

***Generalized models to estimate carbon and nitrogen stocks
of organic soil horizons in Interior Alaska***

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1 **Abstract**

2 Boreal ecosystems comprise one tenth of the world's land surface and contain over 20 % of the
3 global soil carbon (C) stocks. Boreal soils are unique in that its mineral soil is covered by what can be
4 quite thick layers of organic soil. These organic soil layers, or horizons, can differ in their state of
5 decomposition, source vegetation, and disturbance history. These differences result in varying soil
6 properties (bulk density, C concentration, and nitrogen (N) concentration) among soil horizons. Here we
7 summarize these soil properties, as represented by over 3000 samples from Interior Alaska, and examine
8 how soil drainage and stand age affect these attributes. The summary values presented here can be used to
9 gap-fill large datasets when important soil properties were not measured, provide data to initialize
10 process-based models, and validate model results. These data are available at
11 <https://doi.org/10.5066/P960N1F9> (Manies, 2019).

12

13 **1 Introduction**

14 Boreal soils play an important role in the global carbon (C) budget and are estimated to store
15 between 375 - 690 Pg of C (Hugelius et al., 2014; Bradshaw and Warkentin, 2015; Khvorostyanov et al.,
16 2008), which is over 20 % of the global soil C stock (Jackson et al., 2017). A large portion of this C can
17 be found within the organic soil layer (Jorgenson et al., 2013). Although plant inputs into the soil can be
18 relatively high during the summer, C losses from the soil are low, as cool and/or freezing soil
19 temperatures result in low rates of decomposition. The imbalance between C inputs and losses results in
20 organic soils that can be quite thick and store large amounts of C (Jorgenson et al., 2013). There is also
21 considerable C found in the mineral soil of these systems, especially where protected by permafrost
22 (O'Donnell et al., 2011). Thus, both organic and mineral soil play an important role determining the
23 amount of C stored in boreal ecosystems.

24 Nitrogen (N) also plays an important role in boreal ecosystems due to N limitations on plant
25 growth (Herndon et al., 2020). N inputs to boreal ecosystems often begin with N fixation from
26 cyanobacteria, usually associated with mosses, or symbiotic actinomycetes, mainly the genus *Frankia*.

27 Net N mineralization increases over the course of upland succession, until the oldest state, black spruce
28 (*Picea mariana*) forest, when rates drop sharply (Kielland et al., 2006). Boreal ecosystems can have N
29 restricted by certain species, such as *Sphagnum* spp., through competitive interactions and slow rates of
30 turnover (Malmer et al., 2003). In addition, N cycling can become limited due to environmental factors
31 such as permafrost or anerobic conditions (Limpens et al., 2006; Bonan, 1990). Once released, N
32 availability impacts decomposition and plant growth and, therefore, can also influence rates of C
33 accumulation and loss.

34 Boreal organic soils are unique when compared to soils from other regions. These organic soils
35 can be thick, ranging from several centimeters to several meters (Ping et al., 2006). They are also
36 comprised of layers, or horizons, which as they deepen and increase in age also increase in their degree of
37 decomposition. These organic soil horizons are also influenced by the vegetation from which they formed
38 (Deluca and Boisvenue, 2012). Vegetative history is usually determined by post-disturbance plant
39 succession. Age and vegetative history not only affect the soil density, but also C and N concentrations,
40 resulting in large differences in C and N storage among horizons.

41 The main disturbances that affect boreal soil properties are fire and permafrost thaw. Fires affect
42 boreal soils through the combustion of litter and surface organic layers (as ground fuel; Harden et al.,
43 2000), with the amount and depth of combustion regulated by fire severity (Turetsky et al., 2011). Fire
44 directly effects surface organic soils, both in elemental composition and structure (Neff et al., 2005). In
45 addition, there are indirect effects of fire on soil properties. The loss of insulating organic soil results in a
46 darkened soil surface, which in turn warms post-fire soils, increasing decomposition rates from the
47 surface downward (Genet et al., 2013; O'Neill et al., 2002). In addition, both fire return interval and fire
48 severity influence post-fire vegetation and the re-accumulation of organic soil layers. As different tree and
49 understory species have different amounts of C and N in their tissues (Van Cleve et al., 1983), changes in
50 post-fire vegetation affect soil C and N accumulation rates and thus, the concentration of these elements
51 in surface soil. Permafrost thaw also affects soil properties in several ways. By definition, thaw exposes
52 older, previously sequestered C to warmer soil temperatures (Osterkamp et al., 2009), increasing rates of

53 decomposition (Mu et al., 2016; Schadel et al., 2016). In well drained sites post-thaw conditions usually
54 result in water draining from the soil, resulting in oxic conditions (Estop-Aragonés et al., 2018). In
55 lowlands, permafrost thaw often results in subsidence and inundation, changing the ecosystem from a
56 forested permafrost plateau to a thermokarst wetland (Schuur et al., 2015). Fire can often be a trigger for
57 this rapid permafrost thaw (Myers-Smith et al., 2008). Post fire vegetation changes affects both C and N
58 inputs, again affecting the concentration of these elements within surface organic soil layers. As both fire
59 frequency and permafrost thaw are expected to increase in the future (Hinzman et al., 2005),
60 biogeochemical models have a need to characterize how these disturbances will impact C and N stocks.
61 To accurately represent future scenarios, models need to include the distinct properties of organic soil
62 horizons found in the boreal region (Flato et al., 2013).

63 Despite the need to accurately portray the state and dynamic nature of boreal organic soil
64 properties, these soils have not been widely characterized nor compiled into a common framework.
65 Instead, much of the work regarding boreal soils has focused on predicting C and N stocks for combined
66 organic and mineral soil horizons to a predetermined depth (Johnson et al., 2011; Bauer et al., 2006). Ping
67 (2010) examined organic soils for Alaska, but only focused on black spruce (*Picea mariana*) forests. In
68 addition, Michaelson et al. (2013) compiled a great deal of Alaskan-based soil data, although they present
69 these data for the organic soil layer as a whole. Therefore, there is currently no source of summarized data
70 of soil properties by organic soil horizon. To fill this gap, we summarized soil properties from a database
71 of over 3000 observations from Interior Alaska (Figure 1). Soil properties were categorized by degree of
72 decomposition (via classification into distinct organic soil horizons), soil drainage, and stand age. This
73 data set can be used in many ways including field comparisons, models construction, and model
74 validation.

75

76

77 **2 Methods**

78 **2.1 Field site classifications**

79 Soil cores were sampled at 58 different sites located within several areas of Interior Alaska (Figure 1).
80 Several different ecosystem types were sampled, including black spruce forests (~50%), wetlands (~26%),
81 and deciduous and mixed forests (~16%). Between 1 and 14 soil profiles were sampled at each site, for a
82 total of 292 soil profiles. Sampling took place over a 15-year period from 2000-2015. We examined the
83 effect of fire or permafrost thaw disturbance on soil properties by categorizing each of the soil profiles in
84 relation to time since the last disturbance, which we divided into three age classes: new (<5 yrs old), young
85 (5 – 50 yrs old), and mature (> 50 yrs old). All new sites were recently burned and thus had lost some
86 portion of their surface organic horizons (Harden et al., 2000), while young sites experienced either fire or
87 permafrost thaw.

88 In addition, sites were classified according to their soil drainage. Although classifications of soil
89 drainage have been established for many soil types (Soil Survey Division Staff, 1993), the presence of
90 permafrost, and its effect on drainage and soil moisture, necessitates modifications of this system (Expert
91 Committee on Soil Survey, 1982). Although generally described (Harden et al., 2003; Johnstone et al.,
92 2008), a soil drainage classification for permafrost landscapes is lacking. Here we present a soil drainage
93 classification decision tree, developed over the past two decades, for areas of discontinuous permafrost
94 (Figure 2). Well drained sites are similar to traditional drainage classifications, in that water moves through
95 the soil rapidly. However, moderately well drained drainage sites have permafrost between 75 – 150 cm,
96 which increases soil moisture of surface organics. Somewhat poorly, poorly, and very poorly drained sites
97 have some factor (permafrost, soil texture, or landscape position) that inhibits drainage and causes
98 redoximorphic features such as blue-grey colors in the mineral soil to appear. Somewhat poorly drained
99 sites have a shallow active layer (often around 50 cm), which affects soil moisture and surface vegetation.
100 Poorly drained sites experience saturated surface conditions only while seasonal ice is present (usually May
101 through early July), while very poorly drained sites have saturated surface soils during the entire growing
102 season.

103 Modification of the drainage class occurs when sites are on a slope. When sites are located on a slope
104 of greater than 5 %, drainage increases (Woo, 1986; Carey and Woo, 1999), and therefore drainage class
105 designation (Figure 2) is increased by one step. This is called the hillslope modifier. In addition, because
106 burning increases active layer thickness (Gibson et al., 2018), recently burned sites may have deeper
107 permafrost or no permafrost at all. Because the effects of these drier soil properties may not have yet
108 propagated through factors such as thickness of the deeper organic layers, for many analyses, including this
109 paper, it makes more sense to ascribe their soil drainage using nearby unburned sites.

110

111 **2.2 Soil sampling methodology**

112 Soil cores were obtained using several different methods. The first method, most often used with
113 surface horizons, involved cutting soil blocks to a known volume. Another method often used involves a
114 coring device inserted into a hand drill (4.8 cm diameter; Nalder and Wein, 1998). Wetter sites were
115 sometimes sampled while frozen using a Snow, Ice, and Permafrost Research Establishment (SIPRE)
116 corer (7.6-cm diameter; Rand and Mellor, 1985). Alternatively, if wetter sites were sampled unfrozen we
117 used a ‘frozen finger’. This coring method uses a thin-walled, hollow tube (~6.5 cm diameter), sealed at
118 one end, which is inserted into the ground until it hits mineral soil. A slurry of dry ice and alcohol is then
119 poured into the corer, freezing the unfrozen material surrounding the corer to the outside. The corer is
120 removed and the exterior of the core is scraped to remove any large roots or material that stuck to the
121 sample during removal. Another method occasionally used in unfrozen saturated soils involves the
122 insertion and careful removal of PVC tubing sharpened on one end. Finally, a variety of commercially- or
123 home-made soil corers were used to obtain volumetric samples for ~6% of these data, usually for mineral
124 soil samples. For some soil profiles, two coring methods were combined to create continuous samples
125 from the surface to the mineral soil. While most cores were sampled into the mineral soil, some cores
126 ended at or before the organic/mineral interface due to the presence of permafrost without proper
127 sampling equipment or because the cores were collected for the purpose of only studying surface
128 organics. All sampling methods were volumetric, providing the basis for bulk density calculations (g/cm^3)

129 Organic soil layers or horizons were described and then subdivided according to field-based
130 visual and tactical factors such as level of decomposition, color, and root abundance, regardless of region
131 or soil drainage. These horizons provided the basis for our analyses and are based on Canadian (Soil
132 Classification Working Group, 1998) and U.S. Department of Agriculture's Natural Resource
133 Conservation Service (Soil Survey Staff, 1998) soil survey techniques. A description of the horizons and
134 the codes we used to represent them are found in Table 1, but in summary there are six main horizons:
135 live moss (L), dead moss (D), fibric (mostly undecomposed; F), mesic (more decomposed; M), humic
136 (very decomposed; H), and mineral soil (Min).

137 To aid researchers who may need to have these properties summarized in a more simplified scheme (as
138 in Yi et al., 2009; O'Donnell et al., 2009), we also combined horizons post-hoc into a simplified scheme.
139 Here, the fibrous horizon consists of both the dead moss (D) and fibric (F) horizons, while the amorphous
140 horizon combined the mesic (M) and humic (H) horizons. These combinations were based on similarities
141 in decomposition state and depth within the organic soil profile. We also present data for several types of
142 surface horizons that are only found a small fraction of sites; those data are presented separately. Ash and
143 burned organic surface horizons are only found in recently burned sites. Lichen and litter dominated
144 horizons are only found on the surface of ~16 % of profiles and related to well drained forest conditions.
145 Our field studies also found several horizon types (buried wood, grass, etc.) for which we had few
146 observations (5 or less), and, thus, were not included in our analyses.

147

148 **2.3 Laboratory methodology**

149 Once returned from the field soils horizon samples were weighed and air-dried at room temperature
150 (20 °C to 30 °C) to a constant mass, then oven-dried for 24-48 hours in a forced-draft oven. Organic soils
151 (live moss, dead moss, fibric, mesic, and humic horizons) were oven-dried at 65 °C to avoid the alteration
152 of organic matter chemistry. Mineral soils were oven-dried at 105 °C. Mineral soil samples were gently
153 crushed using a mortar and pestle, with care to break only aggregates, and then sieved through a 2-mm
154 screen. Soil particles that did not pass through the screen were removed, weighed, and saved separately;

155 soil that passed through the screen was then ground by using a mortar and pestle to pass through a 60-mesh
156 (0.246-mm) screen. The ground material was mixed and placed in a labeled glass sample bottle for
157 subsequent analyses. Organic soil samples were weighed, and roots wider than 1 cm in diameter were
158 removed, weighed, and saved separately. The remaining sample material was then milled in an Udy Corp.
159 Cyclone Sample Mill to pass through a 0.25-mm screen and placed in a labeled glass vial.

160 We analyzed soil samples for total C and N using a Carlo Erba NA1500 elemental analyzer
161 (Fisons Instruments). Samples were combusted in the presence of excess oxygen. The resulting sample
162 gases were carried by a continuous flow of helium through an oxidation furnace, followed by a reduction
163 furnace, to yield CO₂, N₂, and water vapor. Water was removed by a chemical trap and CO₂ and N₂ were
164 chromatographically separated before the quantification of C and N (Pella, 1990a,b). We assumed that
165 mineral soil samples below pH 7, which are common to Interior AK, had no inorganic carbon (IC)
166 present, and thus total C represents total organic C. For mineral-soil horizons where IC was present, we
167 removed carbonates using the acid fumigation technique (Komada et al., 2008) prior to running samples.
168 To do this, we preweighed samples in silver capsules and transferred them to a desiccator. Samples were
169 wetted with 50 µL of deionized water and then exposed to vaporous hydrochloric acid (1 N) for a
170 minimum of 6 hours, during which carbonates degassed from samples as carbon dioxide.

171

172 **2.4 Data quality and statistical methodology**

173 Often the soil descriptions at the interface of the organic and mineral soil included notations
174 indicating that these horizons consisted of mixed organics and mineral soil. Using visual and textural cues
175 the field, horizons were categorized as either mineral (< 20 % C) or organic (≥ 20 % C). However,
176 chemistry data sometimes shows these horizons were miscategorized due to slight under or over
177 estimations of OM content (for example, a mineral soil with 22 % C). We used C chemistry to remove
178 organic soils with < 20 % C from our analyses.

179 All statistical analyses were run using the R program (R Core Team, 2017). Data were
180 transformed to meet assumptions of normality (Table S1). The effects of drainage and age class for all

181 soil horizons with the exception of the fibrous and amorphous horizons, was tested for significant
182 difference among the different soil horizons using the mixed-effects model command *lmer* (lme4; Bates et
183 al., 2015), using soil profile (or soil core) as the random effect. When significant, differences among
184 drainage types or age class were determined using estimated marginal means (Least-squares means;
185 *emmeans*) (Lenth et al., 2020). No interactions were examined. Evaluation for the fibrous and amorphous
186 horizons, because all samples were within a single soil profile, was done using the analysis of variance
187 model (*aov*) with the Tukey honestly significant difference (*TukeyHSD*) function.

188

189 **3. Dataset Review**

190 **3.1 Bulk density**

191 Bulk density varied by depth and was significantly different ($p < 0.05$) among all horizon types
192 (live moss, dead moss, fibric, mesic, humic, and mineral soil; Table 1). Surprisingly, as they are
193 comprised of very similar material, even the live and dead moss horizons had significantly different bulk
194 densities. Bulk density increases ~10-fold from one organic horizon to the next down the soil profile
195 (from 0.022 g cm^{-3} for live moss to 0.215 g cm^{-3} for the humic horizon). These differences are likely
196 related to the length of time each soil horizon has had to decompose. As soil horizons become older, plant
197 fibers break down physically and biologically, becoming smaller and more compressed.

198 Bulk density also varied by drainage class, particularly at the deeper depths. Well drained sites
199 tended to have higher bulk densities than other poorer soil drainage classes, especially for the deeper soil
200 horizons (e.g. fibric and mesic; Table S2). Higher bulk densities with better drainage is likely related to
201 two factors: 1) the influence of lichens and litter, which often found at well drained sites, and have higher
202 bulk densities than moss (Table 4), and 2) the influence of mineral soil, which, due to shallower organic
203 soils, is more likely to be incorporated into fibric (F) and mesic (M) horizons. Greater mineral
204 incorporation into organic layers of shallow well drained soils is supported by the lower % C values also
205 found within well-drained F and M horizons (Table S3). New (< 5 yr old) sites often had higher bulk

206 densities than the older age classes (Table S2). There were, however, very few significant differences in
207 bulk density by age class, so this factor does not appear to play strong role in determining bulk density.

208

209 **3.2 Carbon**

210 Upper, shallow organic soil horizons (live moss, dead moss, and fibric horizons) differ from
211 deeper horizons (mesic and humic horizons) in several respects. Shallow horizons are consistently higher
212 in % C than deeper horizons (Table 2). However, upper, shallow horizons are lower in bulk density than
213 deeper horizons (Table 2), so that C density values (g cm^{-3}) increase dramatically with depth (Figure 3).
214 Therefore, even though the deeper organic horizons (M and H) have slightly lower C concentrations than
215 the shallow horizons, their high bulk densities result in large amounts of C at depth. In fact, given average
216 thickness, bulk density, and % C (Table 2), approximately 75% of the soil C is stored in the mesic and
217 humic soil horizons.

218 There were few clear trends with C concentration with drainage class, although moderately well
219 drained sites usually had higher C concentrations than the other drainage classes, especially somewhat
220 poorly drained sites (Table S3). Lower C values for the fibric and mesic well-drained sites are likely due
221 to the inclusion of mineral soil material into these horizons. While this difference is likely due in large
222 part to natural process such as cryoturbation or aeolian contributions, these horizons are thinner in well
223 drained sites (Table 3), so any accidental inclusion of mineral soil within these horizons during sampling
224 would have more of an effect.

225 C concentration increased with increasing age class for all organic horizons but the humic horizon
226 (Table S3). Since all sites classified as ‘new’ were recently disturbed by fire, this increase could be due to
227 the inclusion of more live roots and/or the loss of ash in older stands. Ash has a lower C content (Table 4)
228 and is a component of recently burned soil’s surface horizons.

229

230

231 **3.3 Nitrogen**

232 N concentration within the organic horizons increased with depth and then declined again in the
233 mineral soil (Table 2). There was significant variability in N by drainage class for each horizon type
234 (Table S4). The poorly and very poorly drained sites had greater concentrations of N than other drainage
235 classes for the fibric (F), mesic (M), and humic (H), and mineral horizons, and lower concentrations of N
236 in the dead moss (D) horizon. These higher values are likely because N builds up under saturated
237 conditions, due to low rates of microbial activity, limiting decomposition (Limpens et al., 2006). There
238 was also more N in the live and dead moss horizons of the new and younger stands (Table S3). These
239 differences are likely related to differences in N quality of early succession litterfall (Bonan, 1990).

240

241 **3.4 C:N ratio**

242 C:N ratios patterns followed those of C and N, with the surface organic horizons (live moss, dead
243 moss, and fibrics) having more similar values than the deeper soil organic horizons (Table 2). Well
244 drained sites tended to have lower C:N ratios (Table S5), likely caused by the lower C concentrations
245 found there (see section 3.2). C:N ratio increased with age class, but only in the surface organic horizons
246 (live moss, dead moss, and fibrics). These trends appear to be more influenced by differences in N by age
247 class than changes in C.

248

249 **3.5 Soil horizon thickness**

250 The factor that varied the most by horizon was the thickness of each horizon type (Table 2), and,
251 unlike most of the other factors, the standard deviation was often greater than the mean. There was a very
252 strong effect of drainage on horizon thickness, with the well-drained sites having much thinner soil
253 horizons (and no humic horizon) than the other drainage classes and the very poorly drained sites having
254 much thicker soil horizons than the other drainage classes (Table 3). Age class also plays a role in horizon
255 thickness: new sites (<5 yrs old) had much thinner organic soil horizons than young or mature sites (Table
256 3). Since new sites recently burned, these thin soil horizons are the result of the loss of organics due to

257 combustion. Both fire return interval and fire severity impact the amount of legacy soil remaining
258 (Harden et al., 2012), therefore fire history likely plays a large role in horizon thickness.

259 Vegetation could also influence horizon thickness. An examination of these data that included
260 current surface vegetation found greater thicknesses for sites with *Sphagnum* sp. and sedges, although this
261 factor usually was not statistically significant. Historical vegetation could also influence horizon
262 thickness. For instance, if a site was *Sphagnum* dominated in the past, even if it is not the current surface
263 vegetation, the soil profile is more likely to have thicker soil horizons due to the slow decomposition rate
264 of *Sphagnum* (Turetsky et al., 2008). Because such historical factors are difficult to measure and predict,
265 we recommend that researchers obtain their own measurements of organic horizon thickness whenever
266 possible and, if using the thickness data presented in Table 3, account for the variability found for
267 thickness estimates in their analyses.

268

269 **4.0 Discussion of the data set**

270 **4.1 Comparison to other data sets**

271 Our data are the first of its kind to present organic horizon data across a range of Alaskan boreal
272 ecosystems. Other studies have examined organic soil as a separate entity from mineral soil but with
273 certain limitations. Michaelson et al. (2013) used Alaskan USDA-NRCS soil pedon data to examine soil
274 properties of both organic and mineral soil but present these data for the organic portion as a whole. This
275 study shows that there is significant variation in bulk density and C and N concentration across organic
276 horizons, and therefore, one should not disregard these horizon-based variations. In a separate study,
277 Ping et al. (2010) separated the organics into two horizons from boreal black spruce stands (O_{surface} ,
278 O_e/O_a). Our study supports the results of Ping et al., (2010), which found a decrease in C:N ratios with
279 increasing depth. Moreover, our study provides data from a fuller suite of soil horizons and includes data
280 from bogs, fens, and deciduous forests.

281

282

283 4.2 How well do these values represent other data?

284 We tested how well our data from Interior AK can predict C and N stocks in other studies. Our
285 first test was for 142 samples taken from two fire chronosequences located near Thompson, Manitoba
286 (Manies et al., 2006). Each chronosequence represents a different drainage class: moderately well drained
287 versus somewhat poorly drained. These data were based on the same methods of sampling and describing
288 soil horizons. Using the horizon designations (Table 1) and horizon thickness (cm) from the Canadian
289 data, we assigned bulk density, C, and N values (Table 2). These predicted horizon-based C and N stocks
290 were summed for each soil profile and compared to the measured values. We found our predicted stocks
291 were relatively evenly distributed between being lower or higher than measured stocks (Figure S1), with
292 the majority of estimated stocks (>85%) within 50% of measured stocks and over 60% within 20% of
293 measured stocks. Soil profiles with much higher predicted than measured stocks were due to very low
294 measured bulk densities (e.g., a measured bulk density for a fibric horizon of 0.01 g/cm^3 , as compared to
295 the predicted value of 0.06 g/cm^3). The differences we found between measured and predicted stocks
296 could be due to regional differences between the Alaskan and Canadian sites in factors, such as
297 disturbance history or vegetation composition. In addition, accurate bulk density measurements is time
298 consuming to do correctly (Nalder and Wein, 1998) and could also play a role.

299 To further explore the predictive capabilities of our data, we also compared predicted versus
300 measured C stocks for a second study, this one located within Alaska (Kane and Ping, 2004), in which
301 horizon thickness (all samples), % C (all samples), and bulk density (one 5.08 cm diameter sample per
302 horizon per site) for soil profiles were measured along a continuum of tree productivity. To calculate
303 predicted C stocks we used their thickness values with bulk density and % C values from Table. 1.
304 However, Kane and Ping (2004) used the US Soil System to describe and sample their soils, dividing the
305 organic soil profile into O_i and O_e/O_a horizons. We chose to represent their O_i data, which they described
306 as slightly decomposed moss, using our fibrous horizon and their O_e/O_a data, which they described as
307 intermediately decomposed moss with rare saprics, as our amorphous horizon. Predicted C stocks were
308 higher than measured stocks (Figure S2). This result was mostly due to differences in bulk density values

309 between our amorphous horizon and their O_e/O_a horizon. Their study had O_e/O_a bulk density values that
310 ranged between 0.06 and 0.12 g/cm², which is typical of our fibric (F) and mesic (M) horizons (Table 2).
311 When we model their O_e/O_a data using F values, we slightly underestimate stocks, while if we model their
312 O_e/O_a data using M values we slightly overestimate their stocks (Figure S2). Thus, bulk density
313 measurements play a role in these differences. These results also demonstrate that soil description
314 protocols play an important role in characterizing C and N stocks and, in this case, the different system
315 used to identify and sample organic soil horizons may not be equivalent.

316

317 **4.3 Caveats and suggestions for use**

318 One of the important uses of this dataset is the potential for estimating C and N stocks based on
319 simple field characterizations of organic soil horizons of North American boreal forests and wetlands.
320 Because soil sampling and processing is quite time intensive, researchers may decide to measure
321 thicknesses of the various soil horizons within their sites, using the descriptors in Table 1, and then
322 calculate C and N stocks using the average values presented in Tables 2, S2, S3, or S4. This approach
323 minimizes errors associated with the high variability found for horizon thicknesses, due to variable site
324 histories.

325 While C stocks of mineral soils were not evaluated in this study, this region contains large
326 amounts of C within mineral soils, especially within Yedoma deposits (Hugelius et al., 2014; O'Donnell
327 et al., 2011). The mineral soil data presented here represent mostly the uppermost mineral soil. Additional
328 examinations into bulk density and C concentrations of Alaskan mineral soil can be found in Ping et al.
329 (2010), Michaelson et al. (2013), and Ebel et al. (2019).

330 Although our data provide an important resource for several properties of organic horizons, we
331 acknowledge that our samples are dominated by mature sites from areas that are not well drained.
332 Therefore, as additional soil horizon data is sampled, we encourage researchers to expand upon the work
333 presented here.

334

335 **5 Data Access**

336 All data used in this manuscript are available from <https://doi.org/10.5066/P960N1F9> (2019).
337 This publication includes both .csv data files as well as metadata. A short description of these files and the
338 data found within them can be found in Tables 5 and 6. In addition, many additional soil attributes not
339 included in that publication, such as von Post decomposition index and additional soil chemistry
340 information, can be found for the majority of these data through various USGS Open-File Reports
341 (Manies et al., 2017; Manies et al., 2016; Manies et al., 2014; O'Donnell et al., 2013; O'Donnell et al.,
342 2012; Manies et al., 2004).

343

344 **6 Conclusions**

345 Boreal ecosystems are especially sensitive and vulnerable due to climate change. Models may not
346 accurately forecast high latitude biogeochemical processes for many reasons (Flato et al., 2013). One
347 reason for the discrepancies between model results and data is that many models lack the input data
348 required, including important factors for modeling soil thermal dynamics like bulk density (Koven et al.,
349 2013; Khvorostyanov et al., 2008). While these processes are starting to be incorporated into land surface
350 and regional models (see, for example, Genet et al., 2013; Koven et al., 2011), currently few models
351 include the distinct properties of organic soils that are found in the boreal region (Flato et al., 2013). The
352 >3,000 soil samples, from >290 soil profiles, presented in this paper provide information regarding the
353 important soil properties of bulk density, C concentration, N concentration, C:N ratios, and thickness by
354 organic soil horizon. Such data are needed for initializing and validating models related to boreal organic
355 soils. In addition, these data can be used by scientists to calculate C and N stocks where researchers only
356 have soil horizon thickness data or to address shortcomings of missing data in instances when an
357 important soil property was not measured.

358

359

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366

367 **Author contribution**

368 KM prepared the manuscript with the help of MW and JH. All authors were involved in supporting the
369 collection of these data.

370

371 **References Cited**

- 372 Bates, D., Maechler, M., Bolker, B., and Walker, S.: Fitting Linear Mixed-Effects Models Using lme4,
373 *Journal of Statistical Software*, 67, 1-48, doi:10.18637/jss.v067.i01, 2015.
- 374 Bauer, I. E., Bhatti, J. S., Cash, K. J., Tarnocai, C., and Robinson, S. D.: Developing Statistical Models to
375 Estimate the Carbon Density of Organic Soils, *Canadian Journal of Soil Science*, 86, 295–304, 2006.
- 376 Bonan, G. B.: Carbon and Nitrogen Cycling in North American Boreal Forests. I. Litter Quality and Soil
377 Thermal Effects in Interior Alaska, *Biogeochemistry*, 10, 1-28, 1990.
- 378 Bradshaw, C. J. A., and Warkentin, I. G.: Global estimates of boreal forest carbon stocks and flux, *Global
379 and Planetary Change*, 128, 24-30, 10.1016/j.gloplacha.2015.02.004, 2015.
- 380 Carey, S. K., and Woo, M. K.: Hydrology of two slopes in subarctic Yukon, Canada, *Hydrological
381 Processes*, 13, 2549-2562, 10.1002/(SICI)1099-1085(199911)13:16<2549::AID-HYP938>3.0.CO;2-
382 H, 1999.
- 383 Deluca, T. H., and Boisvenue, C.: Boreal forest soil carbon: distribution, function and modelling,
384 *Forestry: An International Journal of Forest Research*, 85, 161-184, 10.1093/forestry/cps003, 2012.
- 385 Ebel, B. A., Koch, J. C., and Walvoord, M. A.: Soil Physical, Hydraulic, and Thermal Properties in
386 Interior Alaska, USA: Implications for Hydrologic Response to Thawing Permafrost Conditions,
387 *Water Resources Research*, 55, 4427-4447, 10.1029/2018wr023673, 2019.
- 388 Estop-Aragonés, C., Cooper, M. D. A., Fisher, J. P., Thierry, A., Garnett, M. H., Charman, D. J., Murton,
389 J. B., Phoenix, G. K., Treharne, R., Sanderson, N. K., Burn, C. R., Kokelj, S. V., Wolfe, S. A.,
390 Lewkowicz, A. G., Williams, M., and Hartley, I. P.: Limited release of previously-frozen C and
391 increased new peat formation after thaw in permafrost peatlands, *Soil Biology and Biochemistry*, 118,
392 115-129, <https://doi.org/10.1016/j.soilbio.2017.12.010>, 2018.
- 393 Expert Committee on Soil Survey: The Canada Soil Information System (CanSIS): Manual for describing
394 soils in the field, LRR Contribution No. 82-52, 175 pp., 1982.
- 395 Flato, G., Marotzke, J., Abiodun, B., Braconnot, P., Chou, S. C., Collins, W., Cox, P., Driouech, F.,
396 Emori, S., Eyring, V., Forest, C., Gleckler, P., Guilyardi, E., Jakob, C., Kattsov, V., Reason, C., and
397 Rummukainen, M.: Evaluation of Climate Models, in: *Climate Change 2013: The Physical Science
398 Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental*

399 Panel on Climate Change, edited by: Stocker, T. F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S. K.,
400 Boschung, J., Nauels, A., Xia, Y., Bex, V., and Midgley, P. M., Cambridge University Press,
401 Cambridge, United Kingdom and New York, NY, USA, 741–866, 2013.

402 Genet, H., McGuire, A. D., Barrett, K., Breen, A., Euskirchen, E. S., Johnstone, J. F., Kasischke, E. S.,
403 Melvin, A. M., Bennett, A., Mack, M. C., Rupp, T. S., Schuur, A. E. G., Turetsky, M. R., and Yuan,
404 F.: Modeling the effects of fire severity and climate warming on active layer thickness and soil carbon
405 storage of black spruce forests across the landscape in interior Alaska, *Environmental Research*
406 *Letters*, 8, 45016-45016, 2013.

407 Gibson, C. M., Chasmer, L. E., Thompson, D. K., Quinton, W. L., Flannigan, M. D., and Olefeldt, D.:
408 Wildfire as a major driver of recent permafrost thaw in boreal peatlands, *Nature Communications*, 9,
409 3041, 10.1038/s41467-018-05457-1, 2018.

410 Harden, J. W., Trumbore, S. E., Stocks, B. J., Hirsch, A., Gower, S. T., O'Neill, K. P., and Kasischke, E.
411 S.: The role of fire in the boreal carbon budget, *Global Change Biology*, 6, S174–S184, 2000.

412 Harden, J. W., Meier, R., Silapaswan, C., Swanson, D. K., and McGuire, A. D.: Soil drainage and its
413 potential for influencing wildfires in Alaska, in: *Studies by the U.S. Geological Survey in Alaska*,
414 2001, edited by: Galloway, J., U.S. Geological Survey Professional Paper 1678, 139–144, 2003.

415 Harden, J. W., Manies, K. L., O'Donnell, J., Johnson, K., Froking, S., and Fan, Z.: Spatiotemporal
416 analysis of black spruce forest soils and implications for the fate of C, *Journal of Geophysical*
417 *Research*, 117, G01012, 10.1029/2011JG001826, 2012.

418 Herndon, E., Kinsman-Costello, L., and Godsey, S.: Biogeochemical Cycling of Redox-Sensitive
419 Elements in Permafrost-Affected Ecosystems, in: *Biogeochemical Cycles*, edited by: Dontsova, K.,
420 Balogh-Brunstad, Z., and Le Roux, G., 245-265, 2020.

421 Hinzman, L. D., Bettez, N. D., Bolton, W. R., Chapin, F. S., Dyrgerov, M. B., Fastie, C. L., Griffith, B.,
422 Hollister, R. D., Hope, A., Huntington, H. P., Jensen, A. M., Jia, G. J., Jorgenson, T., Kane, D. L.,
423 Klein, D. R., Kofinas, G., Lynch, A. H., Lloyd, A. H., McGuire, A. D., Nelson, F. E., Oechel, W. C.,
424 Osterkamp, T. E., Racine, C. H., Romanovsky, V. E., Stone, R. S., Stow, D. A., Sturm, M., Tweedie,
425 C. E., Walker, M. D., Walker, D. A., Webber, P. J., Welker, J. M., Winker, K. S., and Yoshikawa, K.:
426 Evidence and implications of recent climate change in northern Alaska and other arctic regions,
427 *Climatic Change*, 72, 251–298, doi: 10.1007/s10584-005-5352-2, 2005.

428 Hugelius, G., Strauss, J., Zubrzycki, S., Harden, J. W., Schuur, E. A. G., Ping, C. L., Schirmer, L.,
429 Grosse, G., Michaelson, G. J., Koven, C. D., O'Donnell, J. A., Elberling, B., Mishra, U., Camill, P.,
430 Yu, Z., Palmtag, J., and Kuhry, P.: Estimated stocks of circumpolar permafrost carbon with quantified
431 uncertainty ranges and identified data gaps, *Biogeosciences*, 11, 6573-6593, 10.5194/bg-11-6573-
432 2014, 2014.

433 Jackson, R. B., Lajtha, K., Crow, S. E., Hugelius, G., Kramer, M. G., and Piñeiro, G.: The Ecology of
434 Soil Carbon: Pools, Vulnerabilities, and Biotic and Abiotic Controls, *Annual Review of Ecology*,
435 *Evolution, and Systematics*, 48, 419-445, 10.1146/annurev-ecolsys-112414-054234, 2017.

436 Johnson, K. D., Harden, J. W., McGuire, A. D., Bliss, N. B., Bockheim, J. G., Clark, M., Nettleton-
437 Hollingsworth, T., Jorgenson, M. T., Kane, E. S., Mack, M., O'Donnell, J., Ping, C. L., Schuur, E. A.
438 G., Turetsky, M. R., and Valentine, D. W.: Soil carbon distribution in Alaska in relation to soil-
439 forming factors, *Geoderma*, 167-168, 10.1016/j.goederma.2011.10.006, 2011.

440 Johnstone, J. F., Hollingsworth, T. N., and Chapin, F. S., III: A key for predicting postfire successional
441 trajectories in black spruce stands of interior Alaska, *USDA Forest Service*, 37 pp., 2008.

442 Jorgenson, M. T., Harden, J. W., Kanevskiy, M., O'Donnell, J. A., Wickland, K. P., Ewing, S. A., Manies,
443 K. L., Zhuang, Q., Shur, Y., Striegl, R. G., and Koch, J. C.: Reorganization of vegetation, hydrology,
444 and soil carbon after permafrost degradation across heterogeneous boreal landscapes, *Environmental*
445 *Research Letters*, 8, 035017, 10.1088/1748-9326/8/3/035017, 2013.

446 Kane, E. S., and Ping, C.-L. L.: Soil carbon stabilization along productivity gradients in interior Alaska:
447 Summer 2003. Fairbanks, B. C. L.-U. o. A. (Ed.), <http://www.lter.uaf.edu/data/data-detail/id/132>,
448 2004.

449 Khvorostyanov, D. V., Krinner, G., Ciais, P., Heimann, M., and Zimov, S. A.: Vulnerability of permafrost
450 carbon to global warming. Part I: model description and role of heat generated by organic matter
451 decomposition, *Tellus B*, 60, 250-264, 10.1111/j.1600-0889.2007.00333.x, 2008.

452 Kielland, K., Olson, K., Ruess, R. W., and Boone, R. D.: Contribution of winter processes to soil nitrogen
453 flux in taiga forest ecosystems, *Biogeochemistry*, 81, 349-360, 10.1007/s10533-006-9045-3, 2006.

454 Komada, T., Anderson, M. R., and Dorfmeier, C. L.: Carbonate removal from coastal sediments for the
455 determination of organic carbon and its isotopic signatures, $\delta^{13}\text{C}$ and $\Delta^{14}\text{C}$: comparison of
456 fumigation and direct acidification by hydrochloric acid, *Limnology & Oceanography: Methods*, 6,
457 254-262, 2008.

458 Koven, C. D., Ringeval, B., Friedlingstein, P., Ciais, P., Cadule, P., Khvorostyanov, D., Krinner, G., and
459 Tarnocai, C.: Permafrost carbon-climate feedbacks accelerate global warming, *Proceedings of the
460 National Academy of Sciences*, 108, 14769-14774, 10.1073/pnas.1103910108, 2011.

461 Koven, C. D., Riley, W. J., and Stern, A.: Analysis of permafrost thermal dynamics and response to
462 climate change in the CMIP5 Earth System Models, *Journal of Climate*, 26, 1877-1900, 2013.

463 Lenth, R., Signmann, H., Love, J., Buerkner, P., and Herve, M.: emmeans: Estimated Marginal Means,
464 aka Least-Squares Means, 2020.

465 Limpens, J., Heijmans, M. M. P. D., and Berendse, F.: The Nitrogen Cycle in Boreal Peatlands, in: *Boreal
466 Peatland Ecosystems*, edited by: Wieder, R. K., and Vitt, D., *Ecological Studies*, Springer Berlin
467 Heidelberg, 195-230, 2006.

468 Malmer, N., Albinsson, C., Svensson, B. M., and Wallén, B.: Interferences between Sphagnum and
469 vascular plants: effects on plant community structure and peat formation, *Oikos*, 100, 469-482,
470 10.1034/j.1600-0706.2003.12170.x, 2003.

471 Manies, K.: Data Supporting Generalized models to estimate carbon and nitrogen stocks of organic layers
472 in Interior Alaska. Survey, U. G. (Ed.), 2019.

473 Manies, K. L., Harden, J. W., Silva, S. R., Briggs, P. H., and Schmid, B. M.: Soil data from *Picea
474 mariana* stands near Delta Junction, Alaska of different ages and soil drainage types, U.S. Geological
475 Survey, Menlo Park, CA, Open File Report 2004-1271, 19 pp., 2004.

476 Manies, K. L., Harden, J. W., and Veldhuis, H.: Soil data from a moderately well and somewhat poorly
477 drained fire chronosequence near Thompson, Manitoba, Canada, U.S. Geological Survey, Menlo
478 Park, CA, Open File Report 2006-1291, 17 pp., 2006.

479 Manies, K. L., Harden, J. W., and Hollingsworth, T. N.: Soils, Vegetation, and Woody Debris Data from
480 the 2001 Survey Line Fire and a Comparable Unburned Site, US Geological Survey, 36 pp., 2014.

481 Manies, K. L., Harden, J. W., Fuller, C. C., Xu, X., and McGeehin, J. P.: Soil Data for a Vegetation
482 Gradient Located at Bonanza Creek Long Term Ecological Research Site, Interior Alaska, US
483 Geological Survey, 20 pp., 2016.

484 Manies, K. L., Fuller, C. C., Jones, M. C., Waldrop, M. P., and McGeehin, J. P.: Soil data for a
485 thermokarst bog and the surrounding permafrost plateau forest, located at Bonanza Creek Long Term
486 Ecological Research Site, Interior Alaska, Reston, VA, Report 2016-1173, 1-11 pp., 2017.

487 Michaelson, G. J., Ping, C.-L., and Clark, M.: Soil Pedon Carbon and Nitrogen Data for Alaska: An
488 Analysis and Update, *Open Journal of Soil Science*, 3, 11, 10.4236/ojss.2013.32015, 2013.

489 Mu, C., Zhang, T., Zhang, X., Li, L., Guo, H., Zhao, Q., Cao, L., Wu, Q., and Cheng, G.: Carbon loss and
490 chemical changes from permafrost collapse in the northern Tibetan Plateau, *Journal of Geophysical
491 Research: Biogeosciences*, 121, 1781-1791, 10.1002/2015JG003235, 2016.

492 Myers-Smith, I. H., Harden, J. W., Wilking, M., Fuller, C. C., McGuire, A. D., and Chapin III, F. S.:
493 Wetland succession in a permafrost collapse: Interactions between fire and thermokarst,
494 *Biogeosciences*, 5, 1273-1286, 2008.

495 Nalder, I. A., and Wein, R. W.: A new forest floor corer for rapid sampling, minimal disturbance and
496 adequate precision, *Silva Fennica*, 32, 373-381, 1998.

497 Neff, J. C., Harden, J. W., and Gleixner, G.: Fire effects on soil organic matter content and composition in
498 boreal Interior Alaska, *Canadian Journal of Forest Research*, 35, 2178-2187, 2005.

499 O'Donnell, J. A., Romanovsky, V. E., Harden, J. W., and McGuire, A. D.: The effect of soil moisture
500 content on the thermal conductivity of soil organic horizons from black spruce ecosystems in Interior
501 Alaska, *Soil Science*, 646–651, 10.1097/SS.0b013e3181c4a7f8, 2009.

502 O'Donnell, J. A., Harden, J. W., McGuire, A. D., Kanevskiy, M. Z., and Jorgenson, M. T.: The effect of
503 fire and permafrost interactions on soil carbon accumulation in an upland black spruce ecosystem of
504 interior Alaska: Implications for post-thaw carbon loss, *Global Change Biology*, 1461–1474,
505 10.1111/j.1365-2486.2010.02358.x, 2011.

506 O'Donnell, J. A., Harden, J. W., Manies, K. L., and Jorgenson, M. T.: Soil data for a collapse-scar bog
507 chronosequence in Koyukuk Flats National Wildlife Refuge, Alaska, 2008., U.S. Geological Survey,
508 Open-File Report, 14 pp., 2012.

509 O'Donnell, J. A., Harden, J. W., Manies, K. L., Jorgenson, M. T., Kanevskiy, M., and Xu, X.: Soil data
510 from fire and permafrost-thaw chronosequences in upland black spruce (*Picea mariana*) stands near
511 Hess Creek and Tok, Alaska, U.S. Geological Survey 16 p pp., 2013.

512 O'Neill, K. P., Kasischke, E. S., and Richter, D. D.: Environmental controls on soil CO₂ flux following
513 fire in black spruce, white spruce, and aspen stand of interior Alaska, *Canadian Journal of Forest
514 Research*, 32, 1525–1541, 2002.

515 Osterkamp, T. E., Jorgenson, M. T., Schuur, E. A. G., Shur, Y. L., Kanevskiy, M. Z., Vogel, J. G., and
516 Tumskey, V. E.: Physical and ecological changes associated with warming permafrost and
517 thermokarst in Interior Alaska, *Permafrost and Periglacial Processes*, 20, 235–256, 2009.

518 Ping, C., Boone, R. D., Clark, M. H., Packee, E. C., and Swanson, D. K.: State factor control of soil
519 formation in Interior Alaska, in: *Alaska's changing boreal forest*, edited by: Chapin Iii, F. S., Oswood,
520 M. W., Van Cleve, K., Viereck, L. A., and Verbyla, D. L., Oxford University Press, Oxford, 21-38,
521 2006.

522 Ping, C. L., Michaelson, G. J., Kane, E. S., Packee, E. C., Stiles, C. A., Swanson, D. K., and Zaman, N.
523 D.: Carbon Stores and Biogeochemical Properties of Soils under Black Spruce Forest, Alaska, *Soil
524 Science Society of America Journal*, 74, 969–978, 2010.

525 R Core Team: R: A language and environment for statistical computing, R Foundation for Statistical
526 Computing, Vienna, Austria, 2017.

527 Rand, J., and Mellor, M.: Ice-coring augers for shallow depth sampling, U.S. Army Cold Regions
528 Research and Engineering Laboratory, Hanover, New Hampshire, CRREL Report 85-21, 27 pp.,
529 1985.

530 Schadel, C., Bader, M. K. F., Schuur, E. A. G., Biasi, C., Bracho, R., Capek, P., De Baets, S., Diakova,
531 K., Ernakovich, J., Estop-Aragones, C., Graham, D. E., Hartley, I. P., Iversen, C. M., Kane, E.,
532 Knoblauch, C., Lupascu, M., Martikainen, P. J., Natali, S. M., Norby, R. J., O'Donnell, J. A.,
533 Chowdhury, T. R., Santruckova, H., Shaver, G., Sloan, V. L., Treat, C. C., Turetsky, M. R., Waldrop,
534 M. P., and Wickland, K. P.: Potential carbon emissions dominated by carbon dioxide from thawed
535 permafrost soils, *Nat. Clim. Chang.*, 6, 950-953, 10.1038/nclimate3054, 2016.

536 Schuur, E. A., McGuire, A. D., Schädel, C., Grosse, G., Harden, J. W., Hayes, D. J., Hugelius, G., Koven,
537 C. D., Kuhry, P., Lawrence, D. M., Natali, S. M., Olefeldt, D., Romanovsky, V. E., Schaefer, K.,
538 Turetsky, M. R., Treat, C. C., and Vonk, J. E.: Climate change and the permafrost carbon feedback,
539 *Nature*, 520, 171-179, 10.1038/nature14338, 2015.

540 Soil Classification Working Group: Canadian System of Soil Classification, 3rd ed., National Research
541 Council Canada Research Press, Ontario, 188 pp., 1998.

542 Soil Survey Division Staff: Examination and Description of Soil Profiles, in: *Soil survey manual*, edited
543 by: Ditzler, C., Scheffe, K., and Monger, H. C., USDA, Government Printing Office, Washington,
544 D.C., 1993.

545 Soil Survey Staff: Keys to soil taxonomy, 8th ed., Pocahontas Press, Blacksburg, Virginia, 599 pp., 1998.

546 Turetsky, M. R., Crow, S. E., Evans, R. J., Vitt, D. H., and Wieder, R. K.: Trade-offs in resource
547 allocation among moss species control decomposition in boreal peatlands, *Journal of Ecology*, 96,
548 1297-1305, 10.1111/j.1365-2745.2008.01438.x, 2008.

- 549 Turetsky, M. R., Kane, E. S., Harden, J. W., Ottmar, R. D., Manies, K. L., Hoy, E., and Kasichke, E. S.:
550 Recent acceleration of biomass burning and carbon losses in Alaskan forests and peatlands, *Nature*
551 *Geosciences*, 4, 27–31, 10.1038/NGEO1027, 2011.
- 552 Van Cleve, K., Oliver, L., Schlentner, R., Viereck, L. A., and Dyrness, C. T.: Productivity and nutrient
553 cycling in taiga forest ecosystems, *Canadian Journal of Forest Research*, 13, 747–766, 1983.
- 554 Woo, M. k.: Permafrost hydrology in North America, *Atmosphere-Ocean*, 24, 201-234,
555 10.1080/07055900.1986.9649248, 1986.
- 556 Yi, S., Manies, K., Harden, J., and McGuire, A. D.: Characteristics of organic soil in black spruce forests:
557 Implications for the application of land surface and ecosystem models in cold regions, *Geophysical*
558 *Research Letters*, 36, L05501, 10.1029/2008GL037014, 2009.

559

Figure 1. Location of the sites used in this study, all located within Interior Alaska. Regions, as ascribed in the dataset, are noted in red. Cities are written in yellow. (Map data: Google, 2020.)

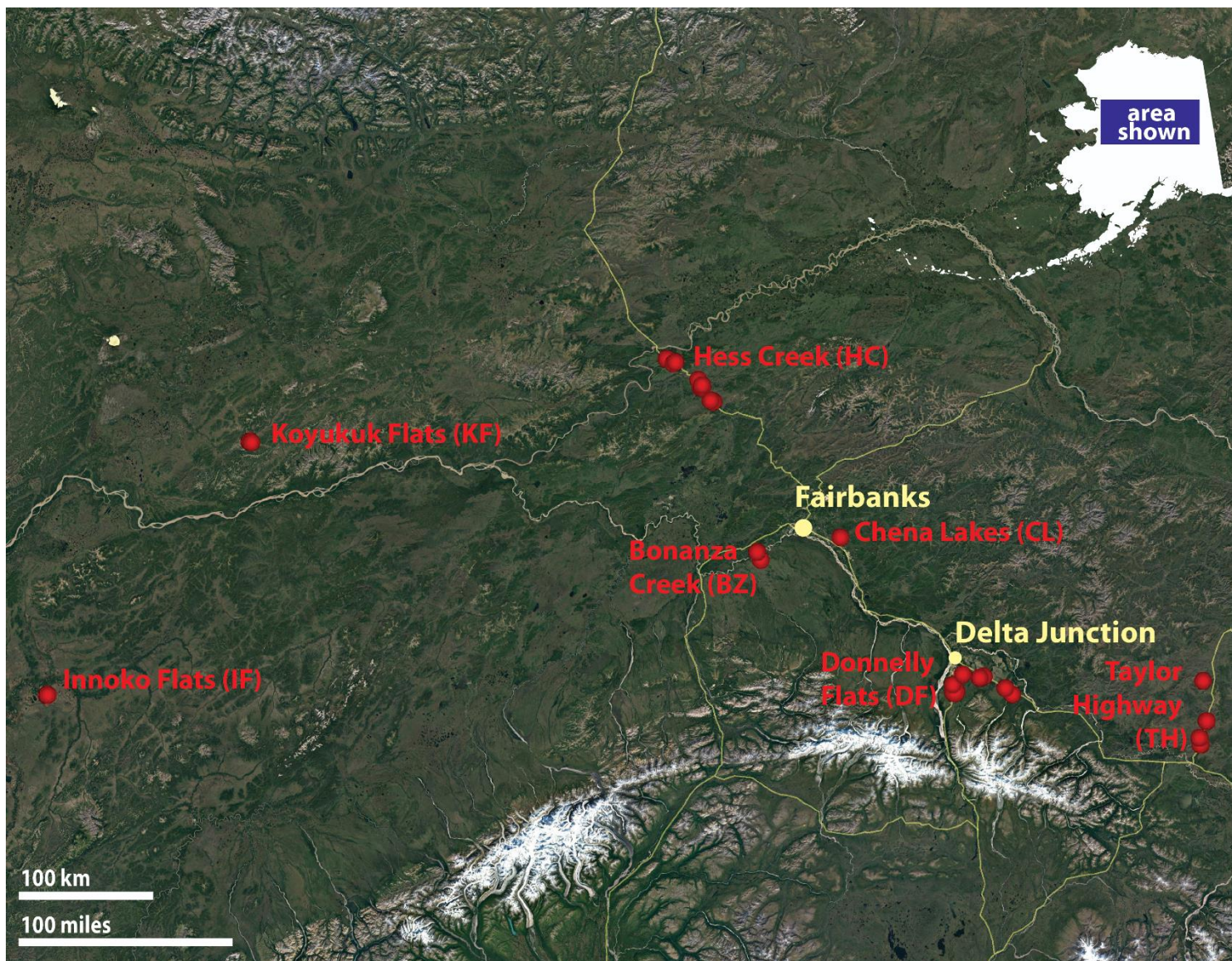
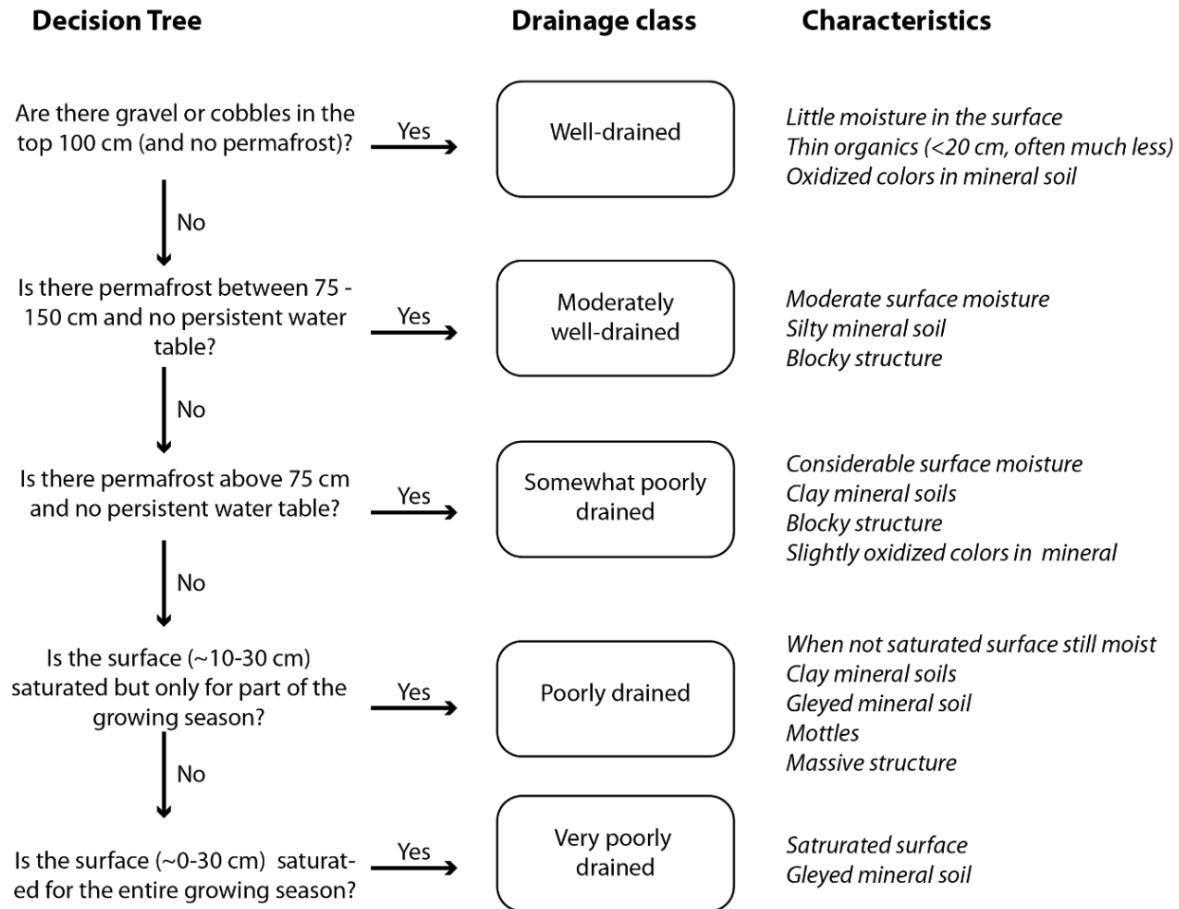


Figure 2. Soil drainage class decision tree. Beginning in the top left, if the soil meets the criteria, one has found the designated drainage class, having the characteristics located on the right. If the soil of interest does not meet the criteria, one moves down to the next drainage class to determine if its criteria is met. Drainage classes are also modified by slopes of greater than 5 % by moving up one drainage class.



Slope modifier:

If the slope of the site is greater than 5% the site should be better drained by one drainage class. For example, a somewhat poorly drained site would become a moderately well drained site.



well-drained



somewhat poorly drained



gleyed soil found in frozen poorly and very poorly drained soils

Figure 3. Trends in carbon and nitrogen density (g cm^{-3}) by horizon type using average values for bulk density, carbon, and nitrogen (Table 2). Horizon designations: L = live moss, D = dead moss, F = fibric, M = mesic, H = humic, Min = mineral.

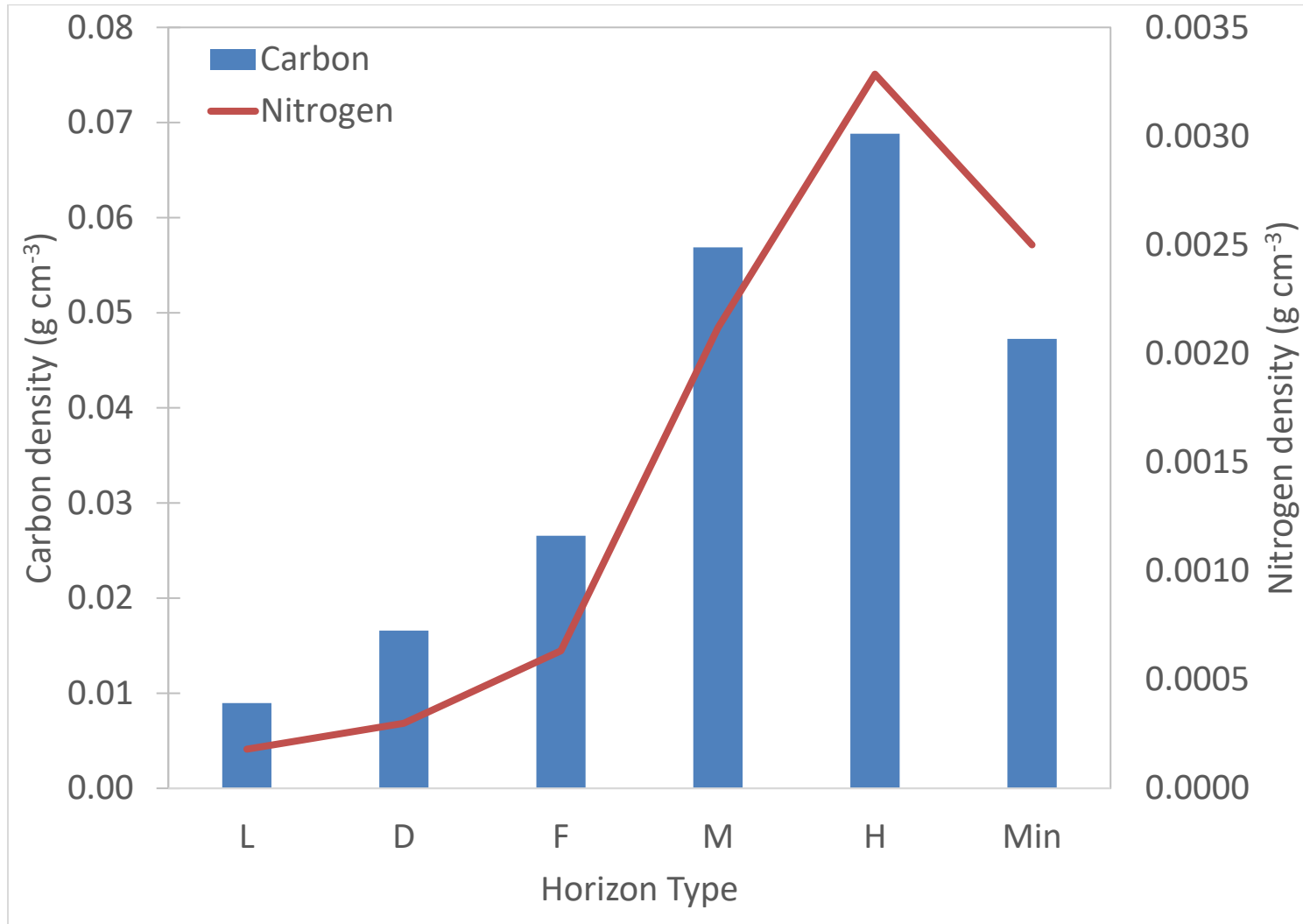


Table 1. A description of the soil horizons, as assigned by examining the composition of the soil horizon, including the degree of decomposition, color, and root abundance.

Horizon Type	Horizon Code	Description
Live moss	L	Live moss, which is usually green. This horizon generally also contains a small amount of plant litter. Plant material is completely undecomposed.
Dead moss	D	Moss that is dead and either undecomposed or slightly decomposed. Plant parts are easily identifiable. This horizon would be considered an O _i horizon in the U.S. soil system.
Fibric	F	Fibrous plant material that varies in the degree of decomposition (somewhat intact to very small plant pieces), but there is no amorphous organic material present. Very fine roots often make up a large fraction of this horizon. This horizon would be considered an O _i horizon in the U.S. soil system.
Mesic	M	This horizon is comprised of moderately decomposed material, with few, if any, recognizable plant parts other than roots. There is amorphous present within this horizon to varying degrees, but it is not smeary. This horizon is often considered an O _e horizon (U.S. soil system).
Humic	H	This organic horizon is highly decomposed and is mostly amorphous material. The soil in this horizon smears when rubbed and contains little to no recognizable plant parts. The H horizon is generally considered an O _a horizon (U.S. soil system).
Mineral	Min	Classified as an A, B, or C mineral soil (U.S. soil system), it contains less than 20-volume-percent organic matter, as judged in the field.

Table 2. Bulk density (g/cm³), C (%), N (%), C:N ratio, and thickness (cm) for the main horizon codes averaged across all drainage and age classes. Significant differences (p < 0.05) among the main six horizon codes are indicated with different letters. There are no thickness values for mineral soil because these results would reflect the thickness sampled, not the actual thickness of this horizon. Stdev is one standard deviation.

Horizon Code	Bulk Density (g/cm ³)			Carbon (%)			Nitrogen (%)			C:N			Thickness (cm)		
	mean	stdev	n	mean	stdev	n	mean	stdev	n	mean	stdev	n	mean	stdev	n
live moss (L)	0.022 ^a	(0.018)	138	41.7 ^a	(3.8)	145	0.84 ^a	(0.25)	145	53.8 ^a	(16)	141	2.5 ^a	(1.6)	136
dead moss (D)	0.039 ^b	(0.026)	540	42.6 ^a	(3.8)	538	0.77 ^a	(0.27)	537	62.1 ^a	(23)	541	13.9 ^b	(24.2)	157
fibric (F)	0.065 ^c	(0.041)	552	41.0 ^a	(5.6)	566	0.98 ^a	(0.42)	564	47.6 ^a	(17)	552	12.8 ^{bc}	(17.9)	221
mesic (M)	0.149 ^d	(0.077)	634	38.2 ^b	(6.8)	650	1.42 ^b	(0.54)	651	30.6 ^b	(13)	634	20.4 ^c	(40.3)	208
humic (H)	0.215 ^e	(0.096)	160	32.1 ^c	(6.6)	164	1.53 ^c	(0.44)	164	22.2 ^c	(6)	160	9.7 ^b	(11.3)	74
mineral (Min)	0.731 ^f	(0.380)	584	6.5 ^d	(6.2)	674	0.34 ^d	(0.32)	673	18.0 ^d	(7)	603	--	--	--
fibrous (D & F)	0.052	(0.037)	1092	41.8	(4.8)	1104	0.88	(0.37)	1101	54.6	(21)	1101	22.8	(41.1)	220
amorphous (M & H)	0.162	(0.085)	794	36.9	(7.2)	814	1.44	(0.52)	815	28.9	(12)	813	19.7	(27.7)	263

Table 3. Thickness (cm) of the main horizon codes by soil drainage and age class. The mineral soil horizon was not included in this table because the way in which we sampled mineral soil led to arbitrary thicknesses. Significant differences ($p < 0.05$) for horizon codes among drainage classes are indicated with different letters. Stdev is one standard deviation.

Horizon		Drainage					Age		
		Well drained	Moderately Drained	Somewhat Poorly Drained	Poorly Drained	Very Poorly Drained	New	Young	Mature
live moss (L)	mean stdev n	2.2 ^a (1.0) 6	2.2 ^a (0.8) 11	2.1 ^a (1.1) 75	1.5 ^a (0.7) 18	4.3 ^b (2.1) 26	1.0 ^{ab} (-) 2	2.6 ^a (2.) 42	2.4 ^b (1.2) 92
dead moss (D)	mean stdev n	3.3 ^a (1.6) 6	7.4 ^a (6.8) 19	7.6 ^a (10.8) 77	6.5 ^a (6.5) 21	38.1 ^b (40.8) 34	6.3 ^{ab} (4.5) 17	16.3 ^a (20.0) 42	14.1 ^b (27.5) 98
fibric (F)	mean stdev n	3.1 ^a (3.0) 11	9.6 ^{bc} (5.0) 18	7.9 ^b (5.2) 121	13.7 ^c (10.9) 45	40.2 ^d (38.7) 26	6.4 ^a (5.7) 65	19.8 ^b (31.8) 39	14.0 ^c (14.7) 117
mesic (M)	mean stdev n	2.8 ^a (1.3) 5	13.3 ^{ab} (17.6) 15	13.2 ^{ab} (38.0) 112	14.4 ^b (21.8) 39	57.2 ^c (53.4) 33	6.3 ^{ab} (3.5) 53	21.3 ^a (33.2) 50	27.2 ^b (50.8) 101
humic (H)	mean stdev n	none -- --	12.1 ^{ab} (15.4) 7	5.6 ^a (7.4) 38	7.4 ^a (3.4) 13	20.2 ^b (14.7) 16	4.3 ^a (3.2) 24	13.1 ^{ab} (12.7) 17	11.9 ^b (13.1) 33
fibrous (D & F)	mean stdev n	4.5 ^a (4.4) 12	14.8 ^b (8.1) 21	11.3 ^b (9.9) 135	14.8 ^b (11.2) 51	58.5 ^c (47.8) 40	7.1 ^a (6.3) 73	27.0 ^a (36.7) 54	22.9 ^b (26.9) 132
amorphous (M & H)	mean stdev n	2.8 ^a (1.3) 5	15.8 ^a (25.3) 18	14.3 ^a (38.4) 118	16.1 ^a (21.0) 41	63.2 ^b (51.5) 35	7.7 ^a (4.5) 56	23.0 ^a (33.5) 56	29.9 ^b (51.9) 105

Table 4. Physical and chemical properties of additional surface horizons. Number of observations, bulk density (g/cm³), C (%), N (%), C:N ratio, and thickness (cm) of non-main horizon codes. Values in parenthesis are standard deviations.

Horizon	N	Bulk density (g/cm³)	Carbon (%)	Nitrogen (%)	C:N Ratio	Thickness (cm)
ash	14	0.183 (0.155)	38.0 (14.4)	0.84 (0.34)	49 (20)	0.1 (-)
burned organics	99	0.122 (0.142)	38.6 (8.9)	1.07 (0.32)	99 (38)	1.8 (1.0)
lichen	31	0.034 (0.019)	40.3 (5.9)	0.76 (0.41)	69 (37)	4.1 (3.1)
litter	16	0.044 (0.018)	41.2 (3.1)	1.55 (0.52)	29 (10)	1.6 (0.9)

Table 5. Data columns found in megaAlaska_v11-2 for ScienceBase.csv. This datafile can be found at <https://doi.org/10.5066/P960N1F9>: Data Supporting Generalized models to estimate carbon and nitrogen stocks of organic layers in Interior Alaska.

Column Name	Units	Column Description
sampleID	--	The first four characters are based on the region and site. Then there is a space. Next the soil core number, followed by a period, and then the basal depth of the soil horizon.
depth	cm	Basal depth of the soil horizon
Hcode	--	Horizon code as determined from Table 1
Sample	--	Qualitative description of the soil horizon
date	mm/dd/yy	Date sample was taken
thickness	cm	Thickness of the soil horizon
BDall	g/cm ³	Bulk density, all soil
BDfine	g/cm ³	Bulk density, fines (soil particles > 2 mm and roots > 1 cm diameter excluded)
HtAboveMin	cm	Height of each basal depth above the organic-mineral soil boundary
carbon	%	Carbon concentration
nitrogen	%	Nitrogen concentration
13C	‰	Per mil (‰) value of delta ¹³ C
14C	‰	Per mil (‰) value of delta ¹⁴ C for bulk soil sample
LOI	%	Loss-on-ignition value
volume_method	--	Method used to sample soils volumetrically
region	--	Region within Alaska where the site is located (Figure 1)
site	--	Site where the core was taken
profile	--	Soil profile, or core, number
drainage	--	Soil drainage category (Figure 2)
standage	yrs	Age from last disturbance (fire or thaw)
ageclass	--	N = newly burned (< 5 yrs), Y= young (5-50 yrs), M = mature (>50 yrs)
SurfaceVeg	--	Types of vegetation found on the soil surface
SubbedBD	--	If Y the bulk density is not a measured value. Instead an average value was used.
SubbedC	--	If Y the carbon concentration is not a measured value. Instead an average value was used.
SubbedN	--	If Y nitrogen concentration is not a measured value. Instead an average value was used.
GroupedHcode	--	Horizon codes grouped into fewer categories
GroupedVeg	--	Surface vegetation grouped into fewer categories

Table 6: Data columns found in Site_GPS_coordinates_v2. This datafile can be found at <https://doi.org/10.5066/P960N1F9>: Data Supporting Generalized models to estimate carbon and nitrogen stocks of organic layers in Interior Alaska.

Column Name	Description
Region	Region within Alaska where the site is located (Figure 1)
Region Code	Two letter code for the region
Site	Site where the core was taken
Profile	Which soil profiles are located at this location - all indicates general coordinates for all soil profiles
Latitude	Latitude in decimal degrees
Longitude	Longitude in decimal degrees
Datum	Datum of the coordinates