



1 **Long-term vegetation monitoring in Great Britain - the**
2 **Countryside Survey 1978-2007 and beyond**

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30 **Abstract**

31 The Countryside Survey (CS) of Great Britain provides a unique series of datasets, consisting
32 of an extensive set of repeated ecological measurements at a national scale, covering a time
33 span of 29 years. CS was first undertaken in 1978 to monitor ecological and land use change
34 in Britain using standardised procedures for recording ecological data from representative 1
35 km squares throughout the country. The same sites, with some additional squares, were used
36 for subsequent surveys of vegetation undertaken in 1990, 1998 and 2007, with the intention of
37 future surveys. Other data, for example regarding soils, freshwater and habitat diversity and
38 extents, have also been sampled in the same locations on analogous dates. However, the
39 present paper describes only the vegetation surveys.

40 The survey design is based on a series of gridded, stratified, randomly selected 1 km squares
41 derived from a statistical environmental classification of Britain. 256 1 km sample squares
42 were included in the 1978 survey, 506 in 1990, 569 in 1998 and 591 in 2007. Initially each
43 square contained up to 11 dispersed vegetation plots but additional plots were later placed in
44 different features so that eventually up to 36 additional sampling plots were recorded, all of
45 which can be relocated where practically possible, providing a total of 16,992 plots by 2007.
46 This database of vegetation plots is a unique national resource providing the only
47 comprehensive quantitative ecological coverage of Britain, with a time-series of vegetation
48 samples dating back to 1978. Plots in different habitats, land cover types and landscape
49 features are included.

50 Although a range of analyses have already been carried out, with changes in the vegetation
51 being related to a range of drivers at local and national scales, there is a major potential for
52 further analyses, for example in relation to climate change.

53 Data from each of the survey years (1978, 1990, 1998, 2007) are available via the following
54 DOIs: *Countryside Survey 1978 vegetation plot data* (<https://doi.org/10.5285/67bbfabb-d981-4ced-b7e7-225205de9c96>), *Countryside Survey 1990 vegetation plot data*
55 (<https://doi.org/10.5285/26e79792-5ffc-4116-9ac7-72193dd7f191>), *Countryside Survey 1998*
56 *vegetation plot data*. (<https://doi.org/10.5285/07896bb2-7078-468c-b56d-fb8b41d47065>),
57 *Countryside Survey 2007 vegetation plot data* (<https://doi.org/10.5285/57f97915-8ff1-473b-8c77-2564cbd747bc>).

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61 **1 Introduction**

62

63 The Countryside Survey (CS) of Great Britain was initiated in the late 1970s to monitor
64 ecological and land cover change using quantitative and repeatable methods. The history of
65 the development of the methodology is given by Sheail and Bunce (2003). The survey is
66 based on 1 km squares as a convenient sized unit, which had previously been tested in
67 Cumbria (Bunce and Smith, 1978) and Shetland (Wood and Bunce, 2016) in the years
68 preceding the first survey in 1978. The survey design is based on a series of stratified,
69 randomly selected 1 km squares from across Britain, which numbered 256 in 1978, 506 in
70 1990, 569 in 1998 and 591 in 2007. The stratification used was the statistical land
71 classification of 1 km squares in GB as described in Bunce et al. (1996a), Bunce et al. (1996c)
72 and summarized in Section 2.

73 In the first survey, data were recorded from up to 11 vegetation plots of four different types,
74 distributed through each of the squares (which form the main subject of the present paper),
75 along with soil samples and land cover maps using standard classes which were later
76 converted into standard habitat categories (Wood et al., 2012). Subsequent surveys including
77 the vegetation component were undertaken in 1990, 1998 and most recently, 2007 (with an
78 additional habitat survey in 1984). During this period, additional vegetation plots have been
79 placed in different habitats and landscape features for policy objectives, eventually giving up
80 to 36 more plots per square. Varying numbers of each vegetation plot type were initially
81 placed in locations across each survey square according to rules outlined in Section 3. In
82 subsequent surveys, these plots are repeated in these same fixed positions, except those such
83 as on field margins, which are based on rules applied in the field. Details of the types of plot
84 employed are described below, with an average of 29 plots being completed in each sample
85 square.

86 The survey as a whole provides a wealth of unique ecological data, consisting of an extensive
87 range of measurements at a national scale, covering a time span of 29 years. In addition to the
88 vegetation plots described here, data are also recorded from linear features such as hedgerows,
89 landscape elements such as veteran trees, areal habitats, soils and aquatic invertebrates (see
90 (Carey et al., 2008).

91



92 **2 Survey design: site selection and stratification**

93

94 As a result of the earlier work carried out in the 1970s on a regional scale (Wood and Bunce,
95 2016; Bunce and Smith, 1978), a sample unit of 1km square was found to be an appropriate
96 size for CS. 1km square is small enough to survey in a reasonable period of time (1 week or
97 less) and yet large enough to contain sufficient environmental features to allow differentiation
98 between squares. 1 km square is also widely used in other European projects (for example in
99 Spain (Ortega et al., 2013)), although in countries with small scale landscapes, for example
100 Northern Ireland, 0.25 km square has been adopted (Cooper et al., 2009).

101 With over 240,000 1km squares in Great Britain, a sampling approach was essential and a
102 statistical environmental classification was constructed, from which the stratified, random
103 samples of 1km squares were taken. Due to the limitations of computing power and lack of
104 readily available data at the time, the classification initially only covered a subset of 1212
105 1km squares, rather than the whole of Britain, based on a 15x15 km grid drawn from the
106 National Grid defined by Britain's national mapping agency, the Ordnance Survey.

107 Altitude, climate, geology, human geography and location variables from each 1 km square
108 were analysed using Indicator Species Analysis (ISA, now TWINSPAN (Hill and Šmilauer,
109 2005) and stopped at 32 classes. These strata were then described based on the average
110 values of the environmental characteristics of the initial database, for example, altitude and
111 rainfall (Bunce et al., 1996b; Bunce et al., 1996c).

112 Having constructed this initial stratification, the number of samples to be surveyed in the first
113 (1978) survey was considered. Ideally, this number would depend on the size of the stratum
114 (i.e. how many 1 km squares of the class occurred in GB) and on the ecological variability
115 within the stratum. Preliminary work had suggested that for ecological surveys of this type, at
116 least eight samples per stratum were necessary. Eight 1 km squares were therefore selected at
117 random from each of the classes from the grid of classified squares. Thus the final sample for
118 the first GB survey was a gridded, stratified, random sample of 256 x 1 km squares.
119 Surveying commenced in 1977, although the majority of squares were surveyed in 1978.
120 Note that the location of the 1 km sample squares is not disclosed by agreement with land-
121 owners. This policy ensures that the squares do not attract additional research or land
122 management activity that could potentially undermine their status as an unbiased,



123 representative sample of the British countryside. The majority of the land in the sample
124 squares is in private ownership.

125 A primary objective of the methodology is to reduce variation, as the classification divides
126 the population into discrete strata which are then used to derive samples from which
127 ecological parameters such as vegetation can be recorded. By using this statistically robust
128 method, it is then possible to scale up the results from the sample sites to describe the entire
129 population.

130 By 1990, all 1 km squares in GB were classified into the same set of strata, which was not
131 considered possible at the start of the 1970s. Known as the ‘Institute of Terrestrial Ecology
132 (ITE) Land Classification of Great Britain’ (Bunce et al., 1990, 1996b, a), it has developed
133 over the 30 year period (Sheail and Bunce, 2003). The most recent modifications largely
134 concern the incorporation of the requirement for country level reporting, separating Scotland
135 (in 1998) and Wales (in 2007). The basic stratification still underpins the CS and the latest
136 development of the original Land Classification (Bunce et al., 2007) consists of 45 classes (or
137 strata), and is illustrated in Figure 1, along with the sampling framework (Figure 2).

138

139 *Figure 1. ITE Land Classification, 2007*

140 *Figure 2. Map of Countryside Survey sampling locations across Britain*

141

142 **3 Sampling sites and plots**

143

144 Initially, vegetation and soil data were recorded from five dispersed random (‘X’) plots in
145 each 1 km square, which were located using a restricted randomization procedure designed to
146 reduce auto-correlation. Vegetation was sampled from a further six plots placed along linear
147 features (two Hedgerow (‘H’) plots, two Streamside (‘S’) plots, two Roadside (‘R’) plots).
148 Plots have never been placed in built-up areas or below the Mean High Water mark, and are
149 only sited where the landowner has given permission. The types and total numbers of plots
150 have increased over time from 1978 to 2007 along with the total number of CS 1 km squares
151 surveyed. The total number of plots within squares varies depending on the landscape type
152 and range of landscape features. Plots differ in size depending upon their type (Table 1). By
153 2007, the mean number of plots per square was 29 (min. 2, max. 47). The locations of all
154 plots were mapped, together with measurements to local features thus allowing them to be



155 found again and re-recorded in the same place. Additional information ensuring the highest
 156 degree of accuracy when re-finding plots began to be recorded in the 1990 survey, as
 157 described in Section 4.2. The same plot locations have been repeated in all subsequent
 158 surveys (where appropriate), with additions.

159

160 *Table 1. Summary of vegetation plot types, sizes and numbers.*

161

Code	Name	Other names	Where	Size	No. per square
<i>Areal plots</i>					
X ¹	Large	'Wally plot' Main	Random points in open polygons	200 m ²	5
Y ²⁺⁴	Small	Targeted Habitat	Uncommon vegetation types and in 2007, Priority Habitats	4 m ²	Up to 5
U ³	Unenclosed		Unenclosed Broad Habitats	4 m ²	Up to 10
<i>Linear plots</i>					
B ²	Boundary		Adjacent to field boundaries and paired with X plots	10 x 1 m	5
A ³	Arable		Arable field edges centred on each B plot	100 x 1 m	Up to 5
M ⁴⁺⁵	Margin		Field margins	2 x 2 m	Up to 15
H ¹	Hedgerow		Alongside hedgerows	10 x 1 m	2
D ³	Hedgerow diversity		Hedgerows/Woody Linear Features	30 x 1 m	Up to 10
S ¹ /W ²	Streamside		Alongside water courses	10 x 1 m	5
R ¹ /V ²	Roadside		Alongside roads and tracks	10 x 1 m	5

162 ¹ first recorded in 1978, ² first recorded in 1990, ³ first recorded in 1998, ⁴ first recorded in 2007, ⁵ if
 163 there are 5 A plots in a square and wide margins

164

165 4 Data collected

166 The vegetation survey involves recording plant species presence and abundance in different
 167 sizes and types of vegetation plot. In each vegetation plot, a complete list of all vascular
 168 plants and a selected range of the more easily identifiable bryophytes and macro-lichens is



169 made (with the exception of D plots, which record woody species only). Predetermined
170 combinations of species are recorded as aggregates reflecting known difficulties in their
171 separation in the field (refer to (Maskell et al., 2008)). Cover estimates are made to the nearest
172 5% for all species reaching at least an estimated 5% cover. Presence is recorded if cover is
173 less than 5%. Canopy cover of overhanging trees and shrubs is also recorded even if
174 individuals were not rooted within the plot. Additionally, general information about the plot
175 is recorded to provide a unique reference for analytical purposes as well as habitat
176 characteristics.

177

178 **4.1 Plot Types**

179

- 180 • ***X Plots - Large or Main Plots***

181 The X plots are large nested plots designed to provide a random sample in proportion to the
182 extent of the different vegetation types in each square and therefore in the wider countryside.
183 X plots were pre-located before going into the field, with one plot being randomly placed into
184 one of 5 equal sectors dividing the 1km survey square. X plots typically sample the most
185 common vegetation types. The X plot is 200m² (14.14 x 14.14m); the large size was adopted
186 to obtain the maximum number of species within the study area and is not therefore intended
187 to fit into phytosociological criteria of homogeneity. The methodology was originally
188 produced for woodlands as described by Bunce and Shaw (1973) and was also used and found
189 appropriate for strategic ecological survey, as described by Bunce and Smith (1978). The
190 design of plot not only aids a systematic search of the vegetation present but is
191 straightforward to set out in the field, and ensures a standard area of the plot is covered on
192 every occasion. The plot is set up by using a centre post and four corner posts, with a set of
193 four strings tagged with markers at specified distances. The tagged strings form the diagonals
194 of the square (as shown in Figure 3). The diagonals should be orientated carefully at right
195 angles and the plot should be orientated with the strings on the north/south, east/west axes.
196 The different nested plots shown in Figure 4 are marked by different coloured strings on the
197 appropriate position of the diagonal. The design is to ensure that the whole plot is covered, as
198 tests had shown that a major difference between observers was their search routine.

199 Within the plot shown in Figure 4, the first nest of the plot (2x2m) is searched first. This
200 procedure is then repeated for each nest of the quadrat, increasing the size each time and only



201 recording additional species discovered in each larger nest (Figure 4). In the final nest (the
202 whole 200m² plot), the percentage cover (to the nearest 5%) of each species is also estimated.
203 Estimates of cover for litter, wood, rock and bare ground are also included where present. In
204 2007, an additional 1m² nest (not shown in Figure 4) was introduced, in order allow joint
205 analysis of 1m² plots being recorded in parallel as part of agri-environment scheme
206 monitoring programs. This nest is located in the northernmost corner of the inner 4 m² nest
207 (named nest '0'). Vegetation height, aspect and slope are also recorded. Soil samples are
208 also taken at the same time, at the site of these plots; the procedure used for recording soil
209 samples is given by Emmett et al. (2008) and is outwith the remit of the present paper.

210 In arable fields where full access is not possible, species records are made from plots taken
211 from an estimated 14m square, starting three metres into the crop in order to avoid any edge
212 effect. Access is made using drill lines where possible in order to causing minimum
213 disturbance to the crop. Overall cover is also estimated as in other habitat types. The relative
214 uniformity of species within crops led to the adoption of this approach and the subsequent
215 changes observed in species numbers in arable fields justified its use.

216

217 *Figure 3. X plot construction*

218 *Figure 4. Layout of vegetation X plot.*

219

220 • **Y Plots - Small, Targeted or Habitat Plots**

221 These are small (2 x 2m) plots located in less common habitat types, usually of conservation
222 interest, often occurring in small patches not sampled by other plot types. In 2007, additional
223 Y plots also were placed in Priority Habitat (Maddock, 2008) patches that had also not been
224 sampled by any other plot in the square. The Y plots are therefore important in sampling
225 fragments of semi-natural habitat particularly in lowland landscapes where patches may be
226 small and embedded in a matrix of intensive farmland. These plots are placed randomly by
227 surveyors in suitable patches of habitat (based on rules described in Maskell et al. (2008)). Of
228 all the plots recorded, they are most similar to the approach taken when positioning quadrats
229 during National Vegetation Classification (NVC) (Rodwell, 2006) survey where the location
230 of the plot is designed to represent a vegetation unit perceived to be floristically distinct and
231 homogenous. However, protocols for locating Y plots from 1998 onwards stipulated random
232 location from within a larger extent of habitat in the 1 km square or from a number of patches



233 representing the mapped habitat type. The validity of statistically analysing plots located with
234 a degree of subjectivity is an ongoing matter of debate (see for example Lájér (2007);Palmer
235 (1993) for an illustration of analytical problems but also Ross et al. (2010) for counter-
236 argument and examples of analysing temporal change in subjectively located samples).

237

238 • ***U Plots - Unenclosed Plots***

239 These plots were introduced into the CS methodology for the first time in the 1998 survey to
240 characterise the unenclosed Broad Habitats (Jackson, 2000) - these being Calcareous and
241 Acid grasslands; Bracken; Dwarf shrub heath; Fen, marsh and swamp; Bog; Montane; Supra
242 littoral rock and sediment; and Inland rock. Up to ten plots were established in any
243 unenclosed habitat types that occurred within the square (proportional to area), again placed
244 randomly by surveyors. The plots are 2x2 m in all instances, regardless of the habitat in which
245 they are located.

246

247 • ***B Plots - Boundary Plots***

248 Boundary linear plots are recorded at a position on the boundary closest to each X plot and on
249 a cardinal axis from it (i.e. north, south, east or west). A boundary is taken to be any linear
250 physical feature that has a length greater than 20m and which is an interface between the land
251 cover of the 200m² X plot and any other land cover type. This might include a hedge, wall,
252 fence, ditch or embankment. These are linear 10 x 1m plots.

253

254 • ***A Plots - Arable Field Margin Plots***

255 Arable field margin plots were recorded for the first time in the 1998 survey. The purpose of
256 establishing the plots was to record the arable weed population at the edge of cultivated fields
257 and any subsequent changes. It is established that non-crop plant diversity increases towards
258 the edge of a field. Field edges therefore contribute an important source of biodiversity not
259 present in the arable X plots, which cover the overall composition of arable crops because as
260 described above, the margin is specifically excluded. The uptake of ‘conservation headland’
261 options for arable field management under agri-environment schemes may further enhance
262 species diversity in A plots. The plots are 100m long by 1m wide and located adjacent to B



263 plots which border arable fields, up to a maximum of five per square. They always sample
264 the first 1m of cultivated land moving away from the perennial dominated margin.

265

266 • ***M Plots - Margin Plots***

267 M plots are small (2 x 2m) square plots and were new in the 2007 survey. They are
268 associated with B plots where an A plot is present, and the number depends on the widths of
269 the margins present, with up to three per field. They are designed to record the quality of new
270 arable field margins that form part of the agri-environment agreements on farms and other
271 margins put in without agri-environment support. These margins are additional to the cross
272 compliance margin (not relevant in Wales) which is a 2m margin measured from the centre of
273 the hedge. The most common types of margin likely to be encountered are perennial grass
274 margins, with or without supplementary wildflowers. Other rarer types include, uncropped
275 strips, usually cultivated each year (regenerating from the seedbank); wild bird seed cover e.g.
276 kale, quinoa; pollen and nectar mixes, usually with a high proportion of legumes.

277

278 • ***H Plots - Hedgerow Plots***

279 H plots are linear 10 x 1m plots running alongside managed woody linear features ('WLF's,
280 hedgerows). Within H plots, species associated with the managed woody linear features are
281 recorded. When first recorded in 1978, the plot positions were located as close as possible to
282 the two X plots in each square which were furthest apart.

283

284 • ***D Plots - Hedgerow Diversity Plots***

285 Hedgerow diversity plots were recorded for the first time in 1998. The overall purpose was to
286 set up a baseline of plots to monitor woody species diversity in woody linear features. One D
287 plot is placed on each woody linear feature in the square, up to a maximum of ten plots. As
288 well as providing information on woody species diversity, the data collected in D plots also
289 helps to provide an assessment of the condition of hedgerows and other woody linear features
290 by providing vital information about the size of the woody linear features, gappiness, levels of
291 disturbance and species composition. Each plot is 30 m long and includes the full width of the
292 WLF.

293



294 • ***S/W Plots - Streamside Plots***

295 The term “Streamside plot” denotes linear plots which lie alongside running water features
296 (mainly rivers and streams, but also canals and ditches). The S and W prefixes refer to the
297 different origins of the plots: two Streamside (S) plots were established in the 256 1 km
298 squares in 1978, located as close as possible to the two X plots in each square which were
299 furthest apart. W plots were up to three additional Waterside plots, placed in all squares in
300 1990 to increase representation of other waterside types. These are linear 10 x 1m plots.

301

302 • ***R/V Plots - Roadside and Verge Plots***

303 The term “Roadside plot” denotes those linear plots which lie alongside transport routes
304 (mainly roads and tracks). The R and V prefixes refer to the different origins of the plots: two
305 Roadside (R) plots were established in the 256 1 km squares in 1978, located as close as
306 possible to the two X plots in each square which were furthest apart. V plots are up to three
307 additional Verge plots first placed in 1 km squares in 1990 to increase representation of other
308 transport types. These are linear 10 x 1m plots.

309

310 **4.2 Plot Relocation**

311

312 To analyse change, it is important to relocate the exact same sampling plot locations in
313 successive surveys. The data from the repeated vegetation plots provide a globally unique
314 dataset allowing large-scale yet fine-grained change in overall vegetation and the state or
315 condition of the Broad and Priority Habitats over time to be documented at four points over
316 the last 29 years. There are no other national data sets that cover entire landscapes, including
317 constituent habitats over such a long period of time. In practice, there are actually very few
318 long term studies of vegetation change. Those existing are usually either opportunistic,
319 because some local recording has given a precise location, for example Dunnett et al. (1998)
320 on a roadside verge in Bibury in Gloucestershire; or pertain to specific habitats, such as the
321 Park Grass Experiment at Rothamsted (Silvertown et al., 2006).

322 During the surveys, plot locations have been recorded on paper using a sketch map with
323 measurements from distinguishing landscape features (Figure 5), and by taking at least two
324 photographs (see Figure 6 for an example), preferably also including key landscape features



325 in proximity to the plot. In addition to these, permanent metal plates or wooden stakes were
326 introduced in the 1990 survey. In 1998, a GPS position was recorded in some remote squares,
327 which assisted locating plots again in 2007. In 2007, the plot locations were recorded via the
328 ruggedized field computers using the in-built GPS (where a GPS signal was available).
329 Surveyors are also able to record whether the plots have been re-found adequately or
330 otherwise. Circumstances where a plot may not be repeated might include an area becoming
331 built-up, or a feature having been removed.

332

333 *Figure 5. Example of a plot sketch map*

334 *Figure 6. Example of a plot photograph*

335

336 **5 Data quality**

337

338 Each field survey was carried out by teams of experienced botanical surveyors, and was
339 preceded by an intensive training course, ensuring high standards and consistency of
340 methodology, identification and recording across CS according to criteria laid out in the field
341 handbooks (Maskell et al., 2008; Barr, 1998; Barr, 1990; Bunce, 1978). During the surveys,
342 survey teams were initially supervised and later monitored by experienced project staff in
343 order to control data quality.

344 Data were recorded on waterproof paper sheets in 1978, 1990 and 1998 and were
345 consequently transferred from the original field sheets to spreadsheets, using a ‘double-punch’
346 method to minimise errors in data entry. They were checked using range and format checks,
347 and corrected to produce a final validated copy. In 2007, a new electronic data capture
348 method was developed by the Centre for Ecology & Hydrology and used in CS for the first
349 time. The move to electronic methods created greater efficiency in terms of data entry and
350 also eliminated a potentially significant source of error. Improvements to data quality also
351 resulted from the inclusion of mandatory data entry fields for each plot.

352 In terms of assessing the actual level of botanical expertise in the field surveys, Quality
353 Assurance (QA) reports were completed by independent botanists for the surveys in 1990,
354 1998 and 2007 (Prosser and Wallace, 2008, 1999, 1992). These reports have been a vital tool
355 in assessing and validating the quality of the botanical record in each CS. Paired species
356 records from a subset of plots (the QA plots) have been analysed in a number of ways to



357 measure the consistency of recording effort within each survey. In all three surveys the QA
358 assessors found more species than the CS field teams, yet in both the 1990 and 1998
359 assessments, the results showed that there was no bias in the species composition of the
360 vegetation recorded, as described by DCA analysis, despite differences in species richness. In
361 2007, the QA analysis appeared to show a decline in the quality of botanical recording.
362 However this was possibly due to the much less comprehensive recording of common
363 bryophytes than in previous surveys, and subsequent analyses determined the bias was not
364 significant (Scott et al., 2008). Errors attributable to use of the electronic data capture
365 software were minor and not significant (Prosser and Wallace, 2008).

366

367 **6 Methodological Development**

368

369 The now established method of CS, using a stratified random series of samples, was
370 developed over two decades by what was then the Institute of Terrestrial Ecology as described
371 by Sheail & Bunce (2003). The first national series of stratified random samples was the
372 1971 Woodland Survey (Wood et al., 2015) and strategic sampling at the landscape level was
373 then tested successfully in regional surveys in Cumbria and Shetland (Bunce and Smith,
374 1978; Wood and Bunce, 2016). These methods have now been proven as a successful national
375 vegetation monitoring strategy incorporating four surveys across nearly 30 years. Minor
376 modifications to the methods have more recently been used for a comprehensive ecological
377 survey of Wales (2013-), the Glastir Monitoring and Evaluation Programme (GMEP)
378 (Emmett and GMEP team, 2014).

379 Since the first survey in 1978, the methods have gradually developed to incorporate
380 contemporary technologies, for example, the introduction of GPS in 1998, and the use of
381 ruggedised field computers with internal GPS to record the location and species composition
382 of the vegetation plots in 2007. Over time, the development of Geographical Information
383 Systems (GIS) has greatly facilitated both the efficiency of storage of ecological spatial data,
384 and also the types of analyses that can be undertaken. It is now possible to perform much
385 wider analyses than previously, using a range of ancillary explanatory datasets, as described
386 in the Integrated Assessment Report for the Countryside Survey (Smart et al., 2010a). The
387 underlying principles of the Countryside Survey methodology provide an ideal framework for



388 the planning of large scale monitoring, not only in Britain, but across Europe and worldwide,
389 as discussed in Wood and Bunce (2016).

390 It has now been a decade since the last survey, and current funding constraints mean that the
391 traditional cycle of large one-off national surveys taking place roughly one year in every
392 decade is likely to need revising. Various options are available for repeating all or parts of the
393 survey. A rolling program over several years is attractive because it spreads the financial
394 load. It also allows inter-annual effects of differences in the weather and variation in recorder
395 effort to be more robustly estimated and separated from long-term trends. A Markov chain
396 approach could be used to examine possible outcomes from the time series of plots (for
397 example, Balzter (2000)). Between 2013 and 2016, CS methods have already been applied in
398 an annual rolling program to monitor the effects of the Glastir agri-environment scheme
399 across Wales (<https://gmep.wales/>).

400 Plot numbers could be rationalised according to the desired results. Using previous data, it is
401 possible to identify the optimal numbers of plots required by plot type, habitat type and region
402 in order to provide data on specific criteria, for example, species richness change at GB level
403 by plot type. Less costly options for maximising the use of the existing surveys in future
404 surveillance have been suggested as part of the Future Options review for national monitoring
405 in Wales (Emmett et al., 2016a). However, the feasibility of these options has yet to be
406 determined.

407

408 **7 Use of the data**

409

410 The Countryside Survey provides a unique and well utilised resource, offering potential for a
411 wide range of analyses at different temporal and spatial scales. A major benefit of the
412 programme is the co-registration of a wide range of recorded ecological variables (i.e. soil,
413 vegetation, habitats, freshwater). In parallel to its direct policy application, a vibrant and
414 productive research agenda has used CS vegetation data often in combination with other
415 datasets to produce improved understanding about the significance and causes of large-scale
416 but finely resolved ecological change in Britain. Questions can be broadly categorised as:
417 ‘What has changed and where?’, ‘What are the drivers of change?’, ‘Is the change important?’
418 and ‘Can we use forecast future change?’.

419

420 **7.1 Key findings**

421

422 Key findings and fundamental questions about the extent and condition of terrestrial habitats
 423 are addressed in the reporting round to policy makers that has followed each survey (e.g.
 424 (Haines-Young et al., 2000; Carey et al., 2008; Smart et al., 2009 ;Norton et al., 2009; Emmett
 425 et al., 2010; Bunce, 1979; Barr et al., 1993) and elsewhere (Smart et al., 2003; Norton, 2012).
 426 Overall changes in plant species richness formed part of a trend in species loss (8%) across
 427 Great Britain between 1978 and 2007. Woody species increased in vegetation associated with
 428 landscape boundaries by 14% between 1998 and 2007 and by nearly 80% in Great Britain
 429 between 1978 and 2007.

430 The most commonly recorded species in CS, ryegrass (*Lolium perenne*), was the same in
 431 2007 as in both 1998 and 1990. The top ten most commonly recorded species in 2007 also
 432 included stinging nettle (*Urtica dioica*), hawthorn (*Crataegus monogyna*), and bramble
 433 (*Rubus fruticosus*) all of which increased between 1998 and 2007.

434 Long-term change in vegetation from 1978 to 2007 has also been assessed using a range of
 435 condition measures (Table 2). In open countryside in Great Britain, between 1998 and 2007
 436 plant species that prefer wetter conditions increased while those preferring fertile soils and
 437 high pH decreased. In the period 1978 to 2007, an increase in species preferring wetter
 438 conditions was the most consistent signal in plots sampling different parts of the landscape
 439 across all countries.

440

441 *Table 2. Change in the characteristics of all types of vegetation in 200m² Main Plots in Great*
 442 *Britain between 1978 and 2007. Arrows denote significant change (p<0.05) in the direction*
 443 *shown.*

444

Vegetation Condition Measures	Mean Values				Direction of significant changes 1978 - 2007
	1978	1990	1998	2007	GB
Species Richness (No. of Species)	17.1	16.5	16.2	15.7	↓
Light Score	6.98	6.95	6.95	6.95	
Fertility Score	4.53	4.64	4.61	4.55	
Ellenberg pH Score*	5.07	5.17	5.14	5.09	



	Moisture Score	5.75	5.71	5.77	5.82	↑
445	*Ellenberg (1988)					
446						
447	7.2 Wider uses of data to date					
448						
449	After CS in 2007, the data continued to have a substantial impact, contributing to many areas					
450	of the UK National Ecosystem Assessment (NEA) (Watson et al., 2011), which articulated					
451	ecological status and change in terms of ecosystem services (ES). This was the first analysis					
452	of the UK's natural environment in terms of the benefits it provides to society and continuing					
453	economic prosperity. Soils, vegetation, headwater stream and land-cover data from					
454	Countryside Survey were also jointly analysed with a range of explanatory variable datasets					
455	to produce new indicators and analysis of potential ecosystem service delivery in the					
456	Integrated Assessment project that marked the final phase of reporting after the 2007 survey					
457	(Smart et al., 2010a;Norton et al., 2012).					
458	Subsequently, CS plot data has been used in conjunction with Land Cover Map data (Morton					
459	et al., 2011) and wider environmental datasets as part of a natural capital mapping tool which					
460	has been used, alongside other modelling techniques, to produce maps of natural capital for					
461	policy makers (https://eip.ceh.ac.uk/naturalengland-ncmaps) and to help in understanding the					
462	factors influencing spatial differences in ES delivery (Henrys et al., 2015a;Norton et al.,					
463	2016). Analysis demonstrated fundamental trade-offs between ecosystem productivity and					
464	soil carbon concentration while a range of biodiversity indicators appeared to peak at					
465	intermediate levels of productivity (Maskell et al., 2013). The novel inclusion of dynamic					
466	ecosystem model estimates of productivity provided both the foundation and research					
467	direction for ongoing work that has sought to develop dynamic models of natural capital and					
468	ecosystem service delivery (Emmett et al., 2016b;Smart et al., 2017;Rowe et al., 2016).					
469	CS datasets have also made a unique contribution to the development of plant species niche					
470	models for ecosystem dominants and many rare species in Britain (Hill et al., 2017;Henry et					
471	al., 2015b;Smart et al., 2010c;Smart et al., 2010b). The policy motivation for this originally					
472	was detection and modelling of the effects of atmospheric pollutant deposition (De Vries et					
473	al., 2010;Stevens et al., 2016).					
474	The statistically robust, national scale of the CS vegetation data makes it ideally placed to					
475	detect realistically scaled relationships between global change drivers, such as pollutant					



476 deposition (for example van den Berg et al. (2016);Maskell et al. (2010);Smart et al.
477 (2005a);Stevens et al. (2009)) as well as other drivers of eutrophication and land management
478 change (Smart et al., 2012;Smart et al., 2002;Smart et al., 2003;Smart et al., 2005b;Smart et
479 al., 2006a;Smart et al., 2006b). While research into the causes and consequences of
480 eutrophication was a response to clear policy interest, analysis of CS vegetation data has also
481 contributed evidence in response to concerns over the causes and consequences of loss of
482 pollinators in North West Europe and Britain (Smart et al., 2000;Carvell et al., 2006;Baude et
483 al., 2016).

484 Habitat specific studies, such as those relating to woodlands (for example (Petit et al.,
485 2004;Kimberley et al., 2013;Kimberley et al., 2016) and hedgerows (McCollin et al.,
486 2000;Garbutt and Sparks, 2002;Critchley et al., 2013)) have been facilitated through the use
487 of CS data. Interesting conclusions have been made through use of the data with regard to
488 increasing numbers of non-native invasive species (Chytrý et al., 2008;Maskell et al., 2006).

489 The results and information derived from CS can often be set into wider contexts, for
490 example, European (Chytrý et al., 2008;Metzger et al., 2013), or in relation to other
491 monitored datasets (Rose et al., 2016;Scott et al., 2010;Carey et al., 2002;Rhodes et al., 2015).
492 Data from the 1990 survey were used in the development of a statistically based British
493 vegetation classification, termed the Countryside Vegetation Classification (CVS) as
494 described in Bunce et al. (1999). This led to the development of a computer system termed
495 MAVIS (Modular Analysis of Vegetation Information System), enabling classification of any
496 lists of species from plots into the CVS but also into the phytosociological classes of the
497 National Vegetation Classification Rodwell (Rodwell, 2006). The software is publicly
498 available (Smart and DART Computing, 2017).

499

500

501 **8 Data availability**

502

503 The datasets have been assigned Digital Object Identifiers and users of the data must
504 reference the data as follows:

505



- 506 • Barr, C.J.; Bunce, R.G.H.; Smart, S.M.; Whittaker, H.A. (2014). Countryside Survey
507 1978 vegetation plot data. NERC Environmental Information Data Centre.
508 <https://doi.org/10.5285/67bbfabd-d981-4ced-b7e7-225205de9c96>
509
- 510 • Barr, C.J.; Bunce, R.G.H.; Gillespie, M.K.; Hallam, C.J.; Howard, D.C.; Maskell,
511 L.C.; Ness, M.J.; Norton, L.R.; Scott, R.J.; Smart, S.M.; Stuart, R.C.; Wood, C.M.
512 (2014). Countryside Survey 1990 vegetation plot data. NERC Environmental
513 Information Data Centre. [https://doi.org/10.5285/26e79792-5ffc-4116-9ac7-
514 72193dd7f191](https://doi.org/10.5285/26e79792-5ffc-4116-9ac7-72193dd7f191)
515
- 516 • Barr, C.J.; Bunce, R.G.H.; Gillespie, M.K.; Howard, D.C.; Maskell, L.C.; Norton,
517 L.R.; Scott, R.J.; Shield, E.R.; Smart, S.M.; Stuart, R.C.; Watkins, J.W. ; Wood, C.M.
518 (2014). Countryside Survey 1998 vegetation plot data. NERC Environmental
519 Information Data Centre. [https://doi.org/10.5285/07896bb2-7078-468c-b56d-
520 fb8b41d47065](https://doi.org/10.5285/07896bb2-7078-468c-b56d-fb8b41d47065)
521
- 522 • Bunce, R.G.H.; Carey, P.D.; Maskell, L.C.; Norton, L.R.; Scott, R.J.; Smart, S.M.;
523 Wood, C.M. (2014). Countryside Survey 2007 vegetation plot data. NERC
524 Environmental Information Data Centre. [https://doi.org/10.5285/57f97915-8ff1-473b-
525 8c77-2564cbd747bc](https://doi.org/10.5285/57f97915-8ff1-473b-8c77-2564cbd747bc)
526

527 The datasets are available from the CEH Environmental Information Data Centre Catalogue
528 (<https://catalogue.ceh.ac.uk>). Datasets are provided under the terms of the Open Government
529 Licence ([http://eidchub.ceh.ac.uk/administration-folder/tools/ceh-standard-licence-texts/ceh-
530 open-government-licence/plain](http://eidchub.ceh.ac.uk/administration-folder/tools/ceh-standard-licence-texts/ceh-open-government-licence/plain) , [http://www.nationalarchives.gov.uk/doc/open-government-
531 licence/version/3/](http://www.nationalarchives.gov.uk/doc/open-government-licence/version/3/)). The metadata is stored in the ISO 19115 (2003) schema (International
532 Organization for Standardization, 2015) in the UK Gemini 2.1 profile (UK GEMINI, 2015).
533 Users of the datasets will find the following documents useful (supplied as supporting
534 documentation with the datasets): The Sampling Strategy for Countryside Survey (Barr and
535 Wood, 2011) and the Field Survey Handbooks (Barr, 1990;Barr, 1998;Bunce, 1978;Maskell
536 et al., 2008).



537

538 **9 Conclusions**

539 The vegetation data recorded during the Countryside Survey of Great Britain are an
540 invaluable national resource, which, over the years, has been exploited in a large number of
541 ways. The data are collected in a statistically robust and quality controlled manner, follow
542 standard, repeatable methods and cover wide temporal and spatial scales. The intention is that
543 a repeat survey will be undertaken in the near future (indeed a sub-sample of plots (the
544 majority being located in Wales), have already been surveyed in the summer of 2016, largely
545 as part of the Glastir Monitoring and Evaluation Programme (Emmett and GMEP team,
546 2014)). As a decade has now passed since the most recent full survey, an addition to this
547 long-term national resource is becoming increasingly timely, particularly in these current
548 times of political, socio-economic and climatic change.

549

550 **Author Contributions.** C. M. Wood prepared the manuscript with significant contributions
551 from all co-authors, and is the current database manager for the Land Use Research Group at
552 CEH Lancaster. R. G. H. Bunce designed the sampling framework and survey strategy in
553 1978. R.G.H. Bunce, S. M. Smart, L. C. Maskell, L. R. Norton and D. C. Howard have all
554 been part of the Countryside Survey co-ordination team for at least one survey, with W. A.
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556

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565

566



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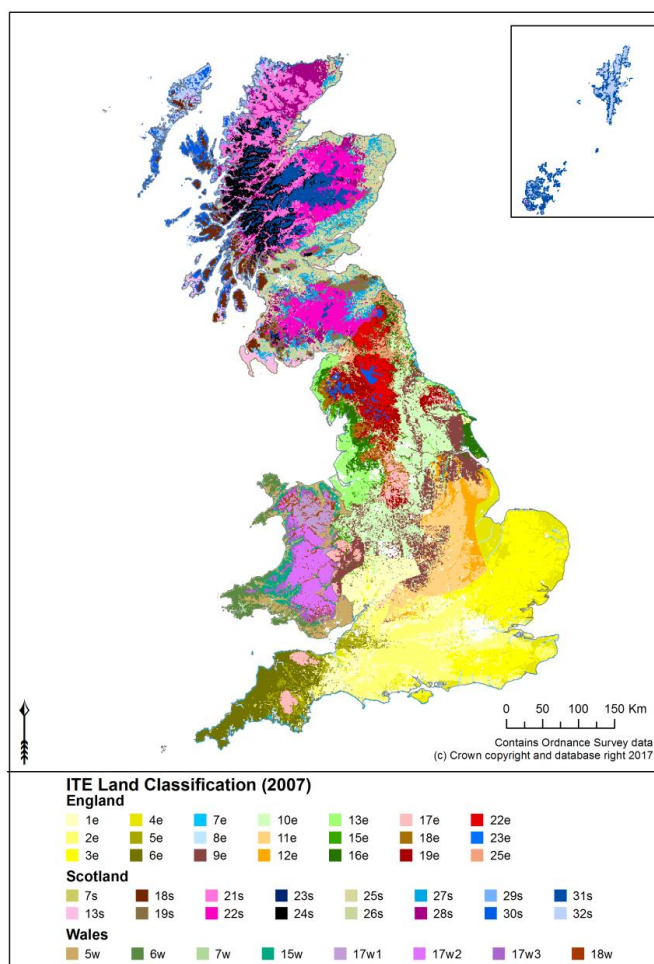
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851 *Figure 1. ITE Land Classification, 2007*



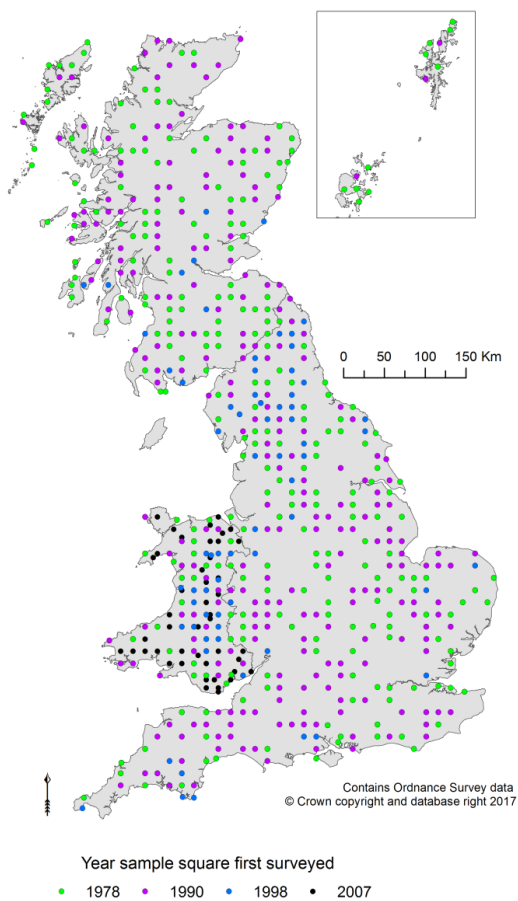
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860 *Figure 2. Map of Countryside Survey sampling locations across Britain*

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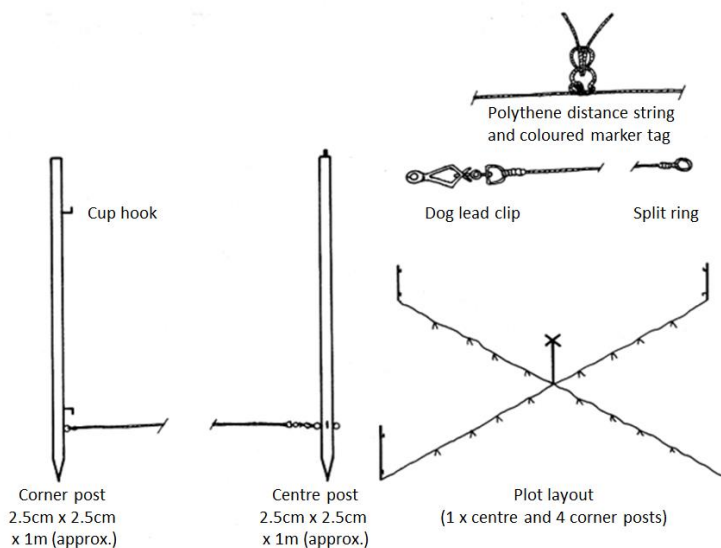
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870 *Figure 3. X plot construction*

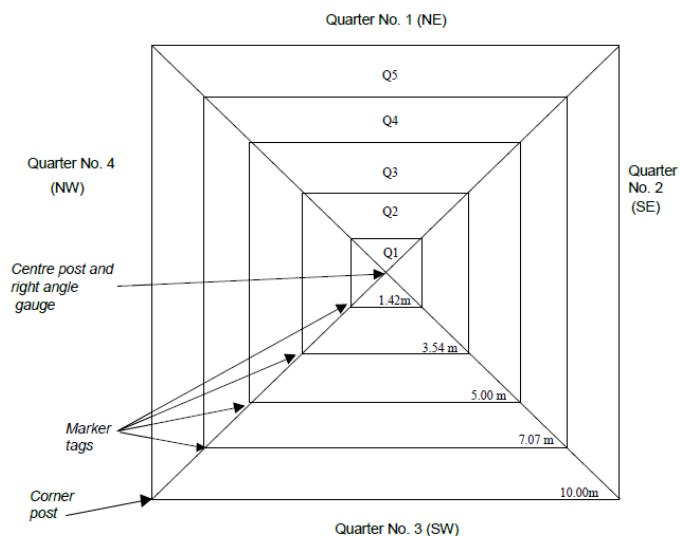


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889 *Figure 4. Layout of vegetation X plot.*

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Distance string position from centre - 1/2 diagonal:

Q1 = 4 m² quadrat (2 m x 2 m) = 1.42 m diagonal

Q2 = 25 m² (5.00 x 5.00 m) = 3.54 m

Q3 = 50 m² (7.07 x 7.07 m) = 5.00 m

Q4 = 100 m² (10.00 x 10.00 m) = 7.07 m

Q5 = 200 m² (14.14 x 14.14 m) = 10.00 m

Not to scale

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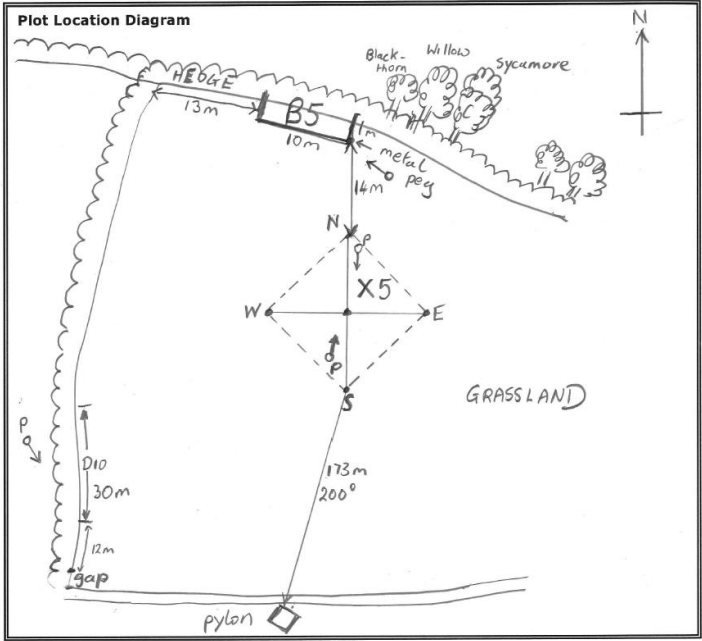
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901 *Figure 5. Example of a plot sketch map*
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905 *Figure 6. Example of a plot photograph*
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