Global variability of belowground autotrophic respiration in terrestrial ecosystems

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15 Abstract

Belowground autotrophic respiration (RA) is one of the largest, but highly uncertain carbon flux components in terrestrial ecosystems. However, RA has not been explored globally before and still acted as a "black box" in global carbon cycling currently. Such progress and uncertainty motivate a development of global RA dataset and understand its spatial and temporal patterns, causes and responses to future climate change. We applied Random Forest (RF) algorithm to upscale an updated dataset from Global Soil Respiration Database (v4) – covering all major ecosystem types and climate zones with 449 field observations, using globally gridded temperature, precipitation, soil and other environmental variables. We used a 10-fold cross-validation to evaluate the performance of RF to predict the spatial and temporal pattern of RA. Finally, a globally gridded RA dataset from 1980 to 2012 was produced with a spatial resolution of 0.5° × 0.5° (longitude × latitude) and a temporal resolution of one year, expressed by g C m⁻² yr⁻¹ (gram carbon for per square meter per year).

Globally, mean RA was 43.8±0.4 Pg C yr⁻¹ with a temporally increasing trend of 0.025±0.006 Pg C yr⁻² from 1980 to 2012. Such increment trend was widely spread with 58% global land areas. For each 1 °C increase in annual mean temperature, global RA increased by 0.85±0.13 Pg C yr⁻², and it was 0.17±0.03 Pg C yr⁻² for 10 mm increase in annual mean precipitation, indicating a positive feedback of RA to future climate change. Precipitation was the main dominant climatic driver controlling RA, accounting for 56% of global land areas with wide spread globally, particularly in dry or semi-arid areas, followed by shortwave radiation (25%) and temperature (19%). Different temporal patterns for varying climate zones and biomes indicated uneven responses of RA to future climate change, challenging the perspective that the parameters of global carbon stimulation independent on climate zones and biomes. The developed RA dataset, the missing carbon flux component that is not constrained and validated in terrestrial ecosystem models and earth system models, will provide insights into understanding mechanisms underlying the spatial and temporal variability of belowground vegetation carbon dynamics. The developed RA dataset also has great potentials to serve as a benchmark for future data-model comparisons. The developed RA dataset in a common netCDF format is freely available at https://doi.org/10.6084/m9.figshare.7636193 (Tang et al., 2019).

40 1 Introduction

Belowground autotrophic respiration (RA) mainly originated from plant roots, mycorrhizae, and other micro-organisms in the rhizosphere directly relying on labile carbon component leaked from roots (Hanson et al., 2000; Tang et al., 2016; Wang and Yang, 2007). Thus, RA reflects the photosynthesis derived carbon respired back to the atmosphere by roots and regulates the net photosynthetic production allocation to belowground tissues (Högberg et al., 2002). RA is one main component of soil respiration (Hanson et al., 2000), and soil respiration represents the second largest source of carbon fluxes from soil to the atmosphere (after gross primary production, GPP) in the global carbon cycle (Raich and Schlesinger, 1992). Globally, RA could amount roughly up to 54 Pg C yr⁻¹ (1 Pg = 10¹⁵ g, calculating RA as an approximate ratio of 0.5 of soil respiration, more details in Hanson et al., 2000) according to different estimates of global soil respiration (Bond-Lamberty, 2018), which is almost 5 times of the carbon release from human activities (Le Quéré et al., 2018). However, the contribution of RA to soil respiration varied greatly from 10% to 90% across biomes, climate zones and among years (Hanson et al., 2000), leading to the strong spatial and temporal variability in RA. Thus, whether RA varies with ecosystem types or climate zones remains an open question at the global scale (Ballantyne et al., 2017). Consequently, an accurate estimate of RA and its spatial-temporal dynamics are critical to understand the response of terrestrial ecosystems to climate change.

Due to the difficulties in separation and measurement of RA at varying spatial scales and its diurnal, seasonal and annual variabilities, RA becomes one of the largest but highly uncertain carbon flux components in terrestrial ecosystems. Although individual site measurements of RA have been widely conducted across ecosystem types and biomes, the globally spatial and temporal patterns of RA have not been explored and still act as a "black box" in global carbon cycling (Ballantyne et al., 2017). This "black box" is not well constrained and validated, because most terrestrial ecosystem models and earth system models were commonly calibrated and validated against eddy covariance measurements of net ecosystem carbon exchange (Yang et al., 2013). Such progress and uncertainty motivate a development of global RA dataset from observations and understand its spatial and temporal patterns, causes and responses to future climate change. Despite of the general agreement that global soil respiration increased during last several decades (Bond-Lamberty et al., 2018; Bond-Lamberty and Thomson, 2010b; Zhao et al., 2017), how global RA responding to climate change is far from certain because of different temperature sensitivities of RA across terrestrial ecosystems (Liu et al., 2016; Wang et al., 2014). Therefore, reducing RA uncertainty and clarifying its response to climate change, particularly to temperature and precipitation, is essential for global carbon allocation and future projection of the impact of climate change on global terrestrial carbon cycling.

Although several studies have globally estimated soil respiration and its response to climate variables (Bond-Lamberty and Thomson, 2010b; Hursh et al., 2017; Zhao et al., 2017), such efforts have not been conducted for global RA directly. Hashimoto et al. (2015) indirectly derived RA via the difference between total soil respiration and heterotrophic respiration, however, it might lead to uncertainties due to the inclusion of the temperature and precipitation as the only model drivers and a low model

efficiency (32%). Besides temperature and precipitation, other variables, e.g. soil water, carbon and nitrogen content, are additionally critical factors regulating RA, and those factors generally varied with biomes and climate zones. Consequently, Hashimoto et al. (2015) may not reflect the key processes affecting RA, such as soil nutrient constrains.

On the other hand, the climate-derived models usually explain < 50% variability of soil respiration (Bond-Lamberty and Thomson, 2010b; Hashimoto et al., 2015; Hursh et al., 2017), which might be another uncertainty source. Recent studies have included more variables and field observations to improve the prediction ability of linear and non-linear models (Jian et al., 2018b; Zhao et al., 2017), however, it may propagate error because of the overfitting and autocorrelation among these variables (Long and Scott, 2006). Random Forest (RF, Breiman, 2001), a machine learning approach, could overcome these issues based on the hierarchical structure, and be insensitive to outliers and noise compared to single classifiers (Breiman, 2001; Tian et al., 2017). RF uses a large number of ensemble regression trees but a random selection of predictive variables (Breiman, 2001). RF only requires two free parameter settings: the number of variables sampled as candidates for each split and the number of trees. The performance of the RF model usually is not sensitive to the number of trees and number of variables. Moreover, RF regression can deal with a large number of features, which could help feature selection based on the variable importance and can avoid overfitting (Jian et al., 2018b). Consequently, it has been widely used for carbon fluxes modelling in recent years (Bodesheim et al., 2018; Jung et al., 2017).

Therefore, we applied RF algorithm to retrieve global RA based on the updated RA field observations from the most updated Global Soil Respiration Database (SRDB v4, Bond-Lamberty and Thomson, 2018) with the linkage of other global variables (see "materials and methods" part) for the first time, aiming to: (1) develop a globally gridded RA dataset using field observations (named RF-RA); (2) estimate the spatial and temporal patterns of RA at the global scale; (3) identify the dominant driving factors of the spatial and temporal variabilities of RA; (4) compare RF-RA dataset with the previous RA estimates from Hashimoto et al. (2015). The developed RF-RA dataset will advance our understanding of global RA and its spatial and temporal variabilities. The RF-RA is expected to serve as a benchmark for global vegetation models and future data-model comparison, which further advance our knowledge of the co-variation of RA with climate, soil and vegetation factors, linking the empirical observations temporally and spatially to bridge the knowledge gap among local, regional and global scales.

2 Material and methods

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2.1 Development of RA observational dataset

First, RA observational dataset was developed based on SRDB (v4) across the globe, which is publicly available at https://daac.ornl.gov/cgi-bin/dsviewer.pl?ds_id=1578 (Bond-Lamberty and Thomson, 2018). Then, we further updated the dataset using observations collected from Chinese peer-review literatures from China Knowledge Resource Integrated Database (www.cnki.net) up to March 2018, which followed the identical criteria applied in SRDB development. To control

the data quality, annual RA observations were filtered that: (1) annual RA was directly reported in publications indicated by "years of data" of SRDB; (2) the start and end years were recorded in literatures or expanded from "years of data" of SRDB; (3) soil respiration measurements with Alkali absorption and soda lime were not included due to the potential underestimate of respiration rate with the increasing pressure inside chamber (Pumpanen et al., 2004); (4) observations with treatments of nitrogen addition, air/soil warming, and rain/litter exclusion were not included, except cropland; (5) potential problems observations labelled by "Q10" (potential problem with data), "Q11" (suspected problem with data), "Q12" (known problem with data), "Q13" (duplicate) and "Q14" (inconsistency) were excluded. Finally, this study included a total of 449 field observations (Fig. 1), including 68 observations from CNKI. RA observations were absolutely dominated by forest ecosystems (379 observations) with globally unevenly distributed, mainly from China, America and Europe. Although there was a lack of RA observations in Australia, Russia, Africa, and South America, our dataset covered all major ecosystem types and climate zones across the globe.

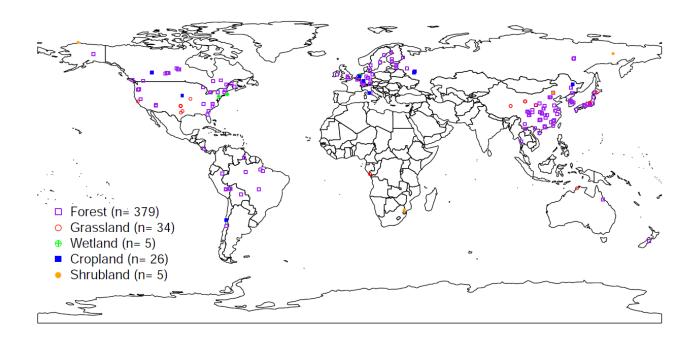


Figure 1 Distribution of observational sites used to develop the globally gridded RF-RA dataset

2.2 Vegetation, climate and soil data

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A total of 11 environmental variables were used to model global RA (Table 1). Specifically, global land cover with a half degree resolution was obtained from MODIS land cover (MCD12Q1 v5, Friedl et al., 2010). Monthly gridded temperature, precipitation, diurnal temperature range, potential evapotranspiration, and self-calibrated Palmer Drought Severity Index (PDSI) at 0.5° resolution were obtained from Climatic Research Unit (CRU) time-series (TS) Version 4.01 from 1901 to 2016 (Harris et al., 2014; van der Schrier et al., 2013). Monthly shortwave radiation (SWR, Kalnay et al., 1996), and soil water content (van den Dool et al., 2003) at 0.5° resolution were from National Oceanic and Atmospheric Administration/Earth System Research Laboratory (NOAA/ESRL) at Physical Sciences Division. Soil organic carbon content with a resolution of

250 m was downloaded from soil grid data (Hengl et al., 2017), and soil nitrogen density was from 'Global Soil Data Task', the Program 'International Geosphere-Biosphere Programme (IGBP)' (Global Soil Data, 2000), while monthly nitrogen deposition data with a resolution of 0.5° were downloaded from the Earth System Models of GISS-E2-R, CCSM-CAM3.5 and GFDL-AM3, covering since 1850s (Lamarque et al., 2013). The monthly global variables were first aggregated to year scale and then resampled to a 0.5° resolution using bilinear interpolation for those variables without a 0.5° resolution. These variables could represent different aspects controlling RA variability. For instance, temperature, precipitation and soil water content are most important variables controlling plant photosynthesis, which is the primary carbon source of RA (Högberg et al., 2002; Högberg et al., 2001). Finally, global variables of each given site extracted by coordinates corresponding with annual RA estimates from the SRDB.

Table 1 Global variables used for producing the global RH dataset

	Variables	Туре	Type of variability	Sources
Climate	Mean annual temperature (°C)	Split	Yearly	
	Mean annual precipitation (mm)	Split	Yearly	https://crudata.uea.ac.uk/cru/data/hrg/cru_ts
	Diurnal temperature range (°C)	Split	Yearly	4.01/(Harris et al., 2014) https://crudata.uea.ac.uk/cru/data/drought/ (van der Schrier et al., 2013)
	Potential evapotranspiration (mm)	Split	Yearly	
	Palmer Drought Severity Index	Split	Yearly	
	Nitrogen deposition (g N m ⁻² yr ⁻¹)	Split	Yearly	https://www.isimip.org/gettingstarted/availab ility-input-data-isimip2b/ (Lamarque et al., 2013)
	Downward Shortwave radiation (W m ⁻²)	Split	Yearly	ftp://ftp.cdc.noaa.gov/Datasets/ncep.reanalys is.derived/surface_gauss/dswrf.sfc.mon.mean .nc (Kalnay et al., 1996)
Soil	Soil carbon content (g kg ⁻¹)	-	Static	https://soilgrids.org/#!/?layer=TAXNWRB_2 50m (Hengl et al., 2017)
	Soil nitrogen density (g m ⁻²)	1	Static	https://webmap.ornl.gov/ogc/dataset.jsp?ds_i d=569 (Global Soil Data, 2000)
	Soil water content (mm)	Split	Yearly	https://www.esrl.noaa.gov/psd/data/gridded/data.cpcsoil.html (van den Dool et al., 2003)
Land cover	MODIS land cover	-	Static	https://glcf.umd.edu/data/lc/ (Friedl et al., 2010)

2.3 Random Forest-based RA Modelling

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A RF model was trained with the 11 variables listed in Table 1 by *caret* by linking *RandomForest* package in R 3.4.4 (Kabacoff, 2015), then the trained model was implemented to estimate grid RA at $0.5^{\circ} \times 0.5^{\circ}$ resolution over 1980-2012. The performance of RF was assessed by a 10-fold cross-validation (CV). A 10-fold CV suggested that the whole dataset was subdivided into 10 parts with approximately an equal number of samples. The target values for each of these 10 parts were predicted on the

training using the remaining nine parts. Two statistics were employed in model assessment: modelling efficiency (R^2) and root mean square error (RMSE) (Yao et al., 2018). The 10-fold CV result showed that RF performed well and could capture the spatial- and temporal-pattern of RA (Fig. S1 in supplementary materials).

2.4 Temporal trend analysis

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We applied Theil-Sen linear regression to estimate temporal trend analysis of RA and its driving variables for each grid cell. The Theil-Sen estimator is a median-based non-parametric slope estimator, which has been widely used for spatial analysis of time series carbon flux analysis (Forkel et al., 2016; Zhang et al., 2017). Mann-Kendall non-parametric test was applied for the significant change trend in RA and its driving factors for each grid cell (p < 0.05).

2.5 Relationships between RA and climate variables

Mean annual temperature, precipitation and shortwave radiation were considered as the most important proxies driving RA. The relationships between RA and temperature, precipitation and shortwave radiation were analyzed by partial correlation for each grid cell. The absolute value of the correlation coefficient of these three variables was used in RGB combination to indicate the dominant factors of RA.

2.6 Cross comparisons with Hashimoto2015-RA

To further compare the differences between RF-RA dataset and RA developed by Hashimoto et al. (2015) (termed as Hashimoto2015-RA), the comparison map profile (CMP) method was applied. Hashimoto developed a climate-driven model by updating Raich's model, which stimulated soil respiration as a function of temperature and water (precipitation) at a monthly time step (Hashimoto et al., 2015; Raich et al., 2002). Therefore, to get a global estimate to soil respiration at a monthly scale, the globally gridded air temperature and precipitation with a spatial resolution of 0.5° were derived from University of East Anglia CRU 3.21 (Harris et al., 2014), and 1638 field observations were taken from SRDB (v3) for model parameterization (Hashimoto et al., 2015). Monthly soil respiration was summed to a yearly scale. Furthermore, annual soil respiration was divided into autotrophic and heterotrophic respiration using a global relationship between soil respiration and heterotrophic respiration derived from a meta-analysis (Bond-Lamberty et al., 2004). This global relationship can be expressed by:

$$\ln(RH) = 1.22 + 1.73 \times \ln(RS) \tag{1}$$

Therefore, global Hashimoto2015-RA was derived by the difference between soil respiration and heterotrophic respiration. The monthly or annual Hashimoto2015-RA dataset can be freely accessed from

Where RH means annual heterotrophic respiration, and RS stands for annual soil respiration, expressed by g C m⁻² yr⁻¹.

(http://cse.ffpri.affrc.go.jp/shojih/data/index.html, Hashimoto et al., 2015).

CMP was developed based on absolute distance (D) and cross-correlation coefficient (CC) through multiple scales (Gaucherel et al., 2008). D and CC reflect the similarity of data values and spatial structure of two images with the same size, respectively (Gaucherel et al., 2008). Low D and higher CC reflects goodness between the compared images, and vice versa. The D among moving windows of two compared images was calculated by equation (2) (Gaucherel et al., 2008):

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$$D = abs(\bar{x} - \bar{y}) \tag{2}$$

 \overline{x} and \overline{y} are averages calculated over two moving windows (3×3 to 41×41 pixels in this study). Finally, the mean D was averaged for different scales.

The CC was calculated by equation (3) (Gaucherel et al., 2008):

$$CC = \frac{1}{N^2} \sum_{i=1}^{N} \sum_{j=1}^{N} \frac{(x_{ij} - \bar{x}) \times (y_{ij} - \bar{y})}{\sigma_x \times \sigma_y}$$
(3)

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$$\sigma_x^2 = \frac{1}{N^2 - 1} \sum_{i=1}^N \sum_{j=1}^N (x_{ij} - \bar{x})$$
 (4)

Where x_{ij} and y_{ij} are the pixel values at row i and column j of two moving windows of the two compared images, respectively. N represents the number of pixels for each moving window, while σ_x and σ_y are the standard deviation calculated from the two moving windows. Finally, like D calculations, CC was calculated as the mean of different scales.

3 Results

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3.1 Spatial patterns of RA

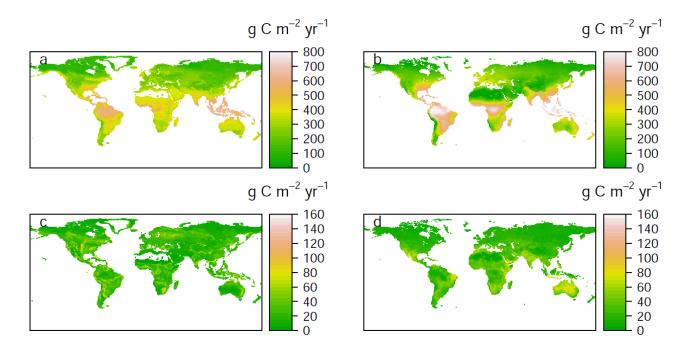


Figure 2 Spatial patterns of annual mean and standard deviation for RF-RA (a, c) and Hashimoto2015-RA (b, d) from 1980 to 2012, respectively. The standard deviation was applied to characterize the inter-annual variability following Yao et al. (2018).

The RF-RA dataset presented a great globally spatial variability during 1980-2012 (Fig. 2a and Fig. 3). Largest RA fluxes commenced from tropical regions, particularly in Amazon tropical and Southeast areas, where generally have a high RA > 700 g C m⁻² yr⁻¹ (gram carbon for per square meter per year). Following the tropical areas, subtropics, e.g. South China, East America, and humid temperate areas, e.g. North America, West and Middle Europe, had typical moderate RA fluxes of 400-600 g C m⁻² yr⁻¹. By contrast, the relative low RA fluxes occurred in the areas with sparse vegetation cover, cold and dry climate, e.g. boreal and tundra, which had low temperature and short growing season. Besides, dry or semi-arid areas, e.g. Northwest China and Middle East, also had typical low RA fluxes below 200 g C m⁻² yr⁻¹, where were often limited by water availability.

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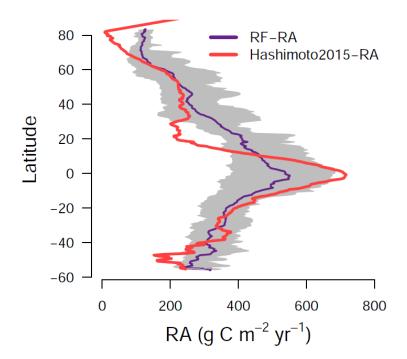


Figure 3 Latitudinal pattern for RF-RA and Hashimoto2015-RA. The grey area means 2.5 to 97.5 percentile ranges of the RF-RA.

The most significant RA inter-annual variability (expressed by standard deviation, Fig. 2c) was found in topical or subtropical regions with above 80 g C m⁻² yr⁻¹, while most areas remained less variable with less than 40 g C m⁻² yr⁻¹. Latitudinally, zonal mean RA increased from cold and dry biomes (tundra and semi-arid) to warm and humid biomes (temperate and tropical forests, Fig. 3), reflecting from more to less environmental limitations. RA varied from 112±21 g C m⁻² yr⁻¹ at about 70°N to 552±101 g C m⁻² yr⁻¹ at equator. Within in 10°S -25°S and 15°N -20°N, due to the limitation of water, zonal mean RA experienced a slight decrease. Therefore, with the increase of water availability, RA led to a second peak in around 20°N and 40°S, respectively.

Compared to RF-RA, Hashimoto2015-RA presented a similarly latitudinal pattern, with highest RA fluxes in tropical regions

characterized by warm and humid climate, followed by subtropical regions, and lowest RA in boreal areas featured by cold and dry climate (Fig. 2b). The most significant change occurred in tropical areas and middle Australia. However, it is worth noted that some clear differences between data-derived and Hashimoto2015-RA existed (Fig. 4): specifically, there was a remarkable difference of above 300 g C m⁻² yr⁻¹ for South Amazon and 200 g C m⁻² yr⁻¹ for subtropical China. Although most areas between RF-RA and Hashimoto2015-RA expressed high and positive correlations, some areas, such as Middle East, West Russia and East America and North Japan, showed negative correlations.

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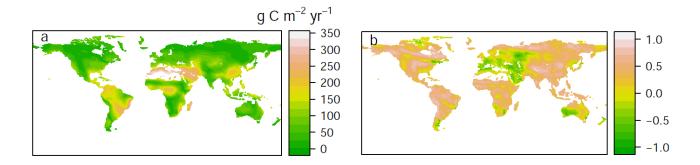


Figure 4 Comparison of RF-RA with Hashimoto2015-RA based on absolute distance (a) and cross-correlation (b)

3.2 | Spatial pattern of RA trend

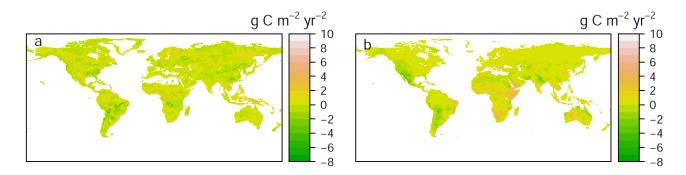


Figure 5 Spatial patterns of the temporal trend for RF-RA and Hashimoto2015-RA during 1980-2012

The trend of RF-RA showed heterogeneous patterns in spatial (Fig. 5). A total of 58% of global areas experienced an increasing trend during 1980-2012 (calculating from cell areas), and 33% of these areas showed a significant change (p < 0.05). Generally, the change trend for the majority areas was from -4 to 4 g C m⁻² yr⁻², while the most striking increasing change occurred in East Russia, tropical, and Eastern regions in Africa with an increasing trend of above 5 g C m⁻² yr⁻². Similarly, 77% of global areas of Hashimoto2015-RA had an increasing trend, 46% of which were statistically significant (p < 0.05).

3.4 | Total RA and its temporal trend

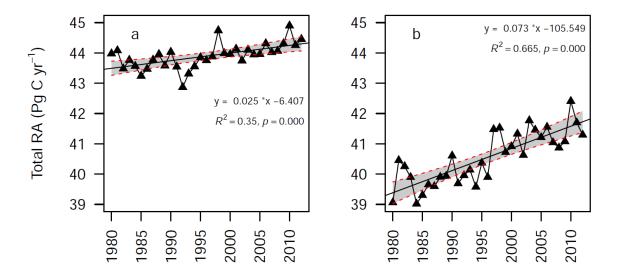


Figure 6 Annual variability of RF-RA (a) and Hashimoto2015-RA (b) from 1980 to 2012. The grey area represents 95% confidence interval.

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Mean global RA was 43.8 ± 0.4 Pg C yr⁻¹ during 1980-2012, varying from 42.9 Pg C yr⁻¹ in 1992 to 44.9 Pg C yr⁻¹ in 2010, with a significant trend of 0.025 ± 0.006 Pg C yr⁻² despite of high annual variabilities (0.06% yr⁻¹, p < 0.001, Fig. 6a). Similarly, a rising trend was also observed for Hashimoto2015-RA (0.073 ± 0.009 Pg C yr⁻², p < 0.001, Fig. 6b), which was higher than that of RF-RA. Annual mean of Hashimoto2015-RA was 40.5 ± 0.9 Pg C yr⁻¹.

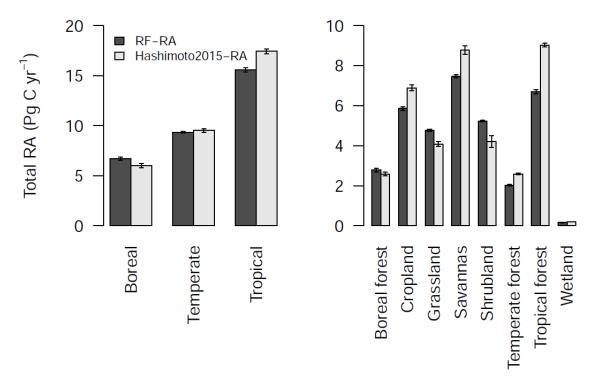


Figure 7 Total amount of RF-RA and Hashimoto2015-RA for three climate zones and eight biomes during 1980-2012. Three climate zones defined as boreal, temperature and tropical regions according to Peel et al. (2007), while eight biomes include boreal forest, cropland, grassland, savannas, shrubland, temperate forest, tropical forest and wetland. The error bars indicated standard deviation.

RA and its trend was also evaluated for three climate zones (boreal, temporal and tropical areas based on Köppen-Geiger climate classification) and eight major biomes (boreal forest, cropland, grassland, savannas, shrubland, temperate forest, tropical forest and wetland, Fig. 7). Tropics had highest RA of 15.6 ± 0.2 Pg C yr⁻¹, followed by temperate regions with 9.3 ± 0.1 Pg C yr⁻¹, and boreal areas represented the lowest RA of 6.7 ± 0.1 Pg C yr⁻¹. These three climate zones were main contributors of global RA, accounting for 72%. Temporally, considerable RA inter-annual variability of these three climate zones existed (Fig. S2). Specifically, RA in tropical and boreal zones showed a significantly increasing trend from 1980 to 2012, with an increasing rate of 0.013 ± 0.003 and 0.008 ± 0.002 Pg C yr⁻², respectively. However, RA in temperate zones presented a slightly decreasing trend of -0.003 ± 0.001 Pg C yr⁻² (p = 0.048) although strong variability was observed.

In terms of biomes, tropical forests had the highest RA, followed by the widely distributed cropland and savannas (Fig. 7), while wetland had the lowest RA due to its limited land cover. RA showed a significantly increasing trend during 1980-2012 ($p_s < 0.01$) in majority biomes, except temperate forests, savannas and wetland. RA in tropical forests, boreal forests and cropland increased by 0.0076 ± 0.0015 , 0.0047 ± 0.0016 , 0.0036 ± 0.0014 Pg C yr⁻², respectively. Compared to RF-RA, Hashimoto2015-RA for the three climate zones and eight biomes generally produced similar change patterns, although the magnitude difference existed (Figs. 7, S2 and S3). However, there were significant increasing trends of total RA in temperate zones, temperate forests, savannas and wetland of Hashimoto2015-RA, which were not observed in RF-RA.

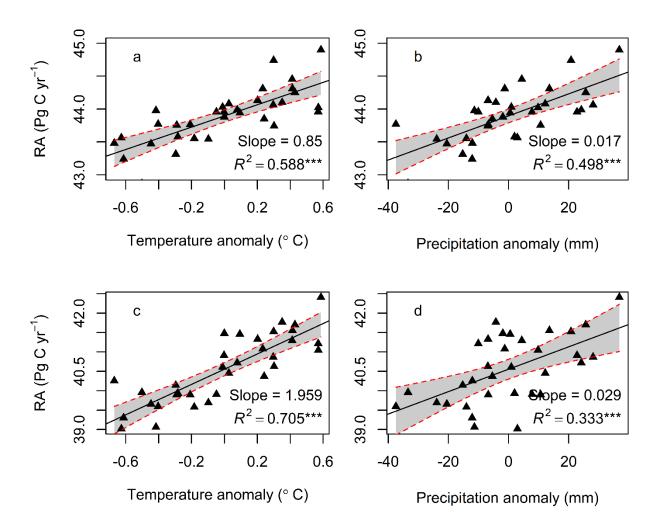


Figure 8 The relationships between RF-RA (a, b)/Hashimoto2015-RA (c, d) and temperature/precipitation anomalies. The anomaly was calculated as the difference between temperature or precipitation of corresponding year to the mean of 1980-2012. *** means significant level at 0.001.

RA was significantly correlated with temperature anomaly ($R^2 = 0.59$, p < 0.001) and precipitation anomaly ($R^2 = 0.50$, p < 0.001, Fig. 8). On average, RA increased by 0.85 ± 0.13 Pg C yr⁻² for 1°C increment in mean annual temperature, and 0.17 ± 0.03 Pg C yr⁻² for 10 mm increase in mean annual precipitation. However, different biomes and climate zones showed uneven responses to the temperature and precipitation changes (Fig. S4 and S5). For example, no significant correlations were found between RA in temperate zone/savannas/wetland and temperature anomaly, while other climate zones and biomes were significantly correlated with temperature/precipitation anomaly.

4. Dominant factors for RA variability

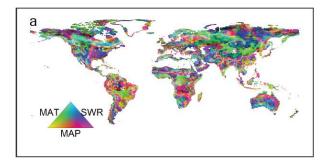
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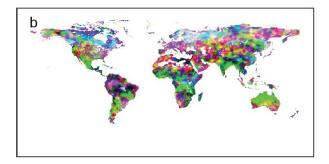


Figure 9 Dominant driving factors for RF-RA (a) and Hashimoto2015-RA (b). MAT = mean annual temperature, MAP = mean annual precipitation; SWR = shortwave radiation.

The dominant environmental factor was examined with partial regression coefficients when regressing RA against annual mean temperature, precipitation and shortwave radiation. Latitudinally, higher mean annual temperature, precipitation and shortwave radiation were associated with higher RA in the major latitudinal gradients (positive partial correlations, Fig. S6). Spatially, the dominant environmental factor varied greatly globally (Fig. 9). Precipitation was the most important dominant factor for the spatial pattern of RA among the three environmental controls, covering about 56% of global land areas (Fig. 10), which was widely distributed globally, particularly in dry or semi-arid areas, such as Northwest China, Southern Africa, Middle Australia and America. Temperature dominated about 19% of global land areas, which mainly occurred in tropical Africa, Southern Amazon rainforests, Siberia and partly tundra. The rest land area (25%) was dominated by shortwave radiation, primarily covering boreal areas above 50°N, Eastern America and middle and Eastern Russian. Similarly, precipitation was also the most important dominant factor for Hashimoto2015-RA, dominating about 77% land areas, while temperature and shortwave radiation dominated 13% and 10% land areas. However, their spatial patterns varied greatly compared to RF-RA. For example, temperature was the main dominant factor for most areas in Australia for Hashimoto2015-RA, while RF-RA indicated that precipitation and shortwave radiation dominated such areas (Fig. 9).

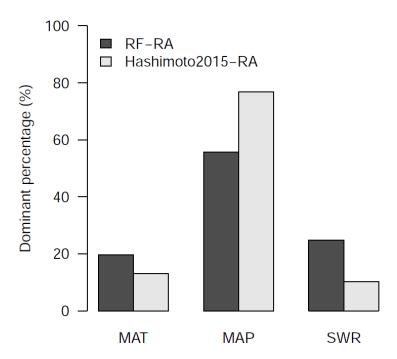


Figure 10 The percentage of land areas (calculated from cell areas) dominated by mean annual temperature (MAT), precipitation (MAP) and shortwave radiation (SWR) for RF-RA and Hashimoto2015-RA.

4 Discussion

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4.1 Global RA

Despite of great efforts to quantify global soil carbon fluxes and their spatial and temporal patterns (Bond-Lamberty and Thomson, 2010b; Hursh et al., 2017; Jian et al., 2018b), to our knowledge, no attempt tried to assess RA using machine learning approach by linking a large number of empirical measurements, and RA's spatial and temporal patterns remain large uncertainties. Such uncertainties justify a development of global RA dataset derived from observations to understand its spatial and temporal patterns, causes and responses to future climate change. Based on the most updated observations from SRDB (Bond-Lamberty and Thomson, 2018) and Chinese peer-review literatures, we for the first time applied RF algorithm to develop RF-RA dataset and estimate the temporal and spatial variability of global RA and its response to environmental variables, which indeed can contribute to reduce RA uncertainties.

Globally, mean annual RA amounted to 43.8±0.4 Pg C yr⁻¹ from 1980 to 2012 (Fig. 6). It was slightly higher than Hashimoto2015-RA (40.5±0.9 Pg C yr⁻¹), and there was great divergence of spatial and temporal patterns (see discussion part in "Comparison with Hashimoto2015-RA"). Due to no direct estimate on global RA, the RF-RA dataset was compared with other RA estimates using total soil respiration multiplied by the proportion of RA or heterotrophic respiration. Bond-Lamberty et al. (2018) proposed that the global average proportion of heterotrophic respiration ranged from 0.54 to 0.63 over 1990-2014 and global soil respiration was 67 to 108 Pg C yr⁻¹ according to different estimates (Bond-Lamberty and Thomson, 2010b; Hashimoto et al., 2015; Hursh et al., 2017; Jian et al., 2018b), thus global RA varied from 25 to 51 Pg C yr⁻¹. The developed

RF-RA dataset fell in this range. Similarly, RA increased by 0.025±0.006 Pg C yr⁻² during 1980-2012. Such increase may be related to the increasing photosynthesis due to global warming and CO₂ fertilization effects, which could increase carbon availability in plant-derived substrate inputs into the soil (e.g. root exudates and biomass) for both root metabolism (Piñeiro et al., 2017; Zhou et al., 2016). Such annual increase accounted for about 25% of global soil respiration increase (0.09 and 0.1 Pg C yr⁻²) (Bond-Lamberty and Thomson, 2010b; Hashimoto et al., 2015), suggesting that about one quarter of the total soil respiration increment due to climate change came from RA.

With 1 °C increase in global mean temperature, RA will increase by 0.85±0.13 Pg C yr⁻² and 0.17±0.03 Pg C yr⁻² for 10 mm increase in precipitation, indicating that carbon fluxes from RA might positively feedback to future climate change, which was typically characterized by increasing temperature and changes in precipitation (IPCC, 2013). However, RA increment varied with climate zones and ecosystem types (Figs. S2 and S3), which was similar to previous findings that total soil respiration or RA varied with climate zones or ecosystem types (Ballantyne et al., 2017; Jian et al., 2018a). These differences may be related to regional heterogeneity and plant functional trait. For example, regional temperature significantly differed from global averages (Huang et al., 2012), with much faster change in high-latitude regions (Hartmann et al., 2014), and semi-arid dominated the trend and variability of global land CO₂ sink (Ahlström et al., 2015). Therefore, the regionally uneven responses of RA to climatic variables highlight the urgent need to account for regional heterogeneity when studying the effects of climate change on ecosystem carbon dynamics in future.

RF-RA also has important indications of carbon allocation from photosynthesis. The immediate carbon substrates for RA were primarily derived from recent photosynthesis (Högberg et al., 2001; Subke et al., 2011). Strong correlation between photosynthesis and RA demonstrated the evidence for their close coupling relationships (Chen et al., 2014; Kuzyakov and Gavrichkova, 2010). Globally, GPP was about 125 Pg C yr⁻¹ during last few decades (Bodesheim et al., 2018; Zhang et al., 2017). Thus, roots respired more than one third carbon from GPP, suggesting that except the carbon used for constructing belowground tissues, a large proportion of carbon will be returned back to atmosphere respired by roots. However, it should be noted that through root respiration, soil nutrients for vegetation growth will be required, which may affect the RA flux.

4.2 Dominant factors

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Spatially, the dominant driving factors for RA varied greatly. Temperature and shortwave radiation were the main driving factors for high latitudinal areas above 50°N (Figure 9a). This result was not surprising because RA was positively correlated with temperature or photosynthesis (indirectly reflecting the solar radiation) (Chen et al., 2014; Tang et al., 2016), and high latitudinal regions was always limited by temperature or energy, leading to low RA as well (Fig. 3a).

Globally, precipitation was the most important factor, covering about 56% of land area (Figs. 9a and 10). Precipitation was always considered as a proxy for soil water content (Hursh et al., 2017; Yao et al., 2018), and such wide dominance of

precipitation on RA was related to the mechanisms of soil water availability driving RA. First, soil water exists in form of ice when temperature is below zero, that plant and soil microbes could not directly use for growth or respiration. This could be observed in some boreal areas where precipitation was the dominant factor of RA (Fig. 9a). Second, too high or too low soil water content (e.g. flooding and drought) could limit the mobility of substrates and carbon input to belowground, which could affect RA. Yan et al. (2014) found that soil respiration decreased once soil water content was below a lower (14.8 %) or above an upper (26.2%) threshold in a poplar plantation. Similarly, Gomez-Casanovas et al. (2012) also found that RA decreased when soil water content was above 30%. These results seemed to support our findings. Third, the relationship between soil water content and RA or total soil respiration is more complex than the relationship between temperature and soil respiration. Numerous formula, such as linear (Tang et al., 2016), polynomial (Moyano et al., 2012), logarithmic (Schaefer et al., 2009), quadratic (Hursh et al., 2017) models have been widely applied to describe the relationship between soil water content and soil respiration. The multifarious relationships between soil water content and RA may occur because soil water content affect RA in multiple ways. Meanwhile, seasonal variability of precipitation and soil water content is often correlated with temperature (Feng and Liu, 2015), making the relationship between soil water content and RA more complex.

Similarly, the dominance of precipitation in Hashimoto was also widely observed (Fig. 8), dominating 77% of land areas (Fig. 10). Although this percentage was 17% higher than RF-RA, both results demonstrated that the global RA in majority land areas was dominated by precipitation. However, it is noticeable that dominant environmental factor controlling spatial carbon fluxes gradient may differ among different years (Reichstein et al., 2007), e.g. climate extreme and disturbance.

4.3 Comparison with Hashimoto2015-RA

Globally, total RF-RA was slightly higher than Hashimoto2015-RA, however, great divergence was observed both spatially and temporally (Fig. 6), particularly in tropical regions, where RF-RA was much lower than Hashimoto2015-RA (Fig. 3). These differences could be attributed to several reasons. First, two RA datasets had different land cover areas, especially in desert areas in North Africa, where existed very sparse or no vegetation. If RF-RA was masked by Hashimoto2015-RA, global RA was 39.6±0.4 Pg C yr⁻¹, which was pretty close to Hashimoto2015-RA (Fig. S8). Second, different predictors and algorithms were applied for RF-RA and Hashimoto2015-RA prediction. Besides temperature and precipitation, RA was also affected by soil nutrient, carbon substrate supply, belowground carbon allocation, site disturbance and other variables (Chen et al., 2014; Hashimoto et al., 2015; Tang et al., 2016; Zhou et al., 2016). Hashimoto2015-RA was calculated from the difference between total soil respiration and heterotrophic respiration, which were predicted by a simple climate-driven model using temperature and precipitation only (Hashimoto et al., 2015). Thus, Hashimoto2015-RA could not reflect its soil nutrient and other environmental constrains. To overcome such limitations, besides temperature and precipitation, we included soil water content, soil nitrogen and soil organic carbon as proxies for environmental and nutrient constraints of RA and considered the interactions among these variables using RF, achieving a model efficiency to 0.52 for RA prediction (Fig. S1), which was

higher than that for Hashimoto soil respiration with a model efficiency of 0.32 (Hashimoto et al., 2015). The simple climate-model for Hashimoto soil respiration could be its advantages and limitations (Hashimoto et al., 2015). Third, the empirical model (the relationship between total soil respiration and heterotrophic respiration) deriving Hashimoto2015-RA originated from forest ecosystems (Bond-Lamberty et al., 2004; Hashimoto et al., 2015), which may bring uncertainties to other ecosystems. For example, the difference between RF-RA and Hashimoto2015-RA varied up to 350 g C yr⁻¹ in South, North Amazon areas and Madagascar, where the savannas widely distributed (Fig. 4), thus Hashimoto2015-RA might not capture the spatial and temporal pattern of RA for non-forest ecosystems. Including more environmental variables and improving algorithm could be a great advantage to reduce the uncertainty in modelling RA.

4.4 Advantages, limitations and uncertainties

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Generally, the developed RF-RA dataset had four main advantages to estimate global RA: first, the RF-RA dataset, to our knowledge, was the first attempt to model RA using a large number of empirical field observations, and the spatial and temporal patterns of RA were investigated globally. While most previous studies mainly focused on global soil respiration, which was not partitioned into RA and heterotrophic respiration globally (Hursh et al., 2017; Jian et al., 2018b; Zhao et al., 2017). Second, we used an up-to-date field observational dataset developed from SRDB up to November, 2018 (Bond-Lamberty and Thomson, 2018), and updated it by including 68 observations from Chinese peer-review literatures. This new updated dataset included a total of 449 field observations (Fig. 1). These observations had a wide coverage range of global terrestrial ecosystems and represented all major biomes and climate zones. Third, the global terrestrial ecosystems were separated into eight biomes, including boreal forest, cropland, grassland, savannas, shrubland, temperate forest, tropical forest and wetland. The total RA and its inter-annual variability were evaluated for each of the eight biomes (Fig. S3 and S4). Besides, total RA and its interannual variability was also assessed for three climate zones – boreal, temperate and tropical zones (Figs. S2 and S5), according to the Köppen-Geiger climate classification system (Peel et al., 2007). Different temporal change trends across biomes and climate zones also further indicated uneven responses of RA to climate change across the globe. Fourth, we used a RF algorithm to model and map global RA with the linkage of climate, soil and other environmental predictors. The results showed that RF could well capture the spatial and temporal variability of RA (Fig. S1). Compared to linear regressions for soil respiration prediction (because no global RA prediction before this study) with a model efficiency less than 35% (Bond-Lamberty and Thomson, 2010b; Hashimoto et al., 2015; Hursh et al., 2017), RF algorithm achieved a much higher model efficiency to 52%, which indeed improved the RA modelling and reduced the uncertainties.

Although data-derived global RA could serve as a benchmark for global carbon cycling modelling, and the RF-RA had filled the data-gaps of global RA, limitations and uncertainties still remained in few aspects. First, although we conducted a data quality control to develop the RF-RA dataset, a lack of reliable approach to separate RA and heterotrophic respiration may lead to an important uncertainty of RA estimates. There are several approaches, e.g. trenching, stable or radioactive isotope,

gridding, to partition soil respiration (Bond-Lamberty et al., 2004; Högberg et al., 2001; Hanson et al., 2000), however, each of these approaches has its own limitations. For example, trenching has been widely applied to partition RA and heterotrophic respiration due to easy operation and low cost, on the other hand, heterotrophic respiration may be increased due to the termination of water uptake by roots and the decomposition of remaining dead roots in trenching plots (Hanson et al., 2000; Tang et al., 2016). Commonly, RA was calculated from the difference between total soil respiration and heterotrophic respiration, thus the trenching approach might lead to an underestimation of RA. In our dataset, a total of 254 RA observations were estimated by trenching approach, while the rest RA observations were estimated by other separation approaches, e.g. isotope, radiocarbon, mass balance. Thus, inconsistent separation approaches could also be another source of uncertainty of RA values.

Second, due to the limited observations of RA at a daily or monthly scale, the RF-RA dataset was produced at an annual scale. Although there was no direct study to compare the difference of RA upscaling from daily or monthly and annual scale, substantial differences of soil respiration upscaling from daily or monthly and annual scales (Jian et al., 2018b) indirectly illustrated the potential difference of RA upscaling from different timescales.

Third, the effects of rising atmospheric CO₂ on root growth was not explicitly represented when developing the RF-RA dataset, although CO₂ fertilization effects could partly be represented in the increase temperature. While the magnitude of CO₂ fertilization effects on photosynthesis is still uncertain (Gray et al., 2016), RF or other machine learning approaches are encouraged to quantify the uncertainties due to CO₂ fertilization.

Fourth, we did not consider the effects of human activities and historical changes in biomes on RA. However, important changes may occur in tropical forest, grassland and cropland during last several decades due to human activities (Hansen et al., 2013; Klein Goldewijk et al., 2011). Thus, changes in biomes should be included in future global RA and carbon cycling modelling. However, the lack of such data is the main constrain of detecting the effects of biome change on RA.

Finally, uneven coverage of observations in the updated dataset would be another source of uncertainties. Although our dataset had a wide range of land cover, the observational sites mainly distributed in China, Europe and North America and were dominated by forests. There was a great lack of observations in areas, such as Africa, Austria and Russia, and biomes, such as tropical forest, shrubland, wetland and cropland. However, our dataset covered all major ecosystem types and climate zones across the globe. RA observations caused bias of RF model toward the regions with more observations. Therefore, including more observationss in these areas and biomes without observations should largely increase our capability to assess the spatial and temporal patterns of global RA and contribute to improve the global carbon cycling modelling to future climate change.

5 Data availability

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The developed RF-RA dataset is freely downloadable from https://doi.org/10.6084/m9.figshare.7636193 (Tang et al.,

2019), named as "Respiration_autotrophic_belowgroud_glob_1980_2012_yr_half_dgree_TangX.nc", which is a globally gridded RA dataset from 1980 to 2012 with a spatial resolution of 0.5 degree at an annual scale, expressed by g C m⁻² yr⁻¹ (gram carbon for per square meter per year). The RA dataset is provided in Netcdf format (Network Common Data Form).

6 Conclusions

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Although data-derived RA may serve as a benchmark for ecosystem models, no such study has assessed the global variability in RA with a large number of empirical observations that can help bridge the knowledge gap between local, regional, and global scales. The RF-RA dataset has filled this knowledge gap by linking field observations and globally gridded environmental variables using RF algorithm, providing a global RF-RA dataset at a spatial resolution of 0.5° × 0.5° (longitude × latitude) at an annual scale from 1980 to 2012. Currently, robust findings include: (1) Annual mean RA was 43.8±0.4 Pg C yr⁻¹ with a temporally increasing trend of 0.025±0.006 Pg C yr⁻² over 1980-2012, indicating an increasing carbon return from the roots to the atmosphere; (2) unevenly temporal and spatial variabilities in varying climate zones and biomes indicated their uneven responses to future climate change, challenging the perspective that the parameters of global carbon stimulation independent on climate zones and biomes; (3) precipitation dominated RA for most of land areas globally; (4) The RF-RA dataset could serve as a has great potentials to serve as a benchmark for future data-model comparisons to understand the mechanisms of belowground vegetation carbon allocation and its dynamics. However, further improvements in modelling algorithms and including more observations in areas without field measurements should overcome shortcomings from reduced data availability and the mismatch in spatial resolution between covariates and *in situ* RA.

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Competing interests. The authors declare that they have no conflict of interest.

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