



1 **The integrated water balance and soil data set of the Rollesbroich hydrological observatory**

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19



20 **Abstract**

21 The Rollesbroich headwater catchment located in Western Germany is a densely instrumented  
22 hydrological observatory and part of the TERENO (Terrestrial Environmental Observatories)  
23 initiative. The measurements acquired in this observatory present a comprehensive dataset that  
24 contains key hydrological fluxes in addition to important hydrological states and properties.  
25 Meteorological data (i.e. precipitation, air temperature, air humidity, radiation components, and  
26 wind speed) are continuously recorded and actual evapotranspiration is measured using the eddy  
27 covariance technique. Runoff is measured at the catchment outlet with a gauging station. In  
28 addition, spatio-temporal variations in soil water content and temperature are measured at high  
29 resolution with a wireless sensor network (SoilNet). Soil physical properties were determined  
30 using standard laboratory procedures from samples taken at a large number of locations in the  
31 catchment. This comprehensive data set can be used to validate remote sensing retrievals and  
32 hydrological models; to improve the understanding of spatial temporal dynamics of soil water  
33 content; to optimize data assimilation and inverse techniques for hydrological models; and to  
34 develop upscaling and downscaling procedures of soil water content information. The complete  
35 data set is freely available online (<http://www.tereno.net>).

36



## 37 **1. Introduction**

38 Climate and land use changes are taking place on different spatial and temporal scales affecting  
39 all environmental compartments. Soil water content is known to be a major control for  
40 evapotranspiration, precipitation-runoff response, and heat transfer between soil and atmosphere,  
41 and it plays an essential role for climate projections, weather and flood forecasting, water and soil  
42 resources management, agriculture, and water quality control (Albertson and Kiely, 2001; Betts  
43 et al., 1996; Crow et al., 2005; Robinson et al., 2008; Vereecken et al., 2008; Western et al.,  
44 2002). However, the highly heterogeneous pattern of soil water content leading to complex and  
45 scale-dependent patterns of water, energy, and matter fluxes makes it challenging to predict  
46 terrestrial system responses for both scientists and policymakers (Jaeger and Seneviratne, 2011;  
47 Seneviratne et al., 2010). Therefore, integrated observations of soil water content and the  
48 exchange of water and heat between the soil, vegetation, and atmosphere are critical to improve  
49 our understanding of the terrestrial system response to changes in climatic conditions and land  
50 management (Dirnbock et al., 2003; Foley et al., 1998; Hinzman et al., 2005; Refsgaard, 1997;  
51 Seneviratne et al., 2010).

52

53 To this end, a network of integrated observation platforms have been established in the  
54 framework of the Terrestrial Environmental Observatories (TERENO) initiative to investigate the  
55 consequences of global change on terrestrial ecosystems (Bogena et al., 2012; Zacharias et al.,  
56 2011). TERENO aims to collect long-term data series of system states and fluxes using state-of-  
57 the-art monitoring technologies. In previous work, the comprehensive catchment monitoring  
58 within TERENO has enabled to close the local water balance (Graf et al., 2014) and to  
59 investigate of the effects of deforestation on water, energy, and matter fluxes in an integrative  
60 manner (Bogena et al., 2015). One of the highly instrumented sites within TERENO is the



61 Rollesbroich headwater catchment, which is fully covered by grassland. It is located in the  
62 TERENO observatory Eifel/Lower Rhine Valley. All components of the water balance (e.g.  
63 precipitation, evapotranspiration, runoff, soil water content) are continuously monitored in the  
64 Rollesbroich catchment using state-of-the-art instrumentation, providing detailed information  
65 about the spatial and temporal variation of the local water cycle for the evaluation of hydrological  
66 models (Bloschl and Sivapalan, 1995; Thompson et al., 2011). In addition, using water balance  
67 data within the context of hydrological modelling helps to determine measurement errors, to  
68 diagnose such errors, and to avoid misattribution of water balance components (Evetts et al., 2012;  
69 Kampf and Burges, 2010; Vasilenko, 2004). Finally, quantification of water balance components  
70 is helpful for understanding the availability of water resources, the potential of hydrologic  
71 extremes such as floods and droughts, and the interactions between the land surface and the  
72 atmosphere (Flerchinger and Cooley, 2000; Huntington, 2006).

73

74 Here, we present a comprehensive hydrological data set recorded in the Rollesbroich catchment  
75 from 1<sup>st</sup> May 2011 to 31<sup>st</sup> December 2013. The hydrological data set includes time series of  
76 meteorological forcing (i.e. precipitation, air temperature, air humidity, radiation components,  
77 and wind speed), actual evapotranspiration, runoff, as well as soil water content data from a  
78 wireless sensor network (SoilNet). In addition, information on soil physical properties and  
79 vegetation (i.e. Leaf Area Index, LAI) useful for the parameterization of hydrological models is  
80 presented.

81



## 82 **2. Catchment description**

83 The Rollesbroich catchment (50°37'27"N, 6°18'17"E) is located in the Eifel and covers an area of  
84 about 40 ha with altitudes ranging from 474 to 518 m.a.s.l. The catchment mean annual air  
85 temperature and precipitation are 7.7 °C and 103.3 cm, respectively, for the period from 1981 to  
86 2001. These data are recorded by a meteorological station operated by the North Rhine-  
87 Westphalian State Environment Agency with a distance of 4 km from the Rollesbroich  
88 catchment. The dominant soils are Cambisols in the southern part and Stagnosols in the northern  
89 part of the catchment. The grassland vegetation is dominated by perennial ryegrass (*Lolium*  
90 *perenne*) and smooth meadow grass (*Poa pratensis*). The average slope within the hydrological  
91 observatory is 1.63° (min.: 0.35°, max.: 3.12°).

92

## 93 **3. Methods**

### 94 **3.1 Meteorological data**

95 Meteorological data, i.e. precipitation, air temperature, air humidity, radiation components, and  
96 wind speed, were recorded at a micrometeorological tower located in the center of the almost flat  
97 terrain in the southern part of the Rollesbroich catchment (see Figure 1). Wind speed was  
98 obtained with a sonic anemometer at 2.6 m above surface (CSAT3, Campbell Scientific, Inc.,  
99 Logan, USA). The H<sub>2</sub>O concentration was measured using an open-path infrared gas analyzer  
100 (LI7500, LI-COR Inc., Lincoln, NE, USA) at the same height. Air temperature and humidity  
101 (HMP45C, Vaisala Inc., Helsinki, Finland) were measured at 2.6 m height above the ground  
102 surface. Incoming short- and longwave radiation were determined using a NR01 four-component  
103 net radiometer (Hukseflux Thermal Sensors, Delft, Netherlands). Data of all instruments  
104 including diagnostic data was recorded with a logger (CR3000, Campbell Scientific, Logan, UT,



105 USA) at 20 Hz. Precipitation was recorded by a heated Hellmann type tipping bucket rain gauge  
106 (eco-Tech GmbH, Bonn, Germany). In July 2013, a weight-based precipitation gauge (Pluvio<sup>2</sup>,  
107 OTT Hydromet GmbH, Kempten, Germany) was added to the nearby backup climate station,  
108 providing more accurate measurements of all precipitation types. Both precipitation gauges were  
109 installed at a height of 1 m above ground surface as recommended by the German Weather  
110 Service for elevations larger than 500 m a.s.l. and occasional heavy snowfall. All meteorological  
111 measurements were stored at 10 min intervals.

112

### 113 3.2 Actual evapotranspiration

114 Latent heat flux was obtained by the eddy covariance (EC) technique. The EC post processing  
115 software TK3.1 (Mauder and Foken, 2011) was used to calculate latent heat flux from the vertical  
116 wind velocity obtained by the sonic anemometer (CSAT3, Campbell Scientific, Inc., Logan,  
117 USA) and water vapor density obtained by an infrared gas analyzer (LI7500, LI-COR Inc.,  
118 Lincoln, NE, USA). The processing and quality assurance of the EC data followed the  
119 corresponding TERENO strategy presented in Mauder (2013). After receiving latent heat flux,  
120 actual evapotranspiration was calculated using:

$$ET_a = \frac{LH}{\rho_w * L_{water}} \quad \text{Eq. 1}$$

$$L_{water} = 10^{-3} * (2500.8 - 2.36 * T + 0.0016 * T^2 - 0.00006 * T^3) \quad \text{Eq. 2}$$

121 where  $ET_a$  is actual evapotranspiration ( $\text{m s}^{-1}$ ),  $LH$  is latent heat flux ( $\text{W m}^{-2}$ ),  $\rho_w$  is water density  
122 ( $\text{kg m}^{-3}$ ),  $L_{water}$  is latent heat of condensation of water in the temperature range from -25 to 40 °C  
123 ( $\text{J kg}^{-1}$ ), and  $T$  is air temperature (°C).

124



### 125 **3.3 Runoff**

126 Runoff is measured at the catchment outlet using a gauging station equipped with a combination  
127 of a V-notch weir for low flow measurements and a Parshall flume to measure normal to high  
128 flows. Runoff data of the two weir types are combined by using V-notch values for water levels  
129 below 5 cm, Parshall flume values for water levels greater than 10 cm and the weighted mean of  
130 V-notch and Parshall flume values for water levels between 5 and 10 cm, where the water levels  
131 refer to those of the V-notch weir.

132

### 133 **3.4 Soil water content**

134 Soil water content was at 87 SoilNet locations within the southern part of catchment (Figure 1)  
135 using SPADE soil moisture sensors (Qu et al., 2013; Hübner et al., 2009). The SPADE sensors  
136 were vertically installed at 5 cm, 20 cm and 50 cm depth. Two SPADE sensors were installed in  
137 parallel at each depth with a distance of ~10 cm to increase the sensing volume and to allow  
138 examination of inconsistencies in sensor reading. The measurement frequency was 15 min.

139

140 The SPADE sensor is a ring oscillator. The oscillator frequency is a function of the dielectric  
141 permittivity of the surrounding medium (Qu et al., 2013), which strongly depends on the water  
142 content of the soil because of the high permittivity of water ( $\epsilon_w \approx 80$ ) as compared to mineral soil  
143 solids ( $\epsilon_s \approx 2\sim 9$ ), and air ( $\epsilon_a \approx 1$ ). The temperature of the soil was also determined by the SPADE  
144 sensor using a digital thermometer (DS18B20) with an accuracy of  $\pm 0.5$  °C in the range from -10  
145 to 85 °C. The two-step calibration procedure suggested by Jones et al. (2005) was used to relate  
146 sensor reading to soil moisture. In a first step, reference liquids with a known dielectric



147 permittivity were used to calibrate the following empirical model (Eq. 3) that relates sensor  
148 reading to apparent dielectric permittivity:

$$K_a = \gamma + \frac{1}{\alpha + \beta/v} \quad \text{Eq. 3}$$

149 where  $K_a$  is the dielectric permittivity,  $\alpha$ ,  $\beta$ , and  $\gamma$  are the fitting parameters, and  $v$  is sensor output  
150 (unit, V). Prior to installation, 60 SPADE sensors were calibrated in five reference liquids that  
151 covered a permittivity range from 2.2 to 34.8. The outputs for the 60 sensors as well as the fitted  
152 model are shown in Figure 2. The root mean square error (RMSE) between known and predicted  
153 dielectric permittivity was 0.0188, and the best fitting parameters of  $\alpha$ ,  $\beta$ , and  $\gamma$  were -0.1502,  
154 0.3612, and -0.1599, respectively.

155

156 In a second step, a site-specific calibration between dielectric permittivity and soil water content  
157 was obtained using a complex refraction index model (CRIM, Eq. 4) as proposed by Birchak et  
158 al. (1974):

$$\theta = \frac{K_a^{0.5} - (1 - \eta) * K_{solid}^{0.5} - \eta * K_{air}^{0.5}}{K_{water}^{0.5} - K_{air}^{0.5}} \quad \text{Eq. 4}$$

159 where  $\eta$  is the porosity of the soil,  $1 - \eta$  is the solid fraction,  $K_a$  is the permittivity of soil, and  
160  $K_{water}$ ,  $K_{solid}$ , and  $K_{air}$  are the permittivity of water, solids, and air component of soil, respectively.  
161 In order to estimate the unknown value of  $K_{solid}$  and to assess the accuracy of this relationship, 15  
162 undisturbed samples (length = 7.7 cm, diameter = 5 cm) were taken from the two main soil types  
163 in 5, 20 and 50 cm depth. These samples were first saturated with deionized water and then CS  
164 640-L 3-rod TDR probes with a length of 7.5 cm were inserted in the middle of the sample.  
165 These probes were connected to a TDR100 system (Campbell Scientific, Inc., Logan, USA) to





166 determine the dielectric permittivity of the soil samples using a custom-made MatLab algorithm  
167 based on the travel time analysis algorithm (Heimovaara and Bouten, 1990). Subsequently, the  
168 samples were dried and both weight and dielectric permittivity of each sample were determined  
169 in regular time intervals. After drying at room temperature, the remaining water was removed by  
170 oven-drying at 105 °C for 24 hours so that the dry bulk density, porosity, and the volumetric soil  
171 water content could be determined from the recorded weights.

172

173 The measured dielectric permittivity and soil water content and the fitted relationships are shown  
174 in Figure 3. Because of the large difference in porosity at the three depths, the mean porosity at  
175 each depth was determined from the calibration samples and used to parameterize three different  
176 relationships. After fitting the solid permittivity for each depth, the performance of these  
177 calibration relationships was judged by the RMSE (Table 1). It was found that the three  
178 relationships performed well with a RMSE ranging from 0.022 to 0.028 cm<sup>3</sup> cm<sup>-3</sup>.

179

180 As already briefly discussed in Qu et al. (2014), we found that sensor output showed pronounced  
181 diurnal variations after the deployment of the sensor network. This behavior was attributed to a  
182 charging capacitor that affected the first reading of the SPADE sensor. If multiple sensor readings  
183 were made sequentially without turning off the sensor, the stability of the measurement  
184 considerably improved and the temperature dependence of the measurements disappeared (Qu et  
185 al., 2014). To correct these temperature-dependent oscillations in sensor reading, the acquisition  
186 software of SoilNet was changed temporarily so that two measurements were saved from 5<sup>th</sup>  
187 September 2012 to 3<sup>th</sup> March 2013. After 3<sup>th</sup> March 2013, the software was updated again so that  
188 only the second more accurate measurement was saved. Figure 4 shows an example of the



189 measured voltage for the first and second measurements and the soil temperature for a selected  
190 sensor (location 053, 5 cm depth). It can be seen that the difference between the two  
191 measurements ( $\Delta v$ ) is strongly correlated with soil temperature and could be fitted with a second  
192 order empirical polynomial with a RMSE of 5.18 mV (Figure 5). Such second-order polynomial  
193 functions were obtained for all sensors individually and subsequently used to correct  
194 measurements made between April 2011 and September 2012. After the correction,  
195 measurements from the closely-spaced sensors at a single measurement location agreed well with  
196 each other with a RMSE that varied from 0.010 to 0.035 cm<sup>3</sup> cm<sup>-3</sup>. The uncorrected and corrected  
197 voltage as well as the associated soil water content for the selected sensor is plotted in Figure 6. It  
198 is clear that the corrected measurements before September 2012 better match the expected soil  
199 water content after September 2012. On average, the corrected soil water content was 0.07 cm<sup>3</sup>  
200 cm<sup>-3</sup> lower than the uncorrected values. The plausibility of the corrected soil water content values  
201 is further supported by the fact that the unexpected increase in soil water content in the winter of  
202 2012 disappeared after the temperature correction. The corrected soil water content is now  
203 relatively constant in winter and the maximum of the soil water content corresponds well with the  
204 porosity determined from the soil samples (both 0.59 cm<sup>3</sup> cm<sup>-3</sup>).

205

### 206 **3.5 Soil physical properties**

207 Soil cores were taken at all locations where soil water content sensors were installed (length: 100  
208 cm, diameter 8 cm; Carl Hamm GmbH, Essen, Germany). Then, soil samples were taken from  
209 three pedological horizons (0-10, 10-20, and 20-40 cm) within the soil cores (see Figure 1),  
210 which resulted in a total of 282 soil samples. Sand, silt, and clay fractions were determined using  
211 a combination of wet sieving (sand fractions) and sedimentation (silt and clay fraction) following



212 ISO-11277 (2009). Total carbon content (C) and nitrogen content (N) were analyzed on sieved  
213 (<2 mm) and grounded samples by elemental analysis, and complemented with mid-infrared  
214 spectroscopy (ISO-10694, 1995). More details of the soil laboratory analyses can be found at  
215 Schiedung et al. (2015).

216

### 217 **3.6 Leaf area index**

218 The agricultural management of the different fields in the Rollesbroich catchment is very similar.  
219 Heterogeneity of the grass cover is mainly caused by different mowing times, which typically  
220 vary only by a few days. Therefore, we assume that the grass cover is homogeneous on the long-  
221 term in the Rollesbroich catchment. The effective leaf area index ( $LAI_{eff}$ ) that contributes to  
222 actual evapotranspiration was computed from grass height,  $h$ , using the following equations  
223 (Allen et al., 2006; Rochette et al., 1991):

$$LAI = 24 * h \quad \text{Eq. 5}$$

$$LAI_{eff} = \frac{LAI}{0.3 * LAI + 1.2} \quad \text{Eq. 6}$$

224 Average grass height in the Rollesbroich catchment was determined weekly by measuring grass  
225 height at five representative locations in the catchment.

226

227 Time series of LAI were also derived from RapidEye images using the normalized difference  
228 vegetation index (NDVI) approach (Myneni et al., 1997). The NDVI is calculated from infra-red  
229 (NIR) and red-edge spectral bands. The NDVI of a plant with high LAI has a high ratio between  
230 NIR and red reflectance that can be detected by RapidEye. The LAI was computed from the



231 NDVI values using a radiative transfer model (Myneni et al., 1997). Detailed information about  
232 the procedure can be found in Ali et al (2013).

233

#### 234 **4. Data management and quality control**

235 The distributed spatial data infrastructure TEODOOR (<http://www.tereno.net>) was developed to  
236 handle, describe, exchange, and publish all monitored environmental data of the TERENO  
237 project (Kunkel et al., 2013). Each institution hosting an individual observatory maintains its  
238 local data infrastructure. The observatories are connected via OGC-compliant web-services,  
239 while the TERENO Data Discovery Portal (DDP) as a central application enables data searching,  
240 visualization and download. According to the TERENO Policy it is obligated, that each dataset is  
241 described by standardized metadata elements, like ISO19115, OGC SensorML or NetCDF's CF-  
242 Conventions.

243 The observation datasets are processed and assessed within TEODOOR three different ways (for  
244 details see Devaraju et al., 2015). In first the data processing type, the imported data undergoes  
245 automated quality checks (e.g., minimum/maximum thresholds) and subsequently are published  
246 after visual inspection by experts. For example data from weather stations and river gauges are  
247 processed in this way. Secondly, more complex data can be externally processed and assessed by  
248 the principal investigators and subsequently imported into TEODOOR, e.g., eddy flux data. The  
249 third data processing type involves also automatic data import, but in this case the data quality  
250 assessment is executed using an external evaluation method developed by the responsible  
251 scientist, e.g. wireless sensor network data. Subsequently, the flagging information is updated by  
252 TEODOOR after the quality assessment has been completed.



253 Characterization of data quality is done by three descriptors, which are stored together with each  
254 observation: data uncertainty, data processing levels and data quality flags. The observed data  
255 values remain unchanged in any case. Data uncertainty arises from the observation process itself  
256 and is mainly determined by the accuracy of the sensors used. Data processing levels indicate the  
257 status of data handling. For instance, (unpublished) raw data is termed ‘level 1’, ‘level 2’ refers to  
258 data subjected to quality control, whereas the higher levels refer to derived data products. The  
259 flagging scheme consists of two tiers: the first tier includes generic flags, e.g., ‘good’,  
260 ‘unevaluated’, ‘suspicious’ or ‘bad’. The second tier is use-case-specific and indicates either the  
261 result of individual quality tests, e.g., failed gradient checks, or background events affecting data  
262 values, e.g., icing events.

263 In the following, we present in more detail how data was checked for plausibility to derive the  
264 data sets used in this study.

265

#### 266 **4.1 Meteorological data and latent heat flux**

267 Meteorological data was checked for quality by a multi step quality control including the use of  
268 diagnostic information provided by the instruments, the application of site specific plausibility  
269 limits, visual inspections of the data series and cross checks with data from the nearby backup  
270 weather station. The quality control of the latent heat flux was in accordance with the  
271 standardized method for the processing and quality assessment of eddy covariance data as  
272 suggested by Mauder et al. (2013). This scheme includes site-specific plausibility limits and the  
273 application of a spike removal algorithm based on median absolute deviation of raw  
274 measurements. Processed half-hourly fluxes and statistics were checked using three different  
275 flags (high, moderate, and low) based on tests of integral turbulence and stationarity (Foken and



276 Wichura, 1996). In this study, only data of high and moderate quality were used. A more detailed  
277 description of the treatment of eddy covariance data can be found elsewhere (Gebler et al., 2015;  
278 Post et al., 2015).

279

#### 280 **4.2 Runoff**

281 The gauging stations consist of a V-notch weir and a Parshall flume. As a first quality check, time  
282 series of both gauge types were compared for consistency. In addition, both runoff time series  
283 were visually inspected for inexplicable outliers (e.g. runoff peak without preceding rainfall  
284 event) and sensor failures. As outlined above, unreliable data were identified and appropriate  
285 flags were set.

286

#### 287 **4.3 Soil water content**

288 Measurements of soil moisture outside the physical plausibility range ( $0.05$  to  $0.85 \text{ cm}^3 \text{ cm}^{-3}$ )  
289 were identified and flagged. Subsequently, unreliable measurements were identified by analyzing  
290 the first derivative of the soil water content time series. In case an increase larger than two times  
291 the standard deviations of the preceding 24 hours was observed, this measurement was flagged as  
292 an unreliable measurement. In addition, the whole data set was visually inspected to verify the  
293 results of the automatic flagging procedures.

294



## 295 5. Data sets

### 296 5.1 Hydrometeorological data

297 Temporal dynamics of the most important meteorological data (i.e. air temperature, air humidity,  
298 radiation components, wind speed, precipitation, actual evapotranspiration, and runoff) and water  
299 balance components (i.e. precipitation, actual evapotranspiration, runoff, and soil water content)  
300 from 1<sup>st</sup> May 2011 to 31<sup>st</sup> December 2013 are plotted in Figure 7 and Figure 8. The air  
301 temperature, relative humidity, short wave radiation, and evapotranspiration showed a clear  
302 annual pattern. The highest runoff amounts occurred during the winter seasons due to high  
303 precipitation amounts and low evapotranspiration rates, as well as overland flow due to saturation  
304 excess (Figure 8). Generally, soil water content showed a strong dependence on precipitation  
305 events especially at 5 and 20 cm depth. Quick increases in soil water content can be observed  
306 after rainfall events, which were followed by a slow recession during periods without  
307 precipitation.

308

### 309 5.2 Water balance closure

310 The water balance can be written as:

$$P = R + ET_a + \Delta S \quad \text{Eq. 7}$$

311 where P is precipitation, R is runoff,  $ET_a$  is actual evapotranspiration, and  $\Delta S$  is the storage term.  
312 Because of the relatively low hydraulic conductivity ( $10^{-9}$  to  $10^{-7}$   $\text{ms}^{-1}$ ) of the aquifer bedrock  
313 (HK100, 2009), we neglected deep percolation. Mean average annual precipitation was  
314 partitioned into about 57% evapotranspiration and 50% runoff. The residual of the balance was  
315 within 7% of precipitation for the whole time period as shown in Figure 8. This residual is related



316 to measurement uncertainty and soil water storage depletion. As presented by Graf et al. (2014),  
317 soil water storage can be derived from the measured soil water content,  $\theta$ , as follows:

$$S(t) = \int \theta(x, y, z, t) dx dy dz dt \approx \sum_{i=1}^N c_i \theta(i, t) + \varepsilon \quad \text{Eq. 8}$$

318 where  $t$  is time,  $z$  is depth, the integral refers to the 3D domain as defined by the catchment  
319 boundary. The discrete version of storage is expressed by the right-hand side of the equation,  
320 where  $i$  is the number of sensors, and  $c_i$  is the empirical coefficient representing the 3D domain  
321 which is well represented by the sensor. The residual of  $\varepsilon$  is the storage affected by vegetation or  
322 groundwater and not represented well by the sensor. The changes in  $\Delta S$  are expected to  
323 correspond well with the residual of the major water budget component, P-ET<sub>a</sub>-R (Eq. 7). Thus,  
324 the time derivative of soil water storage should be linearly related to this residual, although  
325 measurement errors within the other water balance components and the unaccounted storage  
326 terms (e.g. vegetation storage) will deteriorate this relation. Our results show that there is a linear  
327 relationship between the derivative of soil water content and the water balance residual (Figure  
328 9). The  $R^2$  is 0.60, which means that 60% of the residual of the water balance is explained by soil  
329 water storage changes in the catchment.

330

### 331 **5.3 Soil physical properties**

332 The mean and standard deviation of %sand, %silt, % clay, organic carbon content, and bulk  
333 density are shown in Table 2. The bulk density ranged from 0.94 g cm<sup>-3</sup> to 1.52 g cm<sup>-3</sup> and  
334 generally increased with depth. Porosity ranged from 0.43 cm<sup>3</sup> cm<sup>-3</sup> to 0.65 cm<sup>3</sup> cm<sup>-3</sup> and  
335 decreased with depth. The higher spatial variability of porosity of the subsoil is caused by the  
336 higher and more variable stone content at this depth. In addition, former agriculture land





337 management activities reduced the spatial variability of porosity of the top soil (plough layer, 20  
338 to 30 cm thick). The measured soil texture fractions and soil density were used to estimate the  
339 spatial distribution of soil hydraulic properties with the pedotransfer function Rosetta (Schaap et  
340 al., 2001). Figure 10 shows the spatial distribution of Mualem-van Genuchten soil hydraulic  
341 parameters (van Genuchten, 1980) derived with Rosetta at 5 cm depth. Some soil hydraulic  
342 parameter show a distinct pattern. For instance, the VGM parameter  $\alpha$  is generally larger in the  
343 northern part than in the southern part of the catchment. Such information is important for the  
344 investigation of controlling factors of spatial patterns of soil water content.

345

#### 346 **5.4 Leaf area index**

347 The LAI derived from measured grass height agrees well with the LAI obtained from RapidEye  
348 images (Figure 11). Both LAI time series showed a distinct annual pattern with the highest values  
349 during the summer time. We averaged the monthly LAI derived from measured grass height and  
350 an ANOVA was conducted to test whether there was a significant difference between the LAI  
351 obtained from grass height and RapidEye in the time period of May to December in 2011. The  
352 results of this ANOVA (Table 3) confirmed that there is no significant difference between the  
353 two methods to determine LAI.

354

#### 355 **6 Conclusions and data access**

356 We presented data from the intensively instrumented hydrological observatory Rollesbroich  
357 providing long-term hydrometeorological data with high spatial and temporal resolution. Our  
358 results showed that the catchment water balance is reasonably closed by the provided  
359 measurements, and that 60% of the water balance residual could be related to soil water storage



360 changes within the Rollesbroich catchment. In addition, important soil physical and chemical  
361 properties (e.g. hydraulic properties) have been reported in addition to catchment-scale  
362 information on vegetation. This comprehensive hydrological data set can be used for the  
363 calibration, validation and improvement of hydrological models, e.g. in hydrological model  
364 intercomparison projects (Breuer et al., 2009; Maxwell et al., 2014; Refsgaard, 1997; Smith et al.,  
365 2004) and for the calibration and validation of remote sensing data products (Bastiaanssen et al.,  
366 1998; Jackson et al., 2010; Le Hegarat-Masclé et al., 2002; Njoku et al., 2003). All the data are  
367 freely available from the TERENO data portal (<http://www.tereno.net>). In addition, three  
368 persistent identifiers are associated with the data set described here:

- 369 • Climate/Runoff/Water Quality station: <http://doi.org/10.5880/TERENO.2016.001>
- 370 • EC/Climate station Rollesbroich 3: <http://doi.org/10.5880/TERENO.2016.002>
- 371 • SoilNet Rollesbroich: <http://doi.org/10.5880/TERENO.2016.003>

372

373

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381

382 **References**

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555 *Vadose Zone Journal*, v. 10, no. 3, p. 955-973.

556 **List of Tables**

557 Table 1. Parameters and the RMSE of the CRIM model for 5 cm, 20 cm, and 50 cm depth.

	5cm	20cm	50cm
$K_{\text{water}}$	78.54	78.54	78.54
$K_{\text{solid}}$	2.08	3.78	4.40
$K_{\text{air}}$	1.00	1.00	1.00
$\eta$	0.59	0.49	0.41
RMSE	0.028	0.025	0.022

558

559 Table 2. Descriptive statistics of soil properties determined from 273 soil samples taken in the

560 Rollesbroich catchment.

		Clay %	Sand %	Silt %	Bulk density ( $\text{gcm}^{-3}$ )	Carbon content ( $\text{gkg}^{-1}$ )	Porosity ( $\text{cm}^3\text{cm}^{-3}$ )
5 cm	mean	18.99	19.90	61.10	0.94	54.47	0.65
	std	2.00	3.82	3.79	0.12	15.82	0.05
20 cm	mean	18.03	20.76	61.20	1.28	34.08	0.52
	std	1.99	4.03	3.46	0.15	16.84	0.05
50 cm	mean	16.50	22.00	61.50	1.52	11.22	0.43
	std	2.40	5.68	4.53	0.16	6.01	0.06

561

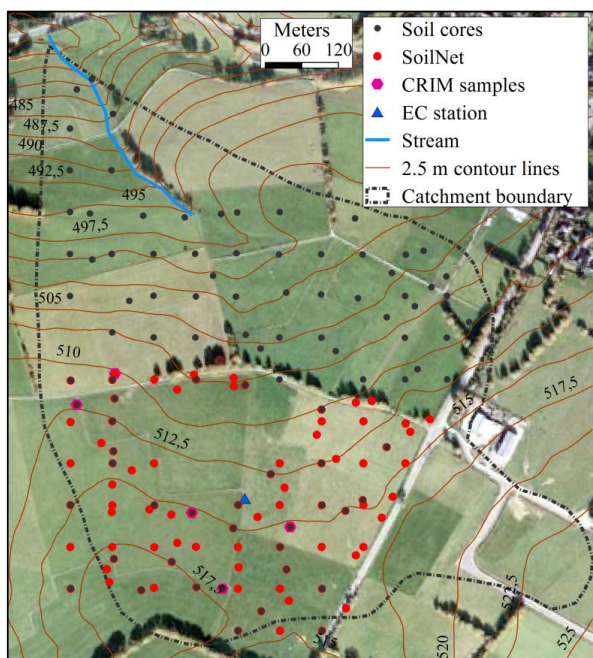
562 Table 3. ANOVA results of LAI determination with grass height and RapidEye from May to

563 December 2011.

Source of Variation	SS	df	MS	F	P-value	$F_{\text{critical}}$
Variability between group	0.311	1	0.311	0.717	0.411	2.145
Variability within group	6.074	14	0.434			
Total	6.385	15				



564 **List of Figures**

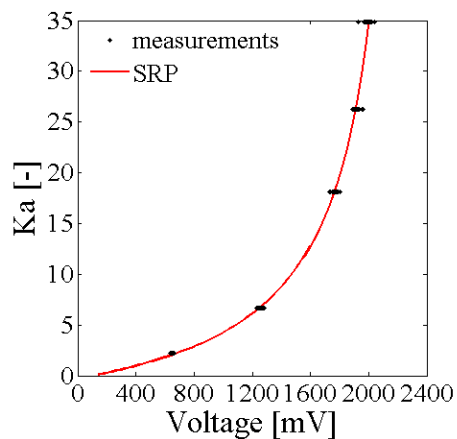


565

566 Figure 1. Map of the Rollesbroich catchment showing locations of the SoilNet sensor network,  
567 locations of the soil samples to determine soil physical and chemical properties, locations of the  
568 soil samples for site-specific calibration of the CRIM model, the location of the eddy covariance  
569 (EC) station, 2.5 m contour lines, and catchment boundary.

570



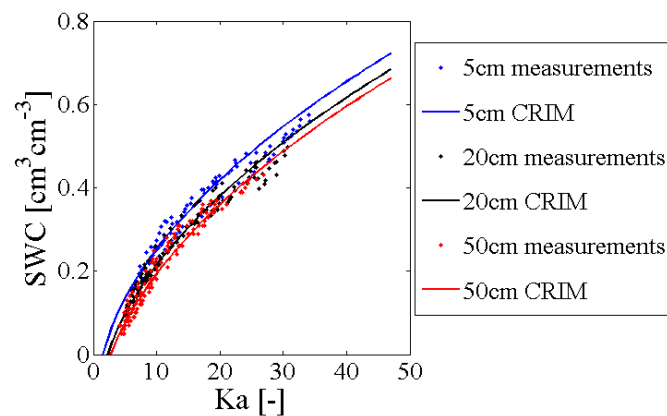


571

572 Figure 2. Sensor output of 60 SPADE sensors in five reference liquids. The fitted ‘universal’  
573 calibration relationship (Eq. 3) is also presented.

574

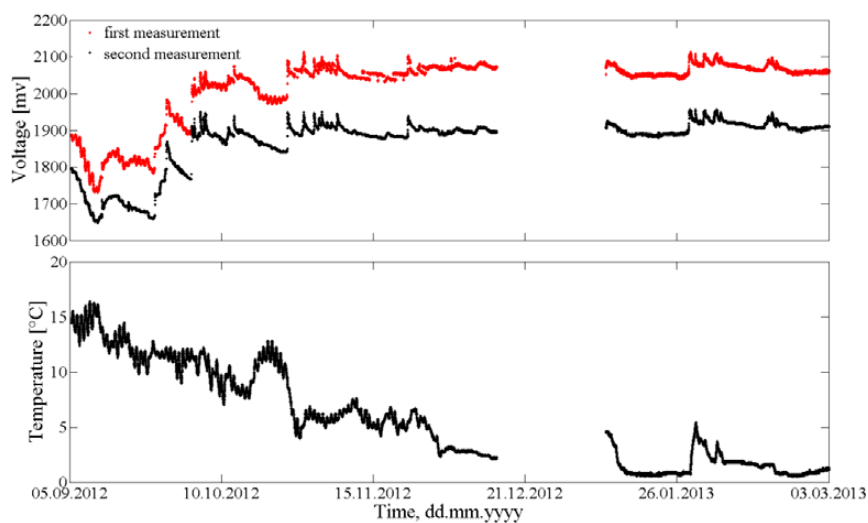
575



576

577 Figure 3. Relationship between dielectric permittivity and soil water content for the Rollesbroich  
578 test site and the derived  $K_a$ - $\theta$  models (CRIM).

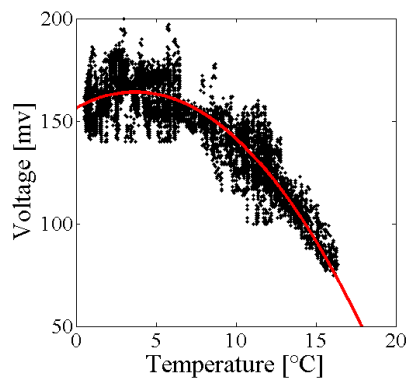
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580

581 Figure 4. Time series of the first and second voltage measured after software update and the  
582 associated temperature.

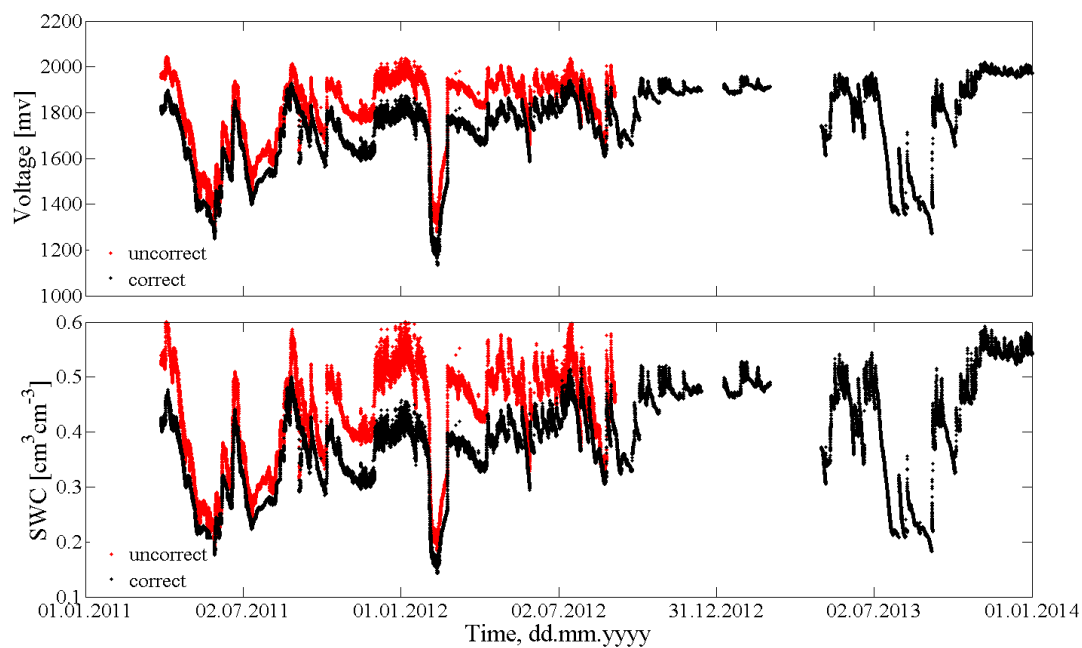
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584

585 Figure 5. Second-order polynomial function fitted to the relationship between soil temperature  
586 and the difference between the first and second voltage measurements.

587

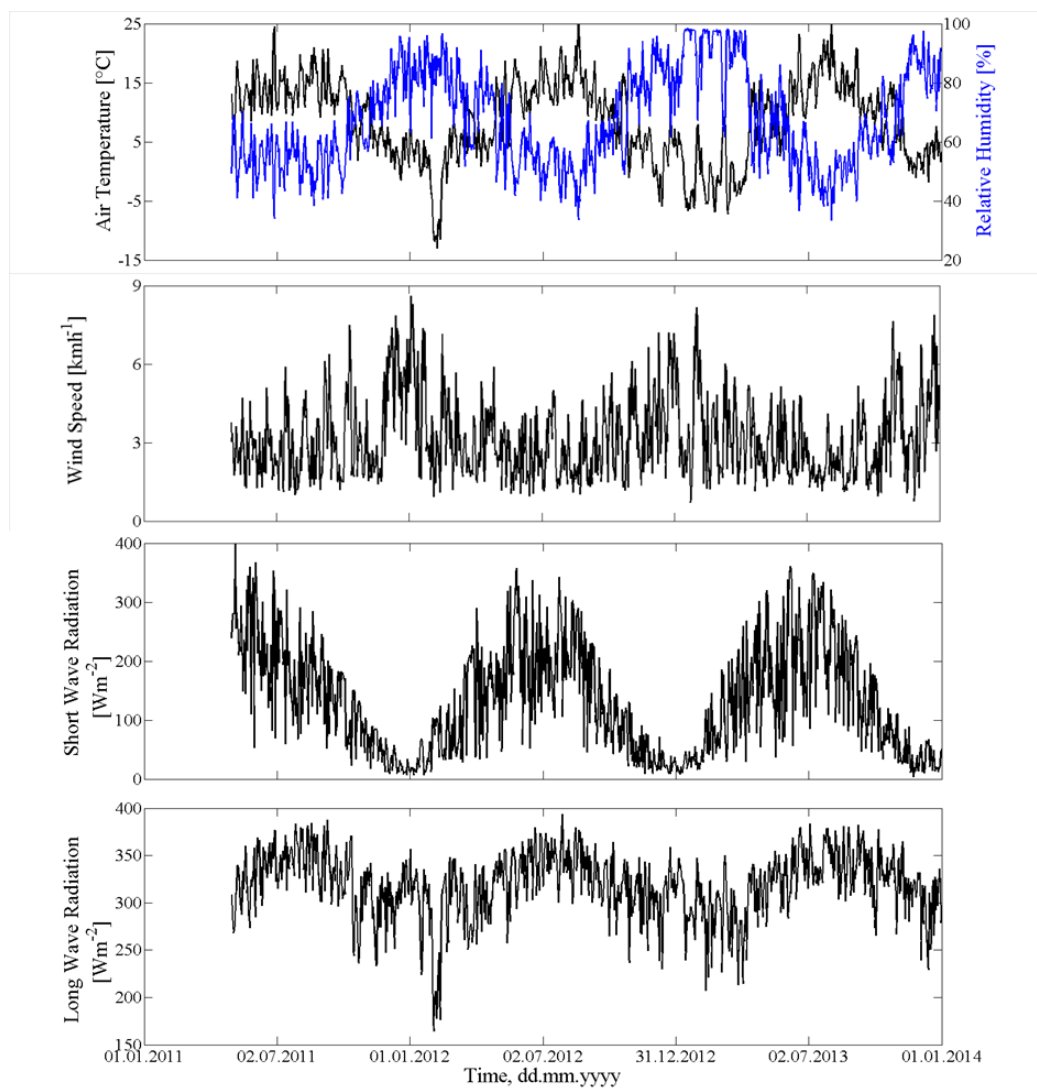


588

589 Figure 6. Uncorrected (red) and corrected (black) voltage and soil water content (SWC)

590 measurements.

591



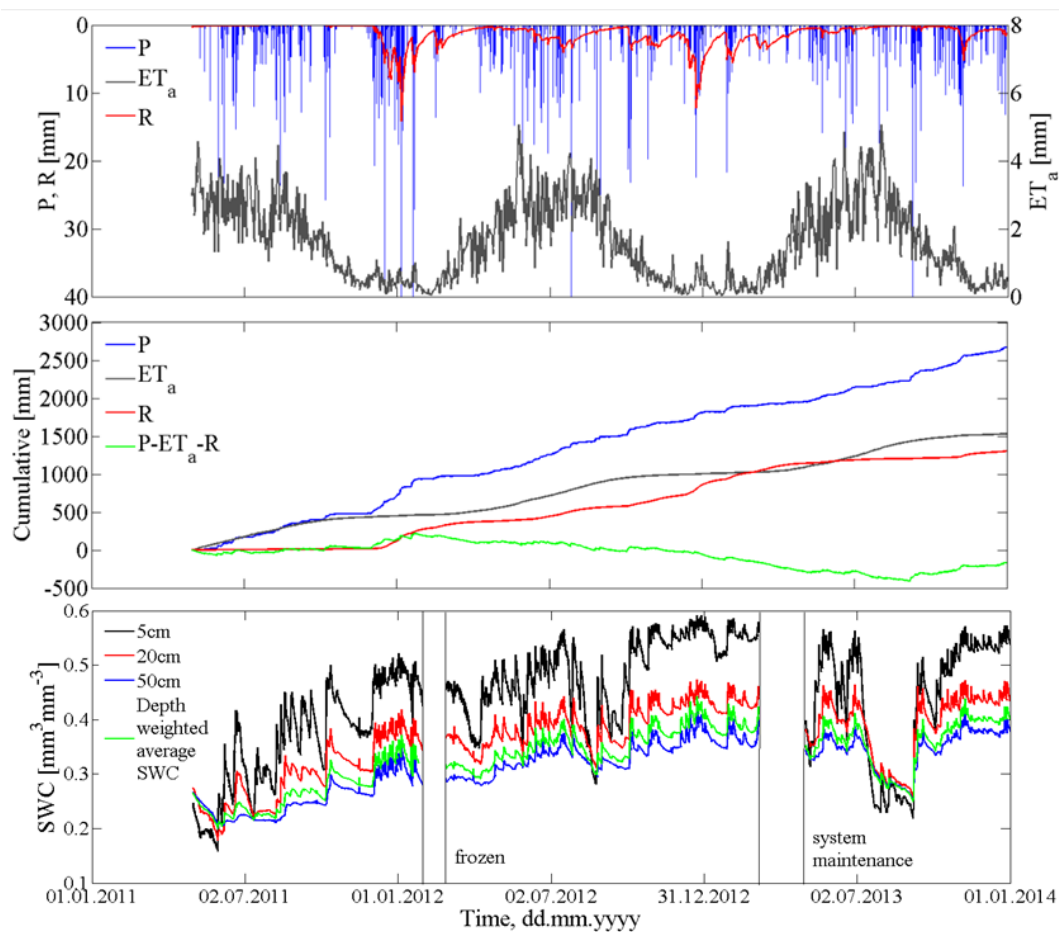
592

593 Figure 7. Daily averages of air temperature, relative humidity, wind speed, incoming short and

594 long wave radiation measured at the eddy covariance station from 1<sup>st</sup> May 2011 to 31<sup>st</sup> December

595 2013.

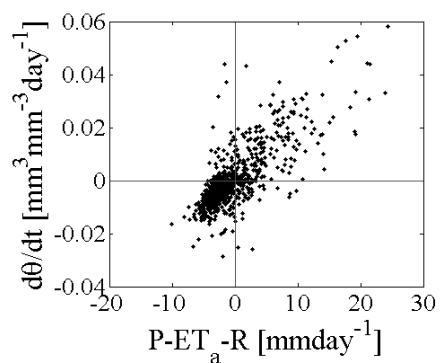
596



597

598 Figure 8. Daily and cumulative time series (1<sup>st</sup> May 2011 to 31<sup>st</sup> December 2013) of precipitation  
 599 (P), runoff (R), actual evapotranspiration (ET<sub>a</sub>), and spatial mean soil water content (SWC) at  
 600 three depths. The SWC data contains two major gaps due to frozen soil conditions and  
 601 maintenance of the SoilNet system.

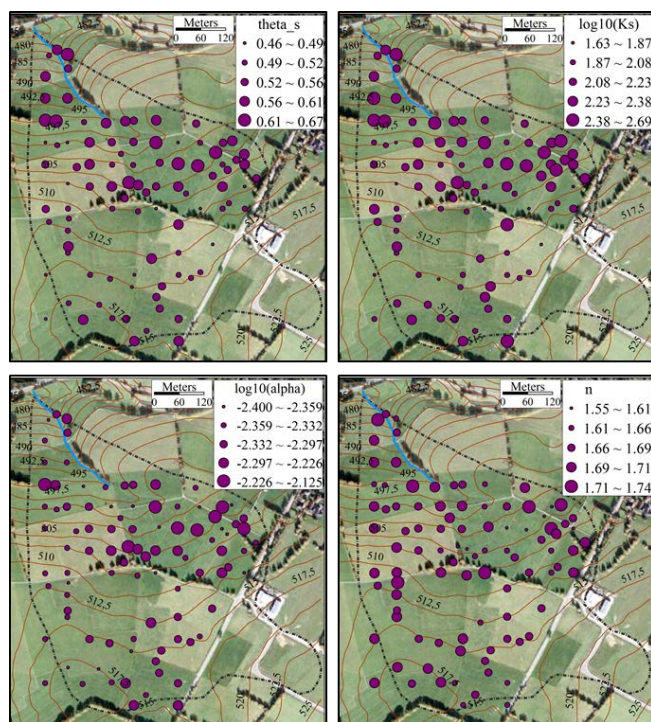
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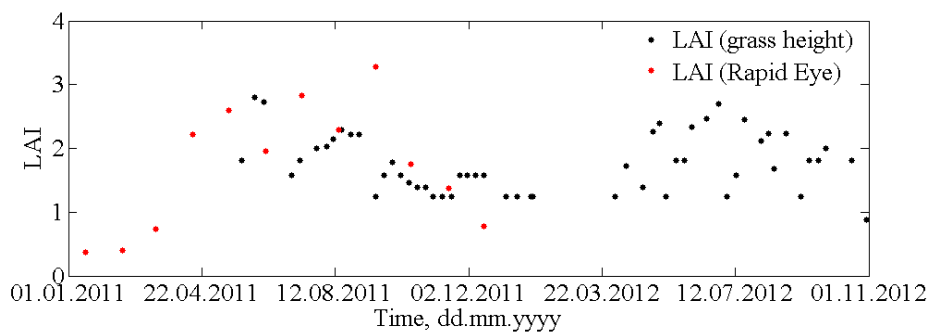
604 Figure 9. Time derivative of volumetric soil water content (average of all depth) versus the water  
 605 balance residual.

606



607

608 Figure 10. Spatial distribution of soil hydraulic properties ( $\theta_s$ ,  $\log_{10}(K_s)$ ,  $\log_{10}(\alpha)$ , and  $n$ ) at 5cm  
 609 depth derived from soil information obtained from the soil cores taken in the Rollesbroich  
 610 catchment.



611

612 Figure 11. Time series of leaf area index (LAI) computed from measured grass height and

613 extracted from the RapidEye images of the Rollesbroich catchment.

614