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Expanding understanding of optical variability in Lake Superior with a four-year dataset

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25 **Abstract**

26 Lake Superior is one of the largest freshwater lakes on our planet, but few optical observations
27 have been made to allow for development and validation of visible spectral satellite remote
28 sensing products. The dataset described here focuses on coincidentally observing inherent and
29 apparent optical properties along with biogeochemical parameters. Specifically, we observe
30 remote sensing reflectance, absorption, scattering, backscattering, attenuation, chlorophyll
31 concentration, and suspended particulate matter over the ice-free months of 2013-2016. The
32 dataset substantially increases the optical knowledge of the lake. In addition to visible spectral
33 satellite algorithm development, the dataset is valuable for characterizing the variable light field,
34 particle, phytoplankton, and colored dissolved organic matter distributions, and helpful in food
35 web and carbon cycle investigations, among others. The compiled data can be freely accessed
36 at: <https://seabass.gsfc.nasa.gov/archive/URI/Mouw/LakeSuperior/>.

37



38 1. Introduction

39 Lake Superior is the largest of the Laurentian Great Lakes, the largest lake on Earth by
40 area and third largest lake on Earth by volume. Direct observations of Lake Superior are rare for
41 nearly half of the year due to the long winter climatic conditions of the region. *In situ*
42 observations are scarce November through April, and in many years October and May are also
43 precluded as seasonal storms and ice coverage hinder ship-based observations for long periods of
44 time. As such, remote sensing is an important observational platform for observing the lake.

45 Remote sensing of inland and coastal waters requires continued investment in *in situ*
46 sampling to fully capture the dynamic range of parameter values and shifting relationships due to
47 response to climate change (Dierssen 2010), anthropogenic stress, and invasive species (Mouw et
48 al., 2015). Remote sensing efforts on the lake initially focused on estimating chlorophyll
49 concentration with band-ratio algorithms (Budd and Warrington, 2004; Li et al., 2004) and met
50 with limited success. More recently, Mouw et al. (2013) evaluated a variety of semi-analytical
51 algorithms to identify the best performing for refinement. They found the retrieval of
52 chlorophyll concentration was not possible via inversion since errors in derived colored
53 dissolved organic matter (CDOM) absorption were greater than the total contribution of
54 phytoplankton to the overall absorption budget. However, they did demonstrate success in
55 retrieving absorption due to CDOM. In light of the challenge to retrieve chlorophyll
56 concentration, Trochta et al. (2015) classified remote sensing reflectance spectra into optical
57 water types and connected the spatial and temporal variability of the water types to physical
58 processes of the lake.

59 *In situ* observations are an essential component of remote sensing satellite algorithm
60 development and validation. At the time that Mouw et al. (2013) completed their evaluation of



61 algorithms, there were only 8 stations with coincident observations of radiometry, absorption,
62 and scattering for Lake Superior (Peng et al., 2009; Effler et al., 2010; O'Donnell et al., 2013).
63 To enable continued development of remote sensing algorithms for Lake Superior, a greater
64 investment in optical observations were needed. Here we describe a dataset of coincident
65 inherent and apparent optical properties collected throughout Lake Superior over the ice-free
66 months of 2013 to 2016. Inherent optical properties (IOPs, i.e. absorption and
67 scattering/backscattering) depend on in-water dissolved and particulate constituents, while
68 apparent optical properties (AOPs, i.e. reflectance and attenuation) are dependent on both in-
69 water constituents and the ambient light field. Collectively, these properties include remote
70 sensing reflectance, absorption, scattering, backscattering, attenuation, pigment concentrations,
71 and total suspended matter. Beyond satellite algorithm development, the dataset is valuable for
72 characterizing the variable light field, particle, phytoplankton, and colored dissolved organic
73 matter distributions, and helpful in food web and carbon cycle investigations, among others.

74

75 **2. Data and Methodology**

76 *2.1 Study Site*

77 Containing approximately 10% of the Earth's surface freshwater, Lake Superior (Figure
78 1) is the largest freshwater body on the planet by area ($8.21 \times 10^4 \text{ km}^2$) and third largest by
79 volume ($1.21 \times 10^4 \text{ km}^3$). The cold temperatures and low nutrient concentrations of Lake
80 Superior (Sterner, 2011), compared to the other Laurentian Great Lakes, result in oligotrophic
81 conditions (e.g. Matheson and Munawar, 1978; Munawar and Munawar, 1973; Weller, 1978;
82 Barbiero and Tuchman, 2001) with low primary production (Sterner, 2010) and low species
83 richness (Hubbs and Lagler, 2004). Superior is the fastest warming of all the Great Lakes



84 (Mason et al., 2016), with summer surface temperatures rising faster than the temperature of the
85 surrounding atmosphere (Austin and Allen, 2011; Austin and Colman, 2007; Lenters, 2004).
86 This trajectory puts Lake Superior close to the upper range of global lake warming trends
87 (O'Reilly et al., 2015). These warming trends impact the lake's energy balance, hydrology, ice
88 cover, and mixing dynamics, which in turn impacts biotic and biogeochemical responses. The
89 previous limited optical observations of Lake Superior reveal that light absorption is dominated
90 by colored dissolved organic matter (CDOM; $\geq 75\%$ of the total absorption at 440 nm) (Effler et
91 al., 2010) and the majority of the backscattering in the lake is attributed to organic rather than
92 minerogenic sources (Peng et al., 2009).

93 2.2 *In Situ Optical Observations*

94 Optical and biogeochemical data are collected in Lake Superior during the ice-free
95 months (May – October) of 2013 through 2016 (Figure 1). The dataset consists of a full suite of
96 coincident IOPs and AOPs. Observed IOPs include absorption due to water itself ($a_w(\lambda)$, m^{-1}),
97 phytoplankton ($a_{ph}(\lambda)$, m^{-1}), non-algal particles ($a_{NAP}(\lambda)$, m^{-1}), colored dissolved organic matter
98 ($a_{CDOM}(\lambda)$, m^{-1}), and backscattering due to water itself ($b_{bw}(\lambda)$, m^{-1}) and particles ($b_{bp}(\lambda)$, m^{-1}).
99 Observed AOPs include downwelling and surface irradiance ($E_d(\lambda)$ and $E_s(\lambda)$, respectively, μW
100 cm^{-2}) along with upwelling radiance ($L_u(\lambda)$, $\mu W cm^{-2} sr^{-1}$), which are used to retrieve remote
101 sensing reflectance ($R_{rs}(\lambda)$, sr^{-1}) and downwelling and upwelling diffuse attenuation coefficients
102 ($K_d(\lambda)$ and $K_u(\lambda)$, m^{-1} , respectively). A summary of parameters and units can be found in Table
103 1.

104 2.2.1 *Apparent Optical Properties*

105 Radiometric measurements are made with three HyperOCR spectral radiometers



106 (Satlantic Inc.) that measure between 350–800 nm with approximately 3 nm resolution (137 total
107 wavelengths). In-water $E_d(\lambda)$ and $L_u(\lambda)$ HyperOCR sensors are attached to a free-falling Profiler
108 II frame (Satlantic Inc.), while the $E_s(\lambda)$ sensor is mounted on top of the ship to allow for
109 correction of the other measurements due to changing sky conditions. At each station, the
110 system is deployed for three cast types: surface, multi- and full profile. To characterize the
111 air/water interface, a floatation collar on the profiler frame enables continuous measurement of
112 $L_u(\lambda)$ approximately 20 cm below the water surface for 5 minutes (surface profile). The floatation
113 collar is removed and the profiler then deployed in free-fall mode, measuring 5 consecutive
114 profiles from the surface to 10 m to characterize the near-surface light field (multi-profile).
115 Finally, the profiler is allowed to free-fall to the 1% light level or to within 10 m of the bottom,
116 whichever is shallower (full profile).

117 All methods and analysis follow the NASA ocean optics protocols for satellite ocean
118 color sensor validation (Mueller et al., 2003a). Briefly, $E_d(\lambda)$, $L_u(\lambda)$ and $E_s(\lambda)$ are dark-corrected
119 and quality controlled to exclude data collected at high instrument tilt ($>4^\circ$). $E_s(\lambda)$ is smoothed
120 over a 15 s interval and $E_d(t,\lambda)$ and $L_u(t,\lambda)$ normalized to the ratio of $E_s(t_0,\lambda)/E_s(t,\lambda)$ to account
121 for variability in surface irradiance throughout each profile measurement. Data from surface,
122 multi- and full profiles are each binned to 1 m.

123 Using binned multi-profile data, normalized $E_d(z,\lambda)$ and $L_u(z,\lambda)$ are natural log
124 transformed and fit as a function of depth to retrieve attenuation coefficients $K_d(\lambda)$ and $K_u(\lambda)$,
125 respectively (Smith and Baker, 1984; 1986):

$$126 \quad \ln[X(z, \lambda)] = \ln[X(z_m, \lambda)] - (z - z_m)K(z_m) \quad (1)$$

127 where X is either E_d or L_u , z is depth, z_m is a central reference depth and K is either K_d or K_u .



128 This method assumes K is uniform across the depth interval used in the regression. For the
129 majority of the Lake Superior dataset, the upper 8 m of the water column is typically utilized for
130 the retrieval of K , as recommended by Smith and Baker (1984; 1986). However, some locations
131 near river outflows show shallow layers with distinct optical changes, reducing the uniform
132 depth interval to about 5 m.

133 Using K -regression results, $E_d(\lambda)$ and $L_u(\lambda)$ are propagated to just below the surface
134 ($E_d(0^-, \lambda)$ and $L_u(0^-, \lambda)$). Water leaving radiance ($L_w(\lambda)$) is calculated from $L_u(0^-, \lambda)$ according
135 to:

$$136 \quad L_w(\lambda) = 0.543 \cdot L_u(0^-, \lambda) \quad (2)$$

137 where 0.543 accounts for the net effects of surface reflectance and the refractive index of
138 seawater as light travels across the sea/air interface (Lee et al., 2002; Mobley, 1994). $R_{rs}(\lambda)$ is
139 then retrieved as:

$$140 \quad R_{rs}(\lambda) = \frac{L_w(\lambda)}{E_s(\lambda)} \quad (3)$$

141

142 2.2.2 *Inherent optical properties*

143 IOPs are collected via a vertically profiled bio-optical package that measures absorption,
144 attenuation (WET Labs AC-S) and backscattering (WET Labs ECO-BB9) along with concurrent
145 temperature, salinity (SeaBird CTD 37SI) and fluorometric chlorophyll a (WET Labs ECO-
146 FL3). All methods and analysis follow the NASA ocean optics protocols for satellite ocean color
147 sensor validation (Mueller et al., 2003b). The ECO-FL3 is calibrated on a yearly basis,
148 regressing *in situ* chlorophyll a fluorescence with extracted chlorophyll a concentration ($[Chl]$,
149 mg m^{-3}) derived from discrete field samples.



150 Total absorption and attenuation ($a_t(\lambda)$ and $c_t(\lambda)$, m^{-1} , respectively) are resolved at 81
151 wavelengths between 400-750 nm. CTD data are used to correct $a_t(\lambda)$ and $c_t(\lambda)$ for temperature
152 and salinity effects using the coefficients of Sullivan et al. (2006). The AC-S is calibrated before
153 and after each cruise with ultrapure water (ion, particle, and bubble free). This pure-water
154 absorption and attenuation ($a_w(\lambda)$ and $c_w(\lambda)$) is subtracted to retrieve the contribution of non-
155 water components ($a_{nw}(\lambda)$ and $c_{nw}(\lambda)$). Data are corrected for scattering effects using the
156 proportional method of Zaneveld et al. (1994). The bio-optical package is deployed for two
157 successive casts: total and filtered. For the filtered cast, the AC-S intake is fit with a 0.2 μm
158 filter (PALL, Maxi Capsule Filter 12112) to allow for the measurement of dissolved components
159 only ($a_{CDOM}(\lambda)$, m^{-1}). During deployment, the package is lowered at 0.2 $m s^{-1}$ or as slow as
160 possible. All data are binned to 1-m during analysis. Particulate absorption ($a_p(\lambda)$, m^{-1}) is
161 retrieved by the difference between total and filtered casts ($= a_{nw}(\lambda) - a_{CDOM}(\lambda)$). Pure-water
162 absorption and attenuation as measured by Pope and Fry (1997) is then added to $a_{nw}(\lambda)$ and
163 $c_{nw}(\lambda)$ to return to total values.

164 The ECO-BB9 measures the volume scattering coefficient (β_t , $m^{-1} sr^{-1}$) for an acceptance
165 angle of 124° at 9 wavelengths: 412, 440, 488, 510, 532, 595, 650, 676 and 715 nm. Instrument
166 calibration was performed at WET Labs in Narragansett, RI using 0.1 μm NIST-traceable beads
167 (Thermo-Scientific, 3100A, bead lot number 43585) at $9^\circ C$ following the procedure of Sullivan
168 et al. (2013). Processed AC-S measurements are interpolated to BB9 wavelengths and
169 backscattering data corrected for absorption:

$$170 \quad \beta = \beta_t e^{(l-a_t)} \quad (4)$$



171 where l is the pathlength of the BB9 (=0.0391 m). The volume scattering function (β_w , $\text{m}^{-1} \text{sr}^{-1}$)
172 and total scattering (b_w , m^{-1}) of pure-water is retrieved from temperature and salinity
173 measurements using the model of Zhang and Hu (2009). Backscattering of pure-water (b_{bw} , m^{-1})
174 is considered half of total scattering ($= b_w/2$). Volume scattering of the particulate fraction (β_p ,
175 $\text{m}^{-1} \text{sr}^{-1}$) can then be retrieved by difference ($= \beta - \beta_w$). Particulate backscattering is calculated
176 by taking acceptance angle into account (Sullivan et al., 2013):

$$177 \quad b_{bp} = 2\pi \chi \beta_p \quad (5)$$

178 where $\chi=1.076$ for a 124° measurement angle. Total backscattering (b_b , m^{-1}) can then be
179 retrieved as the sum of particulate and water components ($= b_{bp} + b_{bw}$).

180

181 *2.2 Laboratory Analysis of Discrete Water Samples*

182 Whole water samples are taken from multiple depths representative of the surface (5 m)
183 and at a mid-depth of the euphotic zone often coincident with the deep chlorophyll maximum if
184 one is present (11-50 m). During well-mixed times of year, when a deep chlorophyll layer is not
185 observed, samples are only taken in the surface layer. Collected water is stored in the dark on ice
186 and is processed within 8 hours. Discrete water samples are used for determining [Chl], $a_p(\lambda)$,
187 $a_{\text{NAP}}(\lambda)$, $a_{\text{ph}}(\lambda)$, $a_{\text{CDOM}}(\lambda)$ and the organic and inorganic components of suspended particulate
188 matter (SPM_0 and SPM_1 , mg L^{-1} , respectively).

189 Samples for [Chl] and $a_p(\lambda)$ are filtered in triplicate onto Whatman glass fiber filters
190 (GF/F, $0.7\mu\text{m}$ nominal pore size), stored in liquid nitrogen while at sea and moved to a -80°C
191 freezer in the lab until analysis. Samples for $a_{\text{CDOM}}(\lambda)$ are filtered through an acid-washed 0.2



192 μm membrane filter (Whatmann Nucleopore, 111106) and the filtrate stored in clear borosilicate
193 bottles (Qorpak, GLC-01151) in a refrigerator until analysis. Directly prior to each cruise,
194 borosilicate bottles are acid-washed and muffled at 450°C for 4 hours to remove any potential
195 contaminants.

196 Spectral CDOM, particulate, non-algal and phytoplankton absorption are measured
197 spectrophotometrically (Perkin-Elmer Lambda 35 UV/VIS dual-beam) for wavelengths between
198 300-800 nm. Absorption of CDOM filtrate is measured in a 10-cm cuvette following NASA's
199 Ocean Optics Protocols (Mueller et al., 2003b) using a slit-width of 2 nm and a scan rate of 240
200 nm min^{-1} . For particulate and non-algal absorption, we follow the transmission-reflectance (T-R)
201 method (Tassan and Ferrari, 2002; Lohrenz, 2000; Lohrenz et al., 2003) that utilizes an
202 integrating sphere to correct measurements for the contribution of scattering. Daily blanks are
203 prepared by filtering 25 mL of ultrapure water (ion and particle free) onto a GF/F and used to
204 normalize T-R data during analysis. Filters are placed on a quartz slide at the entrance
205 (transmittance) and exit (reflectance) of the sphere and scanned at a speed of 120 nm min^{-1} with a
206 slit width of 2 nm both before (a_p) and after (a_{NAP}) exposure to sodium hypochlorite. The
207 difference between these measurements is the contribution due to phytoplankton ($a_{ph} = a_p -$
208 a_{NAP}). Absorption is calculated following Lohrenz et al. (2000).

209 Extracted [Chl] and phaeopigments are determined following the NASA Ocean Optics
210 Protocols (Trees et al., 2003) for both instrument calibration and sample analysis. Working in
211 the dark, filtered samples are placed in pre-chilled 90% acetone and sonicated for 20 to 30s to
212 break cells open. Sonicated samples are stored in the freezer (-20°C) and allowed to extract for
213 24 hours prior to analysis. Samples are brought to room temperature and centrifuged at 3000
214 RPM for 20 minutes to remove filter particulates. Fluorescence is measured before and after the



215 addition of 50 μ L of 10% HCl on a Turner Designs 10-AU fluorometer (Optical Kit 10-037R).
 216 Blanks of 90% acetone are measured daily. The fluorometer is calibrated yearly with a dilution
 217 series of pure chlorophyll *a* (Sigma-Aldrich C6144) extracted in 90% acetone. Field samples are
 218 analyzed within 6 months of fluorometer calibration. [Chl] is retrieved from blank-corrected
 219 fluorescence as:

$$220 \quad [Chl] = \frac{(F_b - F_a)}{F_R} \cdot \frac{\tau}{\tau - 1} \cdot \frac{V_S}{V_F} \quad (6)$$

221 and phaeopigment concentration ([Phaeo], mg m⁻³) calculated as:

$$222 \quad [Phaeo] = \frac{(F_a \tau - F_b)}{F_R} \cdot \frac{\tau}{\tau - 1} \cdot \frac{V_S}{V_F} \quad (7)$$

223 where F_b and F_a (FSU) are fluorescence measured before and after acidification, respectively, F_R
 224 (FSU (mg m⁻³)⁻¹) is the slope of the fluorometer calibration, τ (unitless) is the acid-ratio of the
 225 fluorometer calibration and V_S and V_F (mL) are the volume of solvent and filtered water sample,
 226 respectively.

227 Both organic and inorganic SPM are quantified following the methodology of the
 228 American Public Health Association, American Water Works Association and Water
 229 Environment Federation (2005) and the recommendations of Boss et al. (2009) and Woźniak et
 230 al. (2011). Whatman GF/F filters (47mm, 0.7 μ m pore size) are prepared by washing with
 231 ~100mL ultrapure water to rinse away any loose glass fibers. They are then dried at 103-105°C
 232 for 1 hour and then muffled at 550°C for 4 hours and allowed to cool in a desiccator. Filters are
 233 pre-weighed on a high-precision balance (W_F , mg). Duplicate samples are filtered with a
 234 minimum of 3L of collected water and stored at -80°C until analysis. Samples are removed from
 235 the freezer, dried at 103-105°C for 1 hour and placed in a desiccator to cool. Dried filters are



236 then weighed within 2 hours of removal from the oven (W_R , mg). Filters are muffled at 550°C
237 for 1 hour, returned to the desiccator to cool and re-weighed (W_I , mg). Total and inorganic SPM
238 are retrieved as:

$$239 \quad SPM = (W_R - W_F)/V \quad (8)$$

$$240 \quad SPM_I = (W_I - W_F)/V \quad (9)$$

241 where V is the volume filtered (L). The organic component can be retrieved by difference
242 ($SPM_O = SPM - SPM_I$).

243

244 3. Results

245 Data are collected from 106 stations from May through October of 2013 to 2016.
246 Collection is aimed at maximizing spatial and temporal variability and dynamic range. Thus,
247 repeat stations are only found for 6 locations. Most of the observations are made in July ($n=29$)
248 and August ($n=28$), but also show reasonable distribution in the early spring (May $n=12$; June
249 $n=8$) and fall (September $n=16$; October $n=13$). Data are only collected in U.S. waters, thus the
250 Canadian waters encompassing the northern reaches of the lake are not observed. To some
251 extent, sampling locations are dependent on the home-port location of research vessels. The
252 western basin (41%) and the central basin, near the Keweenaw Peninsula (41%) of the lake are
253 well observed. The eastern basin is observed only during July of 2015 and a single site in July
254 2013, August 2014, and September 2013 (Figure 1).

255 The lake is dimictic (Assel, 1986) and develops a strong thermocline and deep
256 chlorophyll maximum (DCM (Barbiero et al., 2004) in the summer months. The water column is
257 completely mixed in May and June of all years (Figure 2 A, B). A thermocline is evident in



258 nearly all locations by July (Figure 2 C), becoming stronger and deeper in August (Figure 2 D).
259 Depending on location and year, the thermocline either deepened or shallowed in September and
260 October, with evidence of a fully mixed water column in some locations in October (Figure 2 E,
261 F). In the fluorometric [Chl] profiles shown in Figure 2, [Chl] is slightly lower in the near
262 surface water in May and June due to fluorescence quenching when the water column is well
263 mixed (Figure 2 G, H). When [Chl] is derived from absorption, the profiles are uniform at this
264 time of the year (not shown). The formation of a DCM follows the development of the
265 thermocline. As the thermocline sets up in July the DCM develops and becomes more
266 pronounced and deeper in August as the thermocline deepens (Figure 2 I, J). The DCM is still
267 evident in September, but becoming weaker in magnitude and beginning to shallow (Figure 2 K).
268 By October, when the lake is beginning to become fully mixed, [Chl] remains low in the near
269 surface waters and slightly elevated in mid-depths where the DCM was previously found (Figure
270 2 L).

271 Discreet absorption samples partitioned into water, phytoplankton, non-algal particles
272 and CDOM display contributions from riverine inputs, and broad lake processes near surface at
273 mid-depth (Figure 3). The St. Louis River enters Lake Superior in its western arm. Absorption
274 is fairly evenly distributed there between water, phytoplankton, and CDOM, with some
275 contributions of NAP in May. By June and July, the surface waters are dominated by CDOM
276 and a growing contribution from phytoplankton. This is presumably due the spring melt entering
277 the lake. August sees declining overall absorption as this is a time of lower river flow, with a
278 greater transition toward phytoplankton dominance in September and October when fall storms
279 generate more run off and the river serves as a nutrient delivery conduit to the lake (Figure 3).
280 The surface waters of the central region of the lake remain dominated by water year-round with a



281 significant CDOM contribution near the Ontonagon River outlet and within the Keweenaw
282 Peninsula waterway that contains Portage Lake. Samples at mid-depth are not targeted when the
283 water column is well mixed in May and June. The mid-depth becomes progressively dominated
284 by phytoplankton between July and September as the DCM develops and declines. By October,
285 surface and mid-depth absorption are similar (Figure 3). Considering the absorption budget,
286 CDOM is by far the largest contributor (50-80%), followed by phytoplankton, while NAP makes
287 less than a 15% contribution across all stations. Phytoplankton make up a greater percentage of
288 mid-depth absorption, while CDOM is more variable in surface waters. NAP contributions are
289 only slightly less in the mid-depth (Figure 4).

290 To display the spectral variability observed in the absorption budget, three stations are
291 selected in different regions throughout the lake (Figure 5). As $a_{\text{CDOM}}(\lambda)$ absorbs most strongly
292 in the ultraviolet and blue region of the spectrum, its contribution decreases exponentially with
293 increasing wavelength. Phytoplankton also absorb strongly in the blue region of the spectrum,
294 with a peak near 443 nm and a secondary peak at 687 nm. While phytoplankton account for a
295 larger proportion of absorption at mid-depth compared to the surface, they comprise <50% of the
296 total absorption budget in blue wavelengths at all depths (Figure 5). NAP is a larger contributor
297 in the surface, particularly for the sites near river outlets (Figure 5 A, C).

298 The particulate portion of absorption (NAP and phytoplankton) can be further explored
299 by SPM_i and SPM_o . $a_{\text{NAP}}(\lambda)$ includes all non-phytoplankton particles, while SPM_i includes only
300 minerogenic particles. Similar to absorption, greatest SPM loads are found near the outlet of the
301 St. Louis River with magnitude varying with season (Figure 6). The high volumes of spring melt
302 also delivered or stimulated growth resulting in high SPM_o in June. By July, freshwater entering
303 the lake from the river is warm enough to cause a density gradient, where the warm river water



304 sat on top of the colder lake water; thus, SPM is significantly lower at the mid-depth. In August,
305 SPM magnitude is relatively similar across the lake and between surface and mid-depth but
306 SPM_i contributed a greater proportion of SPM at depth. By September and October, the depth
307 distribution of SPM_i and SPM_o are similar, but the magnitude of SPM in the western arm is
308 much greater in October due to runoff and water column mixing from fall storms. At this time of
309 year, the riverine water is cooling faster than the lake, resulting in a submerged river plume; thus
310 greatest SPM is observed in the mid-depth.

311 In addition to the discreet measurements, *in situ* profiles of $a_p(\lambda)$, $a_{CDOM}(\lambda)$ and $b_{bp}(\lambda)$ are
312 collected. Figure 7 displays examples of this spectral data for the same three stations as in
313 Figure 5. The St. Louis River inflow in the western basin is seen in $a_p(\lambda)$ in the surface of the
314 water column (Figure 7A), and the DCM at mid-depth. The presence of the phytoplankton in
315 $a_p(\lambda)$ is evident by a secondary spectral peak near 683 nm. Riverine impact is also seen where
316 $a_{CDOM}(\lambda)$ is elevated in the near-surface waters but slightly deeper than $a_p(\lambda)$, due to CDOM
317 photobleaching (Figure 7B). In the Ontonagon Outflow, the riverine signal is also seen in $a_p(\lambda)$
318 and $a_g(\lambda)$, with a large deep chlorophyll contribution to $a_p(\lambda)$ at mid-depth. $a_{CDOM}(\lambda)$ is elevated
319 below the deep chlorophyll layer, possibly from bloom degradation and sinking (Figures 7 C, D).
320 Along the South Shore, there is no significant riverine source, thus $a_p(\lambda)$ is low in the surface and
321 high at depth associated with the DCM. $a_{CDOM}(\lambda)$ is uniform throughout the water column and
322 significantly lower than the riverine impacted sites (Figures 7 E, F). Spectra are extracted from
323 the profiles at 5 m and 25 m to show an alternative view of the spectral site comparison. Given
324 the low contribution of $a_{NAP}(\lambda)$ to the overall absorption budget (Figure 5), variability in the
325 spectral shape of *in situ* $a_p(\lambda)$ can be generally attributed to differences in the phytoplankton
326 community between the three locations (Figures 7 G, H). The most distinct difference is the



327 presence of a peak at green wavelengths (centered at 550 nm) for the location in the western
328 basin, which is not present at the other two sites. The variability in the slope of $a_{\text{CDOM}}(\lambda)$ is
329 related to the molecular weight of the source carbon (Helms et al., 2008). The slopes of
330 $a_{\text{CDOM}}(\lambda)$ are slightly lower at 5 m than 25 m suggesting proportionately more terrestrial material
331 in surface waters than water at depth (Figures 7 I, J). $b_{\text{bp}}(\lambda)$ magnitude indicates the amount of
332 particulate matter in the water, while the slope is related to the size of the particles, with steeper
333 slopes associated with smaller particles (Reynolds et al., 2001). The magnitude of $b_{\text{bp}}(\lambda)$ at 5 m
334 is directly proportional to the particle delivery to that location, with the St. Louis River, followed
335 by the Ontonagon River outflow, contributing significant particle loads. The $b_{\text{bp}}(\lambda)$ of the
336 western basin and the South Shore are nearly identical at 25 m suggesting a similar size of
337 phytoplankton at these locations. The $b_{\text{bp}}(\lambda)$ at the Ontonagon outflow at 25 m is higher and had
338 a greater slope, suggesting a greater abundance of phytoplankton and also a different community
339 composition (Figure 7 K, L).

340 Absorption and scattering processes impact the underwater light field as well as light
341 exiting the water column (i.e. $R_{\text{rs}}(\lambda)$). $R_{\text{rs}}(\lambda)$ is the parameter that a satellite radiometer observes.
342 Figure 8 displays $R_{\text{rs}}(\lambda)$ from in-water measurements at the surface, across all observation
343 months, and light field variability with depth for the three stations also highlighted in Figures 5
344 and 7. There are groups of similarly shaped $R_{\text{rs}}(\lambda)$, with peaks near 490 or 570 nm and varying
345 magnitudes (Figure 8 A-F). The magnitude of $R_{\text{rs}}(\lambda)$ is related to particulate abundance. The
346 shape of $R_{\text{rs}}(\lambda)$ is primarily related to the variations in the relative contributions and composition
347 of phytoplankton, NAP and CDOM. The spectra with the highest $R_{\text{rs}}(\lambda)$ are found in the St.
348 Louis River outflow. The underwater light field also varies spectrally depending on the



349 abundance and type of optical constituents. The spectrally dependent depth of light penetration
350 (i.e. $E_d(\lambda)$) as well as the spectral light at depth (ratio of $L_u(\lambda)$ to $E_d(\lambda)$), show significant
351 variability across the three example sites (Figure 8 G-L). The least amount of light is available
352 in the western basin with most available light at 510 nm, while light penetrates the deepest near
353 the South Shore and the most available light is at 490 nm. The diffuse attenuation coefficient at
354 490 nm ($K_d(490)$) is mapped to show geographic variability (Figure 9). The shortest attenuation
355 lengths (highest $K_d(490)$) are found in the western basin; the greatest lengths are found in the
356 central basin away from land in August.

357

358 4. Conclusions

359 This dataset has increased the coincident optical observations in Lake Superior from 8
360 observed in July 2006 (Effler et al., 2010) to 106 stations sampled during the ice-free season
361 (May through October) in 2013 to 2016. In addition to greatly expanding the temporal resolution
362 of available data, the spatial (geographic and depth) distribution has grown to cover the western
363 basin, central lake and the southern shore of the eastern basin. In a system that is challenging to
364 observe from direct methods for half the year due to prolonged winter conditions, remote sensing
365 is an important tool. The data presented here are essential for remote sensing algorithm
366 development and validation and this dataset will go a long way toward improving and ensuring
367 science quality satellite products are produced for the interpretation of lake processes.
368 Additionally, this dataset is valuable for characterizing the variable light field, particle,
369 phytoplankton, and colored dissolved organic matter distributions, and helpful in food web and
370 carbon cycle investigations, among others.

371



372 **5. Data Availability**

373 The data set contains coincident observations of total, particulate, phytoplankton and
374 CDOM absorption, total scattering, total and particulate backscattering, attenuation, remote
375 sensing reflectance, chlorophyll concentration and total suspended matter from 106 stations,
376 collected between 2013 and 2016. The compiled data are available from the National
377 Aeronautics and Space Administration's SeaWiFS Bio-optical Archive and Storage System
378 (SeaBASS): <https://seabass.gsfc.nasa.gov/archive/URI/Mouw/LakeSuperior/> (Mouw et al.,
379 2017).

380

381 **Author Contributions.** Mouw secured resources for the purchase of the optical instrumentation
382 and funding supporting the effort and designed the sampling. Ciochetto maintained the optical
383 instrumentation throughout the project period, quality controlled and achieved the dataset. While
384 all authors contributed toward the fieldwork to collect the observations, Grunert and Ciochetto
385 were responsible for the greatest amount of fieldwork. Ciochetto oversaw the laboratory sample
386 analysis, processed all data and prepared figures. Mouw prepared the manuscript with
387 contributions from all co-authors.

388

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397

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525 **Table 1.** Summary of parameters.

526

527 **Found in Dataset**

528	$a_{nw}(\lambda)$	spectral non-water absorption (m^{-1})
529	$a_p(\lambda)$	spectral particulate absorption (m^{-1})
530	$a_{ph}(\lambda)$	spectral phytoplankton absorption (m^{-1})
531	$a_{NAP}(\lambda)$	spectral non-algal particle absorption (m^{-1})
532	$a_{CDOM}(\lambda)$	spectral colored dissolved organic matter absorption (m^{-1})
533	$b_b(\lambda)$	spectral total backscattering (m^{-1})
534	$b_{bp}(\lambda)$	spectral particulate backscattering (m^{-1})
535	$c_{nw}(\lambda)$	spectral non-water attenuation (m^{-1})
536	$E_d(\lambda)$	downwelling irradiance ($\mu W\ cm^{-2}$)
537	$E_s(\lambda)$	surface irradiance ($\mu W\ cm^{-2}$)
538	$L_u(\lambda)$	upwelling radiance ($\mu W\ cm^{-2}\ sr^{-1}$),
539	[Chl]	chlorophyll <i>a</i> concentration ($mg\ m^{-3}$)
540	[Pheo]	pheopigment concentration ($mg\ m^{-3}$)
541	SPM	total suspended particulate matter ($mg\ L^{-1}$)
542	SPM _i	inorganic suspended particulate matter ($mg\ L^{-1}$)
543	SPM _o	organic suspended particulate matter ($mg\ L^{-1}$)

544

545 **Not in Dataset, but used in Calculations**

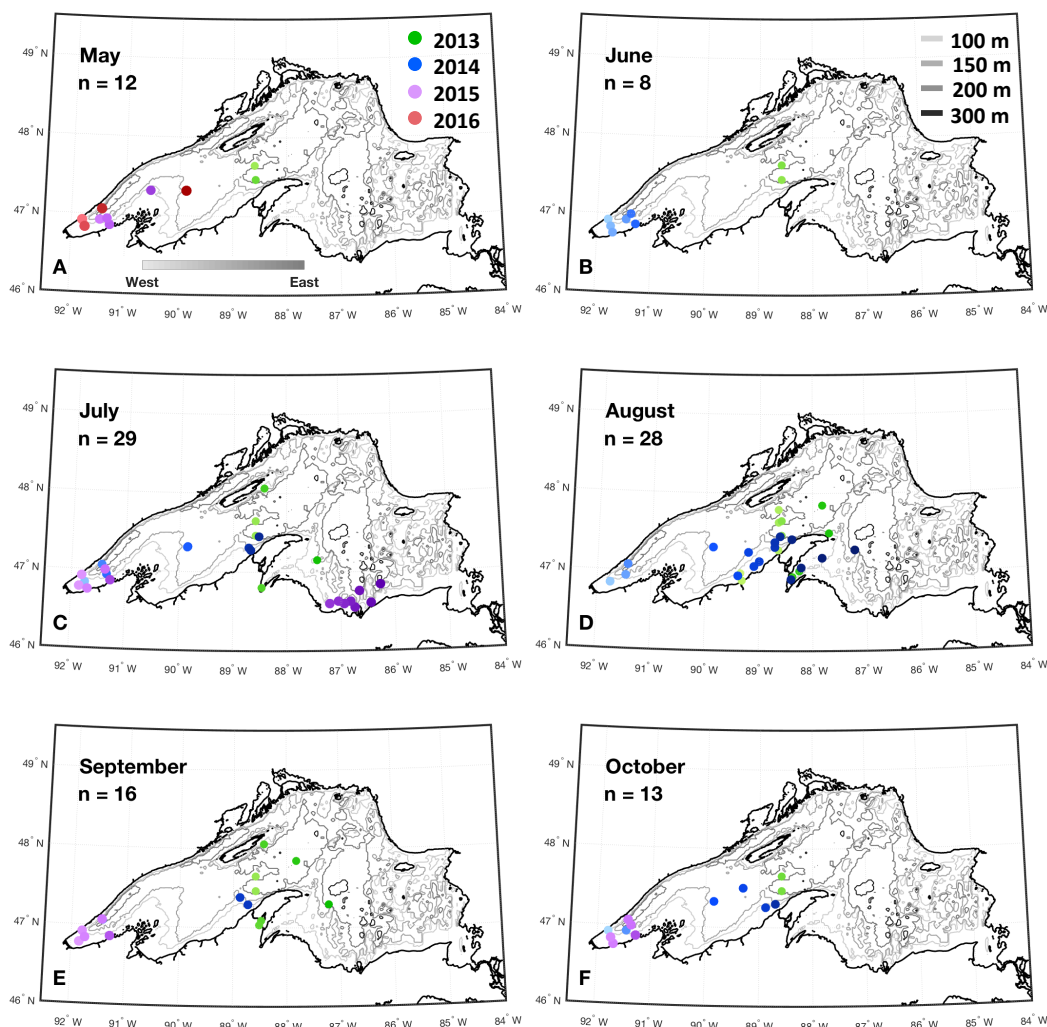
546	$a_t(\lambda)$	spectral total absorption (m^{-1})
547	$a_w(\lambda)$	spectral water absorption (m^{-1})
548	$b_t(\lambda)$	spectral total scattering (m^{-1})
549	$b_w(\lambda)$	spectral water scattering (m^{-1})
550	$b_{bw}(\lambda)$	spectral water backscattering (m^{-1})
551	$c_t(\lambda)$	spectral total attenuation (m^{-1})
552	$L_w(\lambda)$	water-leaving radiance ($\mu W\ cm^{-2}\ sr^{-1}$)
553	$R_{rs}(\lambda)$	spectral remote sensing reflectance (sr^{-1})
554	$R(\lambda)$	spectral reflectance (sr^{-1})
555	$K_d(\lambda)$	spectral downwelling diffuse attenuation coefficient (m^{-1})
556	$K_u(\lambda)$	spectral upwelling diffuse attenuation coefficient (m^{-1})

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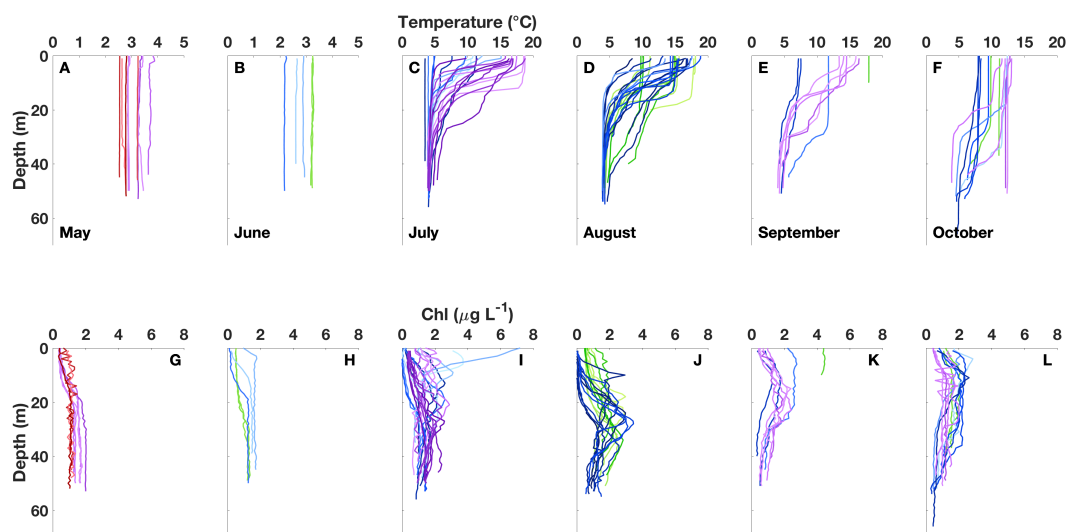


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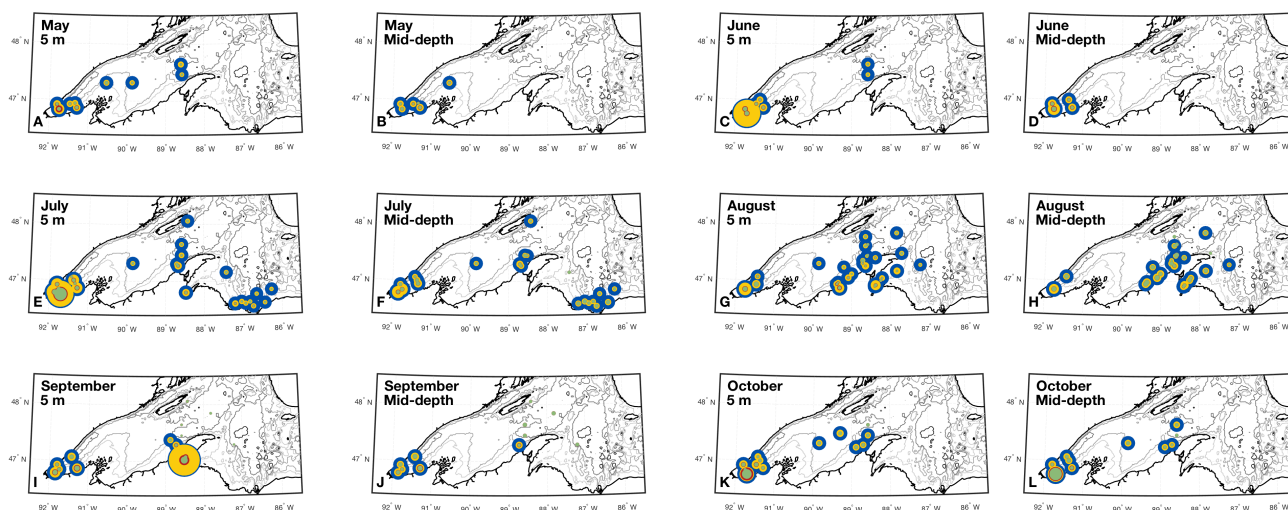
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Figure 1. Station locations for the Lake Superior dataset sampled between 2013-2016 and grouped by month. Locations are colored by year and shaded by longitude from dark to light running east to west as shown in the legend (A). Depth contours shown in grey.



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568 **Figure 2.** Profiles of temperature (A-F) and [Chl] retrieved from *in situ* fluorescence (G-L) for
569 the Lake Superior dataset. Profiles are colored by year and shaded by longitude as in Figure 1.
570 Data are binned by 1-m. Note compressed x-axis limits for temperature profiles in May and June
571 (A-B). Throughout the region, water column stratification is not observed until July or even
572 August (C-D) and is eroded in the fall (F). Absolute [Chl] levels are low throughout the lake,
573 even during spring and summer months (G-L).

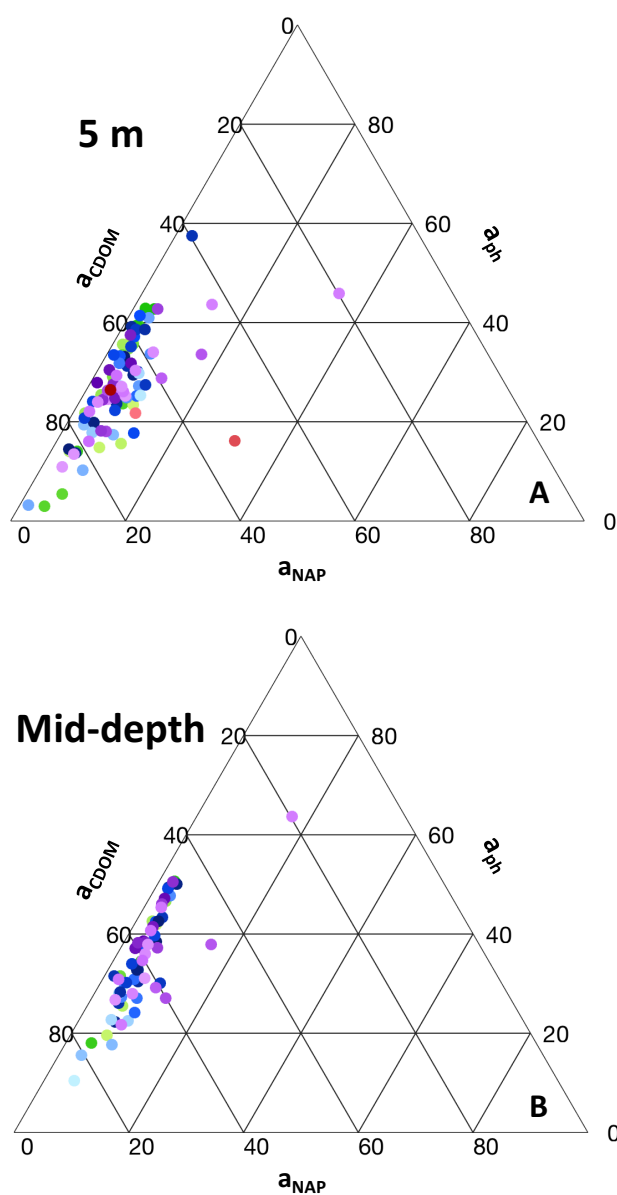


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576 **Figure 3.** Absorption for discrete water samples taken at 5 m and mid-depth throughout Lake Superior grouped by month. The outer
 577 diameter for each station marker is relative to total absorption with the contribution of water, CDOM, non-algal particles and
 578 phytoplankton indicated by the relative size of concentric circles.
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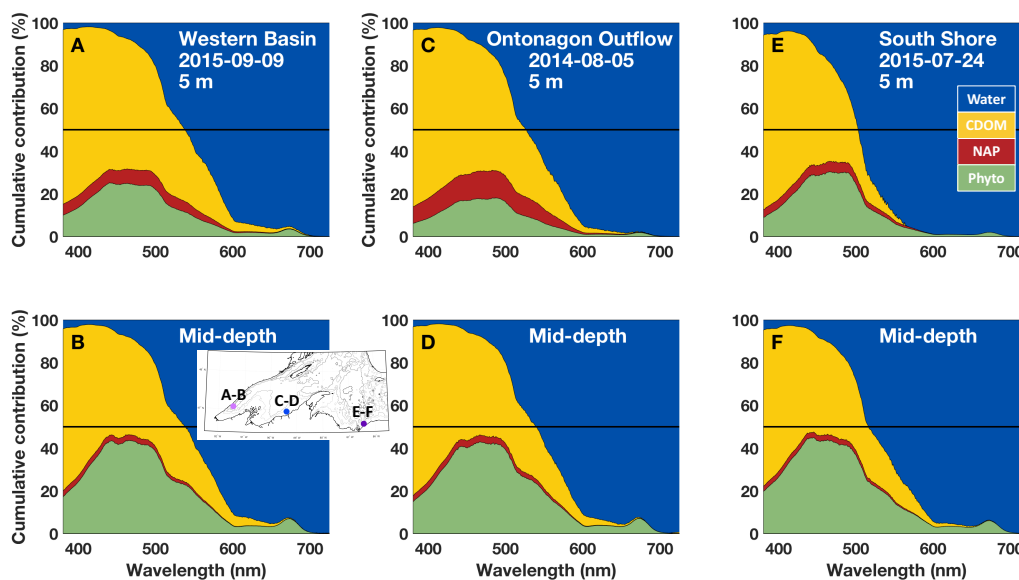


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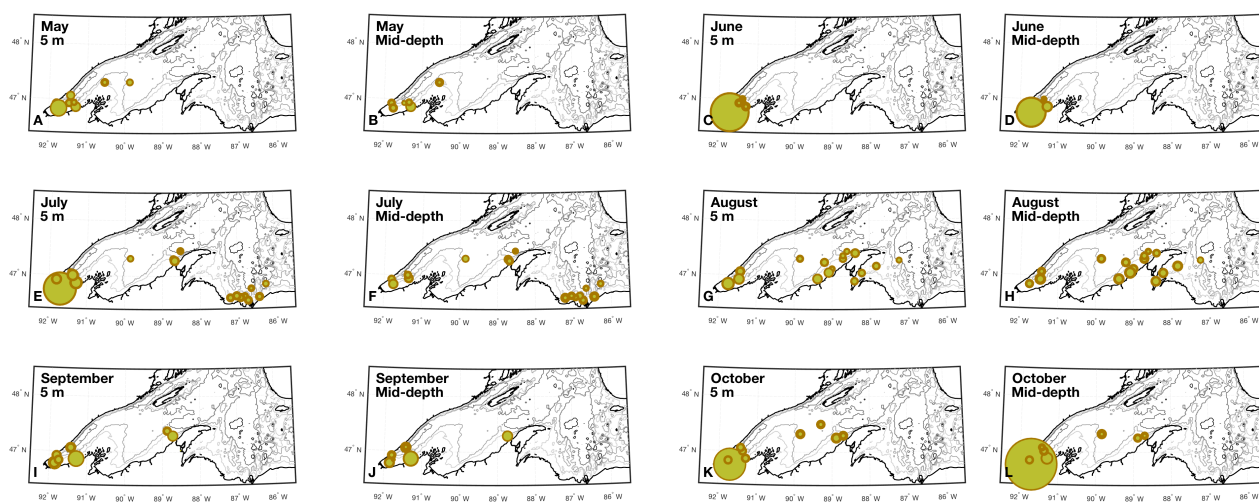
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Figure 4. Ternary plots of the contribution of CDOM, non-algal particles and phytoplankton to total non-water absorption at 443 nm from discrete water samples taken at 5 m (A) and mid-depth (B). Markers are colored by year and shaded by longitude as in Figure 1. The system is CDOM dominated, especially near the surface; phytoplankton contributed more than 50% in only 2 cases for the entire dataset.



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591 **Figure 5.** Spectrally resolved contribution of water, CDOM, non-algal particles and
592 phytoplankton to total absorption for discrete water samples taken at 5m (A, C, E) and at mid-
593 depth (B, D, F) from the Western Basin (A-B, light purple marker on map inset in B), Ontonagon
594 River Outflow (C-D, blue marker on map) and South Shore (E-F, dark purple marker on map) of
595 Lake Superior in the summer of 2014-2015. Black lines indicate 50% cumulative contribution
596 for visual reference.



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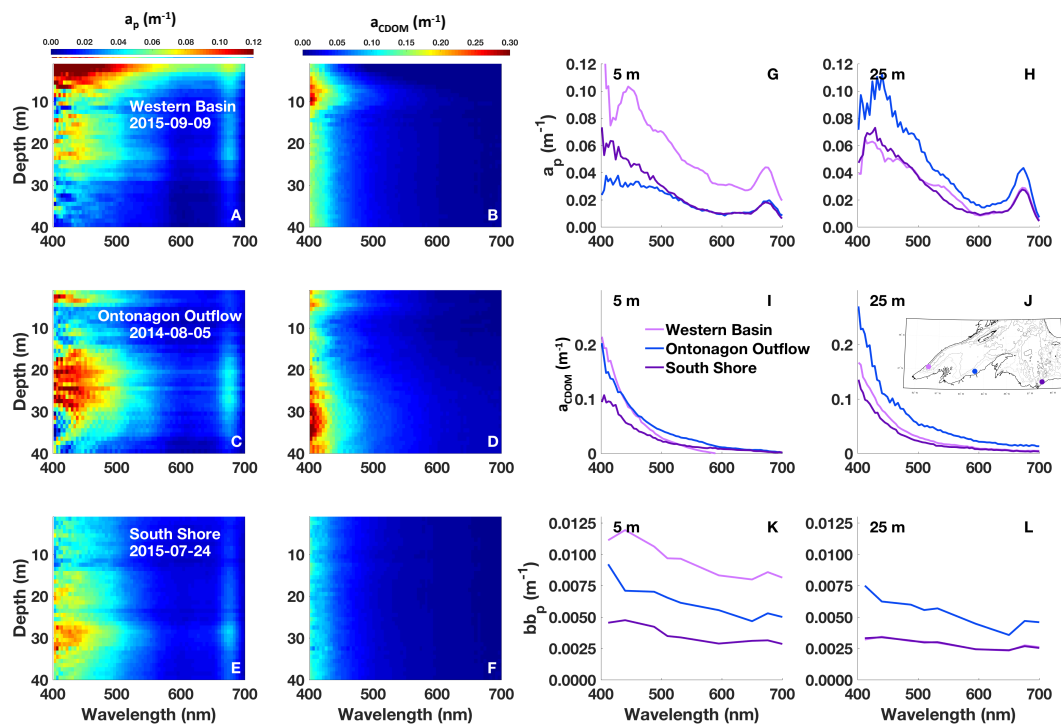
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Figure 6. Suspended particulate matter for discrete water samples taken at 5m and mid-depth for the Lake Superior dataset grouped by month. The outer diameter of each station marker indicates total SPM with concentric circles representing the relative contributions of organic and inorganic components.

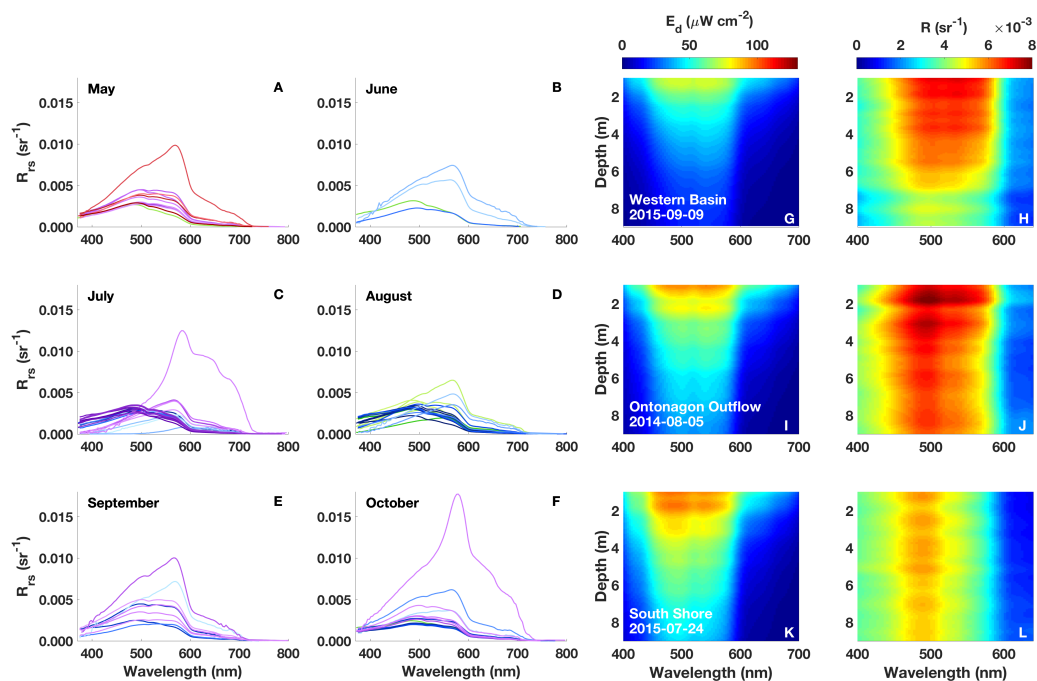


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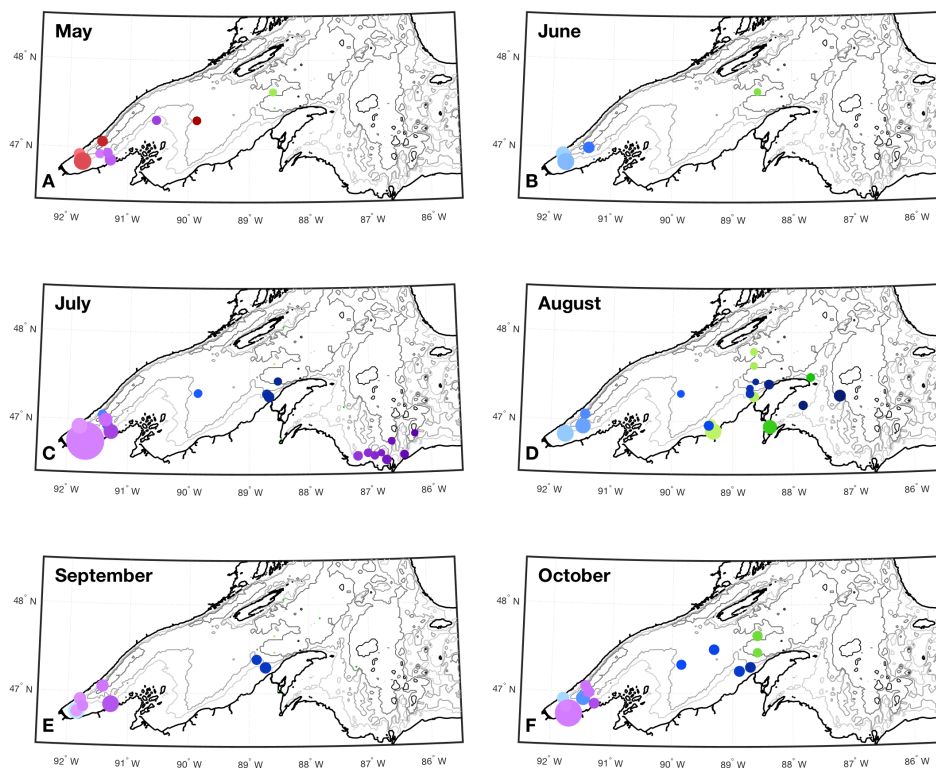
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Figure 7. *In situ* IOP for select stations throughout Lake Superior in the summer of 2014 and 2015 (see inset map in J). Profiles of $a_p(\lambda)$ (A, C, E) and $a_{CDOM}(\lambda)$ (B, D, F) for the Western Basin (A-B), Ontonagon River Outflow (C-D) and South Shore (E-F). Note different color scales for $a_p(\lambda)$ and $a_{CDOM}(\lambda)$ profiles. Examples of individual spectra at 5 m and 25 m are shown in G-J along with *in situ* particulate backscattering (K-L). Data are binned to 1-m.



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Figure 8. $R_{rs}(\lambda)$ ($=L_w(\lambda)/E_s(\lambda)$) retrieved from combined near-surface multi-cast AOP profiles (A-F). Spectra are colored by year and shaded by longitude as in Figure 1. Example spectral multi-cast AOP profiles for downwelling irradiance ($E_d(\lambda)$) and reflectance ($R(\lambda)=L_u(\lambda)/E_d(\lambda)$) for the three locations from Figures 5 and 7 (G-L).



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Figure 9. $K_d(490)$ (m^{-1}) for the Lake Superior dataset grouped by month. Stations are colored by year and shaded by longitude as in Figure 1.