
420 calibrating the corresponding data values in the reference series measured on the same date (if available)
421 to the homogenized datum level. The modified quantile-quantile adjustment (MQQA) method
422 proposed in Bai et al. (2016) was hereby used given its adaptive data adjustment principle. For the
423 predicted values, such MQQA scheme rendered higher accuracy than those interpolated from data
424 values measured on adjacent dates because PM_{2.5} concentration is spatially more correlated than in the
425 temporal domain (Bai et al., 2019b). For the remaining data gaps, those missing values were
426 reconstructed in a similar procedure as the DCCEOF method (Bai et al., 2020b). Note that the matrix
427 used for EOF analysis in the context of DCCEOF was constructed using the neighboring data series
428 measured within a radius of 100 km with a temporal lag of 30 days at most. Finally, all data values
429 were rounded to integer to be in line with the original PM_{2.5} concentration observations.

430 4 Results and discussion

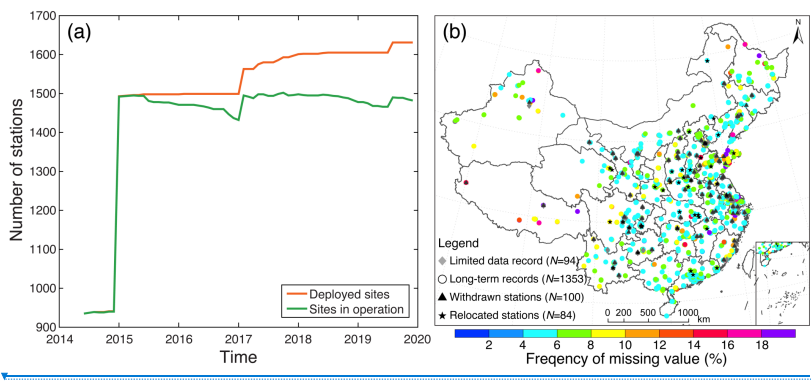
431 4.1 Descriptive statistics

Deleted: It was noted that

Deleted: more correlated

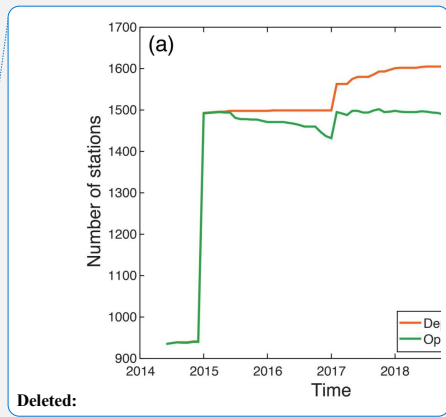
Deleted: ¶

435 Prior to data homogenization, we first need to exclude those short-term and less reliable records.
 436 Figure 2 shows the temporal variations of the number of air quality monitoring stations deployed in
 437 China during 2015–2019 as well as the spatial patterns of the frequency of missing values for each
 438 long-term $PM_{2.5}$ concentration record. It shows that a total of about 1,630 air quality monitoring
 439 stations had been deployed in China before 2020. Nevertheless, about 1,500 sites routinely providing
 440 $PM_{2.5}$ observations were kept up in operation since 2015 (Figure 2a). By referring to the data continuity
 441 of $PM_{2.5}$ observations, it is noticeable that 100 monitoring stations had been withdrawn before 2020
 442 because no $PM_{2.5}$ observations were provided for more than three consecutive months since the release
 443 of their last valid data (Figure 2b). Meanwhile, 42 pairs of stations were found to be relocated since
 444 new stations at nearby started to provide $PM_{2.5}$ observations soon after the suspension of the original
 445 site. This is also corroborated by the temporal lags of $PM_{2.5}$ observations between original and newly
 446 deployed stations as many of them were found to have a time lag less than 15-day. Also, 94 sites were
 447 found with limited data records due to short temporal coverage (newly deployed). Finally, 1,353 long-
 448 term $PM_{2.5}$ concentration records were identified with their first valid data released earlier than 2015.
 449 In regard to the frequency of missing value, it is indicative that data gaps were obvious in these long-
 450 term $PM_{2.5}$ concentration records, with about 6% of hourly data values missed on ~47% of sampling
 451 days on average. This also motivates us to [first](#) fill such data gaps to improve the data integrity.



452
 453 **Figure 2.** Spatial and temporal patterns of air quality monitoring stations in China. (a) Temporal
 454 variations of the total number of air quality monitoring stations in China. (b) Spatial patterns of the

Deleted: even



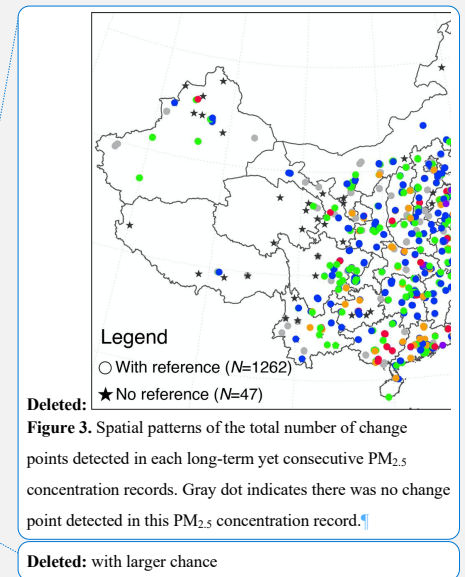
Deleted:

457 frequency of missing value in each long-term hourly PM_{2.5} concentration record measured from
458 January 1, 2015 to December 31, 2019. Stations were categorized into distinct groups according to
459 their data length and temporal continuity. The frequency of missingness was calculated as the ratio of
460 the number of missing values in each PM_{2.5} concentration record to the total number of samplings from
461 the time of the release of the first valid data to December 31, 2019.

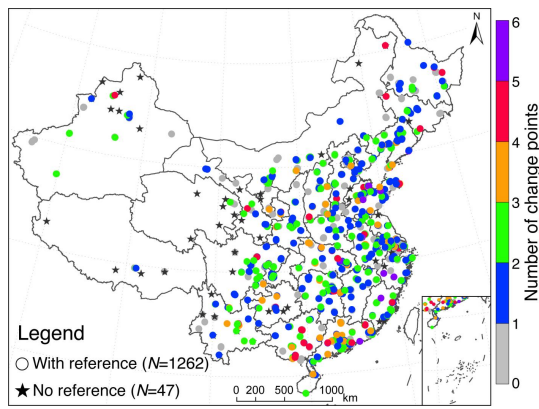
462 4.2 Homogenization of *in situ* PM_{2.5} data

463 A total of 1,395 long-term (with five-year observations) PM_{2.5} concentration records were
464 acquired with the inclusion of 42 temporally merged data series at those relocated stations. After
465 removing those suffering from more than three consecutive months data losses, 1,309 long-term yet
466 consecutive PM_{2.5} concentration records were obtained. The homogeneity test was finally performed
467 on 1,262 records due to the availability of reference series. Figure 3 shows the spatial patterns of the
468 total number of change points detected in 1,262 monthly PM_{2.5} concentration records. The ubiquitous
469 change points imply that there is an obvious inhomogeneity in this *in situ* PM_{2.5} concentration dataset.
470 About 57% (719 out of 1,262) of records failed to pass the homogeneity test due to the presence of
471 change points. Given the overall good agreement between the base and reference series (refer to Figure
472 S1 for the correlation coefficient and root mean square error between them), it indicated that these PM_{2.5}
473 concentration records did suffer from evident discontinuities. Meanwhile, the vast majority (~80%) of
474 the inhomogeneous PM_{2.5} records suffered from no more than two change points (Figure 3), suggesting
475 the mean shift could be the primary reason for the detected discontinuities. Moreover, 20 records were
476 even found suffering from no less than five significant change points, indicating phenomenal
477 discontinuities in these records.

478 Figure 4 shows the temporal variability of the number of change points detected in monthly PM_{2.5}
479 concentration records. As indicated, change points were detected in every specific month of the year
480 from May 2015 to July 2019, especially in late spring (e.g., May), in which change points were more
481 likely to be detected (Figure 4b). This is attributable to the seasonality of PM_{2.5} loading in China as
482 high PM_{2.5} concentrations are always observed in the winter whereas low values in the summer.
483 Consequently, change points were **more likely to be** detected during the chronic transition periods (e.g.,
484 spring to summer). In addition, it is noteworthy that a large volume of change points was detected in

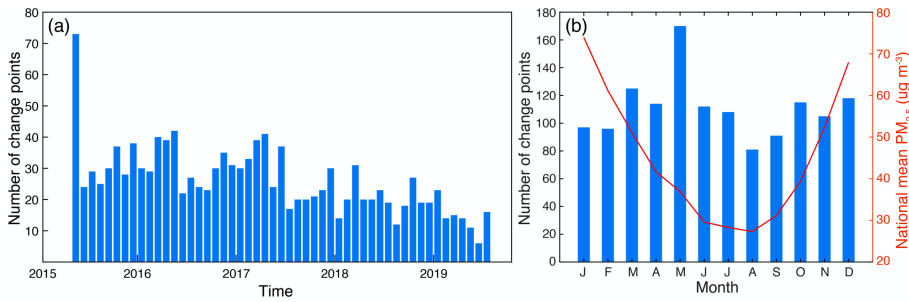


491 early 2015, indicating the existence of phenomenal discontinuities during this period (Figure 4a). After
 492 checking the temporal variations of $PM_{2.5}$ concentration, findings indicate that $PM_{2.5}$ observations
 493 varied with large deviations among each other during this period. This could be linked to the imperfect
 494 instrument calibration or irregular operation in the early stage.



495
 496 **Figure 3.** Spatial patterns of the total number of change points detected in each long-term yet
 497 consecutive $PM_{2.5}$ concentration records. Gray dot indicates there was no change point detected in this
 498 $PM_{2.5}$ concentration record.

Formatted: Indent: First line: 0 cm

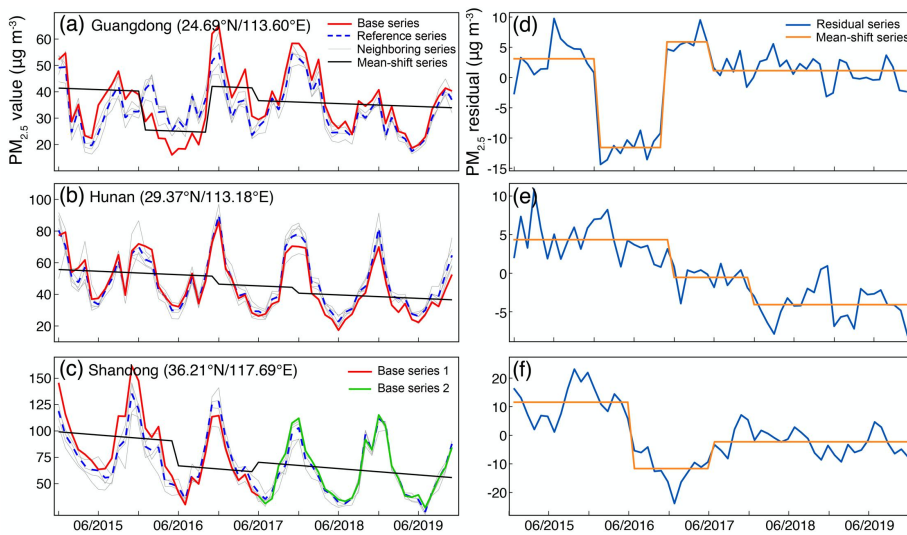


499
 500 **Figure 4.** Temporal variations of the number of change points detected in (a) each specific month from
 501 2015 to 2019 and (b) each month of the year. National mean $PM_{2.5}$ concentration in each month of the
 502 year was calculated based on $PM_{2.5}$ data measured at our selected 1309 sites during 2015–2019.

503 Due to the lack of essential metadata information, it is a challenge for us to verify each detected
 504 change point through a manual inspection. Rather, the variations in the base and reference series was

Deleted: ¶

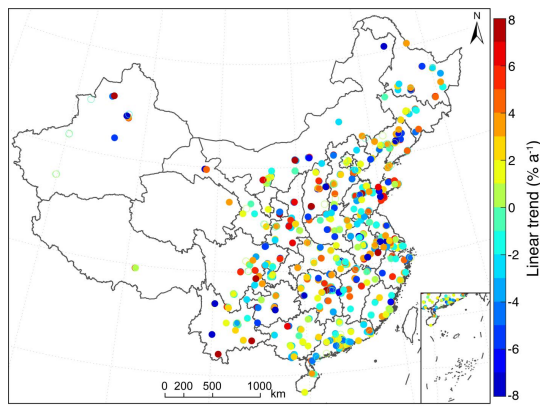
506 explored to identify the possible reasons for the detected discontinuities. Figure 5 presents three typical
 507 inhomogeneous PM_{2.5} time series with different number of change points. The inter-comparisons
 508 between the base and reference series indicate an overall good agreement among them in terms of the
 509 long-term variation tendency. However, obvious drifts were still phenomenal in their residual series,
 510 which were even more evident by referring to their mean-shift series. For example, both the residual
 511 and mean-shift series shown in Figure 5d clearly illustrate a typical discontinuity as there was an
 512 obvious departure of mean PM_{2.5} concentration level during the period of January to October 2016. In
 513 contrast, the Figures. 5b and 5e present another typical inhomogeneity as statistically significant
 514 decreasing trend was found in the residual series with monthly PM_{2.5} concentration deviations
 515 decreased from nearly 5 μg m⁻³ to -4 μg m⁻³ step wise. Such inhomogeneity would undoubtedly result
 516 large bias in the trend estimations over that region. The bottom panel (Figures. 5c and 5f) shows the
 517 change points detected in the merged PM_{2.5} time series at a pair of relocated sites. It is noteworthy that
 518 the detected discontinuity should be largely ascribed to the inconsistency emerged in the first data
 519 series rather than due to the site relocation.



520
 521 **Figure 5.** Temporal variations of three typical inhomogeneous PM_{2.5} concentration records during
 522 2015–2019. (Top) Significant deviations during a short time period, (middle) long-term chronic drifts

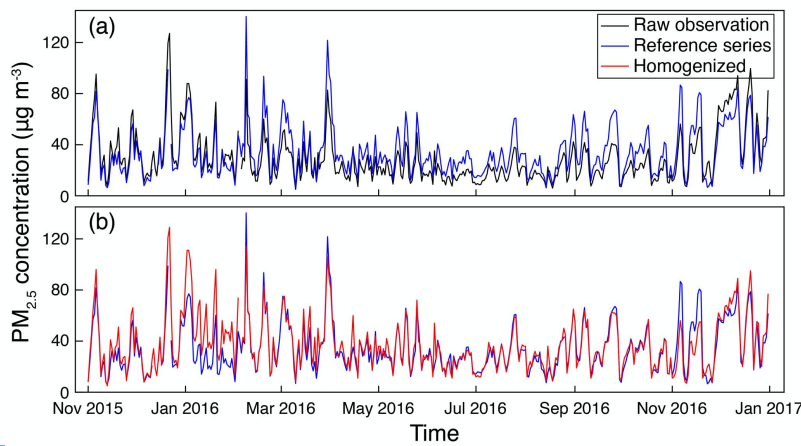
523 with statistically significant varying trend detected in the residual series, (bottom) discontinuity due to
524 site relocation. The left panel compares the base series with the reference and the neighboring series
525 used to compose the reference while the right panel shows the residual series between the base and
526 reference series as well as their mean-shift series.

527
528 Figure 6 shows the estimated linear trends for $PM_{2.5}$ residual series that failed to pass the
529 homogeneity test. Approximately 89% of the residual series were found exhibiting statistically
530 significant linear trends, suggesting the vital importance to homogenize such $PM_{2.5}$ concentration
531 records as the trend estimations at these stations could be prone to large bias if no essential adjustments
532 are performed. Further comparisons of the percentage of data gaps between homogeneous and
533 inhomogeneous records (Figure S2) as well as the spatial distance between the base and the reference
534 series (Figure S3) indicate that both the frequency of data gaps and spatial distance have no obvious
535 impact on the change point detection. In other words, the detected change points have no linkage with
536 neither missing value frequency nor spatial distance between the base and neighboring series,
537 suggesting a high confidence level of the identified discontinuities in these $PM_{2.5}$ concentration records.



538
539 **Figure 6.** Trend estimations for the residual $PM_{2.5}$ concentration data series that failed to pass the
540 homogeneity test during 2015–2019. The solid circles indicate trends are statistically significant at the
541 95% confidence level.

542 Given the emergence of obvious discontinuities in more than half of the selected long-term PM_{2.5}
543 concentration records, the QM adjustment method was applied to correct the discontinuities detected
544 in each PM_{2.5} concentration record. Figure 7 shows an example of homogenization on PM_{2.5}
545 concentration data series that suffered from evident drifts from its reference (large drifts shown in
546 Figure 5d). The inter-comparisons of PM_{2.5} concentration data between the base and reference series
547 indicate that the PM_{2.5} concentration level was obviously underestimated by the raw observations
548 compared with the reference, especially during the middle of 2016 (Figure 7a). Such evident drifts
549 were remarkably diminished after the homogenization (Figure 7b), which shows a good agreement of
550 the mean PM_{2.5} concentration level between the homogenized datum and the reference series.



551
552 **Figure 7.** Comparison of daily mean PM_{2.5} concentration before and after homogenization at one
553 monitoring site in Guangdong province (24.69°N/113.60°E) from November 2015 to December 2016
554 (large drifts shown in Figure 5d).

556 4.3 Validation with independent dataset

557 In this study, PM_{2.5} observations that were collected independently by five consulates of United
558 States distributed in five major Chinese cities between 2015 and 2017 were used to evaluate the
559 consistency of the derived PM_{2.5} concentration records. Figure 8 shows site-specific comparisons of
560 daily PM_{2.5} concentration between homogenized and observed data in Beijing, Shanghai, Chengdu,

Deleted:

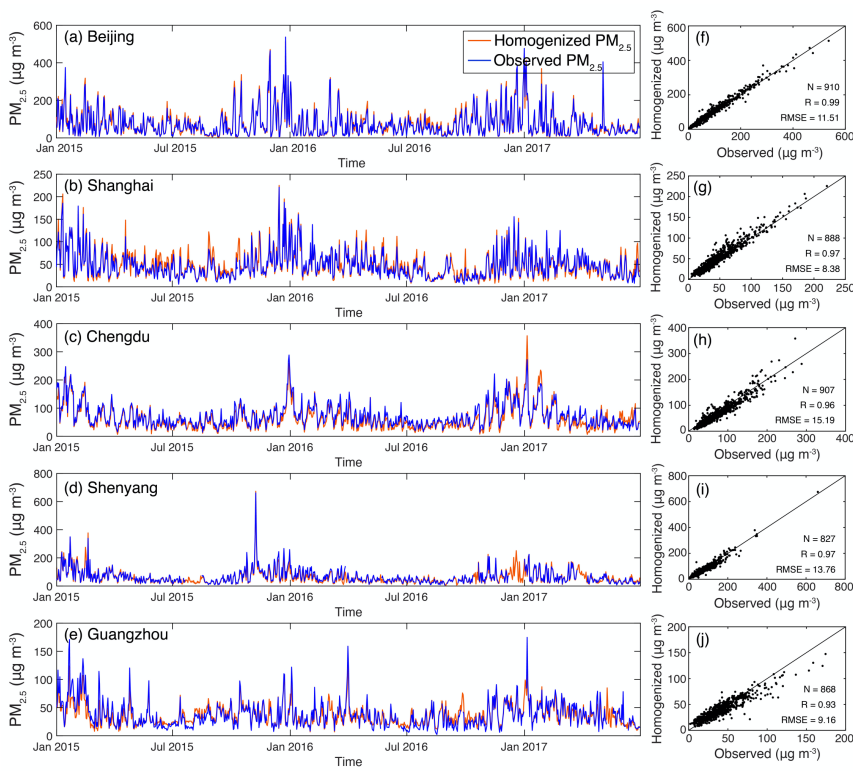
Deleted: ¶

Deleted: ¶

Deleted: ¶

Formatted: Normal, Space Before: 0 pt, Line spacing: single

565 Shenyang, and Guangzhou, respectively. It is indicative that the homogenized daily $PM_{2.5}$
 566 concentration data were in good agreement with $PM_{2.5}$ observations sampled at US consulates, with a
 567 correlation coefficient value of >0.95 and root mean square error of $<15 \mu g m^{-3}$. Given the
 568 independent measurement of $PM_{2.5}$ concentration data at US consulates, we argue that the
 569 homogenized $PM_{2.5}$ records are accurate enough in characterizing the variability of $PM_{2.5}$ loadings in
 570 China. It is also noteworthy that the homogenized $PM_{2.5}$ records are temporally complete whereas
 571 missing values are found in $PM_{2.5}$ observations sampled at US consulates.



572
 573 **Figure 8.** Comparisons of the homogenized $PM_{2.5}$ concentration (red) against $PM_{2.5}$ observations (blue)
 574 measured at five consulates of United States in China from January 2015 to June 2017. (a-e) Temporal
 575 variations of daily $PM_{2.5}$ concentration and (f-j) the associated scatter plots.

Deleted: ~
 Deleted: ~

576

4.4 $PM_{2.5}$ trends estimated from the homogenized dataset

A homogenized data record is essential to trend analysis. Figure 9 presents the annual mean concentration of $PM_{2.5}$ across China between 2015 and 2019. As shown, there is a phenomenal reduction of $PM_{2.5}$ concentration in China in the past five years, especially over North China Plain (the region outlined by a red rectangle shown in Figure 9f) where the annual mean $PM_{2.5}$ concentration decreased from more than $100 \mu g m^{-3}$ in 2015 to about $60 \mu g m^{-3}$ in 2019. Such an evident decrease in $PM_{2.5}$ concentration clearly demonstrates the effectiveness of clean air actions that were implemented in recent years.

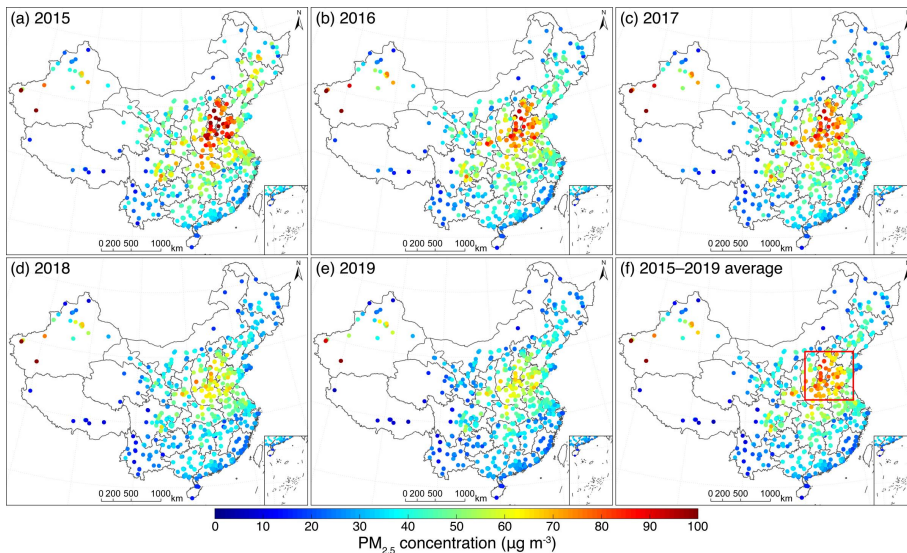


Figure 9. Annual mean $PM_{2.5}$ concentration derived from the homogenized daily $PM_{2.5}$ concentration dataset at 1,309 monitoring stations in China between 2015 and 2019. The North China Plain was outlined by the red rectangle in panel (f).

To evaluate the benefits of data homogenization on $PM_{2.5}$ trend estimations, $PM_{2.5}$ trends estimated from both the raw observations and homogenized dataset were compared. Prior to trend analysis, each $PM_{2.5}$ concentration record was standardized in reference to its mean annual cycle (i.e., $PM_{2.5}$ concentration on the same date of the year between 2015 and 2019 was averaged) to reduce the

Formatted: Subscript

Deleted: T

Deleted: estimations

Deleted: from

Deleted: to

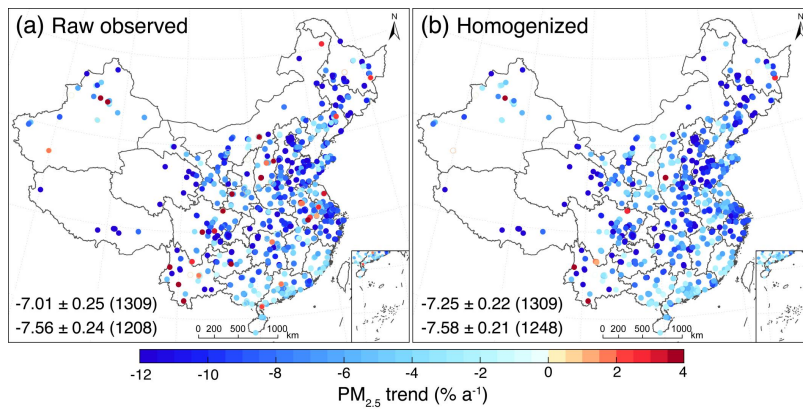
Deleted: in the

Deleted: as

Formatted: Subscript

Formatted: Indent: First line: 0 cm

602 impacts of seasonality and spatial variations. Figure 10 shows a site-specific comparison of PM_{2.5} trend
 603 estimations derived from raw observations and homogenized datasets during 2015–2019. In general,
 604 trend estimations from both datasets showed an evident decreasing tendency of PM_{2.5} concentration
 605 across China during the study period. Nevertheless, noteworthy is that trend estimations derived from
 606 raw PM_{2.5} observations suffered from obvious inhomogeneity over space, being evidenced by
 607 antiphase (positive versus negative) trend estimations even at adjacent stations, especially for those
 608 with positive trends whereas all adjacent neighbors exhibited negative trends. These antiphase trend
 609 estimations over a small region also corroborate the existence of obvious inhomogeneity in raw
 610 observed *in situ* PM_{2.5} concentration dataset.



611
 612 **Figure 10.** Linear trends for (a) raw observed and (b) homogenized daily PM_{2.5} concentration data
 613 during 2015–2019. Solid circles indicate trends are statistically significant at the 95% confidence
 614 interval. Numbers shown in the lower left of each panel indicate the overall trend derived from (top)
 615 all available stations and (bottom) the stations with significant trends at the 95% confidence interval
 616 while the numbers shown in brackets are the corresponding number of data records. Each PM_{2.5} time
 617 series were standardized by its mean annual cycle during the study period to account for spatial
 618 variations of PM_{2.5}.

619
 620 The dotted antiphase trend estimations were substantially diminished after data homogenization,
 621 resulting in a spatially much more homogeneous decreasing tendency of PM_{2.5} concentration across

- Deleted: observed
- Deleted: from 2015 to 2019
- Deleted: However
- Deleted: PM_{2.5}
- Deleted: exhibit
- Deleted: , which is clearly
- Deleted: the
- Deleted:
- Deleted: Such
- Deleted: in
- Deleted: very
- Deleted: demonstrate

- Formatted: Indent: First line: 0 cm
- Deleted: After homogenization,
- Deleted: t
- Deleted: phenomena of
- Deleted: over the local region
- Deleted: as

639 China (Figure 10b). It is indicative that after data homogenization the national mean PM_{2.5} trend was
 640 enlarged from -7.01% a⁻¹ to -7.25% a⁻¹, while the uncertainty was reduced from 0.25% a⁻¹ to 0.22% a⁻¹.
 641 Also, the number of PM_{2.5} records with statistically significant trends was increased from 1,208 to
 642 1,248. These results collectively justify the effectiveness of the QM adjustment method in mitigating
 643 data inhomogeneity in PM_{2.5} observations, which also highlight the critical importance of data
 644 homogenization in accounting for discontinuities in this *in situ* PM_{2.5} concentration dataset. Overall,
 645 our results indicate an obvious decreasing trend of PM_{2.5} concentration in China in the past five years
 646 at a mean rate of $-7.25 \pm 0.22\% \text{ a}^{-1}$. Table 1 further compares the regional mean PM_{2.5} trend between
 647 2015 and 2019. Compared with other regions of interest (ROIs) such as Pearl River Delta (PRD, refer
 648 to Figure S4 for the location) and northern part of Xinjiang (XJ), PM_{2.5} loading over Beijing-Tianjin-
 649 Hebei (BTH), Heilongjiang-Jilin-Liaoning (HJL), and Central China (CC) decreased even more
 650 prominently.

651
 652
 653 Table 1. Regional mean trend for PM_{2.5} concentrations over eight major ROIs in China during 2015–
 654 2019 before and after the data homogenization. Uncertainty in trend estimations were characterized at
 655 the 95% confidence interval. Locations of these ROIs can be found in Figure S4.

ROI	Raw observation (% a ⁻¹)	Homogenized record (% a ⁻¹)
<u>Beijing-Tianjin-Hebei (BTH)</u>	<u>-9.03 ± 0.78</u>	<u>-9.19 ± 0.69</u>
<u>Yangtze River Delta (YRD)</u>	<u>-7.07 ± 0.54</u>	<u>-7.33 ± 0.40</u>
<u>Central China (CC)</u>	<u>-8.47 ± 0.51</u>	<u>-8.58 ± 0.41</u>
<u>Sichuan Basin (SCB)</u>	<u>-7.39 ± 1.02</u>	<u>-7.84 ± 0.89</u>
<u>Pearl River Delta (PRD)</u>	<u>-4.30 ± 0.51</u>	<u>-4.60 ± 0.39</u>
<u>Heilongjiang-Jilin-Liaoning (HJL)</u>	<u>-8.89 ± 0.73</u>	<u>-9.15 ± 0.63</u>
<u>Shaanxi-Gansu-Ningxia (SGN)</u>	<u>-4.85 ± 0.95</u>	<u>-5.30 ± 0.69</u>
<u>North Xinjiang (XJ)</u>	<u>-4.61 ± 1.96</u>	<u>-4.67 ± 1.60</u>

656
 657 To further assess the improvement of the data quality after homogenization, the daily *in situ*
 658 PM_{2.5} concentration records at a 1° × 1° grid cell resolution were grouped across China. In each grid

Deleted: This can be also evidenced by the... it is indicative that after data homogenization enlargement of... the national mean PM_{2.5} decreasing ... trend estimations ... as (increased ... [1])

Deleted:), in particular the decreased variations in trend values... while the (...ncertainty was reduced from 0.25% a⁻¹ to 0.22% a⁻¹)... and ... also, the increased ... number of PM_{2.5} records with statistically significant varying ... trends was increased from (...208 versus ...o 1,248)... These results collectively demonstrate ... justify the effectiveness of the QM adjustment method in mitigating data such ... [2]

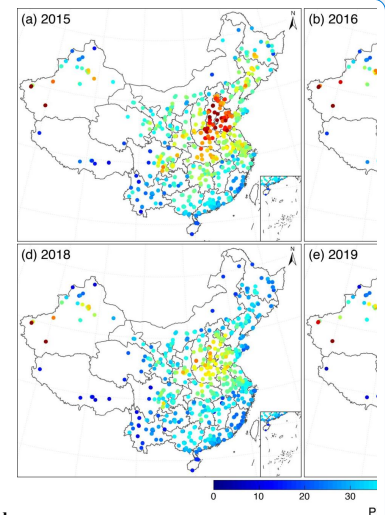
Formatted: Superscript

Formatted: Subscript

Deleted: to ... n accounting for discontinuities in this *in situ* PM_{2.5} concentration dataset. Overall, our results indicate an obvious decreasing trend of PM_{2.5} concentration in China in the past five years at a mean rate of $-7.25 \pm 0.22\%$ per year... [3]

Formatted

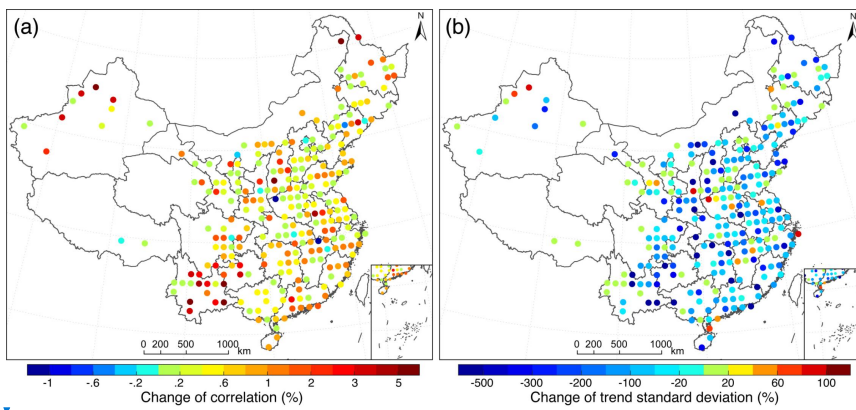
Deleted: Yangtze River Delta (YRD), Sichuan Basin (SCB), ... and Central China (CC) decreased even more prominently (Table 1) ... [5]



Deleted:
Figure 9. Annual mean PM_{2.5} concentration derived from the homogenized daily PM_{2.5} concentration dataset at 1,309 ... [6]

Deleted:

747 cell, the regional mean correlation coefficient among PM_{2.5} concentration time series and standard
 748 deviation of PM_{2.5} trends were estimated from the raw observed and homogenized daily PM_{2.5}
 749 concentration time series, respectively. Their relative differences were then calculated to show the
 750 improvements of data homogeneity within each grid cell. As shown in Figure 11, the correlation among
 751 PM_{2.5} concentration datum was enhanced ubiquitously after homogenization, especially in the
 752 southwest of China (e.g., Yunnan) where obvious inhomogeneity was observed in the raw PM_{2.5}
 753 observations (Figure 10a). Meanwhile, the standard deviation of PM_{2.5} trends within each grid cell was
 754 also substantially reduced, even by more than two folds in the magnitude (Figure 11b). These results
 755 also demonstrate the critical need to homogenize the observed PM_{2.5} concentration data from a large-
 756 scale monitoring network to reduce temporal inconsistency and spatial inhomogeneity that were not
 757 even noticed before.



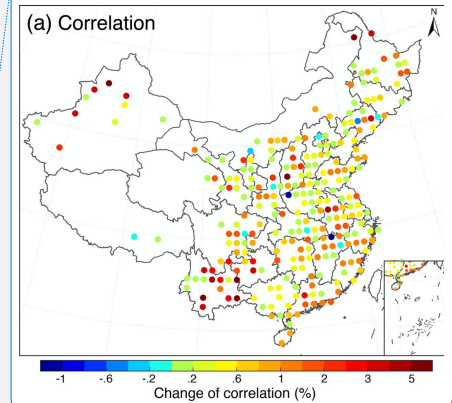
759 **Figure 11.** Spatial distributions of (a) the improvements of mean correlation coefficient among PM_{2.5}
 760 concentration records before and after homogenization at a 1° × 1° grid cell resolution across China,
 761 and (b) their corresponding standard deviations of PM_{2.5} trends.

762 5 Data availability

764 The raw observations of *in situ* PM_{2.5} concentration data in China used in this study were
 765 retrieved via a web crawler from the National Urban Air Quality Real-time Publishing Platform
 766 (<http://106.37.208.233:20035>) between 2014 and 2019. Given the deployment of many new

Deleted:

Deleted: ¶



770 monitoring sites in 2014, we decided to generate a coherent PM_{2.5} concentration dataset starting from
771 2015 to include as many [PM_{2.5} data](#) records as possible. The homogenized daily *in situ* PM_{2.5}
772 concentration dataset developed in this study is publicly accessible at
773 <https://doi.pangaea.de/10.1594/PANGAEA.917557> (Bai et al., 2020a). [To provide a long-term](#)
774 [coherent PM_{2.5} concentration dataset to the scientific community, the homogenized PM_{2.5}](#)
775 [concentration dataset will be regularly updated for each half a year by including new PM_{2.5}](#)
776 [observations that are retrieved during the past six months.](#)

Formatted: Subscript

Formatted: Subscript

Formatted: Subscript

Deleted:

777 6 Conclusions

778 In this study, a homogenized yet temporally complete daily *in situ* PM_{2.5} concentration dataset
779 [was generated based on the discrete hourly PM_{2.5} concentration records that were retrieved from the](#)
780 [China National Urban Air Quality Real-time Publishing Platform using a web crawler during the](#)
781 [period of 2015–2019. To create such a long-term coherent dataset, a set of analytic methods were](#)
782 [geared up seamlessly and applied sequentially to the retrieved raw PM_{2.5} concentration records,](#)
783 [involving quality control, gap filling, data merging, change point detection, and bias correction. This](#)
784 [new dataset would help scientific community better elucidate the temporal and spatial variability of](#)
785 [haze pollution in China in the recent years, which is expected to improve the understanding of](#)
786 [underlying causes.](#)

Deleted: in China

787 The raw PM_{2.5} concentration records were found to be suffering from phenomenal
788 inhomogeneity caused by data [in](#)consistency and temporal [discontinuity](#) as well as the relocation and
789 repeal of a bunch of monitoring stations. [More than half of the long-term PM_{2.5} concentration records](#)
790 [were found failing to pass the homogeneity test due to the presence of substantial change points.](#)
791 Further investigation confirms that large yet short-term mean shifts and chronic drifts are two primary
792 reasons for the detected discontinuities [in raw PM_{2.5} concentration records.](#)

Deleted: coverage

Deleted: It indicated that m

Deleted: ed

Deleted: , given

Deleted: significant

Formatted: Subscript

793 Based on the homogenized dataset, the long-term trends of PM_{2.5} concentration in China were
794 estimated. In contrast to the inhomogeneous trend estimations that were derived from raw PM_{2.5}
795 concentration records, the homogenized dataset yielded a spatially much more homogeneous
796 decreasing tendency of PM_{2.5} [concentration](#) across China at a mean rate of about -7.3% per year. Such

804 an improvement of homogeneity was also evidenced by the enhanced correlation and reduced standard
805 deviation of trend estimations between homogenized PM_{2.5} concentration time series in the
806 surroundings. These results clearly demonstrate the benefits of data homogenization on the
807 improvement of the quality of this PM_{2.5} concentration dataset as evident discontinuities have been
808 removed after homogenization. Overall, our results clearly indicate the presence of discontinuities in
809 the raw in situ PM_{2.5} concentration observations that were measured in China, and the homogenization
810 actions are essential to the acquisition of a long-term coherent PM_{2.5} concentration dataset that can be
811 used to advance PM_{2.5} pollution related policy making and public health risk assessment.

Deleted: work

Deleted: reveals

Deleted: evident

Deleted: records

Deleted: imperative

Deleted: to take in order

Deleted: attain

813 Author contributions

814 The study was completed with cooperation between all authors. JG and KB conceived of the idea
815 behind generating homogenous PM_{2.5} dataset across China; KB and KL conducted the data analyses
816 and KB wrote the manuscript; All authors discussed the experimental results and helped reviewing the
817 manuscript.

818 Competing interests

819 The authors declare that they have no conflict of interest.

Formatted: Justified

Deleted: ¶

820 Acknowledgments

821 This study was financially supported by the National Natural Science Foundation of China (grant
822 41701413), the National Key Research and Development Program of China (grant 2017YFC1501401),
823 and the International Cooperation Platform in Resources, Environment and Ecology, East China
824 Normal University. The authors are grateful to China National Environmental Monitoring Center
825 (<http://www.cnemc.cn/en/>) and the embassy of United States in China (<http://www.stateair.net/>) for
826 releasing the sampled air quality data publicly online. We also want to express our sincere thanks to
827 Dr. Yang Feng in the Expert Team on Climate Change Detection and Indices (ETCCDI)
828 (<http://etccdi.pacificclimate.org/software.shtml>) for providing the RHtestsV4 software package.

Formatted: Justified, Line spacing: 1.5 lines

Deleted: and

Deleted: ¶

839 **References**

- 840 An, Z., Huang, R.J., Zhang, R., Tie, X., Li, G., Cao, J., Zhou, W., Shi, Z., Han, Y., Gu, Z., Ji, Y.,
841 2019. Severe haze in northern China: A synergy of anthropogenic emissions and atmospheric
842 processes. *Proc. Natl. Acad. Sci. U. S. A.* 116, 8657–8666.
843 <https://doi.org/10.1073/pnas.1900125116>
- 844 Anis, A.A., Lloyd, E.H., 1976. The Expected Value of the Adjusted Rescaled Hurst Range of
845 Independent Normal Summands. *Biometrika* 63, 111–116. <https://doi.org/10.2307/2335090>
- 846 Bai, K., Chang, N.-B., Yu, H., Gao, W., 2016. Statistical bias correction for creating coherent total
847 ozone record from OMI and OMPS observations. *Remote Sens. Environ.* 182, 150–168.
848 <https://doi.org/10.1016/j.rse.2016.05.007>
- 849 Bai, K., Chang, N.-B., Zhou, J., Gao, W., Guo, J., 2019a. Diagnosing atmospheric stability effects on
850 the modeling accuracy of PM_{2.5}/AOD relationship in eastern China using radiosonde data.
851 *Environ. Pollut.* 251, 380–389. <https://doi.org/10.1016/j.envpol.2019.04.104>
- 852 Bai, K., Li, K., Chang, N.-B., Gao, W., 2019b. Advancing the prediction accuracy of satellite-based
853 PM_{2.5} concentration mapping: A perspective of data mining through in situ PM_{2.5}
854 measurements. *Environ. Pollut.* 254, 113047. <https://doi.org/10.1016/j.envpol.2019.113047>
- 855 Bai, K., Li, K., Wu, C., Chang, N.-B., Guo, J., 2020a. A homogenized daily *in situ* PM_{2.5}
856 concentration dataset in China during 2015–2019. PANGAEA,
857 <https://doi.pangaea.de/10.1594/PANGAEA.917557>
- 858 Bai, K., Li, K., Guo, J., Yang, Y., Chang, N.-B., 2020b. Filling the gaps of in situ hourly PM_{2.5}
859 concentration data with the aid of empirical orthogonal function analysis constrained by diurnal
860 cycles. *Atmos. Meas. Tech.* 13, 1213–1226. <https://doi.org/10.5194/amt-13-1213-2020>
- 861 Bai, K., Ma, M., Chang, N.-B., Gao, W., 2019c. Spatiotemporal trend analysis for fine particulate
862 matter concentrations in China using high-resolution satellite-derived and ground-measured
863 PM_{2.5} data. *J. Environ. Manage.* 233, 530–542. <https://doi.org/10.1016/j.jenvman.2018.12.071>
- 864 Cai, W., Li, K., Liao, H., Wang, H., Wu, L., 2017. Weather conditions conducive to Beijing severe
865 haze more frequent under climate change. *Nat. Clim. Chang.* 7, 257–262.
866 <https://doi.org/10.1038/nclimate3249>

867 Cao, L.-J., Yan, Z.-W., 2012. Progress in research on homogenization of climate Data. *Adv. Clim.*
868 *Chang. Res.* 3, 59–67. <https://doi.org/10.3724/SP.J.1248.2012.00059>

869 Cao, L., Zhao, P., Yan, Z., Jones, P., Zhu, Y., Yu, Y., Tang, G., 2013. Instrumental temperature
870 series in eastern and central China back to the nineteenth century. *J. Geophys. Res. Atmos.* 118,
871 8197–8207. <https://doi.org/10.1002/jgrd.50615>

872 Ding, A.J., Huang, X., Nie, W., Sun, J.N., Kerminen, V.-M., Petäjä, T., Su, H., Cheng, Y.F., Yang,
873 X.-Q., Wang, M.H., Chi, X.G., Wang, J.P., Virkkula, A., Guo, W.D., Yuan, J., Wang, S.Y.,
874 Zhang, R.J., Wu, Y.F., Song, Y., Zhu, T., Zilitinkevich, S., Kulmala, M., Fu, C.B., 2016.
875 Enhanced haze pollution by black carbon in megacities in China. *Geophys. Res. Lett.* 43, 2873–
876 2879. <https://doi.org/10.1002/2016GL067745>

877 Guo, J.-P., Zhang, X.-Y., Che, H.-Z., Gong, S.-L., An, X., Cao, C.-X., Guang, J., Zhang, H., Wang,
878 Y.-Q., Zhang, X.-C., Xue, M., Li, X.-W., 2009. Correlation between PM concentrations and
879 aerosol optical depth in eastern China. *Atmos. Environ.* 43, 5876–5886.
880 <https://doi.org/10.1016/j.atmosenv.2009.08.026>

881 Guo, J., Xia, F., Zhang, Y., Liu, H., Li, J., Lou, M., He, J., Yan, Y., Wang, F., Min, M., Zhai, P.,
882 2017. Impact of diurnal variability and meteorological factors on the PM_{2.5}-AOD relationship:
883 Implications for PM_{2.5} remote sensing. *Environ. Pollut.* 221, 94–104.
884 <https://doi.org/10.1016/j.envpol.2016.11.043>

885 Guo, S., Hu, M., Zamora, M.L., Peng, J., Shang, D., Zheng, J., Du, Z., Wu, Z., Shao, M., Zeng, L.,
886 Molina, M.J., Zhang, R., 2014. Elucidating severe urban haze formation in China. *Proc. Natl.*
887 *Acad. Sci.* 111, 17373–17378. <https://doi.org/10.1073/pnas.1419604111>

888 He, L., Lin, A., Chen, X., Zhou, H., Zhou, Z., He, P., 2019. Assessment of MERRA-2 Surface PM_{2.5}
889 over the Yangtze River Basin: Ground-based verification, spatiotemporal distribution and
890 meteorological dependence. *Remote Sens.* 11. <https://doi.org/10.3390/rs11040460>

891 He, Q., Huang, B., 2018. Satellite-based mapping of daily high-resolution ground PM_{2.5} in China via
892 space-time regression modeling. *Remote Sens. Environ.* 206, 72–83.
893 <https://doi.org/10.1016/j.rse.2017.12.018>

894 Huang, X., Wang, Z., Ding, A., 2018. Impact of aerosol-PBL interaction on haze pollution: multiyear

895 observational evidences in North China. *Geophys. Res. Lett.* 45, 8596–8603.
896 <https://doi.org/10.1029/2018GL079239>

897 Li, Q., Zhang, H., Chen, J.I., Li, W., Liu, X., Jones, P., 2009. A mainland china homogenized
898 historical temperature dataset of 1951-2004. *Bull. Am. Meteorol. Soc.* 90, 1062–1065.
899 <https://doi.org/10.1175/2009BAMS2736.1>

900 Li, T., Shen, H., Yuan, Q., Zhang, X., Zhang, L., 2017. Estimating ground-level PM_{2.5} by fusing
901 satellite and station observations: A Geo-Intelligent deep learning approach. *Geophys. Res.*
902 *Lett.* 44, 11,985-11,993. <https://doi.org/10.1002/2017GL075710>

903 Li, Z., Guo, J., Ding, A., Liao, H., Liu, J., Sun, Y., Wang, T., Xue, H., Zhang, H., Zhu, B., 2017.
904 Aerosol and boundary-layer interactions and impact on air quality. *Natl. Sci. Rev.* 4, 810–833.
905 <https://doi.org/10.1093/nsr/nwx117>

906 Lin, C., Li, Y., Lau, A.K.H., Li, C., Fung, J.C.H., 2018. 15-Year PM_{2.5} trends in the Pearl River
907 Delta region and Hong Kong from satellite observation. *Aerosol Air Qual. Res.* 1–8.
908 <https://doi.org/10.4209/aaqr.2017.11.0437>

909 Lin, C.Q., Liu, G., Lau, A.K.H., Li, Y., Li, C.C., Fung, J.C.H., Lao, X.Q., 2018. High-resolution
910 satellite remote sensing of provincial PM_{2.5} trends in China from 2001 to 2015. *Atmos. Environ.*
911 180, 110–116. <https://doi.org/10.1016/j.atmosenv.2018.02.045>

912 Liu, D., Deng, Q., Zhou, Z., Lin, Y., Tao, J., 2018. Variation trends of fine particulate matter
913 concentration in Wuhan city from 2013 to 2017. *Int. J. Environ. Res. Public Health* 15, 1487.
914 <https://doi.org/10.3390/ijerph15071487>

915 Luan, T., Guo, X., Guo, L., Zhang, T., 2018. Quantifying the relationship between PM_{2.5}
916 concentration, visibility and planetary boundary layer height for long-lasting haze and fog–haze
917 mixed events in Beijing. *Atmos. Chem. Phys.* 18, 203–225. [https://doi.org/10.5194/acp-18-203-](https://doi.org/10.5194/acp-18-203-2018)
918 2018

919 Ma, Z., Hu, X., Sayer, A.M., Levy, R., Zhang, Q., Xue, Y., Tong, S., Bi, J., Huang, L., Liu, Y., 2015.
920 Satellite-based spatiotemporal trends in PM_{2.5} concentrations: China, 2004–2013. *Environ.*
921 *Health Perspect.* 124, 184–192. <https://doi.org/10.1289/ehp.1409481>

922 Nie, H., Qin, T., Yang, H., Chen, J., He, S., Lv, Z., Shen, Z., 2019. Trend analysis of temperature

923 and precipitation extremes during winter wheat growth period in the major winter wheat
924 planting area of China. *Atmosphere*. 10, 240. <https://doi.org/10.3390/atmos10050240>

925 Peterson, T.C., Easterling, D.R., 1994. Creation of homogeneous composite climatological reference
926 series. *Int. J. Climatol.* 14, 671–679. <https://doi.org/10.1002/joc.3370140606>

927 Rodriguez, D., Valari, M., Payan, S., Eymard, L., 2019. On the spatial representativeness of NO_x
928 and PM₁₀ monitoring-sites in Paris, France. *Atmos. Environ.* X 1, 100010.
929 <https://doi.org/10.1016/j.aeaoa.2019.100010>

930 Shen, H., Li, T., Yuan, Q., Zhang, L., 2018. Estimating regional ground-level PM_{2.5} directly from
931 satellite top-of-atmosphere reflectance using deep belief networks. *J. Geophys. Res. Atmos.*
932 2018JD028759. <https://doi.org/10.1029/2018JD028759>

933 Shi, X., Zhao, C., Jiang, J.H., Wang, C., Yang, X., Yung, Y.L., 2018. Spatial representativeness of
934 PM_{2.5} concentrations obtained using observations from network stations. *J. Geophys. Res.*
935 *Atmos.* 123, 3145–3158. <https://doi.org/10.1002/2017JD027913>

936 Wang, G., Zhang, R., Gomez, M.E., Yang, L., Levy Zamora, M., Hu, M., Lin, Y., Peng, J., Guo, S.,
937 Meng, J., Li, J., Cheng, C., Hu, T., Ren, Y., Wang, Yuesi, Gao, J., Cao, J., An, Z., Zhou, W., Li,
938 G., Wang, J., Tian, P., Marrero-Ortiz, W., Secrest, J., Du, Z., Zheng, J., Shang, D., Zeng, L.,
939 Shao, M., Wang, W., Huang, Y., Wang, Yuan, Zhu, Y., Li, Y., Hu, J., Pan, B., Cai, L., Cheng,
940 Y., Ji, Y., Zhang, F., Rosenfeld, D., Liss, P.S., Duce, R.A., Kolb, C.E., Molina, M.J., 2016.
941 Persistent sulfate formation from London Fog to Chinese haze. *Proc. Natl. Acad. Sci.* 113,
942 13630–13635. <https://doi.org/10.1073/pnas.1616540113>

943 Wang, X., Wang, K., 2016. Homogenized variability of radiosonde-derived atmospheric boundary
944 layer height over the global land surface from 1973 to 2014. *J. Clim.* 29, 6893–6908.
945 <https://doi.org/10.1175/JCLI-D-15-0766.1>

946 Wang, X.L., 2008a. Penalized maximal F test for detecting undocumented mean shift without trend
947 change. *J. Atmos. Ocean. Technol.* 25, 368–384. <https://doi.org/10.1175/2007JTECHA982.1>

948 Wang, X.L., 2008b. Accounting for autocorrelation in detecting mean shifts in climate data series
949 using the Penalized Maximal t or F Test. *J. Appl. Meteorol. Climatol.* 47, 2423–2444.
950 <https://doi.org/10.1175/2008JAMC1741.1>

951 Wang, X.L., Chen, H., Wu, Y., Feng, Y., Pu, Q., 2010a. New Techniques for the Detection and
952 Adjustment of Shifts in Daily Precipitation Data Series. *J. Appl. Meteorol. Climatol.* 49, 2416–
953 2436. <https://doi.org/10.1175/2010JAMC2376.1>

954 Wang, X.L., Chen, H., Wu, Y., Feng, Y., Pu, Q., 2010b. New techniques for the detection and
955 adjustment of shifts in daily precipitation data series. *J. Appl. Meteorol. Climatol.* 49, 2416–
956 2436. <https://doi.org/10.1175/2010JAMC2376.1>

957 Wang, X.L., Wen, Q.H., Wu, Y., 2007. Penalized maximal t test for detecting undocumented mean
958 change in climate data series. *J. Appl. Meteorol. Climatol.* 46, 916–931.
959 <https://doi.org/10.1175/JAM2504.1>

960 Wei, J., Li, Z., Cribb, M., Huang, W., Xue, W., Sun, L., Guo, J., Peng, Y., Li, J., Lyapustin, A., Liu,
961 L., Wu, H., Song, Y., 2020. Improved 1 km resolution PM_{2.5} estimates across China using
962 enhanced space – time extremely randomized trees. *Atmos. Chem. Phys.* 20, 3273–3289.

963 Xin, J., Wang, Y., Wang, L., Tang, G., Sun, Y., Pan, Y., & Ji, D., 2012. Reductions of PM_{2.5} in
964 Beijing-Tianjin-Hebei urban agglomerations during the 2008 Olympic Games. *Adv. Atmos.*
965 *Sci.*, 29(6), 1330–1342. <https://doi.org/10.1007/s00376-012-1227-4>

966 [Xin, J., Wang, Y., Pan, Y., Ji, D., Liu, Z., Wen, T., Wang, Y., Li, X., Sun, Y., Sun, J., Wang, P.,](#)
967 [Wang, G., Wang, X., Cong, Z., Song, T., Hu, B., Wang, L., Tang, G., Gao, W., Guo, Y., Miao,](#)
968 [H., Tian, S., & Wang, L., 2015. The campaign on atmospheric aerosol research network of](#)
969 [China: CARE-China. *Bull. Am. Meteorol. Soc.* 96, 1137–1155. \[D-14-00039.1\]\(https://doi.org/10.1175/BAMS-
970 <a href=\)](#)

971 Xu, W., Li, Q., Wang, X.L., Yang, S., Cao, L., Feng, Y., 2013. Homogenization of Chinese daily
972 surface air temperatures and analysis of trends in the extreme temperature indices. *J. Geophys.*
973 *Res. Atmos.* 118, 9708–9720. <https://doi.org/10.1002/jgrd.50791>

974 Yang, D., Wang, X., Xu, J., Xu, C., Lu, D., Ye, C., Wang, Z., Bai, L., 2018. Quantifying the
975 influence of natural and socioeconomic factors and their interactive impact on PM_{2.5} pollution
976 in China. *Environ. Pollut.* 241, 475–483. <https://doi.org/10.1016/j.envpol.2018.05.043>

977 Yang, Q., Yuan, Q., Yue, L., Li, T., Shen, H., Zhang, L., 2019. The relationships between PM_{2.5} and
978 aerosol optical depth (AOD) in mainland China: About and behind the spatio-temporal

979 variations. *Environ. Pollut.* 248, 526–535. <https://doi.org/10.1016/j.envpol.2019.02.071>

980 You, W., Zang, Z., Zhang, L., Li, Y., Wang, W., 2016. Estimating national-scale ground-level PM_{2.5}
981 concentration in China using geographically weighted regression based on MODIS and MISR
982 AOD. *Environ. Sci. Pollut. Res.* 23, 8327–8338. <https://doi.org/10.1007/s11356-015-6027-9>

983 Zhang, D., Bai, K., Zhou, Y., Shi, R., Ren, H., 2019. Estimating ground-level concentrations of
984 multiple air pollutants and their health impacts in the Huaihe River Basin in China. *Int. J.*
985 *Environ. Res. Public Health* 16, 579. <https://doi.org/10.3390/ijerph16040579>

986 Zhang, T., Zhu, Zhongmin, Gong, W., Zhu, Zerun, Sun, K., Wang, L., Huang, Y., Mao, F., Shen, H.,
987 Li, Z., Xu, K., 2018. Estimation of ultrahigh resolution PM_{2.5} concentrations in urban areas
988 using 160 m Gaofen-1 AOD retrievals. *Remote Sens. Environ.* 216, 91–104.
989 <https://doi.org/10.1016/j.rse.2018.06.030>

990 Zhao, P., Jones, P., Cao, L., Yan, Z., Zha, S., Zhu, Y., Yu, Y., Tang, G., 2014. Trend of surface air
991 temperature in Eastern China and associated large-scale climate variability over the last 100
992 years. *J. Clim.* 27, 4693–4703. <https://doi.org/10.1175/JCLI-D-13-00397.1>

993 Zheng, C., Zhao, C., Zhu, Y., Wang, Y., Shi, X., Wu, X., Chen, T., Wu, F., Qiu, Y., 2017. Analysis
994 of influential factors for the relationship between PM_{2.5} and AOD in Beijing. *Atmos. Chem.*
995 *Phys.* 17, 13473–13489. <https://doi.org/10.5194/acp-17-13473-2017>

996 Zou, B., Pu, Q., Bilal, M., Weng, Q., Zhai, L., Nichol, J.E., 2016. High-resolution satellite mapping
997 of fine particulates based on geographically weighted regression. *IEEE Geosci. Remote Sens.*
998 *Lett.* 13, 495–499. <https://doi.org/10.1109/LGRS.2016.2520480>

999

Page 25: [1] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [1] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [1] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [1] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [1] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**



Page 25: [2] Deleted **Kaixu Bai** **10/12/20 10:30:00 AM**

▼

Page 25: [2] Deleted	Kaixu Bai	10/12/20 10:30:00 AM
----------------------	-----------	----------------------

▼

Page 25: [3] Deleted	Kaixu Bai	10/12/20 10:38:00 AM
----------------------	-----------	----------------------

▼

Page 25: [3] Deleted	Kaixu Bai	10/12/20 10:38:00 AM
----------------------	-----------	----------------------

▼

Page 25: [4] Formatted	Kaixu Bai	10/12/20 10:39:00 AM
------------------------	-----------	----------------------

Superscript

▼

Page 25: [4] Formatted	Kaixu Bai	10/12/20 10:39:00 AM
------------------------	-----------	----------------------

Superscript

▼

Page 25: [5] Deleted	Kaixu Bai	10/12/20 10:43:00 AM
----------------------	-----------	----------------------

▼

Page 25: [5] Deleted	Kaixu Bai	10/12/20 10:43:00 AM
----------------------	-----------	----------------------

▼

Page 25: [6] Deleted	Kaixu Bai	10/12/20 9:33:00 AM
----------------------	-----------	---------------------