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The physical and biogeochemical parameters along the coastal waters of Saudi Arabia during field surveys in summer, 2021

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Abstract. During the last decades, the coastal areas of the Kingdom of Saudi Arabia, on the Red Sea and the Arabian Gulf, have been subjected to intense economic and industrial growth. As a result, it may be expected that the overall environmental status of Saudi Arabian coastal marine waters has been affected by human activities. As a consequence, adequate management of the Saudi Arabian coastal zone requires an assessment of how the various pressures within this zone impact the quality of seawater and sediments. To this end, environmental surveys were conducted over 15 hotspot areas (areas subject to environmental pressures) in the Saudi Arabian coastal zone of the Red Sea and over three hotspot areas in the Saudi Arabian waters of the Arabian Gulf. The survey in the Red Sea, conducted in June/July 2021, acquired measurements from hotspot areas spanning most of the Saudi coastline, extending from near the Saudi-Jordanian border in the north to Al Shuqaiq and Jizan Economic City (close to the Saudi-Yemen border) in the south. The survey in the Arabian Gulf, carried out in September 2021, included the areas of Al Khobar, Dammam and Ras Al Khair. The main objective of both cruises was to record the physical and biogeochemical parameters along the coastal waters of the kingdom, tracing the dispersion of contaminants related to specific pressures. Taken together, these cruises constitute the first multidisciplinary and geographically comprehensive study of contaminants within the Saudi Arabian coastal waters and sediments. The measurements acquired revealed the influence of various anthropogenic pressures on the coastal marine environment of Saudi Arabia and also highlighted a strong influence of hydrographic conditions on the distribution of biochemical properties in the Red Sea and the Arabian Gulf. The data can be accessed at SEANOE https://doi.org/10.17882/96463 (Abualnaja et al., 2023), whereas the details of the sampling stations are available at https://mcep.kaust.edu.sa/cruise-postings (last access: 25 March 2024). The dataset includes the parameters shown in Tables 1a, b and 2a.

1 Introduction

The Red Sea and the Arabian (Persian) Gulf are water bodies of importance to the Middle East and North Africa (MENA) region, particularly to the Kingdom of Saudi Arabia, which has coastlines on both water bodies.

An elongated, marginal oceanic basin, the Red Sea is bordered by northeastern Africa and the Arabian Peninsula. Spanning almost 20° in latitude, the Red Sea is more than 2200 km long and is roughly 200 km wide on average. The bathymetry of the Red Sea is characterized by a deep axial trench, exceeding 3000 m in depth, bordered by shallow (100-200 m) and broad coastal shelf platforms that cover more than 40% of the basin (Rasul et al., 2015). The shallow coastal areas are particularly wide in the southern Red Sea, where they take the form of extended shallow banks riddled with coral reef complexes. The coastal areas in the north are narrower than those to the south and reach depths of more than 200 m. The shallow Gulf of Suez and the deeper and narrower Gulf of Agaba are the natural extensions of the Red Sea to the north. The Red Sea communicates with the Mediterranean Sea through the Suez Canal in the Gulf of Suez; however, this water exchange is practically negligible, and the Red Sea relies on the Indian Ocean for its water renewal through the Bab-el-Mandeb Strait at its southernmost edge. In climatic terms, the Red Sea is divided into two distinct parts: the southern part, which is affected by the Arabian Sea monsoon, and the northern part, where a seasonal cycle of warm and cold periods prevails (Abualnaja et al., 2015; Viswanadhapalli et al., 2017). The elongated shape of the basin, the water exchange with the Gulf of Aden (Indian Ocean) and local atmospheric forcing regulate the water properties and general circulation of the Red Sea. Relatively fresh seawater enters the basin at its southern end, counterbalancing the water deficit produced by excessive evaporation coupled with negligible precipitation and terrestrial runoff. This fresher water from the south, with typical salinity values of 36–37.5, travels northwards through a complicated near-surface circulation (Sofianos and Johns, 2003, 2015; Yao et al., 2014; Zhan et al., 2014). The general northward surface circulation, which becomes stronger in winter, includes northward currents that flow near the east coast of the Red Sea and have been described previously as the Eastern Boundary Current (Sofianos and Jones, 2003). This boundary current affects all coastal regions, creating small eddies and bifurcating currents along natural barriers in the shallow areas near the coasts. During winter, the already hypersaline water in the Gulf of Aqaba and Gulf of Suez, along with the northern part of the Red Sea, cools and sinks to the intermediate and deep layers of the water column, depending on its density (e.g., Sofianos and Johns, 2003, 2015; Papadopoulos et al., 2013, 2015; Zhai et al., 2015; Yao and Hoteit, 2018; Asfahani et al., 2020). It must be mentioned that throughout most of the extent of the Red Sea the tidal component of the flow has a minor contribution on the total large-scale flow in the offshore areas, as well as the nearcoast ones (Churchill et al., 2014a; Pugh et al., 2019). The barotropic tidal flows are larger (up to $\sim 30-50 \text{ cm s}^{-1}$) near the area of the southern end of the Red Sea at the Bab-el-Mandeb Strait and also in the shallow Gulf of Suez (Guo et al., 2018).

Within the Saudi Arabian coastal zone of the Red Sea, the main drivers of pollution include maritime transport; fisheries; aquaculture; oil, gas, and energy infrastructures; coastal industry; coastal and maritime tourism; municipal and industrial discharge; and urban development (Schröder et al., 2021).

In contrast to the Red Sea, the Arabian Gulf is a shallow plateau with relatively smooth bottom topography, especially in its northern and western parts. It is a semi-enclosed oceanic basin that extends between the eastern part of the Arabian Peninsula and the mountainous coastline of Iran. The gulf exchanges water with the Gulf of Oman (Indian Ocean) through the Strait of Hormuz at its south-eastern edge. The mean depth of the Arabian Gulf is 35 m, with its maximum depth of 120 m found close to the Strait of Hormuz. The physical properties of Arabian Gulf water are regulated by the water exchange with the Indian Ocean and the high evaporation rate. It can be considered an idiosyncratic seawater basin, strongly influenced by its surrounding arid-to-hyper-arid areas. It is highly affected by dust storms, which transport large amounts of material and nutrients that influence the local ecosystems (Gherboudj and Ghedira, 2014). The gulf receives a small annual amount of precipitation, around 15 cm yr^{-1} (Reynolds, 1993), and a river discharge estimated at 18 cm yr^{-1} (Sheppard et al., 1992). However, damming of the major rivers in the region has resulted in substantial reductions in freshwater discharge (Sheppard et al., 2010). Evaporation is very high throughout the gulf, especially over the western part, and reaches approximately $200 \,\mathrm{cm}\,\mathrm{yr}^{-1}$, resulting in unusually high salinities compared to the open ocean (Johns et al., 2003). The Arabian Gulf features some of the highest temperatures and salinities observed in any marine water body worldwide



Figure 1. Station maps. (a) Sampling areas within the coastal zone of the Red Sea. A number of sampling stations were located in each area, as shown in Table 1. (b) Sampling areas in the Arabian Gulf. Detailed station grids for each area are shown in Fig. 1b-1, b-2 and b-3. Data for the bathymetry were used from GEBCO Compilation Group (2020).

(Sheppard et al., 2010). It is noteworthy that the maximum salinities within the gulf exceed 45 and are found along the Saudi Arabian, Bahraini and Qatari coastal areas. Regarding the tidal regime of the areas studied in the Arabian Gulf in the frame of this work, the tidal amplitude ranges between 0.5 and 1 m (Neelamani et al., 2021).

The current environmental status of the Arabian Gulf is changing and is increasingly impacted by numerous adverse anthropogenic factors (Al Azhar et al., 2016). Notable among these factors is activity associated with the petroleum industry (Jones et al., 2008). More than 50 % of the world's oil and natural gas reserves are located within the gulf. Other factors include desalination, reduced discharge from the two major rivers of the Tigris and the Euphrates (due to damming for the increased water demands), urbanization, and rapid industrial and residential development. The combination of these factors has substantially altered the environmental status of the Arabian Gulf, especially along the Saudi coast (Vaughan et al., 2019), and these have resulted in the Arabian Gulf being classified among the highest anthropogenically impacted regions in the world (Naser, 2013, 2014, 2015; Schröder et al., 2021).

Despite the economic and environmental importance of the Red Sea and Arabian Gulf to the MENA region, as well as the potential impact of anthropogenic activities on these water bodies, the quality of the marine environment of the Red Sea and the Arabian Gulf has not been extensively studied. Those studies that have been directed at the quality of the Saudi Arabian marine environment have largely been locally focused, e.g., on the marine area near Jeddah and the Jeddah lagoons. A number of studies have focused on trace metals (Al Farawati et al., 2011; Youssef, 2015; Brima and AlBishri, 2017; Al Mur et al., 2017; Ali et al., 2017; Fallatah et al., 2018; Al Mur 2020; Mannaa et al., 2021; Halawani et al., 2022; El Zokm et al., 2022) and/or organic pollutants (PAHs, polycyclic aromatic hydrocarbons) (El Maradny et al., 2023). Very few studies, also related mainly to metals and organic contaminants, are limited to local marine environments such as Al Lith (Abu-Zied and Hariri, 2016; Bantan et al., 2020), Yanbu (El Sorogy et al., 2023), and Jizan (Kahal et al., 2020). The sporadic and geographically confined nature of contaminant measurements acquired in the Red Sea highlight the importance of the dataset obtained in this work, since it extends along the whole Saudi Arabian coastline of the Red Sea.

Similarly, a few studies of contaminants have been conducted in the Saudi Arabian waters of the Arabian Gulf. These deal mainly with metals and petroleum (Freije, 2015; Alharbi et al., 2017; Al Kahtany et al., 2018; El Sorogy et al., 2018; Alharbi and El Sorogy, 2019, 2017; Paparella et al., 2022; Amin and Almahasheer, 2022; Alharbi et al., 2022; Alzahrani et al. 2023; Al-Kahtany et al., 2023; Sohaib et al., 2023).

As part of the Vision 2030 roadmap for economic growth and development in the Kingdom of Saudi Arabia, the Marine and Coastal Environment Protection (MCEP) Initiative for Saudi Arabia was established (https://mcep.kaust.edu.sa/, last access: 25 March 2024). The objective of this project, a collaboration between the National Center for Environmental Compliance (NCEC) and King Abdullah University of Science and Technology (KAUST), was to provide a national

Area	Station	Latitude_North (deg min)	Longitude_East (deg min)	Depth (m)	Date_June 2021
Cement Plant	HB01	24 15 499	37 33 048	603	16
	HB02	24 15.810	37 33.715	170	16
Yanbu	KF9	23 54.663	38 16.905	41	17
	KF8	23 55.278	38 15.550	30	17
	KF7	23 55.860	38 14.136	47	17
	KF5	23 55.197	38 11.796	104	17
	KF4	23 56 255	38 12 655	40	17
	KF6	23 57 383	38 11,590	29	17
	KF1	23 56 963	38 13 530	16	17
	KF3	23 56.500	38 12.865	38	17
Mena Ieddah	IM6	21 27 697	39.06.323	44	18
Wiella Jeddall	IM5	21 27.077	39 08 548	32	18
	IM3	21 27.775	30 00.040	32 26	10
	IM2	21 21.933	39 09.204	20	10
		21 28.073	39 09.018	13	10
	JMI	21 27.327	39 09.940	12	18
	JM4	21 27.220	39 09.318	15	18
Al Khumrah	JS4	21 19.300	39 05.701	81	18
	GH1	21 18.353	39 05.805	71	18
	JS8	21 19.335	39 05.653	80	18
	JS6	21 19.379	39 05.639	80	18
	JS10	21 19.245	39 05.725	80	18
	JS5	21 19.291	39 05.648	82	18
	JS9	21 19.280	39 05.750	75	18
	JS11	21 19.257	39 05.080	398	18
	JS12	21 19.271	39 05.347	295	18
	JS3	21 19.303	39 05.695	81	18
	JS13	21 19.930	39 05.410	80	18
Lagoon	L1	21 29.995	39 08.955	15	19
	L2 (off intended position)	21 29.665	39 09.123	10	19
	L2 (on position)	21 29.737	39 09.886	18	19
	L3	21 29.697	39 09.660	16	19
	L7	21 29.776	39 10.005	1.5	19
	L8	21 29.612	39 10.078	14	19
	L9	21 29.397	39 10.118	6	19
	L10	21 29.222	39 10.304	5	19
	L11	21 29.211	39 10.51	4	19
	L12	21 29 140	39 10 704	5	19
	 L13	21 29 368	39 10.892	5	10
	L14	21 29 50	39 11 008	45	19
	NII	21 28 965	39 10 667	ч.5 Д	10
	NI2	21 28.205	30 10 621	7 2	10
	NI3	21 29.286	39 10.330	3	19
	ΔΙ 7	20.08.500	40.12.000	12	20
AI LIUI		20 08.300	40 12.090	43	20
		20.06.030	40 13.040	18	20
	ALI	20.08.560	40 12.170	60	20
	ALS	20 08.560	40 12./10	24	20
	AL8	20 08.501	40 15.279	9.6	20
	AL4	20 08.972	40 14.140	11	20
	AL3	20 08.967	40 14.831	8.6	20

Table 1. List of sampling sites of the *AEGAEO* cruise in the Red Sea as seen in the National Center for Environmental Compliance (NCEC) database (https://mcep.kaust.edu.sa/cruise-postings, last access: 25 March 2024); list of sampling stations with location, depth and date (conductivity-temperature-depth (CTD) data have been measured at all stations).

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Table 1. Continued.

Area	Station	Latitude_North	Longitude_East	Depth	Date_June
		(deg min)	(deg min)	(m)	2021
Jazan Economic City	JZ1	17 21.610	42 15.590	15	22
	JZ2	17 20.970	42 15.680	16	22
	JZ3	17 20.270	42 15.840	15	22
	JZ4	17 19.600	42 16.000	15	22
	JZ5	17 18.950	42 16.090	14	22
	JZ6	17 18.290	42 16.260	14	22
	JZ7	17 17.610	42 16.400	10	22
	JZ8	17 16.970	42 16.550	12	22
	JZ9	17 16.290	42 16.710	15	22
	JZ10	17 15.640	42 16.850	15	22
	JZ11	17 15.000	42 17.000	12	22
	JZ12	17 14.320	42 17.170	12	22
	JZ13	1/13./00	42 17.300	15	22
Al Shuqaiq	SH8N OR SH8	17 36.588	42 05.061	15	22
	SH7	17 37.594	42 05.135	7	22
	SH6	17 37.928	42 04.390	8	22
	SH4N	17 38.386	42 04.643	4	22
	SHIN	17 39.395	42 04.108	3	22
	SH5	17 38.110	42 03.050	7	22
	SH4	1/ 38.643	42 03.395	8.6	22
RSP West	RSP3	25 27.020	36 42.210	17	27
	RSP5	25 24.860	36 41.470	22	27
	RSP6	25 25.370	36 46.470	20	27
	RSP1	25 33.361	36 40.688	13	27
	RSP2	25 29.181	36 40.776	10	27
	RSP4	25 26.969	36 41.145	29	27
RSP North	N5	25 55.920	36 37.850	20	27
	N1	25 53.736	36 40.507	5	27
	N2	25 53.492	36 39.258	11	27
	N3	25 53.015	36 38.035	15	27
	N4	25.51.758	36 37.147	7	27
Al Wajh	AW2	26 13.190	36 27.240	52	28
	AW3	26 13.710	36 26.890	185	28
	AW4	26 12.740	36 27.510	50	28
	AW1	26 13.479	36 27.652	9	28
	AW5	26 12.990	36 26.850	150	28
Duba Desalination	DBDS	27 21.180	35 39.640	258	28
Magna	MG	28 24.070	34 44.130	80	29
Haql	HQ1	29 17.600	34 55.670	78	29
	HQ2	29 17.650	34 55.310	295	29
	HQ3	29 17.070	34 55.540	81	29
	HQ4	29 18.120	34 55.790	180	29
	HQ5	29 21.120	34 56.700	270	29
Tabuk Fisheries	TB3	27 45.650	35 25.670	50	30
	TB4	27 45.100	35 25.890	100	30
	TB1	27 46.255	35 25.508	26	30
	TB2	27 46.089	35 25.775	21	30

 Table 2. A list of the acoustic Doppler current profiler (ADCP) measurements.

Station	Time (local time)	Latitude_North (deg min)	Longitude_East (deg min)	Depth (m)	Date_June 2021	ADCP type
HAB2	17:45	24 15.000	37 33.000	603	16	SADCP
HAB2	18:47	24 15.862	37 33.636	160	16	SADCP
KF9	03:18	23 54.663	38 16.905	44	17	Portable
KF8	04:03	23 55.284	38 15.455	38	17	Portable
KF7	05:05	23 55.880	38 14.153	49	17	Portable
KF5	06:15	23 55.176	38 11.796	103	17	SADCP
KF4	07:09	23 56.206	38 12.668	48	17	Portable
KF6	08:10	23 57.383	38 11.600	29	17	Portable
KF1	08:49	23 56.963	38 13.530	18	17	Portable
JM6	09:24	21 27.656	38 06.310	50	18	Portable
JM5	10:15	21 27.769	38 08.587	32	18	Portable
JM3	10:50	21 27.934	39 09.209	27	18	Portable
JM2	11:18	21 28.709	39 09.605	17	18	Portable
JMI	11:55	21 27.315	39 09.894	14	18	Portable
JM4	12:34	21 27.215	39 09.302	17	18	Portable
JS4	15:05	21 19.300	39 05.711	81	18	Portable
GHI	15:30	21 18.574	39 05.010	74	18	Portable
128	15:55	21 19.546	39 05.659	82	18	Portable
JS6	16:18	21 19.373	39 05.707	82	18	Portable
JS10	16:42	21 19.240	39 05.738	80	18	Portable
185	16:59	21 19.286	39 05.652	85	18	Portable
JS9	17:23	21 19.214	39 05.746	82	18	Portable
JSTI	18:22	21 19.262	39 05.033	346	18	SADCP
JS3	19:00	21 19.303	39 05.659	//	18	SADCP
AL7	13:14	20 08.509	40 12.096	45	20	Portable
AL2NEW	14:08	20 08.068	40 13.610	20	20	Portable
ALI	15:46	20 06.570	40 12.090	60	20	SADCP
ALS	16:47	20 08.560	40 12.710	26	20	Portable
JZ13	06:00	1/13.68/	42 17.302	17	22	Portable
JZ12	06:38	17 14.333	42 17.160	1/	22	Portable
JZ11 JZ10	06:59	17 15.007	42 17.000	13	22	Portable
JZ10 JZ0	07:16	17 15.646	42 16.865	17	22	Portable
JZ9 170	08:10	17 16.303	42 10.727	17	22	Portable
JZ8 177	08:31	17 16.979	42 16.554	13	22	Portable
JZ/ 17(08:40	17 17.020	42 16.403	14	22	Portable
JZ0 175	09:12	17 18.304	42 10.273	10	22	Portable
JZ3 174	09:30	17 18.962	42 16.102	17	22	Portable
JZ4 172	09:58	17 19.021	42 10.027	17	22	Portable
リム3 170	10:17	17 20.303	42 15.800	1/	22	Portable Dortable
JZム IZ1	10:53	17 21.000	42 13.704	10	22	Portable
5L1 SU5	11:02	1/21.024	42 13.018	18	22	Portable Dortable
505 DSD5	19:50	1/ 38.110	42 03.049	15	22	Portable
RSP6	00:43	25 24.051	30 41.497	20	27	Portable
DCD2	07:44	25 25.560	36 40.471	10	27	Portable
NJF J NJ5	16.27	25 27.022	36 27 000	19	27	Portable
AW5	10.57	25 55,950	26 26 960	156	21	SADCD
AWA	00.20	20 13.029	36 27 480	52	∠0 20	Portabla
ΔW2	07:44	20 12.723	36 27.409	52 50	20 20	Portable
AW3	00.50	20 13.201	36 26 041	180	∠0 28	SADCP
MG	09.19	20 13.074	34 11 155	100	20 20	SADCP
MG	00.10	20 24.077	24 44.133 24 AA 154	90 75	29	Portable
HO5	14.21	20 24.007	34 44.130	270	29 20	SADCP
HO4	14.51	29 21.095	34 55.005	180	29 20	SADCE
HO1	15.45	29 10.073	24 55.475	10U 92	29	Portable
HO1	10:24	29 17.399	34 55 672	03 02	29 20	SADCP
HO3	16.50	29 17.003	34 55 540	92 86	29 20	SADCE
HO3	10.38	29 17.003	24 55 211	202	29	SADCE
TR3	17.50	27 11.313	35 25 658	295 13	29 30	Portable
. Б .	12.43	21 + 5.058	55 45.058	+3	50	I UITAULE

Table 3. List of biogeochemical parameters in the water column of the Red Sea cruise in June 2021 with R/V *AEGAEO* as seen in the NCEC database. DO: dissolved oxygen; BOD: biochemical oxygen demand; TOC: total organic carbon; SPM: suspended particulate matter; TPH: total petroleum hydrocarbon.

						Samp	ling depth	ns in the	water co	lumn (m)				
Area	Station	DO	BOD	Fluo-	Sulf-	TOC	Chl-a	SPM	Chlo-	Cyan-	TPH	Oil	Chloro-	Metals
		and		ride	ides				rine	ide		and	phenols	
		nutrients										grease		
Cement	HB01	2	2	2	2	2	2	2	2	2	2	2	2	2
Plant		20 50	20	20 50	20 50	20 50	20 50	20 50	20 50	20 50	20	20 50	20 50	20 50
					50	50		50			50			
	HB02	2	2	2	2	2	$\frac{2}{20}$	2	2	$\frac{2}{20}$	2	2	2	2
		20 50	20 50	20 50	20 50	20 50	20 50	20 50	20 50	20 50	20 50	20 50	20 50	20 50
Vanhu	VEO	2	2	2	2	2	2	2	2	2	2	2	2	2
Tanou	KI'9	20	20	20	20	20	20	20	20	20	20	20	20	20
		40	40	40	40	40	40	40	40	40	40	40	40	40
	KF8	2	2	2	2	2	2	2	2	2	2	2	2	2
		27	27	27	27	27	27	27	27	27	27	27	27	27
	KF7	2	2	2	2	2	2	2	2	2	2	2	2	2
		20	20	20	20	20	20	20	20	20	47	47	47	20
		47	47	47	47	47	47	47	47	47				47
	KF5	2	2	2	2	2	2	2	2	2	2	2	2	2
		20	20	20	20	20	20	20	20	20	20	20	20	20
		104	104	104	104									
	KF6	2	2	2	2	2	2	2	2	2	2	2	2	2
		27	27	27	27	27	27	27	27	27	27	27	27	27
	KF1	2	2	2	2	2	2	2	2	2	2	2	2	2
		16	16	16	16	16	16	16	16	16	16	16	16	16
	KF3	2	2	2	2	2	2	2	2	2	2	2	2	2
		20 38	20 38	20 38	20 38	20 38	20 38	20 38	20 38	20 38	20 38	20 38	20 38	20 38
Mana	IM6	2	2	2	2	2	2	2	2	2	2	2	2	2
Jeddah	JINIO	43	43	43	43	43	43	43	43	43	43	43	43	43
	IM5	2	2	2	2	2	2	2	2	2	2	2	2	2
	J 1015	33	33	33	33	33	33	33	33	33	33	33	33	33
	IM3	2	2	2	2	2	2	2	2	2	2	2	2	2
	01110	25	25	25	25	25	25	25	25	25	25	25	25	25
	JM2	2	2	2	2	2	2	2	2	2	2	2	2	2
	IM1	2	2	2	2	2	2	2	2	2	2	2	2	2
	51011	12	12	12	12	12	12	12	12	12	12	12	12	12
	IM4	2	2	2	2	2	2	2	2	2	2	2	2	2
	51111	15	15	15	15	15	15	15	15	15	15	15	15	15
Al	189	2	2	2	2	2	2	2	2	2	2	2	2	2
Khumrah		17	17	17	17	17	17	17	17	17	17	17	17	17
	JS11	2	2	2	2	2	2	2	2	2	2	2	2	2
		14	14	14	14	14	14	14	14	14	14	14	14	14
	JS3	2	2	2	2	2	2	2	2	2	2	2	2	2
		25	25	25	25	25	25	25	25	25	25	25	25	25
		36	36	36	36	36	36	36	36	36	36	36	36	36

Table 3. Continued.

						Samp	ling depth	is in the	water col	umn (m)				
Area	Station	DO and nutrients	BOD	Fluo- ride	Sulf- ides	TOC	Chl-a	SPM	Chlo- rine	Cyan- ide	TPH	Oil and grease	Chloro- phenols	Metals
Lagoon	L1	2	2	2	2	2	2	2	2	2	2	2	2	_ 2
										15	15	15	15	
	L3	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15
	L9	2	2	2	2	2	2	2	2	2	2	2	2	2
	L10	2	2	2	2	2	2	2	2	2	2	2	2	2
	L12	2	2	2	2	2	2	2	2	2	2	2	2	2
	L14	2	2	2	2	2	2	2	2	2	2	2	2	2
	NI2	2	2	2	2	2	2	2	2	2	2	2	2	2
Al Lith	AL7	2	2	2	2	2	2	2	2	2	2	2	2	2
		23	23	23	23	23	23	23	23	23	43	43	43	23
		30 43	30 43	30 43	30 43	30 43	30 43	30 43	30 43	30 43				30 43
	AL2	2	2	2	2	2	2	2	2	2	2	2	2	2
		18	18	18	18	18	18	18	18	18	18	18	18	18
	AL1	2	2	2	2	2	2	2	2	2	2	2	2	2
		23 61	23 61	23 61	23 61	23 61	23 61	23 61	23 61	23 61	61	61	61	23 61
	AL5	2	2	2	2	2	2	2	2	2	2	2	2	2
		24	24	24	24	24	24	24	24	24	24	24	24	24
	AL8	2	2	2	2	2	2	2	2	2	2	2	2	2
	AL4	2	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11
	AL3	2	2	2	2	2	2	2	2	2	2	2	2	2
Jizan Econ.	JZ2	2 17	2 17	2 17	2 17	2 17	2 17	2 17	2 17	2 17	2 17	2 17	2 17	2 17
City	JZ5	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14
	JZ7	2 12	2 12	2 12	2 12	2 12	2 12	2 12	2 12	2 12	2 12	2 12	2 12	2 12
	JZ10	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14	2 14
	JZ12	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15
	JZ13	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15
Al	SH8 N	2	2	2	2	2	2	2	2	2	2	2	2	2
Shuqaiq		14	14	14	14	14	14	14	14	14	14	14	14	14
	SH6		2	2	2	2	. 2	2	2	2	2	2	2	
		5	5	5	5	5	-		-	-	-	-		5
	SH4N	2	2	2	2	2	2	2	2	2	2	2	2	2
	SH1N	2	2	2	2	2	2	2	2	2	2	2	2	2
	SH4	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8

Table 3. Continued.

						Samp	ling deptl	ns in the	water co	lumn (m)				
Area	Station	DO and nutrients	BOD	Fluo- ride	Sulf- ides	TOC	Chl-a	SPM	Chlo- rine	Cyan- ide	TPH	Oil and grease	Chloro- phenols	Metals
RSP	RSP3	2 16	2 16	2 16	2 16	2 16	2 16	2 16	2 16	2 16	2 16	2 16	2 16	2 16
	RSP5	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21	2 10 21
	RSP6	2 10 20	2 10 20	2 10 20	2 10 20	2 10 20	2 10 20	2 10 20	2 10 20	2 10 20	2 20	2 20	2 20	2 10 20
	RSP1	2	2	2	2	2	2	2	2	2	2	2	2	2
		12	12	12	12	12	12		12	12	12	12	12	2
	RSP2	9	9	9	9	9	9	- 2	9	9	9	9	9	9
	RSP4	2	2	2	2	2	2	2	2	2	2	2	2	2
	K01 +	28	28	28	28	28	28	- 2	28	28	28	28	28	28
RSP North	N5	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18
	N1	2	2	2	2	2	2	2	2	2	2	2	2	2
	N2	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11	2 11
	N3	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15	2 15
	N4	2	2	2	2	2	2	2	2	2	2	2	2	2
Al Wajh	AW2	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 50	2 50	2 50	2 20 50
	AW3	2 50 100 182	2 50 100 182	2 50 100 182	2 50 100 182	2 50 100 182	2 50 100	2 50 100 182	2 50 100 182	2 50 100 182	2 182	2 182	2 182	2 50 100 182
	AW4	2 30 47	2 30 47	2 30 47	2 30 47	2 30 47	2 30 47	2 30 47	2 30 47	2 30 47	2 47	2 47	2 47	2 30 47
	AW1	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8	2 8
	AW5	2 30 70 145	2 30 70 145	2 30 70 145	2 30 70 145	2 30 70 145	2 30 70	2 30 70 145	2 30 70 145	2 30 70 145	2 145	2 145	2 145	2 30 70 145
Duba Desalination	DBDS	2 50 100 254	2 50 100 254	2 50 100 254	2 50 100 254	2 50 100 254	2 50 100	2 50 100 254	2 50 100 254	2 50 100 254	2 254	2 254	2 254	2 50 100 254
Magna	MG	2 20 80	2 20 80	2 20 80	2 20 80	2 20 80	2 20 80	2 20 80	2 20 80	2 20 80	2 80	2 80	2 80	2 20 80

						Samp	ling deptl	ns in the	water col	lumn (m)				
Area	Station	DO and nutrients	BOD	Fluo- ride	Sulf- ides	TOC	Chl-a	SPM	Chlo- rine	Cyan- ide	TPH	Oil and grease	Chloro- phenols	Metals
Haql	HQ1	2 78	2 78	2 78	2 78	2 78	2 78	2 78	2 78	2 78	2 78	2 78	2 78	2 78
	HO2	2	2	2	2	2	2	2	-					2
	11.22	295	295	295	295	295	100	295	-					295
	HQ3	233 2 81	293 2 81	293 2 81	293 2 81	293 2 81	2 81	293 2 81	2 81	2 81	2 81	2 81	2 81	293 2 81
	HQ4	2 75	2 75	2 75	2 75	2 75	2 75	2 75	2 75	2 75	2 75	2 75	2 75	2 75
	HQ5	2 20 100 270	2 20 100 270	2 20 100 270	2 20 100 270	2 20 100 270	2 20 100	2 20 100 270	2 20 100 270	2 20 100 270	2 270	2 270	2 270	2 20 100 270
Tabuk Fisheries	TB3	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50	2 20 50
	TB4	2 30 70 100	2 30 70 100	2 30 70 100	2 30 70 100	2 30 70 100	2 30 70 100	2 30 70 100	2 30 70 100	2 30 70 100	2 100	2 100	2 100	2 30 70 100
	TB1	2 25	2 25	2 25	2 25	2 25	2 25	2 25	2 25	2 25	2 25	2 25	2 25	2 25
	TB2	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18	2 18

Table 3. Continued.

status quo assessment to set priorities for protection of the kingdom's coastal environment. The project was divided into seven tiered tasks (https://mcep.kaust.edu.sa/, last access: 25 March 2024). Guided by the findings of Task 4, Hotspot Analysis, KAUST's partner - the Hellenic Centre for Marine Research (HCMR) - undertook field surveillance at a number of sites in the Red Sea and Arabian Gulf. The sites were selected in consultation with the NCEC on the basis of the findings from Task 4 (https://mcep.kaust.edu.sa/, last access: 25 March 2024) (Schröder et al., 2021). The field surveillance, Task 6, was designed to trace the discrete sources of pollution in critical hotspot areas. These sources include wastewater treatment plants, desalination plants, ports, industry, petroleum platforms, aquaculture facilities (floating cages and onshore operations) and urban development. A variety of hydrographic and chemical properties in the water column and sediment were measured in each area.

Here, we present physical and biochemical measurements obtained during the two surveillance cruises, conducted in June and September 2021. Together, these cruises constitute the first multidisciplinary and geographically comprehensive survey of contaminants in the Saudi Arabian coastal zone. Here, these data are used to describe hydrographic conditions and the spatial variability of biochemical variables in the Red Sea and the Arabian Gulf. In addition, the baseline values for various pollutants (chlorophyll-*a* (Chl-*a*), nutrients, and rarely measured essential oceanographic variables, such as dissolved organic carbon) are presented along with their spatial distributions.

2 Data provenance

The survey in the Red Sea was carried out from the R/V *AE-GAEO* of the HCMR from 9 June to 6 July 2021, whereas the survey in the Arabian Gulf was carried out using small fishing boats from 17 to 22 September 2021. The cruise in the Red Sea was conducted over a north–south coastal transect from the area located in the northern part of the Gulf of Aqaba close to the Saudi Arabia–Jordan border (Hagl) to the Jizan Economic City area located in the southern region of the Red Sea (Fig. 1a). The hotspot areas surveyed, from north to south, were an area near the Saudi–Jordanian border (Phosphate Terminal in the Port of Aqaba in Jordan, crossborder pollution) and Haql (desalination, power and sewage-

Table 4. List of sediment parameters from the cruise in the Red Sea in June 2021 with R/V *AEGAEO* as seen in the NCEC database. WS Cl: water-soluble chloride; VOC: volatile organic compound; PCB: polychlorinated biphenyl; TN: total nitrogen.

Area	Name	Date_ June	Metals_ Granulo- metry	Hydro- carbons	Cyanide	WS Cl	VOC	Phenols	PCB	Org. C/Carbonate	TN
Yanbu	KF9	17	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	KF8	17		\checkmark		\checkmark		\checkmark		\checkmark	
	KF/ KF4	17		\checkmark	\checkmark	\checkmark	\checkmark	\checkmark		\checkmark	
	KF6	17	\sim	\sim	\sim	\sim	\sim	\sim	\sim	\sim	\sim
	KF1	17		٠ ٧	٠ ٧	٠ ٧	$\overline{\checkmark}$				
	KF3	17	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Mena	JM5	18	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Jeddah	JM3	18	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark		\checkmark	
	JM2 IM1	18 18	\sim	\sim	\checkmark	\sim	$\overline{\mathbf{v}}$	\sim		\sim	
A1 171	102	10	<u>v</u>	<u>v</u>	~ /	<u>v</u>	<u>v</u>	<u>v</u>	<u>v</u>	<u>v</u>	<u>v</u>
AI Knumran	183	18	~	~	~	~	\checkmark	~		~	~
Lagoon	L3	19		\checkmark	\checkmark	\checkmark	/	\checkmark		\checkmark	
	L9 L10	19	\sim	\sim	\sim	\sim	\checkmark	\sim	\sim	\sim	\sim
	L12	19							$\sqrt[n]{}$		
	L14	19		\checkmark	\checkmark	\checkmark	\checkmark	\checkmark		\checkmark	\checkmark
	NI2	19	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Al Lith	AL7	20		\checkmark	\checkmark	\checkmark		\checkmark			
	AL2	20		\checkmark			\checkmark	\checkmark		\checkmark	
	ALI AL5	20	\checkmark	\checkmark	\checkmark	\checkmark	\sim	\checkmark	\checkmark	\sim	\sim
	AL4	20	$\sqrt[n]{}$			$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	\sim	$\sqrt[n]{}$
	AL3	20									
Jazan	JZ2	22	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Economic	JZ5	22	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
City	JZ7	22		\checkmark	\checkmark	\checkmark	\checkmark	\checkmark		\checkmark	
	JZ10 IZ12	22	\checkmark	\checkmark	\checkmark	\checkmark	\sim	\checkmark	\checkmark	\sim	\checkmark
	JZ12 JZ13	22	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	\sim	$\sqrt[n]{}$
Al Shuqaiq	SH8N	22	./	./	./	./	./	./	./	./	./
7 li Shuquiq	SH6	22	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\sqrt[n]{}$	$\tilde{\mathbf{v}}$	$\sqrt[n]{}$
	SH1N	22									
	SH5	22								\checkmark	
	SH4	22	\checkmark							\checkmark	\checkmark
RSP	RSP3	27		\checkmark	\checkmark	\checkmark		\checkmark		\checkmark	
	RSP2 RSP4	27	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\sim	\checkmark
	K51 +	27		~	~	~		~	<u>v</u>	~	<u>v</u>
RSP North	N5A N1	27		\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	N2	27			\checkmark	\checkmark			$\sqrt[n]{}$	\sim	
Al Waih	AW3	28	• /	· · · · · · · · · · · · · · · · · · ·	/	/	/	· · · · · · · · · · · · · · · · · · ·	•		
Ai wajii	AW4	28	\sim	\sim	\sim	\sim	\sim	\sim	\sim	\sim	\sim
	AW1	28		٠ ٧	٠ ٧	٠ ٧	$\overline{\checkmark}$	v			
	AW5	28	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Duba Desalination	DBDS	28	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Magna	MG	29	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Haql	HO2	29	~	2	2	~/	./	2/	~/	./	./
1-	HQ5	29								$\sqrt[n]{}$	
Tabuk	TB3	30	~/	~	~/	~/	~/	~/	~/	~/	~/
Fisheries	TB1	30		Ň	Ň		v	, V	$\sqrt[n]{}$	v V	v
	TB2	30	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark

treatment plants, port activities), Magna (maritime traffic), Tabuk Fisheries (aquaculture activities), Duba (desalination plant), Al Wajh (port facilities, desalination plant), Red Sea Project Lagoon (north and west channels), Yanbu Cement Company (industrial discharge), King Fahad Industrial Port Yanbu (industrial and shipping center -, the largest in Saudi Arabia on the basis of oil and overall cargo volume; Schröder et al., 2021), Jeddah Lagoon system (wastewater inputs), Jeddah Mena (port operations), Al Khurmah (wastewater treatment plant), Al Lith (shrimp and fish farms), Al-Shuqaiq (desalination plant), and Jazan Economic City (expanding industrial facility). During the cruise, measurements of hydrographic and biochemical variables were acquired with traditional techniques (i.e., use of a CTD with companion sensors, acoustic Doppler current profiler (ADCP) and collection of water samples). The sampling strategy was aimed at resolving the dispersion of contaminants related to environmental pressures impacting each area. For example, when sampling near desalination and wastewater treatment facilities, we first conducted a CTD survey, acquiring data on temperature, salinity, dissolved oxygen (DO) and turbidity to identify the signal of the discharge from these facilities. Subsequently, water and sediment samples were obtained within the discharge signal suggested by the CTD data and at a point distant from the signal (reference site). Samples of water and sediment, as well as CTD data, were acquired from both R/V AEGAEO and AEGAEO's support vessel (tender). At deeper (> 10 m) stations, water samples and vertical CTD profiles were acquired using AEGAEO's rosette sampling system, which consists of a CTD profiler with twelve 12L Niskin bottles. Sediment samples at the deeper stations were acquired with the ship's box corer. At shallower stations, the sampling was carried out using the tender. At these shallower stations, sediment samples were acquired with a small grab sampler, and water samples were taken using a single Niskin bottle lowered from the tender. Water samples were taken at discrete depths, such as surface, 10 m, 20 m and near the bottom (roughly 0.5 m from the seabed), as well as from depths of particular interest (e.g., where effluent plumes were identified) if they did not match with the discrete depths.

The shipboard CTD system included a sensor suite for acquiring in situ data of salinity, temperature, DO, pH, specific conductivity, total dissolved solids, turbidity, Chl-*a* and total suspended matter. Onboard analysis was performed on the water samples for ammonium, DO, biological oxygen demand (BOD), sulfides, fluoride and total chlorine. The water samples were also subjected to onboard particle filtration and extraction of organic compounds.

The surveillance cruise in the Arabian Gulf covered the critical hotspot areas predefined by Schröder et al. (2021). Three main areas were sampled: Al Khobar, Dammam and Ras Al Khair. Unfortunately, rough sea conditions prevented surveying more areas, particularly in the northern part of the Saudi Arabian waters. These will hopefully be surveyed in future field surveillance work. The sampling entailed mea-

surements of seawater properties using a portable CTD unit and the collection of seawater and sediment samples from which various types of contaminants and other properties were measured. Altogether, CTD data, as well as seawater and sediment samples, were obtained from 14 coastal stations (Fig. 1b). Similar to the Red Sea surveillance, the sampling strategy was aimed at resolving the dispersion of contaminants related to environmental pressures impacting each survey area. Sediment samples were acquired with a small grab sampler, and water samples were taken by individual Niskin bottles lowered from the boat. Water samples were taken at discrete depths: surface, 10 m and near the bottom (roughly 0.5 m from the seabed). The portable CTD data included in situ measurements of temperature, conductivity and turbidity. On-board chemical analyses were not performed, these were not possible on the small boat employed. Nevertheless, all samples were treated and were subjected to extraction of organic compounds onboard immediately after collection. The water samples were subjected to particle and Chl-a filtration at the ALS Arabia laboratories

Data coverage and parameters measured

Coverage: 17–29° N and 34–42° E Location name: Red Sea Date start: 9 June 2021 Date end: 6 July 2021

A link to the summary page of the Red Sea cruise can be found in the NCEC database under https://mcep.kaust.edu. sa/cruise-postings (last access: 25 March 2024).

Coverage: 26–27° N and 49–50° E Location name: Arabian Gulf Date start: 17 September 2021 Date end: 22 September 2021

3 Methods

3.1 CTD - rosette

In the Red Sea, 96 CTD casts were performed. A total of 71 casts included rosette/Niskin water sampling (full details are given in Tables 1 and 2) while the other 25 acquired only sensor data. A Sea-Bird SBE 9plus CTD underwater unit (Sea-Bird Electronics, Inc.) connected to an SBE 11 deck unit was used. The SBE 9plus CTD system was equipped with a pressure sensor (Digiquartz), temperature sensor (SBE 3), conductivity sensor (SBE 4), DO (SBE 43), transmissometer sensor (Chelsea AlphaTracka Mk II), fluorometer sensor (Chelsea AquaTracka III), pH sensor (SBE 18) and underwater and shipborne photosynthetically active radiation (PAR) sensor (Satlantic). Temperature and conductivity sensors were calibrated before the cruise, in February/-March 2021, at the manufacturer's facilities. Oxygen values were corrected post-cruise using the Winkler oxygen values. The underwater unit was attached to a metallic frame that

Area	Latitude_North (deg min)	Longitude_East (deg min)	Name	Water samples	Sediment sample	Max depth (m)	Samples depth (m)
				(S: near surface) (M: mid-water) (B: near bottom)	(Sed)		
Ras Al Khair	27 33.282	49 08.574	RK1	S	Sed	5.3	2
	27 34.401	49 08.661	RK3		Sed	11	
	27 34.024	49 12.458	RK4	S	Sed	17.6	2
	27 40.434	49 20.584	RK7	S	Sed	13.5	2
	27 33.845	49 08.603	RK2			9	
Dammam	26 29.186	50 08.911	DM1	S	Sed	5.4	2
	26 31.080	50 09.289	DM2		Sed	6.5	
	26 32.109	50 09.738	DM3	S	Sed	7.7	2
	26 43.807	50 17.341	DM4	S, B	Sed	11	2, 10
Al Khobar	26 15.172	50 15.512	RJ1	S	Sed	7.2	2
	26 13.640	50 14.847	RJ2			8.2	
	26 12.315	50 14.310	RJ3	S		6.4	
	26 11.058	50 13.620	RJ4	S		7.2	5
	26 09.757	50 12.355	RJ5			8.2	

Table 5. List of sampling sites from the cruise in the Arabian Gulf in September 2021 with location, depth, date and sampling information (CTD data have been measured at all stations).

Table 6. List of biogeochemical parameters in the water column (sampling depth) from the cruise in the Arabian Gulf in September 2021.

Area	Station	DO and nutrients	Fluoride	TOC	Chl-a	SPM	Cyanide	TPH	Oil and grease	Chlorophenols	Metals
Ras Al Khair	RK1	2	2	2	2	2	2	2	2	2	2
	RK4	2	2	2	2	2	2	2	2	2	2
	RK7	2	2	2	2	2	2	2	2	2	2
Dammam	DM1	2	2	2	2	2	2	2	2	2	2
	DM4	2	2	2	2	2	2	2	2	2	2
	DM4	10	10	10	10	10	10	10	10	10	10
Al Khobar	RJ1	2	2	2	2	2	2	2	2	2	2
	RJ3	2	2	2	2	2	2	2	2	2	2
	RJ4	5	5	5	5	5	5	5	5	5	5

supported 12 Niskin bottles of 12 L volume each. The bottles were connected to a carousel water sampler (SBE 32) linked to the CTD probe via a conductive cable.

In addition, a portable SBE 19 CTD underwater unit equipped with a pressure sensor (strain gauge), unitembedded temperature sensor, unit-embedded conductivity sensor, transmissometer sensor (Chelsea/Seatech) and fluorometer sensor (WET Labs ECO-AFL/FL) was used in the shallow areas. Both CTD sensors recorded measurements in the whole water column from near surface to ~ 0.5 m above the sea bottom. Water samples for the rosette system were taken for the quality control/quality assurance (QC/QA) of the results derived from the CTD probes. Moreover, an inter-calibration test cast was conducted before sampling, in which both the SBE 9 and SBE 19 probes were attached to a frame and lowered in the water column. Minor adjustments were applied to correct the SBE19 sensors. In total, 36 casts were conducted with the tender's CTD unit. Raw CTD data were processed following the manufacturer's recommendations and procedures using the SBE Data Processing software. Accuracies of CTD sensors are shown in Table 8.

The physical properties derived from the CTD and companion sensor data provided indicators of pressure signals. For example, salinity, fluorescence and beam transmission were taken as tracers for brine water (e.g., from desalination plants), wastewater effluent (e.g., from effluent outfalls).

In the Arabian Gulf, data were acquired with a portable Idronaut Ocean Seven CTD unit equipped with pressure, temperature, conductivity, DO, pH and turbidity sensors. All sensors attached to the unit were calibrated before the cruise,

Area	Name	Metals Granulometry	Hydrocarbons	Cyanide	WS Cl	VOC	Phenols	PCB	Org. C/Carbonate	TN
Ras Al Khair	RK1	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	RK3	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	RK4	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	RK7	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Dammam	DM1	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	DM2	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	DM3	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
	DM4	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark
Al Khobar	RJ1	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark	\checkmark

Table 7. List of sediment parameters from the cruise in the Arabian Gulf in September 2021.

Table 8. Accuracy of the CTD sensors.

	SBE 9
Sensor	Accuracy
Temperature (SBE3)	±0.001 °C
Conductivity (SBE4)	$\pm 0.0003 \mathrm{S}\mathrm{m}^{-1}$ ($\sim \pm 0.003$
	in salinity)
Oxygen (SBE43)	$\pm 2\%$ of saturation value
pH (SBE18)	$\pm 0.1 \mathrm{pH}$
	SBE 19
Sensor	Accuracy
Temperature	±0.005 °C
Conductivity	$\pm 0.0005 \mathrm{S} \mathrm{m}^{-1} \ (\sim \pm 0.005 \mathrm{m}^{-1})$
	in salinity)

in September 2021, at the Coastal and Marine Resources (CMR) Core Lab of KAUST according to the manufacturer's instructions. Arabian Gulf water samples were taken using Niskin bottles.

3.2 ADCP

The R/V AEGAEO was equipped with a portable ADCP operating at 300 kHz. The ADCP (Teledyne Workhorse) measured water current magnitude and direction down to a depth of approximately 50 m in optimal conditions. The instrument was deployed on the side of the AEGAEO at shallow stations with the bottom tracking mode enabled. This ADCP was typically set up to measure velocities in 2 m depth bins, collecting 4–5 (ensemble) profiles at 6–8 min intervals. The average velocity profile at each station was computed as a mean of these ensembles with questionable-quality (biased and noisy) measurements removed. At 15 deeper stations where depths ranged from ~ 70 to ~ 600 m, ADCP measurements were collected for 10–15 min via the shipboard ADCP (SADCP Teledyne Ocean Surveyor), a unit installed at the ship's hull and operating at 75 kHz. Averaged velocity profiles were obtained at a total of 59 stations. Unfortunately, ADCP measurements were not taken in the Arabian Gulf.

3.3 DO, nitrogen and phosphorus, Chl-*a*, total organic carbon (TOC) and total suspended solids (TSSs)

Seawater samples were collected from the rosette/Niskin sampling system with the typical precaution to prevent any biological activity and gas exchange with the atmosphere (Strickland and Parsons, 1968). Winkler glass bottles with beveled glass stoppers and a measured capacity specification of ± 0.01 mL were used. Chemical reagents were added immediately after sampling. DO was determined by titration using the Winkler method, according to Carpenter (1965a, b). Reagent 1 (manganese) and reagent 2 (alkaline iodide) were added with semi-automated Eppendorf (Germany) dispensers, and the bottles were kept in the dark. Before titration, the precipitated hydroxides were dissolved with sulfuric acid, and the solution was carefully transferred to a titration beaker. The titration was carried out with a standardized thiosulfate solution using a Metrohm 876 Dosimat (Switzerland). QC/QA was performed daily by standardization of the thiosulfate solution with a reference standard solution of potassium iodide. The precision of the method outlined above is estimated to be 2.2 μ mol L⁻¹.

The onboard determination of DO concentration from the seawater samples provided data for QC/QA of the CTD DO sensor measurements and enabled calibration of the CTD's DO sensor (SBE 43) during the cruise.

For nutrient analysis, seawater samples were collected in tubes amenable to the QuAAtro nutrient autoanalyzer (SEAL Analytical) and were also collected in 125 mL bottles. All sample containers were pre-treated with 10% HCl and distilled water. The samples were collected in triplicate and kept deep-frozen (-20 °C) until their subsequent analysis in the HCMR biogeochemical laboratory, which is certified according to EN ISO/IEC 17025:2017 (2018). In the laboratory, the samples were analyzed using the QuAA-



Figure 2. T-S diagram encompassing all data taken (averaged every 1 m) in the Red Sea. Dots of the same color indicate measurements taken at the same site as shown on the map on the right (© Google Maps 2021).



Figure 3. Vertical profiles of Potential Temperature and Salinity in the study areas of the Red Sea.

tro nutrient autoanalyzer (SEAL Analytical) according to standard methods (Strickland and Parsons, 1968; Murphy and Riley, 1962). Ammonium was determined based on the spectrophotometric measurement of the blue-colored indophenol complex formed by the reaction of phenol and hypochlorite in the presence of the NH_4^+ and NH_3 species (Koroleff, 1970). Absorbance was measured using a UV-2600 Shimadzu UV/Visible spectrometer (USA) with an 8 mL volume, 10 cm path length cell. QC samples were analyzed together with the field samples. Results were calculated according to a calibration curve based on the EN ISO/IEC 17025:2017 standard, followed by the HCMR Laboratory QUASIMEME inter-calibration exercise used for testing the accuracy of the method. The limit of detection (LOD) values of the certified methods were 0.04 μ mol L⁻¹ for nitrate+nitrite, 0.01 μ mol L⁻¹ for nitrite, 0.03 μ mol L⁻¹ for nitrate, 0.01 μ mol L⁻¹ for phosphate and 0.02 μ mol L⁻¹ for ammonium.

For organic phosphorus and nitrogen analysis, seawater was collected in 20 mL Teflon bottles (for phosphorus) and 50 mL glass bottles fitted with screw caps. These digestion bottles were previously washed with 10 % HCl. Preliminary digestion was performed to remove traces of organic matter. The bottles were then kept continuously in a deep freeze $(-20 \,^{\circ}\text{C})$ in the dark until their contents were analyzed in the HCMR laboratory. In the laboratory, digestion was performed by a persulfate wet-oxidation in low-alkaline conditions at 120 $^{\circ}\text{C}$ (1 bar) for 30 min. After cooling at room temperature, the assay mixture was analyzed for nitrate and phosphate on a QuAAtro nutrient autoanalyzer (Seal Analytical) according to the methods for nutrient analysis.

Samples for TOC analysis were collected in 20 mL acidcleaned (10% HCl, 12 h) glass bottles fitted with Teflon cups. Directly after sampling, 50 µL of 2N HCl was added to each bottle. The samples were then refrigerated until analysis. TOC concentrations were determined using a Shimadzu TOC-L organic carbon analyzer following the high-temperature catalytic oxidation (HTCO) method described by Cauwet (1994) and Sugimura and Suzuki (1988). TOC concentration was calculated as the average value of three replicates that yielded a standard deviation < 2%. Analytical precision and accuracy were tested daily against Deep Atlantic Seawater Reference Material provided by the DOC-CRM program (University of Miami – D.A. Hansell). The certified value of the reference material is 0.533–0.574 mg L⁻¹, and the measured values (*n* = 10) dur-



Figure 4. Current vectors at 12 m at specific areas along the Saudi Arabian coast of Aqaba and Red Sea. Scale vector of a westward current with a speed of 15 cm s^{-1} appears in the upper right of the top panel (© Google Maps 2021).

ing the analysis of the samples were between 0.532 and 0.570 mg L^{-1} .

Water samples for the determination of Chl-*a* concentrations (μ g L⁻¹) were collected by Niskin bottles principally near the surface of the water column and at the depth of the deep chlorophyll maximum (DCM) concentrations. For the estimation of the phytoplankton biomass, seawater samples were filtered on board through Whatman GF/F microfiber filters. A volume of 1 or 2L of seawater were filtered depending on the expected concentrations of Chl-*a*. The filters were kept in a deep freezer in the dark at -15 °C, and were than analyzed at the laboratory on a TURNER 00-AU-10 fluorometer. TSSs were determined by a standard method in which 4 L of seawater was collected in polyethylene bottles and filtered on board through pre-weighed glass fiber filters.

3.4 BOD, sulfide, fluorine, cyanide and total chlorine in seawater

The determination of BOD was performed on board immediately after seawater sampling according to Standard Methods for the Examination of Water and Wastewater no. 5210b (EPA 5210B, 2023). Seawater samples were taken from the Niskin bottles with the recommended precaution to prevent any biological activity and gas exchange with the atmosphere (Strickland and Parsons, 1968). A WTW 208262 respirometric BOD measuring system was used for the analysis. The bottles were incubated at 20 °C for 5 d. The accuracy of the BOD method is ± 1 (± 3.55 hPa). The determination of sulfide concentration was performed on board immediately after seawater sampling. Seawater samples were taken from the Niskin sampling bottles, and sulfides were determined by using the photometric method of Cline (1969). Sulfides react with N, N-diethyl-p-phenylenediamine (DPD) sulfate and ferric chloride to produce a blue color measured with a UV/Visible spectrophotometer at 670 nm. The limit of detection (LOD) was $1 \mu g L^{-1}$.

For fluoride measurement, seawater samples were collected from the Niskin bottles in the special tubes equipped for the ECION700/40S Eutech Ion 700 meter with integral electrode holder and 100/240 VAC Adapter CE equipped with Eutech 9609BNWP Fluoride Ionplus Sure-Flow combination electrode with a BNC connector.

For cyanide analysis, sea water samples were collected from the Niskin bottles in 125 mL polyethylene bottles. The samples were preserved by adding NaOH until pH > 12 and stored at 4 °C until their analysis at HCMR labs. The analytical methodology was based on the colorimetric method no. 4500-CN-E in Standard Methods for the Examination of Water and Wastewater (1992). The detection limit is $0.2 \,\mu g \, L^{-1}$.

Chlorine concentration was measured on board immediately (in exactly 2 min) after seawater sampling. Seawater samples were taken from the Niskin bottles, and chlorine was determined by photometry according to the Standard Methods for the Examination of Water and Wastewater no. 4500-Cl G (1992). Chlorine reacts with DPD to produce a red color measured with a UV/Visible spectrometer at 515 nm. The LOD of the method was $10 \,\mu g \, L^{-1}$.

3.5 Total petroleum hydrocarbons (TPHs), oil/grease and chlorophenols in seawater

Total petroleum hydrocarbons (TPH), oil, grease and chlorophenols were analyzed in the HCMR organic chemistry laboratory certified according to EN ISO/IEC 17025:2017. For TPH, oil and grease determination, 2.5 L

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of seawater was collected in glass bottles and, after the addition of deuterated *n*-C24 as an internal standard, was immediately extracted on board with *n*-hexane. For chlorophenols, 1 L of seawater was collected in high-density polyethylene bottles, 2-chlorophenol-3,4,5,6-D4 was added as an internal standard, and then the samples were acidified to pH < 4 and extracted on board with dichloromethane. All extracts were stored in a refrigerator and transferred to the HCMR laboratory.

In the laboratory, TPH was determined by gas chromatography and mass spectrometry with a flame ionization detector (FID) (Agilent 7890A/5975C GC–MS) based on the ISO 9377-2:2000 method (2002), and oil/grease was determined gravimetrically as the hexane extractable material using EPA method 1664 (2010). The hexane was distilled to dryness, and after further drying in an oven at 60 °C, the remaining residue was weighed in a 4-digit analytical balance. The LOD was 0.1 mg L⁻¹.

The analysis of chlorophenols was based on EPA method 528 (2000). In the laboratory, the extracts were dried with sodium sulfate, and their volume was reduced to about 1 mL using a rotary evaporator and finally to 100 µL using a stream of ultraclean nitrogen. Chlorophenols were determined by gas chromatography-mass spectrometry (Agilent 7890A GC - 5975C MS). The following substances were quantified using the internal standard method: 2-chlorophenol, 3-chlorophenol, 4-chlorophenol, 2,3-dichlorophenol, 2,4-dichlorophenol, 2,5-dichlorophenol, 2,6-dichlorophenol, 3,4-dichlorophenol, 3,5-dichlorophenol, 2,3,4-trichlorophenol, 2,3,5-trichlorophenol, 2,3,6-2,4,6trichlorophenol, 2,4,5-trichlorophenol, trichlorophenol, 3,4,5-trichlorophenol, 2,3,4,5-2,3,4,6-tetrachlorophenol, tetrachlorophenol, 2,3,5,6tetrachlorophenol and pentachlorophenol. The LOD for each compound was $0.01 \,\mu g \, L^{-1}$.

3.6 Metals in seawater

For the analysis of dissolved Cd, Co, Cr, Cu, Ni, Pb and Zn, 250 mL of seawater was collected in pre-cleaned polyethylene bottles and stored at -20 °C until analysis in the laboratory. After thawing, the samples were filtered through 0.45 µm filters (Whatman sterile mixed cellulose ester membranes) and acidified with Suprapur HCl to pH < 2. Samples for Cd, Co, Cu, Ni, Pb and Zn determination were pre-concentrated by the Toyopearl AF-Chelate-650M resin to separate these elements from interfering matrix components (Milne et al., 2010; Willie et al., 1998). The trace metals eluted on the resin were collected with 1 M Suprapur HNO₃ and determined using inductively coupled plasma mass spectrometry (ICPMS, Thermo-Elemental X-series II) in a regular laboratory environment. Accuracy and precision were assessed using the class-5 certified reference material for coastal water and acidified seawater samples of the QUASIMEME inter-laboratory exercise (AQ-3, Lab code 122, https://www.wepalquasimeme.nl/en/wepal.htm, last access: 25 March 2024). The results obtained for class 5 were in good agreement with the certified values (Tables S1 and S2 in the Supplement), and the samples of the QUASIMEME tests had acceptable Z scores (-2 < Z < 2).

A co-precipitation method was used for total dissolved Cr (Harper and Riley, 1985). The samples were precipitated with Fe(II) ammonium sulfate. Total dissolved Cr determination was performed by graphite furnace atomic absorption spectrometry (AAS, Shimadzu GFA-7000A) in a regular laboratory environment. The LOD of the method was $0.083 \,\mu g \, L^{-1}$.

For the total Hg, 250 mL of seawater was collected in precleaned glass bottles. After adding 1.2 mL of 37 % HCl, the samples were stored in a refrigeration unit until the analysis in the laboratory. Total Hg was determined by EPA method No 1631 (2002), which consisted of oxidation of all species to Hg(II), purging and trapping onto a gold trap, desorption and cold-vapor atomic fluorescence spectrometry (CVAFS) by a Tekran 2500 mercury analyzer. The LOD was $0.2 \text{ ng } \text{L}^{-1}$, and the limit of quantification (LOQ) was $0.5 \text{ ng } \text{L}^{-1}$. On each day of analysis, the reference material was analyzed at least once daily for recovery estimation (accuracy) (Tables S1 and S2). The reference material was a diluted sample from digested sediment with a certified Hg content of $412 \,\mu g \, kg^{-1}$, an aliquot of which was spiked in purged seawater to reach a concentration of $2.0 \text{ ng } \text{L}^{-1}$. Thus, both accuracy and matrix spike checks were performed.

The samples' pre-treatment, the trace metal preconcentration steps, the trace metal analyses and Hg determination were carried out in US FED-STD Class M5.5 (10000) clean room environments using ultraclean handling techniques (EPA method 1669).

3.7 Sediments

Surface sediment samples were collected for the determination, using EPA methods, of aliphatic hydrocarbons and polycyclic aromatic hydrocarbons (AHCs and PAHs); volatile organic compounds (VOCs) benzene, toluene, ethylbenzene, and xylene (BTEX); polychlorinated biphenyls (PCBs); phenols; organic carbon; and metals. Organic compounds and metals were analyzed in the HCMR organic chemistry laboratories certified according to EN ISO/IEC 17025:2017. An internal quality check was performed by means of analyses of QUASIMEME samples. In order to identify the geochemical background of each area, additional sampling and analysis of sediment cores is needed.

The analysis of AHC and PAH in sediments was based on the recommendations of OSPAR (2013) and UNEP/IOC/I-AEA (1992). The sediments were frozen at -20 °C (UNEP, 1992) until they arrived at the laboratory. In the laboratory, the sediment samples were dried at 40 °C, sieved through a 250 mm sieve and homogenized. After the addition of a mixture of deuterated compounds, used as internal



Figure 5. Average ammonium (left panel) and phosphate (right panel) concentrations in the sampled marine areas along the coastline of the Red Sea (© Google Maps 2021). The mapping presents the average concentrations for each area (average from all stations sampled in the area, average of mean integrated concentrations of the stations included in each area). Ammonium values at Jeddah Lagoon are out of scale.



Figure 6. Average total dissolved nitrogen (left panel) and total organic carbon (right panel) concentrations in the sampled marine areas along the coastline of the Red Sea ($^{\odot}$ Google Maps 2021). The mapping presents the average concentrations for each area (average from all stations sampled in the area, average of mean integrated concentrations of the stations included in each area). Jeddah Lagoon values are out of scale, since they ranged from 11.8 to 167 µmol L⁻¹ for total dissolved nitrogen and from 1 to 5.7 mg L⁻¹ for total organic carbon.



Figure 7. Nitrate+nitrite concentrations in Jeddah Lagoon (© Google Maps 2021).

standards, 0.5-3 g of dried sediment was extracted with a mixture of methanol and dichloromethane (1:2, v/v)using an accelerated solvent extractor (Dionex ASE 350). The extract was saponified with methanolic KOH, and the unsaponified components were extracted with n-hexane and cleaned up and fractionated by passing through a silica column. The final determination of total AHC and PAH was carried out by gas chromatography and mass spectrometry (Agilent 7890A/5975C GC-MS). The quantitation was based on the deuterated internal standards. Blanks were systematically checked to verify the absence of contamination during analyses. Accuracy ranged from 79.8% to 96.5% for individual PAH compounds and was systematically controlled using a reference material (NIST SRM 1941b), and the laboratory also participates in QUASIMEME inter-laboratory exercises. Analytical uncertainties for each PAH ranged from 12.1 % to 37 % (k = 2). The LOD for total AHC and PAH was $0.5 \,\mu g \, g^{-1}$ and $0.1 \, ng \, g^{-1}$, respectively. Individual PAHs determined were naphthalene; acenaphthylene; acenaphthene; fluorene; dibenzothiophene; phenanthrene; anthracene; fluoranthene; pyrene; benzo(*a*)anthracene: chrysene; benzo(*b*)fluoranthene; benzo(k)fluoranthene; benzo(*e*)pyrene; benzo(*a*)pyrene; indeno(1,2,3-*cd*)pyrene; benzo(ghi)perylene; perylene; dibenzo(*ah*)anthracene; and the methylated derivatives of naphthalene, dibenzothiophene, phenanthrene, pyrene, and chrysene.

For the analysis of PCB in sediments, 3 g of dried sediment was extracted, following the addition of a mixture of CB112, CB155 and CB209 (used as internal standards), with a mixture of hexane and dichloromethane (1 : 1, v/v) using an ac-

celerated solvent extractor (Dionex ASE 350). The extract was cleaned on an alumina glass column. The final determination of PCB was carried out by capillary gas chromatography using a non-polar column and an electron capture detector (Agilent 7890A GC). The quantitation was based on the abovementioned internal standards. Accuracy ranged from 69.3 % to 83.5 % for each congener and was controlled using a reference material (NIST SRM 1941b). Analytical uncertainties for each congener ranged from 27.1 % to 39 % (k = 2). The following individual congeners were determined: CB28, CB52, CB101, CB118, CB105, CB138, CB153, CB128, CB156, CB170, CB180, CB183 and CB194. The LOD for each congener was 0.01 ng g⁻¹.

The analysis of BTEX was performed using the equilibrium-based static headspace technique and gas chromatography–mass spectrometry according to EPA methods 5021A (2014) and 8260 (2018). Briefly, ~ 3 g of wet sediment was collected in 20 mL headspace vials and kept at -20 °C until arrival at the HCMR laboratory. In the laboratory, VOC was determined using a headspace autosampler (HTA, HT2800T) and a gas chromatography–mass spectrometry system (Shimadzu GCMS-QP2020 NX). The following substances were quantified using an external standard mixture: benzene; toluene; ethylbenzene; and *o*-, *m*-, and *p*-xylenes. The LOD was 0.05 μ g kg⁻¹.

The analysis of phenols in sediments was based on the EPA (spectrophotometric) method 420.1 (1978). The sediments were frozen at -20 °C until their arrival in HCMR labs. In the lab, 15 g of sediment was put in 500 mL of distilled water, and the pH was adjusted to < 4 using phosphoric acid. The sample was distilled, and total phenols were determined in the distillate by the 4-aminoantipyrine colorimetric method. The dye was extracted with chloroform, and absorbance was measured at 460 nm using a UV-2600 Shimadzu UV/Visible spectrometer and 100 mm cell. The detection limit was 0.1 µg g⁻¹.

Major elements and trace metals in marine sediments were measured by an accredited method (EN ISO/IEC 17025:2017) using a wavelength-dispersive X-ray fluorescence (WDXRF) system. The LOD for this XRF method was calculated from a series of measurements of 5 samples for major elements and 10 samples for trace elements using the certified reference sample PACS-2. The LOD measurements were carried out in the same experimental conditions in which the sediment samples were analyzed (Table S3).

The calibration of the XRF method was carried out by scanning reference samples that contained a wide spectrum of element concentrations. For this calibration analysis, several rock and sediment samples were gathered, mainly from the U.S. Geological Survey and the National Research Council of Canada Reference Materials. Fused beads and powder pellets were prepared carefully, and all samples were scanned for major and trace elements. All measurement parameters were configured through the software, building two distinct "applications" and selecting the optimum settings for each el-



Figure 8. Average Chl-*a* concentrations in the sampled marine areas along the coastline of the Red Sea (left panel) and the indicative assessment of ecological status based on Chl-*a* concentrations according to the five-scale classification scheme for the Eastern Mediterranean (right panel). The mapping presents the average concentrations for each area (average from all stations sampled in the area, average of mean integrated concentrations of the stations included in each area) Chl-*a* values at Jeddah Lagoon (left panel) ranged from 0.61 to $44 \,\mu g \, L^{-1}$ (© Google Maps 2021).

ement separately. Subsequently, element concentrations were plotted against the measured intensities, and a linear fit was generated by regression. Theoretical alpha corrections and possible line overlaps were carefully resolved until the lowest mean error of the fit was obtained. Accredited trace elements are As, Co, Cr, Cu, Mn, Mo, Ni, Pb, Sb, Sn, Sr, V, Zn, Ag, Ba, Bi, Br, Cd, Ce, Cs, Ga, Ge, Hf, Hr, I, La, Nb, Nd, Rb, Sb, Sc, Se, Sm, Sn, Sr, Ta, Te, Th, Tl, U, W, Y, Yb and Zr. Accredited major elements are Fe₂O₃, CaO, TiO₂, Al₂O₃, K₂O, MgO, Na₂O, P₂O₅, SO₃ and Si₂O.

4 Results and discussion

Here, we present the essential physical and biochemical parameters related to the hydrography and the quality of seawater and sediments of the Arabian Gulf and Red Sea study areas (SEANOE, https://doi.org/10.17882/96463, Abualnaja et al., 2023). A principal goal is to trace contaminants from pollution sources along the coastal zone of Saudi Arabia. In addition, the data establish the baseline for the future design of monitoring strategies of the Saudi Arabian coastal marine environment.

4.1 Red Sea

4.1.1 Physical parameters

The water properties measured during the Red Sea cruise would be expected to conform with the general picture of the regional physical oceanography prevailing during the sampling period.

The temperature–salinity (T-S) diagram presenting all the CTD casts of the Red Sea surveys displays the general physical characteristics of the water masses (Fig. 2). Expectedly, general gradients in both temperature and salinity are evident, with temperature increasing and salinity decreasing from north to south, a trend consistent with the previous studies of Neumann and McGill (1962), Maillard and Soliman (1986), Sofianos and Johns (2007), and Ali et al. (2018). In particular, the densest water, with the highest salinity and lowest temperatures, is found within the Gulf of Aqaba (Fig. 3). In general, the potential temperature (hereafter, temperature) values ranged between 21 and 36 °C, which are typical for the mid-summer period (Sofianos and Johns, 2007). The north–south gradient of surface salinity, typical for the Red Sea due to the general northward propagation of rel-



Figure 9. Average total mercury (Hg; ng L⁻¹) (left panel) and petroleum hydrocarbon (TPH) concentrations (μ g L⁻¹) (right panel) at the stations sampled in the Red Sea (© Google Maps 2021). The mapping presents the average concentrations for each area (average from all stations sampled in the area, average of mean integrated concentrations of the stations included in each area).

atively fresher water from the Gulf of Aden (Sofianos and Johns, 2007; Churchill et al., 2014a), is evident, with high surface values (40.7) within the Gulf of Aqaba, decreasing to 38 towards the southern region near the Jizan area. The CTD profiles acquired at shallow stations (< 100 m) show almost constant temperature and salinity with depth, whereas, at the deeper areas, the temperature tends to decrease while salinity increases with depth. Additionally, due to the shallow bathymetry of the coastal areas, they are more susceptible to local atmospheric forcing, with increased temperature compared to offshore during the cruise period.

The T - S properties at the hotspot sites reveal potential signs of pressures on the coastal ecosystem. Low salinity, detected within the Jeddah Lagoon system, indicates the input of fresher water, possibly consisting of sewage effluent as observed by Peña-García et al. (2014). A slight increase in temperature and salinity within the Jeddah Islamic Port, compared to the stations located outside the port, could be attributed to water stagnation in the port. Differences in temperature and salinity values at Al Shuqaiq and Al Lith can be attributed to water discharges linked to desalination plant and aquaculture activities, respectively. In general, the brine discharge at discharging plume in the Red Sea is getting diluted within 10–20 m (Raventós et al., 2006; Fallatah et al., 2018). However, the accurate detection of the plume dispersion area would require a dedicated, denser grid of stations, which was

beyond the implementation of this survey. In addition, an approximate estimation of the plume dispersion would require precise knowledge of the brine water supply from the desalination plants.

Regarding the current measurements, the surface currents range from weak, on the order of $\sim 2-3 \,\mathrm{cm}\,\mathrm{s}^{-1}$, to substantial, reaching values of $17-25 \,\mathrm{cm}\,\mathrm{s}^{-1}$. Dominant northward flow only appears at stations located in the north Red Sea, whereas multi-directional flow seems to be the rule at most stations (Fig. 4). Pavlidou et al. (2021) have found that subsurface currents in the depth range of 18–60 m are mostly in the same direction as the near-surface ones and parallel to the nearby coast. Moreover, it should be mentioned that, in practical terms, our velocity data were collected in June 2021 during lunar days, which were away from the spring tides near the new moon and the full moon days, which were on 10 and 24 June, respectively, and the tidal influence is expected to be very low.

4.1.2 Biogeochemical parameters

The results obtained during the cruise in June 2021 in the Red Sea coastal area showed that the Gulf of Aqaba appears to be better oxygenated through the whole water column, from the surface to the bottom layer, while a decrease in oxygen near the bottom $(115 \,\mu\text{mol L}^{-1})$, which is possibly connected to the organic material accumulated near the



Figure 10. Total aliphatic hydrocarbon concentrations (TAHCs; mg kg⁻¹), total EPA PAH concentrations (\sum PAH_{EPA}), unresolved/resolved (U/R) diagnostic ratio and percentage contribution of pyrolytic PAH in sediments for each site in the Red Sea (\bigcirc Google Maps 2021). The mapping presents the average concentrations for each area (average from all stations sampled in the area, average of mean integrated concentrations of the stations included in each area). For the U/R diagnostic ratio, the color scale indicates average values > 4 (red color) or < 4 (blue color) for all study sites.

bottom, is recorded at the majority of the sampling stations. As recently reported by Povinec et al. (2023), in the Gulf of Aqaba, the first 300 m of the water column indicates stable DO values of about 210 μ mol L⁻¹, whereas DO of about 3.5 μ mol kg⁻¹ was measured below the depth of 500 m. Conversely, in the open Red Sea, DO minimum values from

0.6 to $1.25 \,\mu\text{mol}\,\text{kg}^{-1}$ were observed at depths from 300 to 450 m, showing a decreasing trend below 150–200 m.

The assessment presented herein aims to identify the areas facing eutrophication and/or pollution problems. In general, in the Red Sea, north—south increasing gradients were evident for some of the parameters studied, revealing a link



Figure 11. T-S diagram for all measured stations in the Arabian Gulf. The three colors of dots represent the three hotspot sites. Measurements are averaged every 1 m in the water column up to 0.5 m from the sea bottom.



Figure 12. Vertical profiles of Potential Temperature and Salinity in the study areas of the Arabian Gulf.

between the hydrographic conditions and biogeochemical properties. Nutrients and organic carbon revealed high values at Jeddah Lagoon (Figs. 5–7), which is a unique system with low water renewal and pollution mainly from domestic sewage (Peña-García et al., 2014). The nitrite values at Jeddah Lagoon were high $(3.06-3.90 \,\mu\text{mol}\,\text{L}^{-1})$, and the highest ammonium values $(2.90 \,\mu\text{mol}\,\text{L}^{-1})$ were found at the station (L03), whereas total dissolved nitrogen was extremely high at the same station, reaching $167 \,\mu\text{mol}\,\text{L}^{-1}$. At all other sites of the Red Sea, nitrate+nitrite was low in the euphotic zone and increased with depth. Nitrate+nitrite values in the euphotic layer (approximately 0–100 m) ranged between below the LOQ and 2.30 $\mu\text{mol}\,\text{L}^{-1}$, with the exception

of two relatively high values detected at Al Khumrah, close to the sewage outfall (3.43 μ mol L⁻¹), and station AL1 of the Al Lith grid (5.78 μ mol L⁻¹). These values exceed the commonly observed values in coastal areas and could indicate organic load from anthropogenic activity. Figures 8 and 9 illustrate the distribution of Chl-a and selected pollutants (metals and organic) along the north-south section. Higher Chl-a values recorded in the southern Red Sea correlate with relatively higher nutrient concentrations, which are influenced by the Gulf of Aden Intermediate Water (GAIW), the inflow of nutrient-rich water entering the Red Sea from the Gulf of Aden through the Bab-el-Mandeb Strait (Churchill et al., 2014b). In this survey, the Chl-a concentrations did not exceed $1 \,\mu g \, L^{-1}$ at any site except from Jeddah Lagoon, where extremely high Chl-a values reaching $40 \,\mu g \, L^{-1}$ were sometimes measured within the Jeddah Lagoon system. The Kingdom of Saudi Arabia has set maximum allowable values only for ammonium $(5.5 \,\mu\text{mol}\,\text{L}^{-1})$, whereas it has not set maximum allowable values for Chl-a. The Abu Dhabi Quality and Conformity Council (2017) set 1 mg L^{-1} as the maximum allowable concentration of Chl-a for ambient marine waters, whereas in the Eastern Mediterranean the target values for the "Good" environmental status is $0.53 \,\mu g \, L^{-1}$ (EC Decision 2018/229/EU). Regarding Chl-a distribution, it seems that the ecological status of the water at the southernmost site of the Red Sea, Jizan Economic City, is classified as poor, which corroborates the algal blooms observed in this area during summer and early Autumn 2021 (personal communications with the emirate of Jazan Province). This area may be affected by water intrusion from the south and atmospheric deposition. The deposition of material transferred by dust storm events may also influence the ecological quality of the water in the area, since during the summer period, the Tokar Gap frequently channels strong winds onto the sea surface, causing African dust storms spreading over the southern part of the Red Sea (Jiang et al., 2009; Garrison al., 2010).

Sediments at all of the Red Sea study sites are found to be enriched in some metals (e.g., As). This finding should be further investigated via sampling and analysis of sediment cores in order to define the geochemical background of the region. Moreover, AHC, PAH and PCB, which constitute important classes of organic contaminants that may cause degradation and pose a risk of serious damage in the marine environment, are determined in surface sediments collected from the coastal zone of the Red Sea (Fig. 10). The examination of various indices reveals a chronic petroleumassociated anthropogenic pressure in Jeddah Lagoon system, Jeddah Islamic Port (Mena Jeddah) and Al Khumrah, whereas some petroleum residues were also found at King Fahad Industrial Port Yanbu, Mena Jeddah, shrimp and fish farms near Al Lith, and to a lesser extent at Magna.

Organic pollution in the Jeddah Lagoon system and at King Fahad Industrial Port Yanbu is also confirmed by BOD values (4.5 mg L^{-1} inside the port to 7.7 mg L^{-1} in the lagoon), and fluoride values at the northern part of the Gulf



Figure 13. Average lead (Pb; mg L^{-1}) (left panel) and total mercury (Hg; ng L^{-1}) (right panel) in the Arabian Gulf (© Google Maps 2021). The mapping presents the average concentrations for each area (average from all stations sampled in the area, average of mean integrated concentrations of the stations included in each area).



Figure 14. Average metals in sediments in the Arabian Gulf (© Google Maps 2021). The mapping presents the average concentrations for each area (average from all stations sampled in the area).

of Aqaba confirmed the effect of the phosphate terminal in the Port of Aqaba in Jordan due to cross-border pollution. However, it seems that industrial activities probably enrich the coastal zone of the Red Sea with organic pollutants.

4.2 Arabian Gulf

4.2.1 Physical parameters

Salinity is the main indicator for local differentiations in the Arabian Gulf. A broad range of salinity values is observed, with higher salinity close to the coastline at Ras Al Khair and, especially, at Al Khobar, reflecting the effects of the local desalination plants and their brine discharges. The CTD casts conducted during the survey are plotted in a T-S diagram (Fig. 11) with the different sampling sites distinguished by different colors. In general, temperature values ranged from 31.0 to 33.5 °C. The salinity measurements span a wide range of hypersaline values, from 40 to 52. The lowest salinity was measured at DM4, a station that can be considered a reference for the Dammam and Al Khobar sea areas and as the most representative one for the regional open sea salinity values, which according to earlier studies are around 40 over the western part of the Gulf (John et al., 1990; Chao et al., 1992). Along with the hydrological characteristics of the gulf and the regional morphology, the presence of this desalination plant seems to be the major factor that leads to extremely high salinity values at the study site of Al Khobar (Fig. 12). In the Arabian Gulf, salinity increases > 2 are rarely seen beyond a 400 m radius of the desalination plants's outfalls (Roberts et al., 2010), although in some cases they can extend to several kilometers, as in the case of Az Zour power and desalination plant in southern Kuwait (Uddin et al., 2011).

4.2.2 Biogeochemical parameters

In terms of the eutrophication and pollution status of the three areas in the Arabian Gulf, the Dammam is supposed to be affected by wastewater discharges (Mahboob et al., 2022). The concentrations of inorganic nutrients and organic phosphorus are low at all three sites, indicating that eutrophication is not affecting these specific sites during the sampling period. However, dissolved organic nitrogen and carbon values are relatively high, reaching a maximum value of $18.8 \,\mu\text{mol}\,\text{L}^{-1}$ for nitrogen and 477 μ mol L⁻¹ for organic carbon. Relatively higher Hg and Pb concentrations $(0.55-5.90 \,\mu g \, L^{-1}$ for Hg and 0.18–1.25 μ g L⁻¹ for Pb) were found in the water column (Fig. 13), which is probably linked to the industrial activities in these areas and/or atmospheric deposition (for Pb). However, it should be noted that this analysis only provides a snapshot of the status in the water column at one point in time and that eutrophication-related parameters exhibit strong seasonal variation.

The measurements showed that similar concentration ranges of metals were detected in the Arabian Gulf and Red Sea coastal waters in June 2021. It is noteworthy that TPH was low in the Arabian Gulf. However, sediments at Ras Al Khair were found to be polluted with metals and, in some cases, exceeded the allowable values set in the Abu Dhabi specifications: 7.0 mg L^{-1} for As, 52 mg L^{-1} for Cr, 30 mg L^{-1} for Pb and 125 mg L^{-1} for Zn (Abu Dhabi Quality and Conformity Council, 2017) (Fig. 14). By contrast, TPH was surprisingly low in the Arabian Gulf. However, sediments at Ras Al Khair, and especially at Al Khobar, were found to be polluted with metals. Regarding the degree of contamination, it seems that sediments in the Arabian Gulf are severe-to-heavily (Al Khobar) metal polluted. Thus, an additional, more detailed study of the Arabian Gulf, with finer coverage of the coastal zone from Khafji to Al Khobar, is highly recommended.

5 Data availability

Data described in this manuscript can be assessed at SEANOE: https://doi.org/10.17882/96463 (Abualnaja et al., 2023).

6 Conclusions

This is the first broad coverage study in a one-off sampling campaign in the Saudi Arabian coastal zone. To the best of our knowledge, these cruises constitute the first multidisciplinary and geographically comprehensive survey of contaminants within the Saudi Arabian coastal waters and sediments, extending from near the Saudi-Jordanian border in the north of the Red Sea to Al Shuqaiq and Jizan Economic City (close to the Saudi-Yemen border) in the south, as well as in the Arabian Gulf, including the areas of Al Khobar, Dammam, and Ras Al Khair. The assessment presented in this work, aimed to identify the areas facing eutrophication and/or pollution problems. In general, in the Red Sea, northsouth increasing gradients were evident for some of the parameters studied, revealing a link between the hydrographic conditions and biogeochemical properties. Sediments at all of the Red Sea study sites were found to be enriched in arsenic. In the Arabian Gulf, salinity was defined as the main indicator of local differentiation. A broad range of salinity values were observed, reflecting the effects of the local desalination plants and their brine discharges.

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