



# GLC\_FCS30D: the first global 30 m land-cover dynamics monitoring product with a fine classification system for the period from 1985 to 2022 generated using dense-time-series Landsat imagery and the continuous change-detection method

Xiao Zhang<sup>1,2</sup>, Tingting Zhao<sup>1,4</sup>, Hong Xu<sup>5</sup>, Wendi Liu<sup>1,2,3</sup>, Jinqing Wang<sup>1,2,3</sup>, Xidong Chen<sup>6</sup>, and Liangyun Liu<sup>1,2,3</sup>

<sup>1</sup>International Research Center of Big Data for Sustainable Development Goals, Beijing 100094, China

<sup>2</sup>Key Laboratory of Digital Earth Science, Aerospace Information Research Institute, Chinese Academy of Sciences, Beijing 100094, China

<sup>3</sup>School of Electronic, Electrical and Communication Engineering, University of Chinese Academy of Sciences, Beijing 100049, China

<sup>4</sup>College of Geomatics, Xi'an University of Science and Technology, Xi'an 710054, China

<sup>5</sup>The High-Tech Research & Development Center (HTRDC) of the National Natural Science Foundation of China, Beijing 100044, China

<sup>6</sup>Future Urbanity & Sustainable Environment (FUSE) Lab, The University of Hong Kong, Hong Kong SAR, 999007, China

**Correspondence:** Liangyun Liu (liuly@radi.ac.cn)

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**Abstract.** Land-cover change has been identified as an important cause or driving force of global climate change and is a significant research topic. Over the past few decades, global land-cover mapping has progressed; however, long-time-series global land-cover-change monitoring data are still sparse, especially those at 30 m resolution. In this study, we describe GLC\_FCS30D, a novel global 30 m land-cover dynamics monitoring dataset containing 35 land-cover subcategories and covering the period 1985–2022 in 26 time steps (maps were updated every 5 years before 2000 and annually after 2000). GLC\_FCS30D has been developed using continuous change detection and all available Landsat imagery based on the Google Earth Engine platform. Specifically, we first take advantage of the continuous change-detection model and the full time series of Landsat observations to capture the time points of changed pixels and identify the temporally stable areas. Then, we apply a spatiotemporal refinement method to derive the globally distributed and high-confidence training samples from these temporally stable areas. Next, local adaptive classification models are used to update the land-cover information for the changed pixels, and a temporal-consistency optimization algorithm is adopted to improve their temporal stability and suppress some false changes. Further, the GLC\_FCS30D product is validated using 84 526 globally distributed validation samples from 2020. It achieves an overall accuracy of 80.88 % ( $\pm 0.27$  %) for the basic classification system (10 major land-cover types) and 73.04 % ( $\pm 0.30$  %) for the LCCS (Land Cover Classification System) level-1 validation system (17 LCCS land-cover types). Meanwhile, two third-party time-series datasets used for validation from the United States and Europe Union are also collected for analyzing accuracy variations, and the results show that GLC\_FCS30D offers significant stability in terms of variation across the accuracy time series and achieves mean accuracies of 79.50 % ( $\pm 0.50$  %) and 81.91 % ( $\pm 0.09$  %) over the two regions. Lastly, we draw conclusions about the global land-cover-change information from the GLC\_FCS30D dataset; namely,

that forest and cropland variations have dominated global land-cover change over past 37 years, the net loss of forests reached about 2.5 million km<sup>2</sup>, and the net gain in cropland area is approximately 1.3 million km<sup>2</sup>. Therefore, the novel dataset GLC\_FCS30D is an accurate land-cover-dynamics time-series monitoring product that benefits from its diverse classification system, high spatial resolution, and long time span (1985–2022); thus, it will effectively support global climate change research and promote sustainable development analysis. The GLC\_FCS30D dataset is available via <https://doi.org/10.5281/zenodo.8239305> (Liu et al., 2023).

## 1 Introduction

Land-cover data are important and necessary for supporting sustainable development goals, maintaining biodiversity, and monitoring natural resources (L. Liu et al., 2021; Potapov et al., 2022). Land-cover changes directly or indirectly influence global climate patterns and the speed and magnitude of climate change (Song et al., 2018), and they increasingly affect biogeochemical cycles, the carbon cycle, and the Earth's energy balance (Foley et al., 2005; Hong et al., 2021; Winkler et al., 2021). Since the Industrial Revolution, under the dual pressure from global climate change and human activities, global land cover has undergone drastic changes. According to a Global Carbon Project report in 2020, since the industrialization period, land-cover and land-use changes have contributed to approximately 25 % of all global greenhouse gas emissions (Friedlingstein et al., 2020), and this trend is exacerbated by the ongoing increase in population and per capita energy consumption (Xian et al., 2022). Therefore, understanding and studying land-cover changes is of vital significance for addressing global environmental changes, promoting sustainable development, and safeguarding the Earth's ecological environment.

Remote-sensing techniques, with their periodic Earth observation capability and archived massive long-term observation data since 1972, provide the most cost-effective and practical solutions for long-time-series, large-area land-cover-change monitoring. In the past few decades, with the continuous improvement of remote sensing technology and storage and computing capabilities, global land-cover-change monitoring (GLCCM) has transitioned from 1 km spatial resolution to fine resolutions of 30 or 10 m and from single-phase mapping to long-term monitoring (Ban et al., 2015; Friedl et al., 2010, 2022; Giri et al., 2013). In its early stages, GLCCM mainly relied on time series of MODIS, AVHRR, and Project for Onboard Autonomy (PROBA)-V imagery (Buchhorn et al., 2020; Friedl et al., 2010); for example, Sulla-Menashe et al. (2019) generated a global 500 m annual land-cover product (MCD12Q1) covering the period from 2001 to the present using MODIS time-series imagery with an overall accuracy of 73.6 %. Defourny et al. (2018) integrated PROBA-V and Medium Resolution Imaging Spectrometer (MERIS) time-series observations to develop a global 300 m annual land-cover dynamics dataset (CCI\_LC: Climate Change Initiative Global Land Cover) for

the period from 1992 to 2020 with an overall accuracy of 71.5 %. These coarse land-cover-change products comprehensively captured the spatial patterns of various land-cover types and quantified the global land-cover changes caused by human and natural activities. However, they still had major limitations, especially in regions with intense human activity and high spatial heterogeneity, because these broken and heterogeneous land-cover changes cannot be captured by coarse-resolution satellite observations (Hansen et al., 2013; Herold et al., 2008; L. Liu et al., 2021; Y. Zhang et al., 2021).

Recently, benefiting from the free access to fine-resolution satellite imagery and powerful computing and storage capabilities and the rise of cloud computing (such as the Google Earth Engine (Gorelick et al., 2017) and Microsoft Planetary Computer) in particular, fine-resolution land-cover dynamics monitoring has been experiencing rapid development. Correspondingly, numerous national and global 30 m land-cover dynamics products have been developed (Chen et al., 2015; Homer et al., 2020; H. Liu et al., 2021; Potapov et al., 2022; Yang and Huang, 2021; Zhang et al., 2022). For example, Yang and Huang (2021) used China's land-use/cover datasets (CLUDs) as the prior dataset and then combined multitemporal classification and spatiotemporal consistency post-processing methods to develop an annual 30 m land-cover dataset (CLCD) for China from 1990 to 2019. Similarly, H. Liu et al. (2021) combined pixel-based classification and spatiotemporal consistency post-processing methods to generate the first global 30 m land-cover change products. However, many studies have demonstrated that multiperiod independent classifications lead to significant classification error accumulation in land-cover-change time-series monitoring (Sulla-Menashe et al., 2019; Zhu, 2017). For example, Xian et al. (2022) stated that the independent classification strategy suffered from the constraints of the post-processing requirements that ensure the temporal consistency of land-cover change maps. Therefore, although GLCCM has progressed significantly over the past few decades, an accurate global 30 m land-cover change-detection product generated by an efficient land-cover change method is still urgently required.

One of the greatest challenges in large-area land-cover change detection is to select the optimal algorithm to capture the land-cover changes from time-series observations (Healey et al., 2018; Zhu, 2017). Over the past few decades, a series of change-detection algorithms have been proposed

for monitoring forest disturbance (Huang et al., 2009; Jin et al., 2023; Kennedy et al., 2007, 2010; Qin et al., 2021), urban expansion (Liu et al., 2019; X. Zhang et al., 2021a), cropland dynamics (Dong et al., 2015; Potapov et al., 2021), and land-cover changes (Bullock et al., 2019; Jin et al., 2017; Verbesselt et al., 2010; Zhu et al., 2019). However, most of them are only suitable for regional land-cover change monitoring, and some of the algorithms need prior knowledge (such as that for urban expansion). Zhu (2017) systematically reviewed the performance and limitations of various change-detection methods based on multitemporal satellite data and further explained that the high-temporal-frequency and multivariate change-detection algorithms are more suitable for a long time series of land-cover changes in a large area, provided that the problem of the huge amount of computation involved can be solved. Similarly, Xian et al. (2022) and Liu et al. (2019) concluded that dense and continuous change-detection methods provided higher accuracy and more robustness than traditional change-detection methods for capturing multiple, complicated changes.

The continuous change-detection and classification (CCDC) algorithm, a classical change-detection method based on dense time-series observations proposed by Zhu and Woodcock (2014b), is widely used for regional and national land-cover monitoring (Xian et al., 2022; Xie et al., 2022). It uses all available Landsat observations to build time-series regression models and then captures the outliers by analyzing the differences between the actual observations and model estimations. Zhu and Woodcock (2014b) demonstrated that the CCDC algorithm attained a general accuracy of 90 % and temporal accuracy of 80 % for capturing land-cover changes. Thus, it has been adopted by the United States Geological Survey (USGS) as the official algorithm for monitoring land-cover changes over the United States (Pengra et al., 2016). For example, Xian et al. (2022) implemented the CCDC algorithm and all available Landsat data to develop annual land-cover-change products that cover the contiguous United States (CONUS) for 1985–2017 with an overall accuracy of 82.5 %.

In summary, in recent decades, land-cover mapping and monitoring has made significant progress; however, global 30 m land-cover-dynamics time-series products derived from change-detection algorithms are still lacking. In this study, we had three aims: (1) to use the continuous change-detection algorithm and the full time series of Landsat observations to generate the first global 30 m land-cover dynamics product that covers the period from 1985 to 2022 and uses a fine classification system that contains 35 fine land-cover subcategories with 26 time steps (the maps are updated every 5 years before 2000 and annually after 2000), known as GLC\_FCS30D (it should be noted that GLC\_FCS30D is updated every 5 years before 2000 due to the sparse availability of Landsat 5 imagery; thus, we combine the satellite observations from 2 years before and after the nominal center year from 1985 to 1995 to ensure the mapping ac-

curacy of GLC\_FCS30D before 2000); (2) to quantify the land-cover changes and analyze the spatiotemporal change patterns of various land-cover types based on the developed GLC\_FCS30D dataset; and (3) to quantitatively analyze the performance of the GLC\_FCS30D product using multisourced validation datasets.

## 2 Datasets

### 2.1 Continuous Landsat imagery from 1984 to 2022

All available Landsat imagery from 1984 to 2022 – covering the Landsat 5 Thematic Mapper (TM), Landsat 7 Enhanced Thematic Mapper Plus (ETM+), Landsat 8 Operational Land Imager (OLI), and Landsat 9 OLI missions – were collected via the Google Earth Engine (GEE) cloud-computing platform. Specific measures were taken to build a high-quality, continuous Landsat time-series collection. First, all Landsat images underwent atmospheric correction to convert them to surface reflectance using the Landsat Ecosystem Disturbance Adaptive Processing System (LEDAPS) and Land Surface Reflectance Code (LaSRC) methods (Vermote, 2007; Vermote and Kotchenova, 2008). Then, although the Landsat 5, 7, 8, and 9 missions share similar spectral bands, the wavelength differences between the TM, ETM+, and OLI cannot be ignored. Relative radiometric normalization was applied to the TM and ETM+ imagery using the transformation coefficients suggested by Roy et al. (2016).

### 2.2 Global land-cover dataset at 30 m for the year 2020

The global 30 m land-cover product with a fine classification system for the year 2020 (GLC\_FCS30-2020) is the baseline for generating training samples and identifying land-cover information in the temporally stable regions in Sect. 3. The GLC\_FCS30 dataset was developed using local adaptive classifications and confident and globally distributed training samples, and then validated to reach an overall accuracy of 82.5 % with the basic validation system (X. Zhang et al., 2021b). Cross-comparisons with other land-cover products showed obvious advantages for GLC\_FCS30 in terms of mapping accuracy and diversity of land-cover types. The GLC\_FCS30-2020 dataset is freely available at <https://doi.org/10.5281/zenodo.4280923> (Liu et al., 2020).

### 2.3 Global impervious surface dynamics dataset at 30 m from 1985 to 2022

Many studies found that high spatiotemporal heterogeneity of impervious surfaces and broken impervious surface constructions caused high uncertainty and difficulty when monitoring the dynamics of such surfaces (Gong et al., 2019a; Zhang et al., 2022), and issues with both missed detections and false alarms are encountered when change detection methods are applied to the dynamic monitoring of

heterogeneous impervious surfaces. Thus, we independently produced a global impervious-surface-dynamics time-series dataset at 30 m (GISD30) for 1985–2022 and then overlaid this thematic dataset on the GLC\_FCS30D dataset to ensure high confidence in the impervious surface dynamics. The GISD30 dataset was developed by combining the sample migration, spectral generalization, and local adaptive modeling methods and then optimized by the spatiotemporal-consistency correction method (Zhang et al., 2022). It was validated and found to attain a mean overall accuracy of 90.1 % around the globe and to perform better than other similar products in capturing the changes in impervious surfaces over time and across different types of landscapes. At the same time, third-party validation also indicated that GISD30 exhibited superior performance to similar global 30 m impervious surface products (Wang et al., 2022).

## 2.4 Global 30 m wetland datasets from 1985 to 2022

Like the impervious surface dataset, the global wetland dynamics dataset is independently produced because the reflectance spectra of the wetlands and phenological variations changed daily with the water levels. The continuous change-detection method would suffer from serious commission and omission errors if used for wetland dynamics monitoring (Gallant, 2015). In this study, the GWL\_FCS30 (global 30 m wetland map with a fine classification system) wetland dataset from 1985 to 2022 – developed by integrating the automatic sample extraction method, a stratified classification strategy, and the time series of Landsat observations (Zhang et al., 2023) – is superimposed on the GLC\_FCS30D land-cover dynamics dataset. GWL\_FCS30 was quantitatively assessed as having a mean overall accuracy of 86.44 % using 25 708 validation points, and it demonstrated a higher level of performance than other wetland products when it came to capturing the spatial patterns of wetlands during cross-comparisons (Zhang et al., 2023). GWL\_FCS30 further splits wetlands into seven wetland subcategories (four inland and three coastal subcategories) and is overlaid directly onto the GLC\_FCS30D dataset to not only improve the monitoring accuracy for wetlands but also enrich the number of land-cover types (see Table 1).

## 2.5 Validation datasets

To comprehensively analyze the accuracy metrics for the GLC\_FCS30D dataset, two types of validation datasets were collected: an independent global validation dataset from 2020 and two third-party time-series datasets used for validation for the United States and the European Union.

### 2.5.1 Global validation dataset

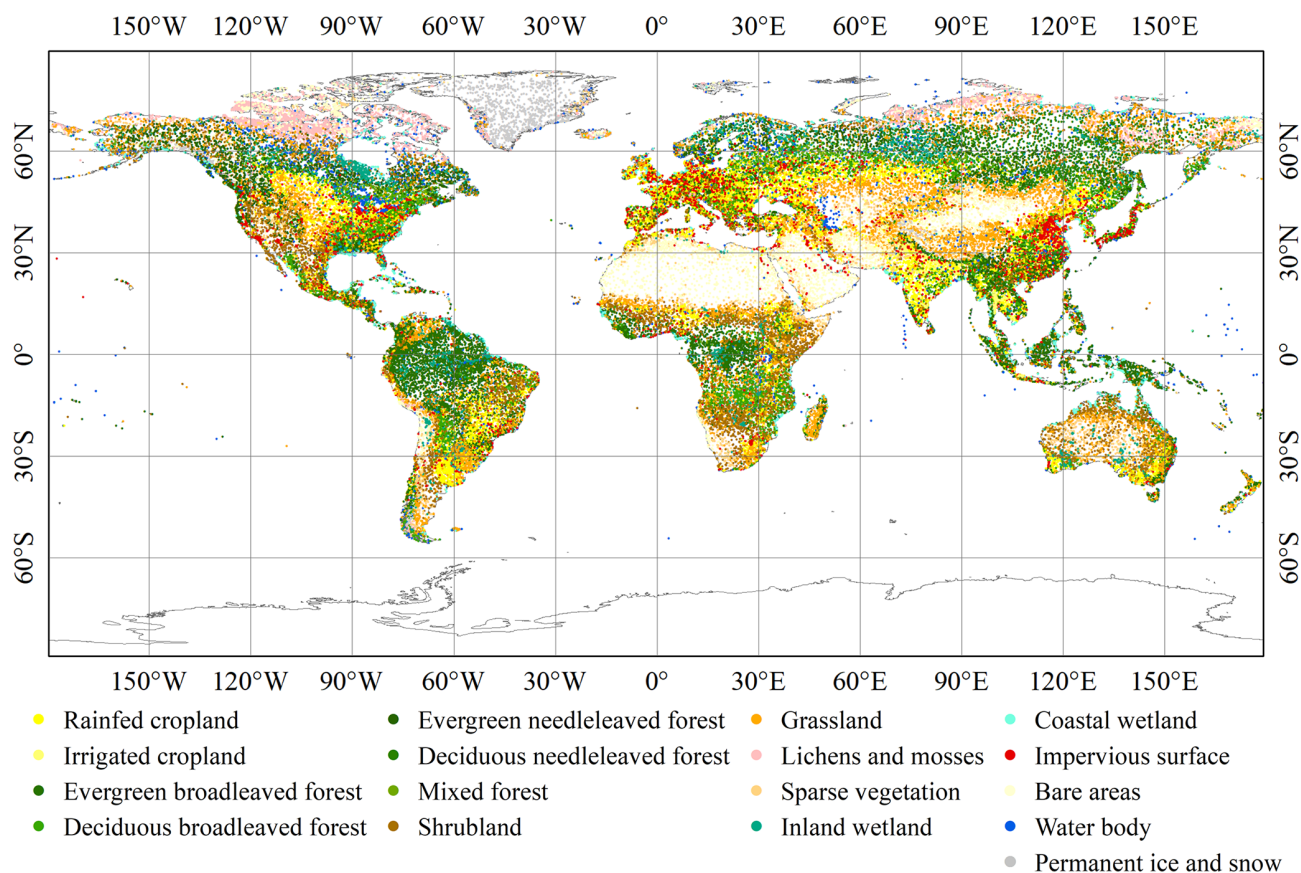
A total of 84 526 globally distributed validation samples are collected to analyze the accuracy metrics for the

GLC\_FCS30D dataset for 2020, and their spatial distribution is illustrated in Fig. 1. Intuitively, the spatial patterns of the global validation dataset are consistent with the actual global land-cover situation. Specifically, to ensure confidence in and the rationality of the validation datasets, several measures are taken, which were explained in detail in our previous work (Zhao et al., 2023). First, a stratified random sampling method is applied by combining the landscape heterogeneity, population density data, and Köppen climate groups, which effectively increases the sample size in the heterogeneous landscapes and for some rare land-cover types (such as impervious surfaces, permanent ice, and snow). Second, for each validation sample, the land-cover type is determined through independent interpretation by trained interpreters after combining high-resolution aerial photography, multitemporal Landsat images, and other relevant ancillary datasets (such as vegetation coverage, tree height, phenological curves, and terrain characteristics). Independent interpretation software based on the GEE platform (<https://eliza-ting.users.earthengine.app/view/crd-vit>, last access: 12 March 2024) for efficiently recognizing the land-cover types of each sample was also developed. Third, a quality-control operation based on duplicate interpretations is used to ensure the confidence level of each validation sample. Each sample is independently labeled by three junior interpreters and then double-checked by the senior experts, and validation samples with huge disparities are discarded. In addition, because the impervious-surface and wetland datasets were produced independently and were validated in our previous works (Zhang et al., 2023, 2022), the corresponding high-quality validation samples of these two thematic types are also imported into the global validation datasets.

### 2.5.2 Third-party regional time-series datasets used for validation

Due to the great difficulty in collecting global long time-series datasets used for validation, we used two third-party regional datasets for the CONUS and the European Union. The first time-series validation dataset assessed the performance of the Land Cover Monitoring, Assessment, and Projection (LCMAP) Collection 1.0 annual land-cover products (Stehman et al., 2021) (called LCMAP\_Val, <https://www.usgs.gov/special-topics/lcmap/validation-data>, last access: 12 March 2024). LCMAP\_Val consisted of 24 971 validation samples with 30 m spatial resolution and covered the time period of 1985 to 2017. It was developed by combining a simple random sampling method and visual interpretation from high-resolution aerial photography, multitemporal Landsat images, and other auxiliary datasets. Meanwhile, to guarantee the reliability of each validation sample, the TimeSync auxiliary tool was also adopted to capture the land-cover changes (Stehman et al., 2021). Quality control was implemented through duplicate interpretations (Xian et al., 2022).





**Figure 1.** Spatial distribution of 84 526 global validation points relating to 17 fine land-cover types in the normal year of 2020.

The second regional validation dataset was the Land Use/Cover Area frame Survey (LUCAS), which is the most comprehensive and largest land-cover validation dataset for the European Union and is freely available at <https://esdac.jrc.ec.europa.eu/projects/lucas> (last access: 12 March 2024). It contains 1 090 863 validation points based on a systematic 2 km × 2 km grid and covers the period from 2006 to 2018 in intervals of 3 years (d’Andrimont et al., 2020). Five LUCAS surveys in 2006, 2009, 2012, 2015, and 2018 assessed the accuracies of the time series of GLC\_FCS30D. LUCAS was developed from a combination of field observations and photo interpretation (Ballin et al., 2018); thus, it can be used with high confidence and also attracts widespread attention for land-cover validations (Gao et al., 2020; Venter et al., 2022).

### 3 Methods

Figure 2 presents a detailed flowchart for monitoring land-cover changes by combining the continuous change-detection (CCD) algorithm proposed by Zhu and Woodcock (2014b) with the local adaptive updating method. Specifically, the flowchart contains four main procedures: (1) detecting the temporally stable pixels and the time points of

abrupt changes in the other land-cover pixels using the continuous change-detection model; (2) deriving the spatiotemporally stable training samples by using the spatiotemporal refinement method from the GLC\_FCS30 land-cover product and temporally stable masks; (3) building the local adaptive classification models for each local region and then updating the land-cover information in the changed pixels; and (4) using the spatiotemporal consistency optimization method to improve the quality of land-cover change maps and suppress false changes.

Before detecting the changed land-cover pixels, all poor-quality pixels (cloud, shadow, and saturated pixels as well as the “scan line corrector off” pixels in Landsat 7) in the continuous-time-series Landsat imagery were first masked using the CFmask algorithm, which has been demonstrated to achieve an overall accuracy of 96.4 % and was adopted by the USGS as its official cloud and shadow detection algorithm (Zhu et al., 2015; Zhu and Woodcock, 2012). Then, in terms of these residual cloud pixels (i.e., those contaminated with light cloud and haze), the Tmask (multiTemporal mask) algorithm, which uses the temporal information from the clear-sky pixels to improve the cloud-detection capability (Zhu and Woodcock, 2014a), was used to mask the residual cloud pixels (note that Tmask

is integrated into the CCD algorithm on the GEE platform as `ee.Algorithms.TemporalSegmentation.Ccdc()`. In other words, the effect of poor-quality pixels was minimized.

### 3.1 The fine classification system used in GLC\_FCS30D

Determining the classification system is usually a prerequisite for land-cover mapping and monitoring. In this study, as we used GLC\_FCS30-2020 as the baseline land-cover product and overlaid the GWL\_FCS30 dataset on GLC\_FCS30D to ensure high accuracy in the wetland areas, the fine classification system used in this study is inherited from those of GLC\_FCS30-2020 and GWL\_FCS30. Table 1 lists the details of the fine classification system. It contains 35 fine land-cover types and has obvious advantages when identifying forest and wetland subcategories.

### 3.2 Detecting changes using the CCD algorithm and continuous Landsat imagery

In general, land-cover changes can be grouped into three categories: periodic changes caused by phenological variability, trend changes driven by natural behavior (such as vegetation growth), and abrupt changes caused by natural or human disturbances (such as deforestation or urban expansion). Thus, capturing these abrupt changes and simultaneously suppressing the periodic and trend changes is the key to land-cover monitoring. In this study, the CCD algorithm (Zhu and Woodcock, 2014b) was used to capture these abrupt changes. This algorithm uses Fourier transformation to fit the time-series observations with the trend term (estimating the trend changes) and harmonic terms (describing the periodic changes) in Eq. (1):

$$\hat{\rho}(i, t) = a_{0,i} + c_{1,i} \times t + \sum_{k=1}^n \left( a_{k,i} \times \cos\left(\frac{2k\pi}{T}t\right) + b_{k,i} \times \sin\left(\frac{2k\pi}{T}t\right) \right), \quad (1)$$

where  $\hat{\rho}(i, t)$  represents the predicted value of the  $i$ th band on the  $t$ th Julian day;  $c_{1,i}$  and  $a_{0,i}$  are the regression slope and intercept of the  $i$ th band, respectively;  $a_{k,i}$  and  $b_{k,i}$  represent the coefficients of the  $k$ th-order harmonic term for the  $i$ th band;  $n$  denotes the number of harmonic terms; and  $T$  is the day number of the year (usually defined as 365). In relation to determining the value of  $n$ , Zhu and Woodcock (2014b) explained that higher-order harmonic terms provided better performance when capturing the periodic variability but caused overfitting in the time-series model and needed more clear-sky observations to initialize the coefficients  $a_{k,i}$  and  $b_{k,i}$ . After balancing the advantages and disadvantages of the higher-order harmonic terms, we finally set  $n$  to be 3, as suggested by other studies (Xian et al., 2022; Xie et al., 2022).

Then, as the CCD is a multivariate change-detection algorithm for capturing the changes in various land-cover types Zhu (2017), five Landsat spectral bands (excluding the blue band for minimizing the effects of the atmosphere and clouds) and three spectral indexes (NDVI (normalized difference vegetation index), NDWI (normalized difference water index), and NBR (normalized burn ratio), as given in Eq. 2) were combined to detect many kinds of changes in the Landsat time series.

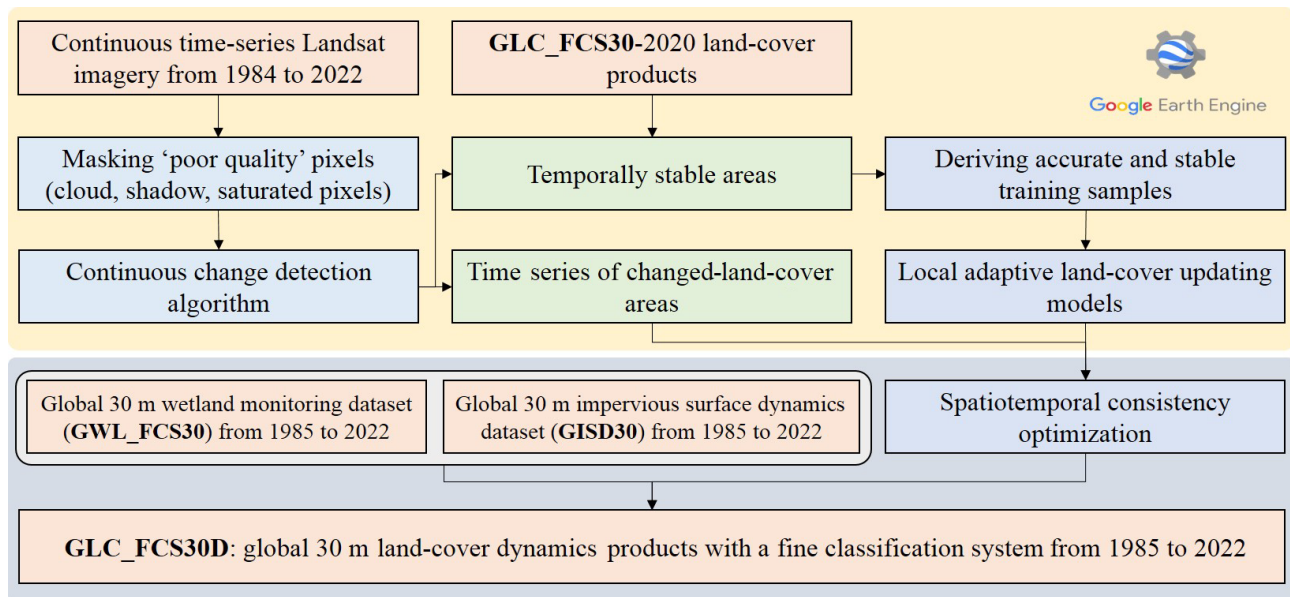
$$\text{NDVI} = \frac{\rho_{\text{nir}} - \rho_{\text{r}}}{\rho_{\text{nir}} + \rho_{\text{r}}}, \quad \text{NDWI} = \frac{\rho_{\text{green}} - \rho_{\text{swir1}}}{\rho_{\text{green}} + \rho_{\text{swir1}}},$$

$$\text{NBR} = \frac{\rho_{\text{nir}} - \rho_{\text{swir1}}}{\rho_{\text{nir}} + \rho_{\text{swir1}}}, \quad (2)$$

where  $\rho_{\text{green}}$ ,  $\rho_{\text{r}}$ ,  $\rho_{\text{nir}}$ , and  $\rho_{\text{swir1}}$  are the surface reflectance of green, red, NIR (near infrared) and SWIR1 (short-wave infrared 1) spectral bands in the Landsat imagery, respectively. Next, to determine the fitted coefficients of the  $k$ th-order harmonic term in Eq. (1), the least absolute shrinkage and selection operator (LASSO) regression algorithm was applied, which demonstrated better performance than the traditional ordinary least squares method in reducing the overfitting problem and dealing with unevenly distributed and sparse Landsat observations (Zhu and Woodcock (2014b)).

Next, the CCD model is a multiparameter change detection model and has been demonstrated to be sensitive to the parameter settings (Xiao et al., 2023; Zhu and Woodcock, 2014b). The CCDC algorithm on the Google Earth Engine platform (`ee.Algorithms.TemporalSegmentation.Ccdc`) contains three key adjustable parameters: `minObservations`, `chiSquareProbability`, and `minNumOfYearScaler`. Zhu et al. (2019) analyzed the relationships of the omission error and commission error of land-cover changes with the variabilities of three parameters in the United States, and they found that the values of the three parameters affected the change detection accuracy. In this study, we also investigated the sensitivity of the change detection accuracy to the parameter settings in Fig. S1 (in the Supplement) using the time-series points from the LCMAP\_Val and LUCAS datasets after partly sampling. Notably, the sensitivity analysis was implemented in two large areas to ensure the feasibility of using the optimized parameters for other areas in land-cover change detection. The results also showed that CCD is a parameter-sensitive algorithm and that the optimal parameter values were 5, 0.95, and 2 years for `minObservations`, `chiSquareProbability`, and `minNumOfYearScaler`, respectively.

After modeling the time-series observations using the CCD algorithm, we analyzed the land-cover changes based on the differences between the actual observations and the predicted values from the time-series fitting models. Figure 3 shows three typical scenarios in which the land-cover dynamics were modeled by the CCD algorithm. Specifically, Fig. 3a illustrates that there was no abrupt break during the whole pe-



**Figure 2.** The flowchart of the proposed method combining the continuous change-detection (CCD) algorithm and a local adaptive updating algorithm.

riod, and thus only a single time-series model was built and the pixel was usually labeled as temporally stable. Figure 3b indicates that the pixel underwent an abrupt change and the time-series observations were split into two segments. The time point of the abrupt change occurred around 1996. Figure 3c gives a complicated time-series disturbance example in which multiple abrupt changes were detected and the time-series observations were split into four segments. The time-series models for segments 1, 2, and 4 showed obvious trend changes.

### 3.3 Updating changed areas using local adaptive classifications

Using the CCD algorithm and continuous Landsat imagery, we identified the temporally stable pixels and the time points of abrupt changes for the land-cover change pixels. Accurately determining the land-cover labels of the changed pixels (or understanding the “from-to” change process) is another key procedure for land-cover time-series monitoring. To achieve this goal, we derived spatiotemporally stable training samples (see Sect. 3.3.1), updated the changed pixels using multitemporal classifications, and finally minimized the cumulative error caused by independent classifications.

#### 3.3.1 Deriving spatiotemporally stable training samples

Numerous studies have demonstrated that the accuracy of the training samples plays a critical role in accurate mapping (Foody and Arora, 2010; Zhang et al., 2020). Visual interpretation can ensure high-confidence samples, but it requires

a lot of manual participation, so it is not suitable for collecting large-area training samples. An alternative option involves generating training samples by refining existing land-cover products through a series of improvement measures (X. Zhang et al., 2021b; Zhang et al., 2023). Inspired by the latter option, we combined the GLC\_FCS30-2020 prior dataset and the change-detection mask (derived using the CCD algorithm described in Sect. 3.2) to obtain the spatiotemporally stable training samples. Specifically, temporally stable areas are known to have higher mapping accuracy (Yang and Huang, 2021; Zhang and Roy, 2017; Zhang et al., 2023); thus, we first used the aforementioned CCD mask to retain the areas that were temporally stable during 1985–2022, and then we overlapped them with the GLC\_FCS30-2020 maps to determine their land-cover labels. Next, because Radoux et al. (2014) emphasized that land-cover transition areas are usually subject to more serious misclassification problems and that pixels with homogeneous land cover have a higher probability of achieving acceptable accuracy, we used a morphological erosion filter of 3 pixels  $\times$  3 pixels to refine these temporally stable areas into spatiotemporally homogeneous areas.

The spatiotemporally stable areas identified using the check for temporal stability during 1985–2022, spatial homogeneity analysis, and the GLC\_FCS30-2020 product (which have an overall accuracy of 82.5 %) were retained to generate the training samples. It should be noted that these spatiotemporally stable areas are not guaranteed to be completely accurate; that is, a small number of the derived training samples may have been mislabeled. Fortunately, previous studies of large-area land-cover mapping demonstrated

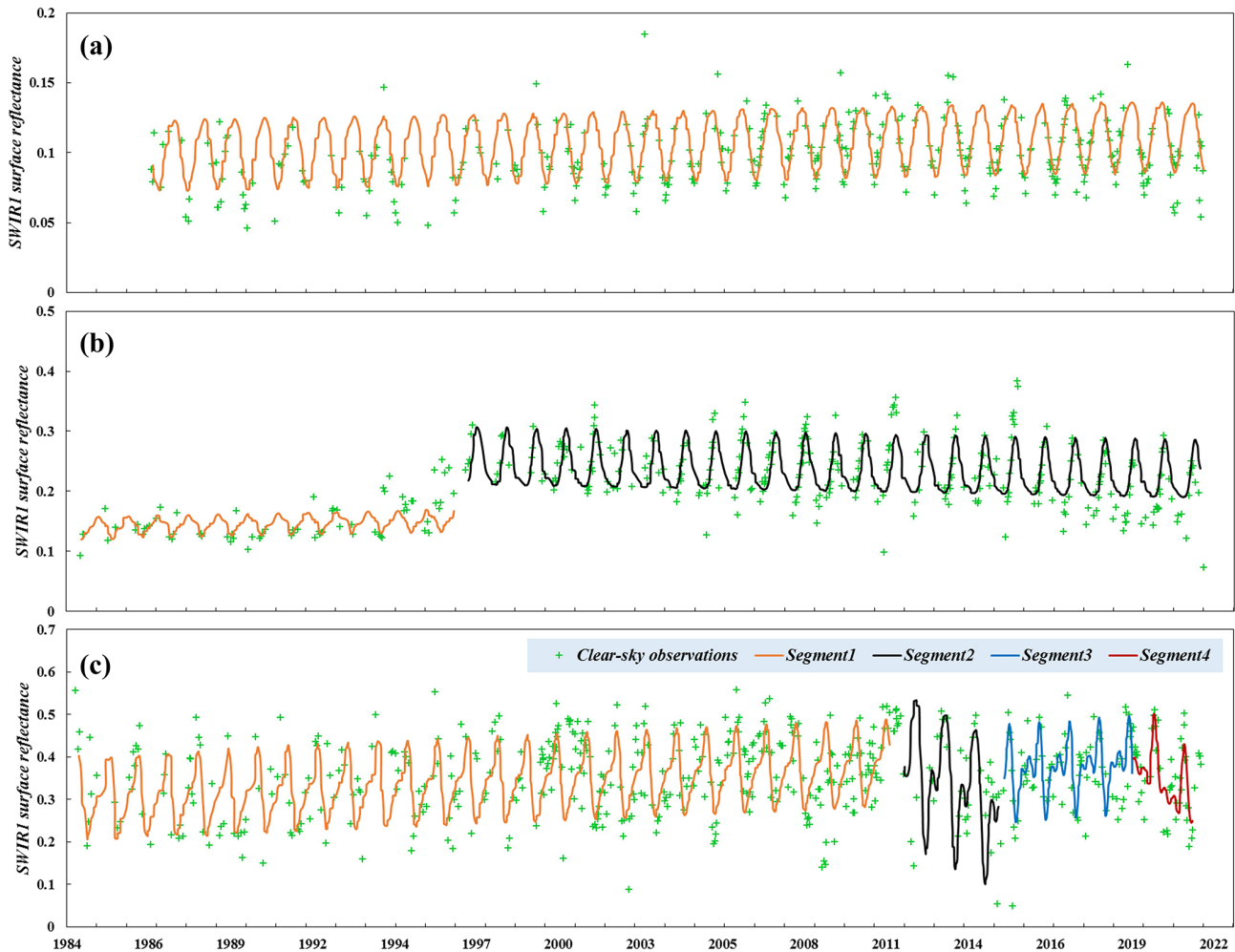
**Table 1.** The details of the fine classification system used in the GLC\_FCS30D land-cover dynamics dataset.

Basic classification system		Level-1 validation system		Fine classification system	ID
Cropland	CRP	Rainfed cropland	RCP	Rainfed cropland	10
				Herbaceous cover cropland	11
				Tree or shrub cover cropland	12
		Irrigated cropland	ICP	Irrigated cropland	20
Forest	FST	Evergreen broadleaved forest	EBF	Closed evergreen broadleaved forest	51
				Open evergreen broadleaved forest	52
		Deciduous broadleaved forest	BDF	Closed deciduous broadleaved forest	61
				Open deciduous broadleaved forest	62
		Evergreen needleleaved forest	ENF	Closed evergreen needleleaved forest	71
				Open evergreen needleleaved forest	72
		Deciduous needleleaved forest	DNF	Closed deciduous needleleaved forest	81
				Open deciduous needleleaved forest	82
		Mixed-leaf forest	MFT	Closed mixed-leaf forest	91
				Open mixed-leaf forest	92
Shrubland	SHR	Shrubland	SHR	Shrubland	120
				Evergreen shrubland	121
				Deciduous shrubland	122
Grassland	GRS	Grassland	GRS	Grassland	130
Tundra	TUD	Lichens and mosses	LMS	Lichens and mosses	140
Wetland	WET	Inland wetland	IWL	Swamp	181
				Marsh	182
				Flooded flat	183
				Saline	184
		Coastal wetland	CWL	Mangrove	185
				Salt marsh	186
		Tidal flat	187		
Impervious surface	IMP	Impervious surface	IMP	Impervious surface	190
Bare areas	BAL	Sparse vegetation	SVG	Sparse vegetation	150
				Sparse shrubland	152
				Sparse herbaceous cover	153
		Bare areas	BAL	Bare areas	200
Consolidated bare areas	201				
Unconsolidated bare areas	202				
Water body	WTR	Water body	WTR	Water body	210
Permanent snow and ice	PSI	Permanent snow and ice	PSI	Permanent snow and ice	220

that the random forest classification model (adopted in this study; see Sect. 3.3.2) is highly robust to erroneous training samples (Gong et al., 2019b; Mellor et al., 2015; X. Zhang et al., 2021b). For example, Gong et al. (2019b) found that the overall accuracy remained relatively stable provided the proportion of erroneous training samples was within 20 %. This provides support for the use of the spatiotemporally stable areas to derive confident training samples and further ensures the quality of land-cover dynamics monitoring.

Numerous studies have highlighted the importance of the training sample balance and distribution, as they significantly influence the mapping performance (Foody, 2009; Jin et al., 2014; Millard and Richardson, 2015). First, there are two options for the training sample distribution: areal-proportional or equal allocation. The former was shown to achieve higher accuracy than the latter option in land-cover mapping, especially in complicated land-cover conditions (Jin et al., 2014). However, when using the areal-proportional sampling strat-





**Figure 3.** Three typical land-cover changes identified using the continuous change-detection (CCD) algorithm and continuous Landsat observations. The times series show (a) a stable land-cover condition, (b) a single abrupt change, and (c) multiple abrupt changes.

egy, the rare land-cover types usually had small sample sizes and were sacrificed because the aim of land-cover mapping was to achieve a global optimum rather than a local optimum. Thus, the study of Zhu et al. (2016) suggested maximum and minimum sample sizes for abundant and rare land-cover types of 8000 and 600, respectively, to avoid the use of extremely large or small sample sizes. Thus, the GLC\_FCS30-2020 product was then split into 961  $5^\circ \times 5^\circ$  geographical tiles, and we used the areal-proportional sampling strategy and two sample balancing thresholds to allocate the training samples from the spatiotemporally stable areas in each  $5^\circ \times 5^\circ$  geographical tile. Lastly, the impervious surface and wetland samples were excluded because both were independently developed as thematic datasets in Sects. 2.3 and 2.4.

### 3.3.2 Updating changed areas using local adaptive classifications

Before building the local adaptive classification models, we had to extract useful spectral features from the Landsat time-series observations. In this study, we used multi-temporal phenological, texture, and topographical features. Specifically, the multitemporal phenological features were extracted by using the percentile-compositing method, which has fewer constraints than other compositing algorithms (such as the season-based compositing method) but achieves similar mapping accuracy (Azzari and Lobell, 2017). The spectral bands for the Landsat time series (five optical bands after excluding the atmospherically sensitive blue band) and the corresponding spectral indexes (NDVI, NDWI, and NBR in Eq. 2) were composited into five percentiles (10th, 25th, 50th, 75th, and 90th). Next, our previous study explained that the texture features made a positive contribution to land-cover mapping (X. Zhang et al., 2021b), so the gray-

level co-occurrence matrix method was used for the 50th-percentile-composited NIR band to extract the homogeneity, entropy, dissimilarity, variance, contrast, and correlation. Lastly, since the land-cover distribution is usually related to the topographical environment (for example, croplands and water bodies are mainly distributed in flat areas), three topographical variables (elevation, slope, and aspect) calculated from a global 30 m DEM dataset named ASTER\_GDEM (Tachikawa et al., 2011) were also imported. In addition, due to the limited storage capacity and satellite–ground data-transmission capacity of early satellites, the density of Landsat imagery is sparse before 2000 (only a single satellite, Landsat 5, acquired data) (Roy et al., 2014b). Thus, we chose a coarse temporal cycle of 5 years to ensure the mapping accuracy before 2000; that is, the satellite observations from 2 years before and after were used for the nominal center year. For example, we updated the land-cover maps in 1995 using all available imagery from 1993 to 1997. In total, there were 49 multisource features, including 40 phenological spectra features, 6 texture features, and 3 topographical variables.

There are two options for global land-cover mapping and updating: global modeling and local adaptive modeling, and our previous studies have found that local adaptive modeling yields superior results compared to global modeling. This is primarily due to the former's capability to take regional characteristics into account more effectively, leading to increased sensitivity in training samples and higher accuracy in land-cover classification (X. Zhang et al., 2021b, 2023, 2022). Thus, we first inherited the regional gridding style used in the GLC\_FCS30 (X. Zhang et al., 2021b); namely, the global land was divided into 961  $5^\circ \times 5^\circ$  geographical tiles. Afterward, the local classification models were independently built to update the land cover in each tile using the corresponding training samples in the neighboring eight surrounding tiles in a  $3 \times 3$  window. The adjacent training samples were imported to increase the continuity of the adjacent land-cover maps.

Lastly, in relation to the selection of the most suitable classification algorithm, the random forest (RF) classifier has significant advantages, including its capacity to accommodate high-dimensional training features and its better ability to deal with the overfitting problem and higher classification accuracy than other widely used classifiers (Belgiu and Drăguș, 2016; Gislason et al., 2006). The RF algorithm was also integrated into the internal function library of the GEE cloud platform as `ee.Classifier.smileRandomForest()`. Thus, the RF algorithm was used to combine the training samples and multisourced features to update the changed pixels. The RF algorithm allows for the adjustment of two key parameters (the number of decision trees (Ntree) and the number of predicted variables (Mtry)), and previous studies have quantitatively analyzed the relationships between classification accuracy and the values of these two parameters. Both theoret-

ical and experimental results indicated that the selection of Mtry and Ntree had little influence on the classification accuracy (Belgiu and Drăguș, 2016; Du et al., 2015). Thus, based on previous studies (Belgiu and Drăguș, 2016; Zhang et al., 2019), the default recommended values of 500 for Ntree and the square of the total number of input features for Mtry were used.

### 3.3.3 Temporal-consistency optimization

To ensure the rationality and consistency of land-cover changes for long time series, the CCD algorithm was applied to capture the time points of land-cover changes, and then the changed pixels were updated using the local adaptive classifications. In this study, despite our best efforts, it was difficult to completely eliminate classification errors, particularly when dealing with changes over time. To address this issue and enhance accuracy in areas with temporal variations, we employed the temporal consistency optimization method described in Eq. (3). This approach incorporates both temporal and spatial neighboring information to assess homogeneity, thereby reducing potential misclassifications of changed areas in the time series.

$$P_{x,y,t} = \frac{1}{N} \left[ \sum_{x'=x-1}^{x'+1} \sum_{y'=y-1}^{y'+1} \sum_{t'=t-1}^{t'+1} \cdot I(L_{x',y',t'} = L_{x,y,t}) \right] \quad (3)$$

Here,  $P_{x,y,t}$  is the homogeneity probability of the pixel at spatial location  $(xy)$  and time point  $t$ ; usually, the higher the value of  $P_{x,y,t}$ , the weaker the classification error effect.  $L_{x,y,t}$  and  $L_{x',y',t'}$  are, respectively, the land-cover labels of the central pixel and the corresponding spatiotemporally neighboring pixels in a local window of  $3 \times 3 \times 3$ , and  $I()$  denotes the indicator function for the equation of the status between two pixels. Namely, if  $L_{x',y',t'}$  is equal to  $L_{x,y,t}$ , then the value of the indicator function is 1; otherwise, it is equal to 0 (Kenny, 2003). In this study, the homogeneity probability was calculated for each changed pixel, and a threshold of 0.5 (as suggested by and used in the studies of Li et al., 2015 and Zhang et al., 2022) was used to judge the rationality of land-cover changes; namely, when  $P_{x,y,t}$  was less than the threshold,  $L_{x,y,t}$  was modified according to the spatiotemporal pixels.

### 3.4 Accuracy assessment

The validation process for the GLC\_FCS30D dataset followed the recommended guidelines proposed by Pontus Olofsson (2014). These guidelines encompass two key components: area estimation (non-site-specific accuracy) and accuracy assessment (site-specific accuracy). The site-specific accuracy assessment mainly focuses on estimating the confusion matrix and calculating some accuracy metrics, including overall accuracy (O.A.), producer's accuracy (P.A.), and user's accuracy (U.A.); and a poststratified estimator is used

to calculate the corresponding standard errors (Pontus Olofsson, 2014).

$$\begin{aligned} P.A._k &= \frac{p_{kk}}{\sum p_k} \\ U.A._k &= \frac{p_{kk}}{\sum p_{.k}} \\ O.A. &= \sum_{k=1}^m p_{kk} \end{aligned} \quad (4)$$

Here,  $p_{kk}$  is the proportion of the area mapped as class  $k$  that had a reference class of  $k$ ;  $\sum p_k$  and  $\sum p_{.k}$  are the proportion of the area mapped as class  $k$  and the proportion of the reference area mapped as class  $k$ , respectively; and  $m$  denotes the number of land-cover types. Afterwards, because there is currently no global long-time-series validation dataset, we used 84 526 global validation points to assess the accuracy metrics of the GLC\_FCS30D dataset in 2020 and used two third-party datasets to analyze the variation across the accuracy time series. GLC\_FCS30D adopts a fine classification system containing 35 subcategories, for which we applied an analysis protocol to the basic classification system and the LCCS level-1 validation system, whose details are explained in the Table 1 and contained 10 major land-cover types and 17 fine land-cover types. Lastly, to quantify the performance of land-cover changed pixels, we followed the proposal of Stehman et al. (2021) for assessing the LCMAP annual land-cover products for 1985–2017; that is, the validation pixels were grouped into “changed” and “unchanged” categories and the corresponding confusion matrix was calculated. Meanwhile, to minimize the imbalance in sample size between the “change” and “no-change” samples, the F1 metric score was calculated as

$$F1 = \frac{P.A. \times U.A.}{P.A. + U.A.} \times 2 \times 100\%. \quad (5)$$

## 4 Results and discussion

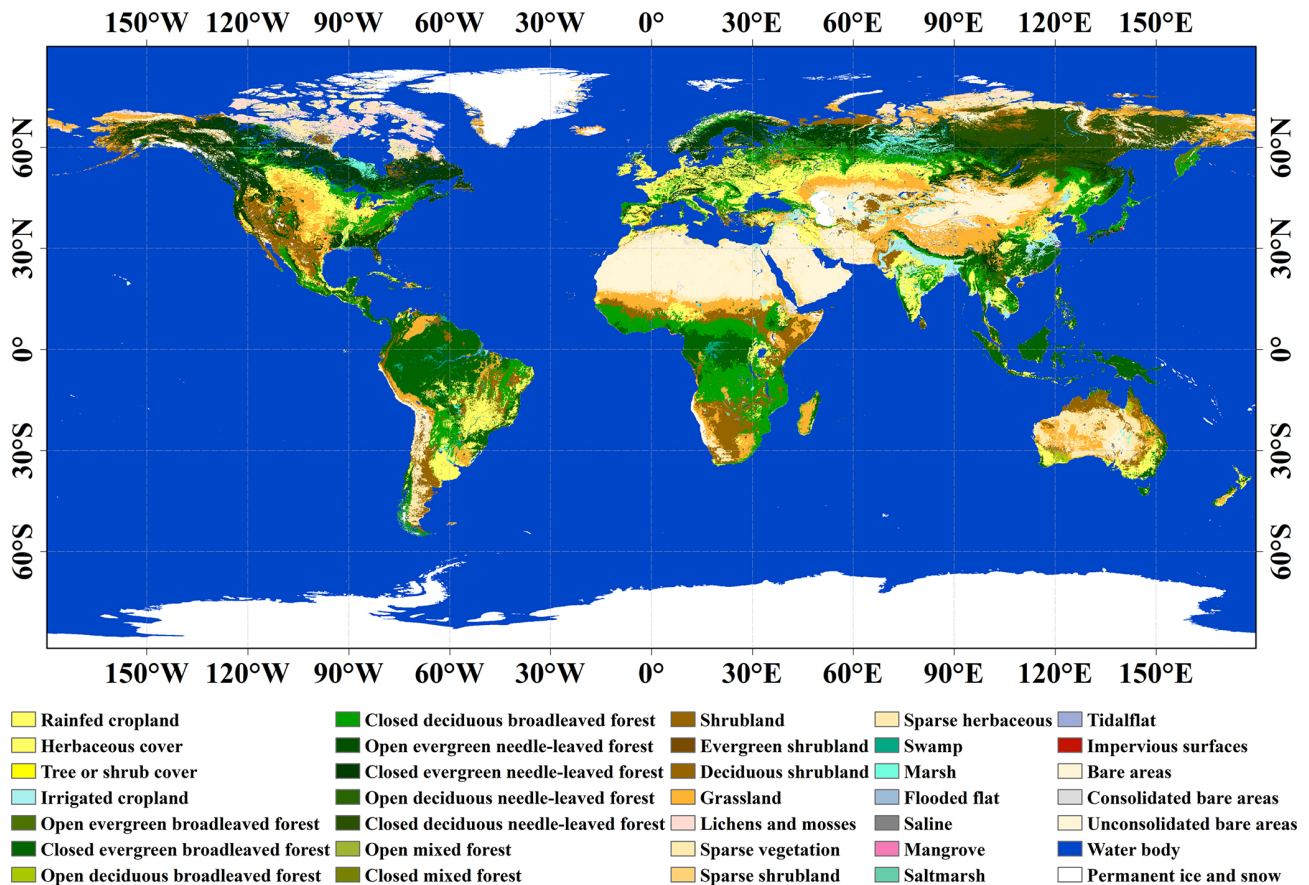
### 4.1 Overview of the GLC\_FCS30D maps and their changes

Figure 4 provides an overview of the GLC\_FCS30D dataset for 2022 (an overview of GLC\_FCS30D for 1985 is given in Fig. S3); overall, it aligns with the real-world land-cover patterns on a global scale. Forest, cropland, barren land, and grassland are the dominant land-cover types, and each of them is distributed in the corresponding ecology subregions. For example, needle-leaved forest is mainly concentrated in the high-latitude cold regions, while broad-leaved forests are mainly distributed in tropical regions; permanent ice and snow are mainly located in Greenland and high-altitude mountains. GLC\_FCS30D has significant advantages over other global land-cover datasets in terms of land-cover type diversity: it contains 35 discrete land-cover types, among which forests and wetlands are subdivided into 10 and 7 land-cover subcategories, respectively.

Figure 5a illustrates the spatial distribution of land-cover change intensity (measuring the proportions of changed pixels in the  $0.05^\circ$  grid) in GLC\_FCS30D from 1985 to 2022 after upscaling to a resolution of  $0.05^\circ$ . Obviously, global land cover has experienced significant changes over the past 37 years, mainly in three areas: (1) areas on the periphery of tropical rainforests in South America and Southeast Asia, with deforestation the dominant cause; (2) areas where wetlands and water bodies intermingle (i.e., water bodies and wetlands transform into one another due to different annual water levels; note that the water body land-cover type in GLC\_FCS30D represents permanent water during the year of interest, although the same area may be a wetland in other years), such as those in North America and northern Asia; and (3) semi-arid areas in Australia, Central Asia, and western Africa, where land cover (such as sparse vegetation or bare land) is directly affected by precipitation and temperature (for example, if there is sufficient precipitation in the year, the sparsely vegetated land and some of the bare land would be covered by grass in semi-arid areas; similarly, the work of Winkler et al., 2021 revealed that these semi-arid areas experienced serious and frequent land-cover changes). Figure 5b quantifies the changed area for 10 major land-cover types from 1985 to 2022. Forest and cropland variations dominated the global land-cover change. The net loss of forests over the past 37 years reached approximately 2.5 million  $\text{km}^2$ , and the decline has been steady over time. Conversely, cropland showed a stable increase, and the net gain in cropland area is approximately 1.3 million  $\text{km}^2$ . Shrubland, wetland, and impervious surface increased in area by 0.45 million, 0.40 million, and 0.37 million  $\text{km}^2$ , respectively. The increased shrubland resulted from the recovery of deforested land, and the wetland gains are due to increases in seasonal water bodies. The work of Pekel et al. (2016) emphasized that global seasonal water bodies (labeled “inland wetlands” in GLC\_FCS30D) showed an overall increase.

Figure 6 further analyzes the net area variations of 10 major land-cover types on six continents. The six continents exhibit various land-cover change characteristics; for example, steady forest loss and cropland gain dominate land-cover change in South America, while the net area variations of most land-cover types fluctuate in Australia. North America has experienced obvious deforestation, and the area of forest loss reaches approximately  $4.5 \times 10^5 \text{ km}^2$ . In contrast, the shrubland, grassland, and impervious surface land-cover types show increasing trends overall, with increases of  $1.4 \times 10^5 \text{ km}^2$ ,  $0.8 \times 10^5 \text{ km}^2$ , and  $0.72 \times 10^5 \text{ km}^2$ , respectively. Similarly, Xian et al. (2022) reported that forest losses and shrubland, grassland, and impervious surface gains were the dominant characteristics of the CONUS from 1985 to 2017. In Europe, the forest area continuously decreased, and the cropland area first decreased and then increased because of the collapse of the Soviet Union in 1990s. Abandoned croplands were transformed into pasture (which also belongs to the cropland land-cover type in GLC\_FCS30D). Among





**Figure 4.** Overview of GLC\_FCS30D for 2022; the color-coded legend is derived from the European Space Agency (ESA) Climate Change Initiative land-cover dataset (Defourny et al., 2018).

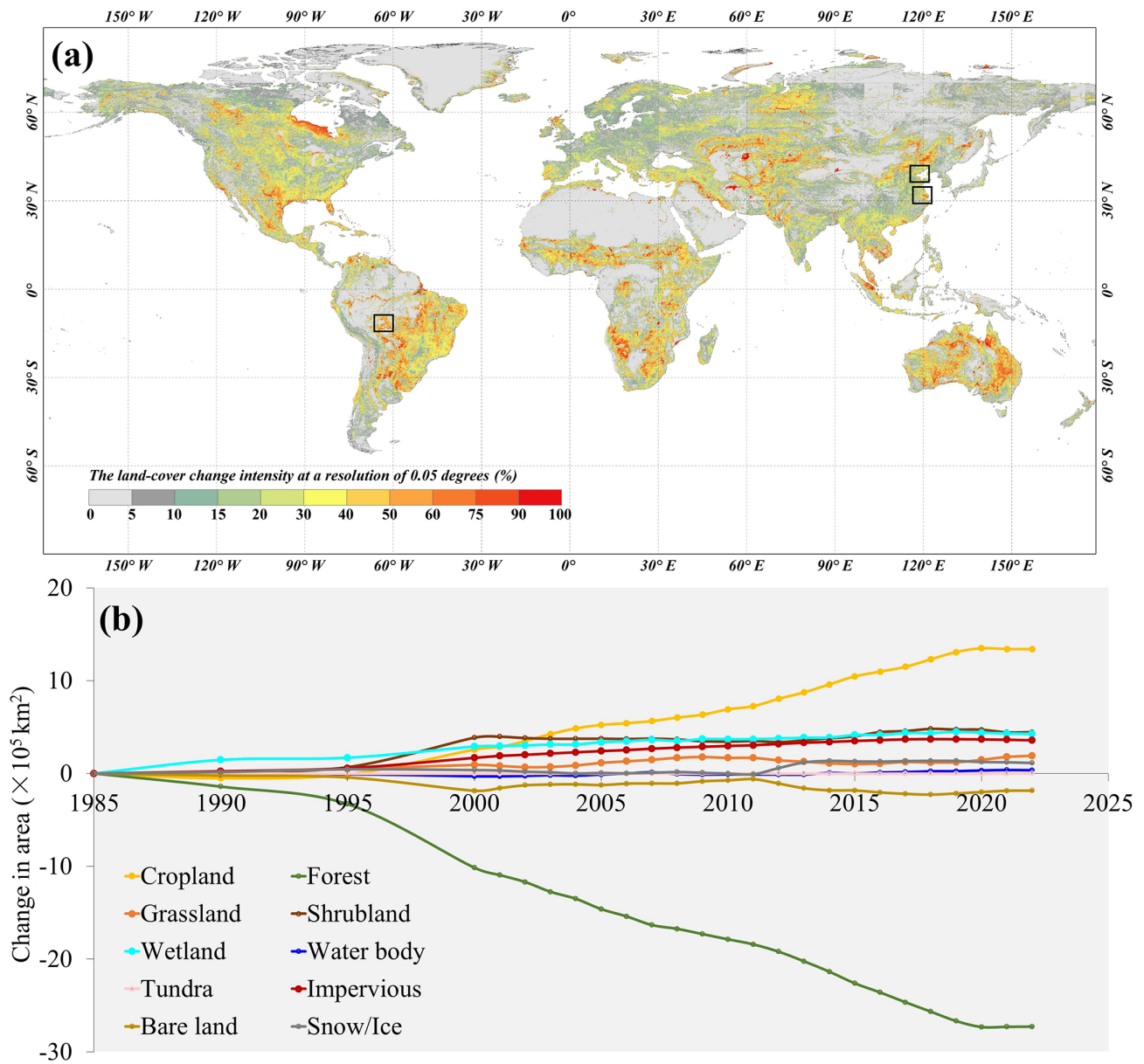
the six continents, the increase in impervious surface is the most significant in Asia, with a net increase of  $1.9 \times 10^5 \text{ km}^2$ ; wetland also shows a large increase of  $1.1 \times 10^5 \text{ km}^2$ . The increased wetland coverage comes from the increase in seasonal water bodies. South America and Africa have experienced similar land-cover change characteristics: they show the most intense deforestation rates and the most significant increases in cropland. According to our statistics, the forest loss on these two continents amounts to  $16.9 \times 10^5 \text{ km}^2$ , and the corresponding increase in cropland is approximately  $11.1 \times 10^5 \text{ km}^2$ . Lastly, Oceania is more sensitive to climate change, especially in terms of precipitation, so fluctuations in shrubland, grassland, and bare land are evident there because the conversion relationship between the three land-cover types is related to annual precipitation.

Figure 7 displays the land-cover transformation relationships from 1985 to 2022 in the GLC\_FCS30D dataset using Sankey diagrams. Global cropland and forest show obvious area changes, with area proportions changing from 12.08 % and 38.26 %, respectively, in 1985 to 12.86 % and 36.48 %, respectively, in 2022. Shrubland changed from 8.70 % in 1985 to 9.03 % in 2022. We mainly focus on forest, cropland,

shrubland, and impervious surface changes, which dominate the land-cover changes in Fig. 5. There are three main causes of forest loss over the past 37 years: (1) 37.58 % of the deforested land was converted to cropland (this transformation was more significant in tropical rainforest areas; Fig. 8a); (2) 26.92 % of the lost forest was regrown as shrubland, which is more common in mountainous areas affected by wildfires; and (3) 13.49 % of the deforested land was converted to grassland. Cropland is converted to forest, grassland, and impervious surfaces. A total of 26.29 % of the lost cropland is converted to grassland due to abandonment, 25.88 % of the lost cropland is covered by forests, and 21.01 % of the lost cropland resulted from urbanization. Lastly, regarding impervious surfaces, our primary focus was on identifying the sources contributing to their expansion. Our findings indicate that approximately 36.24 % of the impervious surface increase can be attributed to the conversion of cropland, while 13.49 % of the increase is a result of deforestation.

To visually understand the land-cover change process captured by the GLC\_FCS30D dataset over the past 37 years, Fig. 8 displays three typical enlargements (the spatial loca-

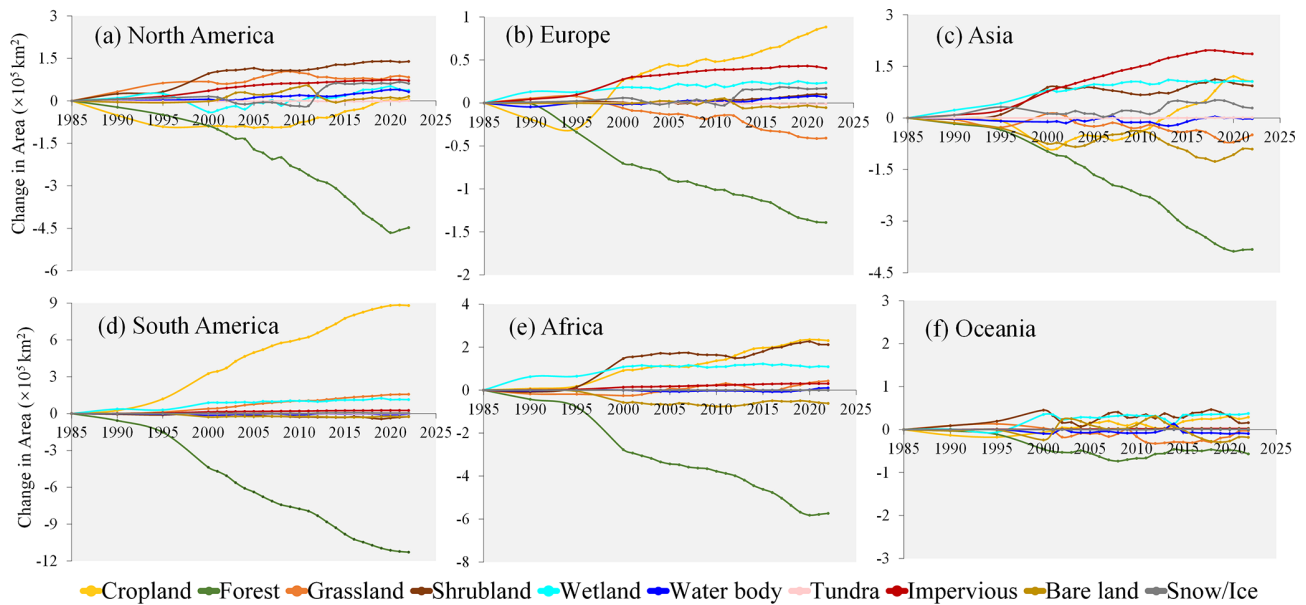




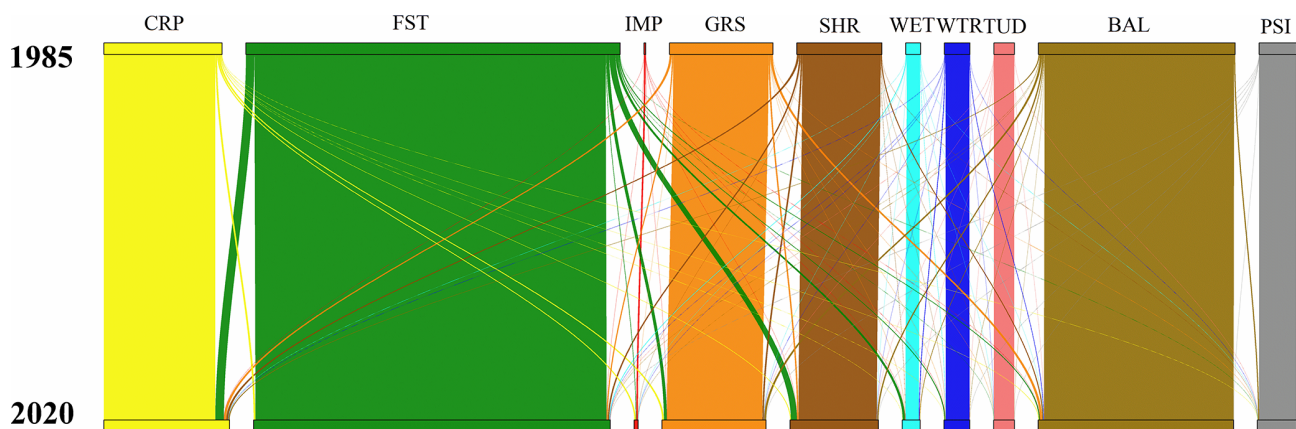
**Figure 5.** (a) The spatial distribution of global land-cover-change intensity from 1985 to 2022 after aggregation to a resolution of 0.05°. (b) The net changed area of 10 major land-cover types in GLC\_FCS30D from 1985 to 2022.

tions of the enlarged areas are illustrated as two black rectangles in Fig. 5a) of the Amazon rainforest (which experienced significant deforestation) and China’s Yangtze River Delta (which underwent rapid urbanization) and Yellow River delta (which showed evident land-cover changes over coastal regions). These three typical areas experienced drastic land-cover changes, and GLC\_FCS30D accurately captures those spatiotemporal changes. Specifically, the deforestation in South America is widely recognized, and GLC\_FCS30D clearly reflects this trend. Namely, the early deforestation showed a grid distribution, and then each grid gradually extended outward and finally connected into patches.

GLC\_FCS30D also shows that deforestation has not stopped in the region in terms of the rate of forest loss, and these findings are in line with the results of earlier research (Harris et al., 2021; Potapov et al., 2022). In the Yangtze River Delta, GLC\_FCS30D depicts that the dominant land-cover change over the enlargement is urbanization, and a large quantity of irrigated cropland has been converted to impervious surfaces. Meanwhile, urban expansion was significantly faster before 2010 than after 2010 according to GLC\_FCS30D. Lastly, the Yellow River delta, a typical coastal region, was selected to assess the ability of GLC\_FCS30D to capture these coastal land-cover changes. Obviously, the land-cover changes in



**Figure 6.** The net area variations of 10 major land-cover types on six continents from 1985 to 2022.



**Figure 7.** Sankey diagrams of the global land-cover changes during 1985–2022 in the GLC\_FCS30D dataset.

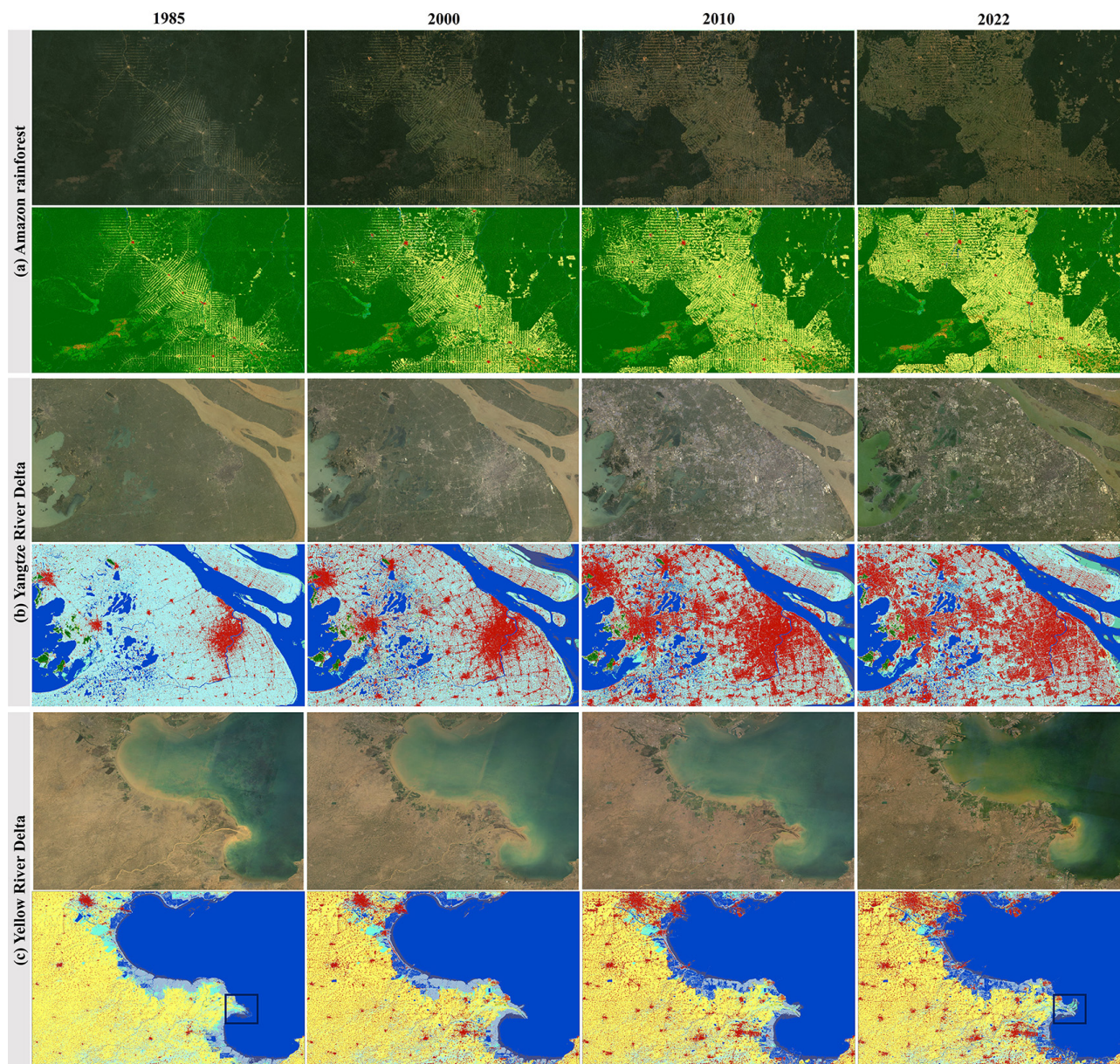
GLC\_FCS30D can be summarized as three aspects: (1) a large number of flooded flats and tidal flats were reclaimed as aquaculture ponds, especially after 2000; (2) the mouth of the Yellow River changed from a southerly orientation to a northerly one (black rectangle), i.e., there were large land-cover changes between tidal/flooded flats, water bodies, and salt marshes; and (3) a lot of impervious surfaces encroached on the coastal water bodies and flats. In short, by combining real remote-sensing observation time series data, GLC\_FCS30D effectively captures the spatiotemporal changes of the land surface.

#### 4.2 Accuracy assessment of GLC\_FCS30D for 2020

Table 2 provides the error matrix and accuracy metrics for the GLC\_FCS30D dataset in the basic classification system con-

taining 10 major land-cover types. The novel GLC\_FCS30D dataset attained an O.A. of 80.88 % ( $\pm 0.27$  %). Cropland, forest, impervious surface, water body, and permanent snow and ice perform better in terms of the P.A. and U.A. than the remaining land-cover types, with the corresponding accuracies exceeding 85 %. The impervious surface and wetland datasets are independently generated and then overlaid on GLC\_FCS30D, helping these complicated land-cover types to achieve high accuracy metrics. Conversely, grassland, shrubland, and tundra have lower accuracies; for example, grassland had the lowest P.A. of 54.41 % and shrubland had the lowest U.A. of 57.63 %. The two reasons that they performed poorly were as follows: (1) these land-cover types usually reflected heterogeneous and varied spectral and spatial characteristics, e.g., the grassland showed similar spectra to cropland and sparse shrubland in the growing season but



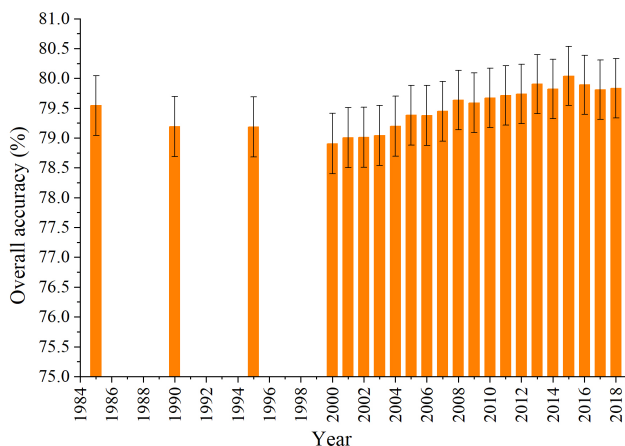


**Figure 8.** Three typical enlargements showing the land-cover changes according to GLC\_FCS30D from 1985 to 2022 in (a) the Amazon rainforest, (b) the Yangtze River Delta in China, and (c) the Yellow River delta in China. The color coding is the same as that used for the global map in Fig. 4. In each case, the natural-color imagery from 1985 to 2022 are composites taken from Landsat imagery.

mimicked bare-land features in harvest season, and (2) all of them were distributed in climate-transition areas with complicated climate variations and landscapes.

Table 3 provides the error matrix of GLC\_FCS30D for 2020 in the LCCS level-1 validation system with 17 land-cover types. The GLC\_FCS30D-2020 dataset achieves an O.A. of 73.04 % ( $\pm 0.30$  %), which is lower than that achieved with the basic classification system because these similar land-cover subcategories suffer more easily from misclassification. For example, forest has a P.A. of 92.83 % ( $\pm 0.31$  %), but the P.A. rapidly decreases to the range of

58.29 % ( $\pm 1.53$  %) to 82.39 % ( $\pm 0.98$  %) when forest is split into five fine subcategories. Cropland, forest, and bare land, which are further divided into multiple subcategories, show obvious decreases in accuracy for their subcategories in terms of P.A. and U.A. Taking cropland and forest as examples, approximately 31.7 % of the irrigated cropland (ICP) is misclassified as rainfed cropland (RCP), and so the U.A. of ICP is only 59.92 %. More than 53.8 % of the mixed-leaf forests (MFT) are wrongly labeled as the other four forest subcategories, and so the mixed-leaf forests have the lowest U.A. of 39.34 % ( $\pm 1.38$  %). Meanwhile, sparse vegetation



**Figure 9.** Time series of the overall accuracy of the GLC\_FCS30D dataset using the LCMAP\_Val annual reference dataset across the contiguous United States (CONUS) from 1985 to 2018. The error bars on the graph show the uncertainty of the data points.

has the second lowest U.A. of 50.63 % ( $\pm 1.47\%$ ) because of the confusion between sparse vegetation, grassland, and bare land. In the basic classification system (Table 1), sparse vegetation is grouped with bare land. A previous study in the European Union proposed grouping it with grassland (Gao et al., 2020). Wetland is further divided into coastal wetland (CWL) and inland wetland (IWL) in Table 3, and CWL has a higher U.A. than that of wetland in Table 2, which is primarily attributable to the significantly more accurate classification of wetland in the CWL subcategory (Zhang et al., 2023).

### 4.3 Accuracy assessment based on two third-party regional validation datasets

#### 4.3.1 Time series of accuracy metrics of GLC\_FCS30D from the LCMAP\_Val dataset

Figure 9 displays a time series of the overall accuracy of the GLC\_FCS30D dataset using the LCMAP\_Val annual validation dataset from 1985 to 2018 over the CONUS. GLC\_FCS30D achieves a mean O.A. of 79.50 % ( $\pm 0.50\%$ ) and the O.A. varies from a high value of 80.04 % ( $\pm 0.49\%$ ) in 2015 to a low value of 78.91 % ( $\pm 0.51\%$ ) in 2000. The overall accuracy of GLC\_FCS30D is slightly lower at the early stage, which might be related to the density of Landsat observations. The early Landsat missions had weaker satellite-to-ground transmission and onboard recording capabilities (Roy et al., 2014a), so the phenological variability and land-cover changes were more difficult to capture at the early stage.

Figure 10 shows time series of the P.A. and U.A. for the GLC\_FCS30D dataset in the CONUS. Visually, the P.A. and U.A. values of the 10 major land-cover types range from 45 % to 100 % and from 35 % to 100 %, respectively, and

**Table 2.** Error matrix of the GLC\_FCS30D dataset for 2020 based on the basic classification system. The reported producer’s accuracy (P.A.) and user’s accuracy (U.A.) come with corresponding standard errors (SE) shown in parentheses.

Reference	Map										Total	P.A. (SE)
	CRP	FST	GRS	SHR	WET	WTR	TUD	IMP	BAL	PSI		
CRP	15.442	0.792	0.679	0.388	0.086	0.027	0	0.174	0.117	0	17.704	87.22 (0.54)
FST	0.513	28.712	0.315	0.811	0.371	0.021	0.008	0.063	0.113	0.002	30.93	92.83 (0.31)
GRS	1.035	1.166	5.906	1.181	0.231	0.011	0.084	0.051	1.181	0.01	10.855	54.41 (1.02)
SHR	0.555	1.798	0.863	5.392	0.161	0.013	0.019	0.05	0.502	0.002	9.356	57.63 (1.09)
WET	0.068	0.465	0.156	0.157	4.047	0.347	0.031	0.021	0.222	0.001	5.516	73.37 (1.27)
WTR	0.04	0.086	0.019	0.017	0.302	3.305	0.008	0.012	0.039	0.002	3.831	86.28 (1.12)
TUD	0.01	0.123	0.168	0.167	0.018	0.03	2.444	0.002	0.473	0.02	3.454	70.76 (1.65)
IMP	0.084	0.058	0.024	0.04	0.001	0.006	0.002	5.043	0.024	0	5.283	95.45 (0.61)
BAL	0.13	0.049	0.783	0.585	0.043	0.045	0.577	0.048	9.239	0.131	11.628	79.45 (0.8)
PSI	0	0.004	0.03	0.005	0	0.023	0.001	0	0.03	1.351	1.443	93.63 (1.38)
Total	17.877	33.251	8.943	8.743	5.259	3.828	3.176	5.464	11.94	1.52		
U.A. (SE)	86.38 (0.55)	86.35 (0.4)	66.05 (1.07)	61.68 (1.11)	76.96 (1.2)	86.33 (1.35)	76.97 (1.6)	92.29 (0.77)	77.38 (0.82)	88.89 (1.72)		

Note: the abbreviations correspond to the 10 categories of the basic classification system in Table 1.



**Table 3.** Error matrix of the GLC\_FCS30D dataset for 2020 based on the LCCS level-1 validation system. The reported producer's accuracy (P.A.) and user's accuracy (U.A.) come with corresponding standard errors (SE) shown in parentheses.

Reference	RCP	ICP	EBF	DBF	ENF	DNF	MFT	SHR	GRS	LMS	SVG	IWL	CWL	IMP	BAL	WTR	PSI	Total	P.A. (SE)
RCP	12.225	1.023	0.239	0.358	0.102	0.016	0.009	0.382	0.66	0	0.078	0.056	0.005	0.124	0.028	0.001	0	15.332	79.7 (0.7)
ICP	0.397	1.932	0.026	0.016	0.005	0	0	0.01	0.025	0	0.012	0.029	0.005	0.052	0	0.018	0	2.527	76.45 (1.81)
EBF	0.2	0.048	9.091	1.098	0.262	0.103	0.151	0.371	0.084	0	0.012	0.136	0.028	0.029	0.001	0.004	0	11.514	78.96 (0.82)
DBF	0.187	0.016	0.632	6.838	0.537	0.294	0.396	0.235	0.144	0.002	0.019	0.077	0.002	0.025	0.005	0.004	0.002	9.054	75.53 (0.97)
ENF	0.046	0.004	0.174	0.316	5.681	0.328	0.439	0.128	0.034	0.006	0.043	0.094	0	0.008	0.01	0.01	0	6.895	82.39 (0.98)
DNF	0.008	0	0.002	0.13	0.245	1.854	0.073	0.071	0.053	0	0.011	0.025	0	0.001	0.007	0.002	0	2.414	76.79 (1.85)
MFT	0.004	0	0.019	0.176	0.234	0.013	0.828	0.014	0.004	0	0	0.010	0.05	0	0.001	0	0	1.308	58.29 (1.53)
SHR	0.518	0.042	0.299	0.9	0.328	0.131	0.034	5.44	0.871	0.019	0.441	0.157	0.005	0.05	0.065	0.013	0.002	9.438	57.63 (1.09)
GRS	0.947	0.097	0.167	0.582	0.209	0.154	0.024	1.191	5.958	0.085	0.974	0.229	0.006	0.052	0.217	0.008	0.01	10.95	54.41 (1.02)
LMS	0.006	0.004	0.001	0.022	0.044	0.053	0.001	0.168	0.169	2.465	0.379	0.02	0.001	0.002	0.098	0.026	0.02	3.484	70.76 (1.65)
SVG	0.064	0.01	0.008	0.006	0.007	0.01	0.001	0.397	0.462	0.025	2.71	0.012	0	0.013	0.643	0.002	0.024	4.399	61.6 (1.57)
IWL	0.01	0.002	0.044	0.029	0.103	0.022	0.002	0.048	0.017	0.008	0.042	2.673	0.024	0.001	0.012	0.224	0	3.263	81.91 (1.45)
CWL	0.004	0.002	0.008	0.002	0.004	0.002	0.004	0.008	0.006	0	0.008	0.188	1.476	0.007	0.007	0.059	0	1.783	82.77 (1.92)
IMP	0.074	0.011	0.008	0.008	0.037	0.002	0	0.041	0.024	0.002	0.014	0.004	0	5.087	0.01	0.004	0	5.329	95.45 (0.61)
BAL	0.048	0.01	0.002	0.004	0.002	0.001	0	0.193	0.328	0.557	0.582	0.043	0.002	0.035	5.384	0.029	0.108	7.33	73.45 (1.11)
WTR	0.014	0.024	0.014	0.014	0.019	0.008	0.006	0.011	0.016	0.007	0.011	0.168	0.114	0.011	0.019	3.054	0.002	3.509	87.04 (1.22)
PSI	0	0	0	0.001	0.002	0	0	0.005	0.03	0.001	0.011	0	0	0	0.019	0.023	1.363	1.455	93.65 (1.37)
Total	14.757	3.224	10.753	10.56	7.833	3.724	1.97	8.711	8.883	3.179	5.353	3.927	1.668	5.497	6.526	3.482	1.532		
U.A. (SE)	82.85 (0.67)	59.92 (1.85)	84.55 (1.75)	64.76 (1)	72.52 (1.08)	49.77 (1.76)	39.34 (1.38)	62.44 (1.11)	67.07 (1.07)	77.55 (1.59)	50.63 (1.47)	68.07 (1.6)	88.49 (1.68)	92.54 (0.76)	82.5 (1.01)	87.73 (1.19)	88.96 (1.72)		
O.A.	73.04% ( $\pm 0.30\%$ )																		

Note: the abbreviations correspond to the 17 categories of the LCCS validation system in Table 1.

the time-series variations are stable. Among them, the water body land-cover type has the highest accuracy metrics, achieving mean P.A. and U.A. values of 95.31 % ( $\pm 1.14$  %) and 98.53 % ( $\pm 0.66$  %), respectively, which benefit from the unique spectral characteristics of this land-cover type. Cropland follows, with mean P.A. and U.A. values of 93.37 % ( $\pm 0.74$  %) and 87.70 % ( $\pm 0.94$  %), respectively. Forest ranks third, with a high P.A. of 97.75 % ( $\pm 0.35$  %) but a relatively low U.A. of 82.42 % ( $\pm 0.82$  %); the unbalanced metrics occur because GLC\_FCS30D and LCMAP\_Val have different definitions for forest. GLC\_FCS30D defines a forest as having a tree cover greater than 15 %, while the threshold setting of LCMAP\_Val is 10 %, so many of the shrublands in GLC\_FCS30D are labeled as forests in LCMAP\_Val. Wetland has a U.A. value of 90.47 % ( $\pm 2.05$  %) but a P.A. value of 57.07 % ( $\pm 2.75$  %), which is also caused by a discrepancy in the definition of wetland. GLC\_FCS30D identifies seasonal water bodies as wetlands while LCMAP\_Val classifies them as water bodies. Impervious surface has a P.A. lower than 60 %, mainly because the GLC\_FCS30D and LCMAP\_Val datasets have different definitions of impervious surface. LCMAP\_Val defines buildings and the surrounding green areas as developed, while GLC\_FCS30D only includes the artificial buildings (houses, roads, squares, and so on). Bare land and shrubland have the lowest U.A. values of 35.58 % ( $\pm 4.39$  %) and 47.29 % ( $\pm 1.56$  %), respectively, mainly because both of them are easily confused with grassland due to their complicated spectral characteristics and because they coexist in climate-sensitive semi-arid regions (e.g., the Midwestern United States). Xian et al. (2022) emphasized that long-term monitoring of shrubs and grasslands presents significant challenges in the CONUS. Permanent snow and ice, which is sparsely distributed in high-elevation mountainous areas of the United States, has unique and specific spectral characteristics, so it achieves a P.A. of 100 % in GLC\_FCS30D. The large fluctuations in U.A. for ice and snow are attributed to (1) the small sample size for ice and snow in the LCMAP\_Val dataset and (2) a few misclassified grass/bare land pixels that are correctly identified as snow and ice during 2005–2014.

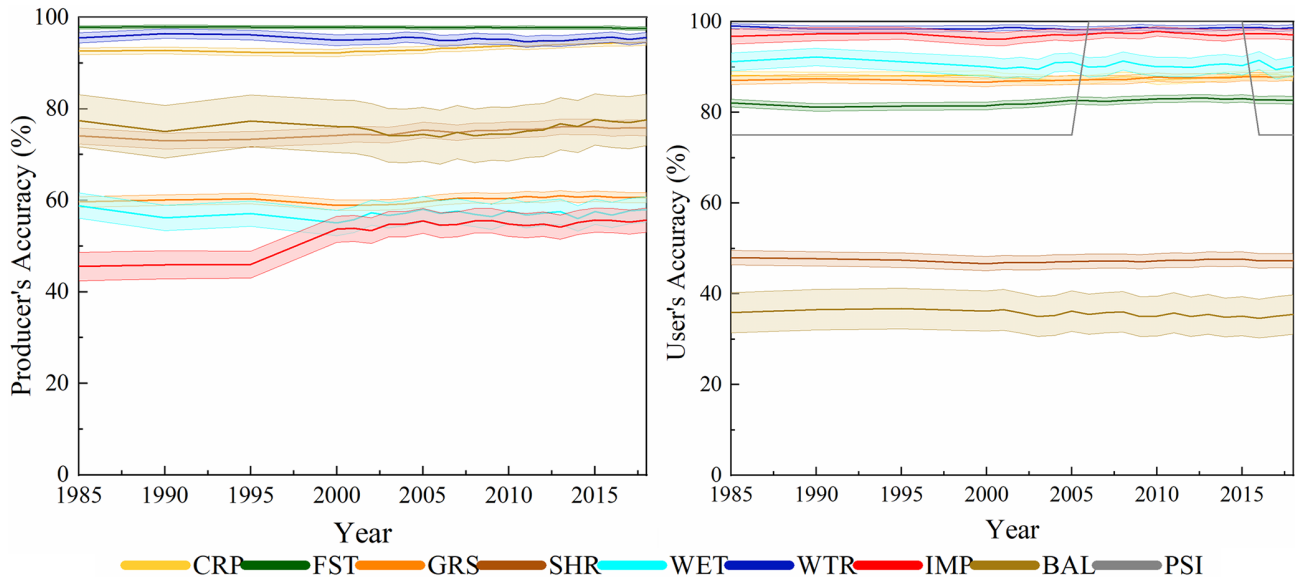
Figure 11 shows the area-bias percentages of eight land-cover types estimated by GLC\_FCS30D and LCMAP\_Val across the CONUS. Intuitively, GLC\_FCS30D and LCMAP\_Val share similar total areas for estimations of cropland, bare land, and water bodies but evident area deviations for estimations of forest, shrubland, and grassland. The deviations in shrubland and grassland occur mainly because these land-cover types coexist in the semi-arid regions of the central United States and share similar spectral characteristics and temporal variability; thus, some grasslands in LCMAP\_Val are considered shrublands in GLC\_FCS30D. Xian et al. (2022) also failed to distinguish grassland from shrubland and combined them into a group when generating LCMAP annual maps. LCMAP\_Val has a broader definition of impervious surface, which results in negative bias, so the

impervious surface area estimated in LCMAP\_Val is larger than the assessment in the GLC\_FCS30D dataset.

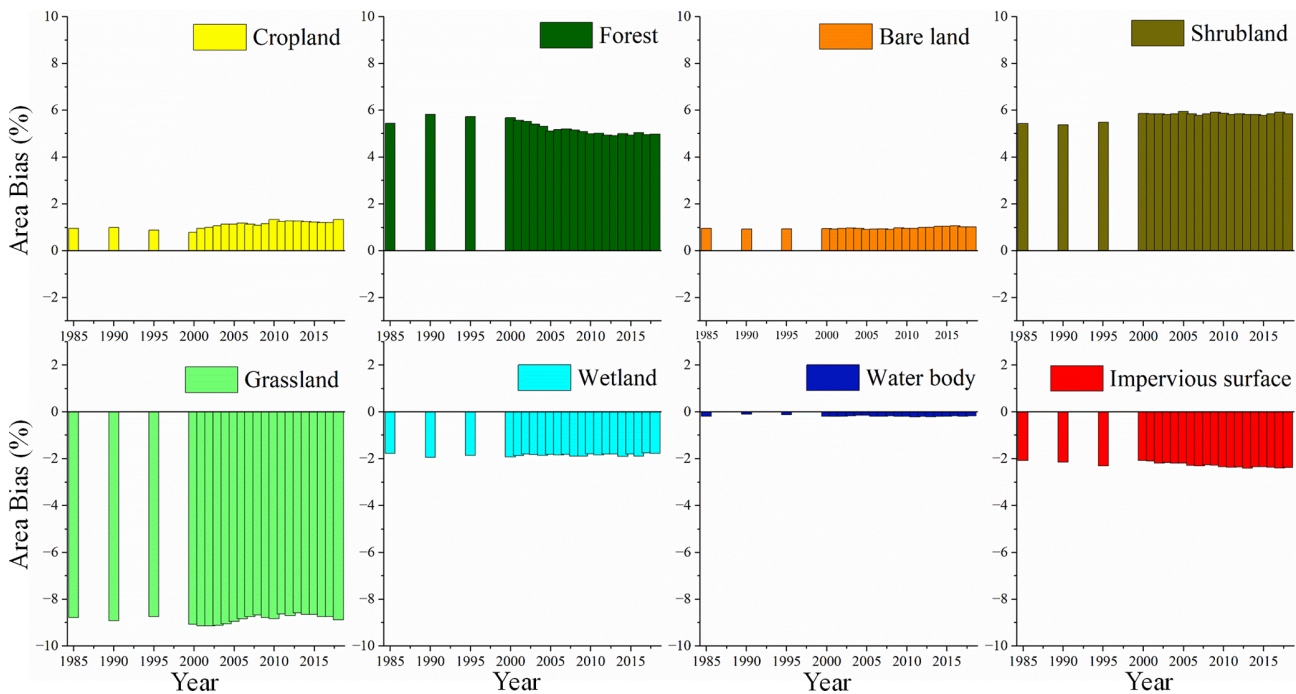
Table 4 analyzes the confusion matrix of changed and unchanged land-cover pixels in GLC\_FCS30D using the LCMAP\_Val dataset. It should be noted that the changed land-cover samples in LCMAP\_Val were still sparse; that is, the size of the changed samples cannot support the analysis of specific land-cover changes. Similarly, Stehman et al. (2021) grouped the land-cover types into “no change” and “change” types for analyzing the land-cover changes. In this study, when using the “changed” and “unchanged” validation points in LCMAP\_Val, the O.A. of GLC\_FCS30D reached  $90.49 \pm 0.45$  %. In particular, the unchanged land-cover pixels played a dominant role and reached a high P.A. of 92.84 % and a high U.A. of 96.28 %. In contrast, the P.A. and U.A. of the changed land-cover pixels were  $72.26 \pm 2.04$  % and  $56.62 \pm 2.00$  %, and the F1 score was 63.49 %.

#### 4.3.2 Time series of accuracy metrics of GLC\_FCS30D from the LUCAS dataset

Table 5 lists the time series of accuracy metrics of the GLC\_FCS30D dataset across the European Union (EU) from 2006 to 2018 using the LUCAS dataset. The GLC\_FCS30D dataset has a mean O.A. of 81.91 % ( $\pm 0.09$  %) and an O.A. range of 81.64 % (0.09 %) to 82.11 % (0.09 %) in the EU. The two dominant land-cover types (cropland and forest), which cover almost 70 % of the entire EU area (Gao et al., 2020), have higher P.A. and U.A. values than the other land-cover types. The P.A. and U.A. of cropland exceed 85 % and 93 %, respectively. Forest has unbalanced P.A. (approximately 95 %) and U.A. (approximately 76 %) values because the LUCAS dataset defines forest more broadly than the GLC\_FCS30D dataset does. In particular, sparse vegetation associated with forest is grouped with forest in LUCAS but with bare land in GLC\_FCS30D. Gao et al. (2020) explained the discrepancy in the definition of forest between LUCAS and GLC\_FCS30D. Shrubland, grassland, and bare land showed inferior performance in terms of both P.A. and U.A. because of their complicated spectral variability and spatial heterogeneity. Gao et al. (2020) also found that three global 30 m land-cover products (GlobeLand30, FROM\_GLC, and GLC\_FCS30) exhibited poor performance for these three land-cover types. Urban green space and discontinuous urban fabric, which are excluded from GLC\_FCS30D, are grouped with impervious surface in LUCAS. Thus, impervious surface also has a low P.A. of approximately 59 %. Lastly, upon investigating the temporal variability of P.A. and U.A. we find that permanent ice and snow and wetland show greater variability and that both are closely related to annual temperature and precipitation; i.e., their spatial distributions are affected by the natural environment.



**Figure 10.** Time series of the producer’s accuracy and user’s accuracy of GLC\_FCS30D based on the LCMAP\_Val dataset from 1985 to 2018 in the contiguous United States (CONUS). The error bands represent  $\pm 1$  standard error.



**Figure 11.** The area-bias percentages of eight land-cover types in the GLC\_FCS30D and LCMAP\_Val datasets from 1985 to 2017 in the contiguous United States (CONUS).

Table 6 shows the area proportions of 10 major land-cover types from both the GLC\_FCS30D dataset (“Map”) and the LUCAS validation dataset (“Ref”). The area bias (“AB”) measures the area deviations between the two different datasets for the same land-cover type. Overall, GLC\_FCS30D overestimates the total area assessments of forest, bare land, and ice and snow and underestimates the re-

maining land-cover types in comparison to the LUCAS estimations. In particular, according to its AB of +7.356 %, forest shows the most significant overestimation, while cropland shows the most underestimation (AB of –4.086 %). Cropland and forest cover together account for approximately 70 % of the total EU area (Gao et al., 2020)); as a result, the area bias (AB) values for these two land-cover types are

**Table 4.** The confusion matrix of changed and unchanged pixels in GLC\_FCS30D when using the LCMAP\_Val datasets.

	Unchanged	Changed	Total	P.A. (SE)	F1
Unchanged	82.21	6.34	88.55	92.84 (0.42)	94.53
Changed	3.18	8.27	11.45	72.26 (2.04)	63.49
Total	85.39	14.61			
U.A. (SE)	96.28 (0.32)	56.62 (2.00)			
O.A. (SE)		90.49 (0.45)			

**Table 5.** Time series of accuracy metrics of the GLC\_FCS30D dataset using the LUCAS validation dataset across the European Union.

	2006		2009		2012		2015		2018	
	P.A. (SE)	U.A. (SE)	P.A. (SE)	U.A. (SE)	P.A. (SE)	U.A. (SE)	P.A. (SE)	U.A. (SE)	P.A. (SE)	U.A. (SE)
CRP	85.49 (0.11)	93.37 (0.08)	85.40 (0.11)	93.31 (0.08)	85.50 (0.11)	93.17 (0.08)	85.47 (0.11)	93.05 (0.08)	85.52 (0.11)	92.82 (0.08)
FST	95.22 (0.08)	76.71 (0.15)	94.97 (0.08)	76.71 (0.15)	94.79 (0.09)	76.82 (0.15)	94.36 (0.09)	76.82 (0.15)	93.71 (0.09)	76.85 (0.15)
GRS	6.13 (0.26)	21.31 (0.83)	6.10 (0.26)	21.13 (0.83)	6.05 (0.26)	20.98 (0.83)	6.08 (0.26)	20.71 (0.82)	5.99 (0.26)	20.74 (0.82)
SHR	8.13 (0.42)	8.93 (0.46)	8.25 (0.43)	8.92 (0.46)	8.02 (0.42)	8.77 (0.46)	7.84 (0.42)	8.60 (0.45)	8.35 (0.43)	8.96 (0.46)
WET	63.10 (0.81)	66.55 (0.81)	61.40 (0.81)	65.55 (0.82)	61.86 (0.81)	66.21 (0.82)	62.64 (0.81)	66.60 (0.81)	62.94 (0.81)	65.34 (0.81)
WTR	89.73 (0.40)	92.44 (0.36)	90.09 (0.40)	92.53 (0.35)	90.28 (0.39)	92.36 (0.36)	90.83 (0.38)	91.63 (0.37)	90.10 (0.40)	91.56 (0.37)
IMP	58.55 (0.56)	72.69 (0.56)	59.21 (0.55)	72.06 (0.56)	59.06 (0.55)	71.72 (0.56)	58.65 (0.55)	70.85 (0.56)	59.01 (0.55)	70.29 (0.56)
BAL	52.77 (1.12)	39.62 (0.95)	52.90 (1.12)	38.44 (0.93)	52.19 (1.13)	37.70 (0.93)	52.07 (1.13)	36.16 (0.90)	52.33 (1.13)	34.69 (0.87)
PSI	86.02 (5.00)	35.01 (4.38)	91.40 (4.04)	36.56 (4.38)	89.25 (4.46)	31.86 (4.00)	96.24 (2.74)	31.40 (3.81)	96.24 (2.74)	31.35 (3.81)
O.A. (SE)	82.11 (0.09)		81.99 (0.09)		81.97 (0.09)		81.82 (0.09)		81.64 (0.09)	

more noticeable or pronounced compared to the AB values of the other land-cover types.

Table 7 presents the confusion matrix of changed and unchanged pixels obtained using the LUCAS validation datasets. The O.A. of GLC\_FCS30D reached  $90.36 \pm 0.38\%$ ; the P.A. and U.A. of the changed pixels were  $52.86 \pm 1.93\%$  and  $73.31 \pm 1.74\%$ , respectively; and the corresponding F1 score was  $61.43\%$ . In contrast, the unchanged land-cover pixels reached high P.A. and U.A. values, with both metrics exceeding  $90\%$ . Thus, the changed land-cover pixels were more difficult to capture compared with these unchanged pixels. Similarly, Stehman et al. (2021) also found that the accuracy metrics of the changed pixels were far lower than those of the unchanged pixels: the producer's accuracy of the changed pixels and unchanged pixels was  $16\%$  and  $99\%$ , respectively.

#### 4.4 Comparisons with other global land-cover dynamics products

Figure 12 gives qualitative comparisons between our GLC\_FCS30D and two widely used land-cover dynamics datasets (CCI\_LC and MCD12Q1) for 2001–2020 in the Indo-China Peninsula, which experienced evident land-cover changes in terms of forest deforestation and urban expansion during that period. In terms of urban expansion, the three datasets revealed rapid urbanization in the megacity of Bangkok, while CCI\_LC underestimated the impervious surface area in 2001 compared with the other two datasets. Meanwhile, GLC\_FCS30D also captured more spatial detail

(such as rural buildings and road networks) than CCI\_LC and MCD12Q1 because of its high spatial resolution of 30 m.

In terms of the most significant deforestation, CCI\_LC showed the worst performance because (1) it underestimated the forest cover in 2001 (rectangular region 1: R1), i.e., some forests were wrongly labeled croplands; (2) some deforested areas were not captured during the period 2001–2020 in rectangular region 2 (R2), so its deforested forest area was less than in GLC\_FCS30D and MCD12Q1; and (3) there was an obvious problem with the misclassification of forest as wetland in 2001 (rectangular region 3: R3). MCD12Q1 also suffered from a forest omission error in R1; namely, the captured forest area in 2001 was lower than the actual forest area based on natural-color imagery. As for the evident deforestation in R2, we find that almost all forest pixels changed to the other land-cover types (savanna and grassland) in MCD12Q1, which obviously deviated from the actual situation; thus, MCD12Q1 overestimated the forest deforestation. Meanwhile, the MCD12Q1 time series showed various land-cover distributions in R3, which indicated that MCD12Q1 has lower mapping accuracy and temporal stability for these wetland areas. In comparison, GLC\_FCS30D achieved the best performance in capturing the spatial distribution of forest in 2001, deforestation during 2001–2020, and wetland stability.

Figure 13 shows another comparison example for these three datasets, focusing instead on Paraguay, South America. The most evident land-cover changes were deforestation and increased cropland according to the time series of Landsat natural-color imagery. In terms of the spa-



**Table 6.** The area proportions and area bias (AB) values of 10 major land-cover types from the GLC\_FCS30D dataset (Map) and the LUCAS validation dataset (Ref).

	2006			2009			2012			2015			2018		
	Map	Ref	AB	Map	Ref	AB	Map	Ref	AB	Map	Ref	AB	Map	Ref	AB
CRP	46.48	50.62	-4.14	46.46	50.64	-4.18	46.59	50.67	-4.08	46.63	50.69	-4.06	46.77	50.74	-3.97
FST	41.39	33.76	7.63	41.28	33.75	7.53	41.14	33.73	7.41	40.96	33.73	7.23	40.66	33.68	6.98
GRS	1.21	4.15	-2.94	1.21	4.15	-2.94	1.21	4.15	-2.94	1.23	4.15	-2.92	1.21	4.15	-2.94
SHR	1.91	2.08	-0.17	1.94	2.08	-0.14	1.92	2.08	-0.16	1.91	2.07	-0.16	1.95	2.06	-0.11
WET	1.70	1.75	-0.05	1.68	1.74	-0.06	1.68	1.73	-0.05	1.69	1.71	-0.02	1.73	1.72	0.01
WTR	2.75	2.85	-0.1	2.76	2.85	-0.09	2.77	2.85	-0.08	2.81	2.85	-0.04	2.79	2.86	-0.07
IMP	3.18	3.82	-0.64	3.25	3.83	-0.58	3.25	3.82	-0.57	3.27	3.82	-0.55	3.32	3.82	-0.5
BAL	1.32	0.95	0.37	1.36	0.95	0.41	1.37	0.95	0.42	1.42	0.95	0.47	1.49	0.95	0.54
PSI	0.06	0.02	0.04	0.06	0.02	0.04	0.07	0.02	0.05	0.07	0.02	0.05	0.07	0.02	0.05

**Table 7.** The confusion matrix of changed and unchanged pixels in GLC\_FCS30D when using LUCAS time-series datasets across the Europe Union.

	Unchanged	Changed	Total	P.A. (SE)	F1
Unchanged	82.69	2.79	85.48	96.73	94.49
Changed	6.84	7.68	14.52	52.86	61.43
Total	89.53	10.47			
U.A. (SE)	92.36(0.36)	73.31(1.74)			
O.A. (SE)		90.36(0.38)			

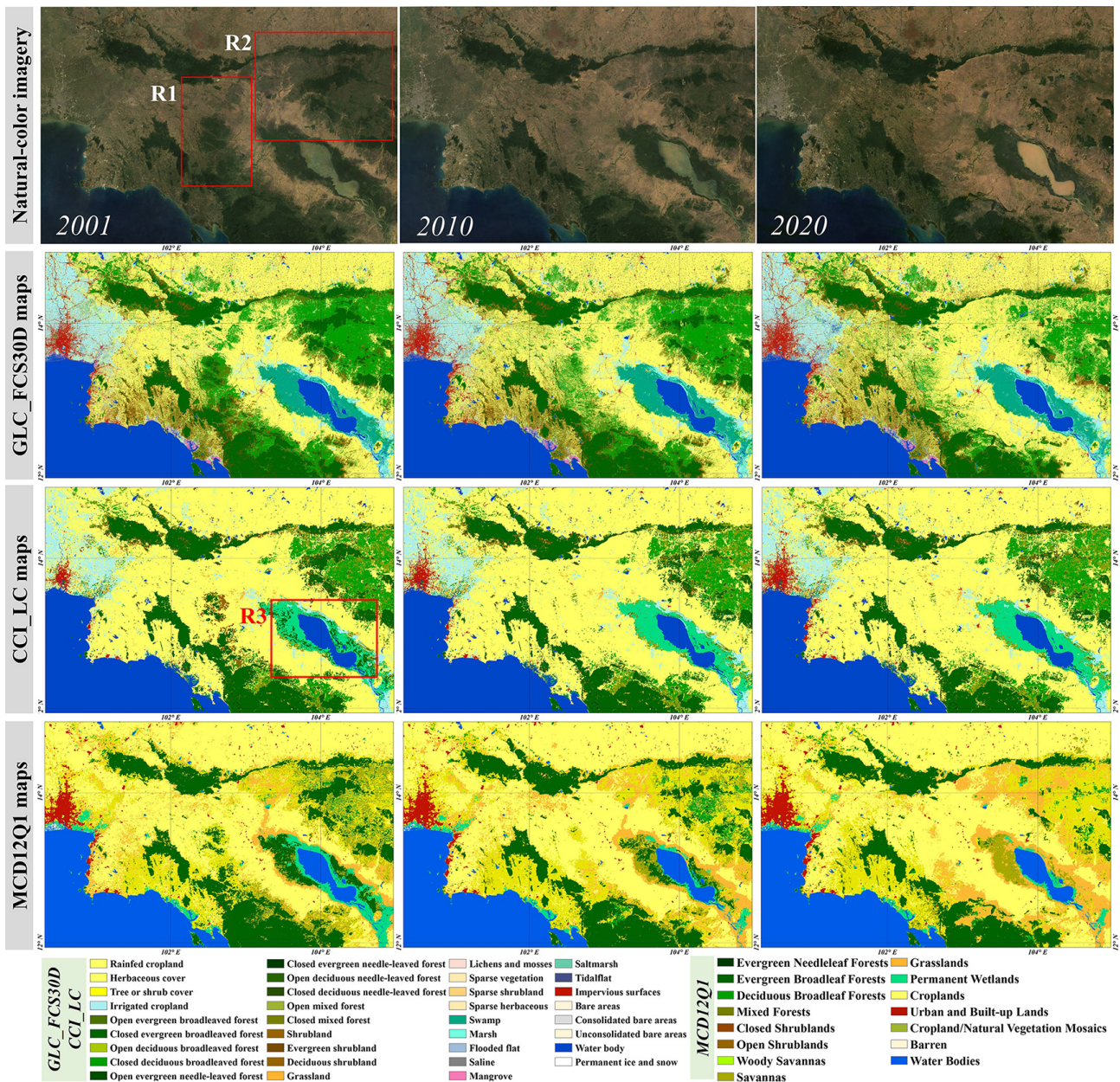
tial distribution, the consistency between GLC\_FCS30D and CCI\_LC was higher, while MCD12Q1 was obviously different from the other two datasets. A large number of deciduous broadleaved forests were labeled as savanna and woody savanna, and most croplands were identified as grasslands in MCD12Q1, mainly because of the difference in classification system. In terms of the changed-land-cover areas, the GLC\_FCS30D showed the highest accuracy and captured richer spatial detail. For example, the deforestation intensity during 2010–2020 was significantly greater than that during 2001–2010, and GLC\_FCS30D also revealed the regular deforestation caused by human factors. In contrast, CCI\_LC and MCD12Q1 captured neither the deforestation during 2010–2020 nor the small and fragmented changes (caused by human activities).

#### 4.5 Limitations of and perspectives on the GLC\_FCS30D dataset

To achieve the goal of accurate and robust monitoring of global land-cover change, four steps are adopted: (1) the advantages of the CCD model and the full time series of Landsat observations are combined to capture the land-cover change time points for any changed pixels; (2) the temporally stable areas are used as prior knowledge to ensure the quality of training samples and local adaptive modeling is adopted to update the land-cover transitions of these changed pixels; (3) global thematic products for two complicated land-cover

types (impervious surface and wetland) are independently developed to improve the reliability of GLC\_FCS30D; and (4) the “spatiotemporal consistency checking” optimization in Sect. 3.3.3 is applied to further guarantee the stability and accuracy of GLC\_FCS30D. The accuracy assessments performed using the developed global validation dataset and two third-party datasets demonstrate that GLC\_FCS30D fulfills the accuracy requirements for a baseline year and for time-series variability at a global or national scale. Comparisons with other land-cover products also highlight the superiority of GLC\_FCS30D in terms of classification system diversity and the monitoring accuracy of these changed areas. However, monitoring global land-cover change across a long time series is an extremely complex and difficult task (Hansen and Loveland, 2012; Song et al., 2018; Winkler et al., 2021; Xian et al., 2022). Although this study uses a series of measurements and methods to achieve global 30 m land-cover change monitoring for the past 37 years, there are still some uncertainties and limitations that need to be resolved in further work.

The CCD algorithm makes full use of dense satellite observations to capture land-cover changes robustly and accurately (Zhu and Woodcock, 2014b; Zhu et al., 2012). However, previous studies have demonstrated that their reliability is highly correlated to the density of valid satellite observations (Bullock et al., 2022; Ye et al., 2021; Zhu et al., 2019). Cloudy and snowy areas lead to greater uncertainty when

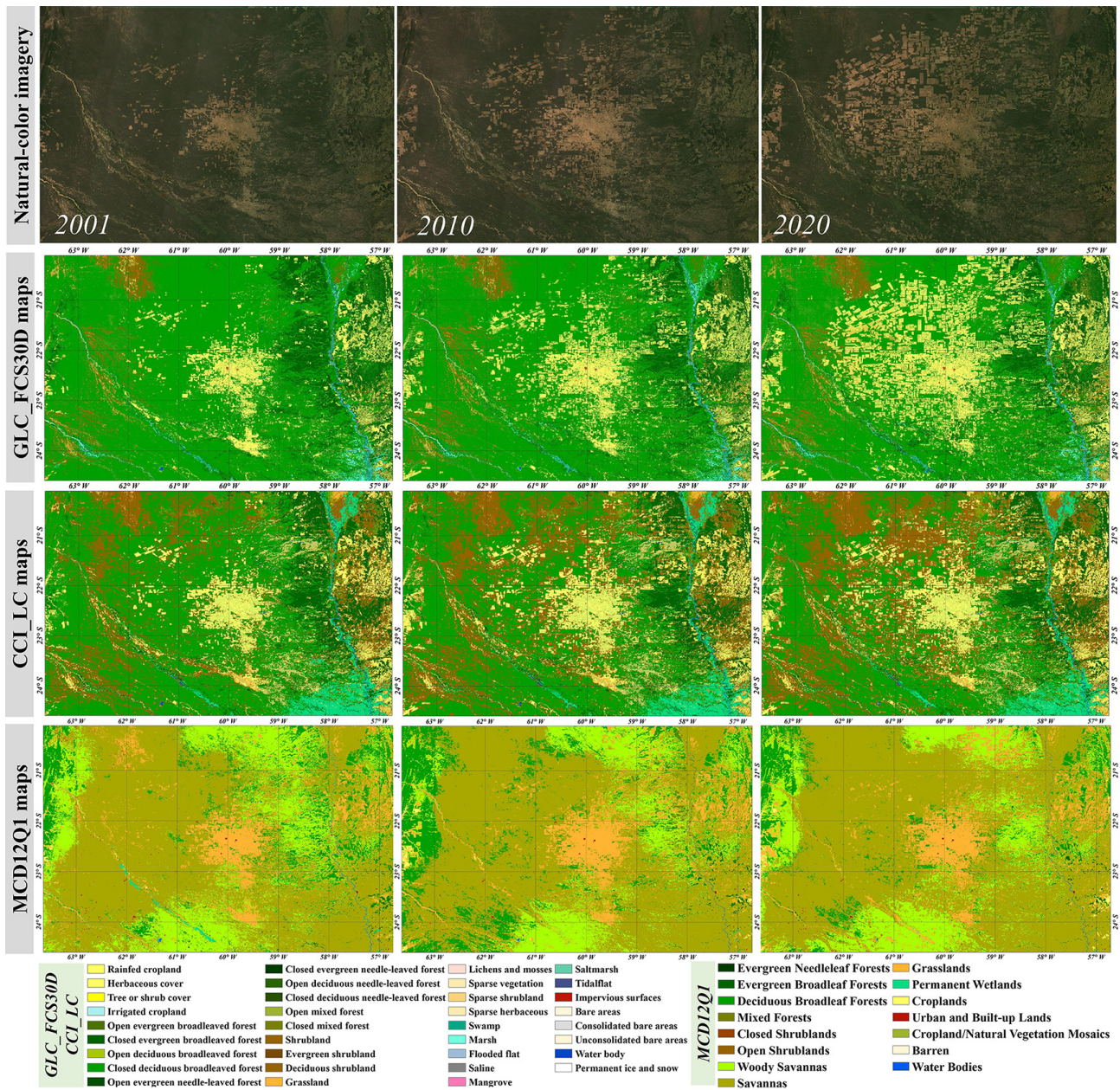


**Figure 12.** Comparisons of GLC\_FCS30D with the CCI\_LC and MCD12Q1 land-cover dynamics products for the Indo-China Peninsula during 2001–2020. The natural-color imagery were composited from the time-series Landsat imagery.

capturing the time points of land-cover change (DeVries et al., 2015; Xian et al., 2022). Additionally, due to the limited storage capacity and satellite–ground data-transmission capacity of early satellites, the density of Landsat imagery is sparse before 2000 (only a single satellite, Landsat 5, acquired data) (Roy et al., 2014b). In this study, we combine the satellite observations from 2 years before and after the nominal center year from 1985 to 1995; for example, we update the land-cover maps in 1995 using all available imagery from 1993 to 1997. However, a previous study found that north-eastern Asia did not have any valid Landsat observations be-

fore 2000 (Zhang et al., 2022), which means that some land-cover changes could not be captured in these areas before 2000 in GLC\_FCS30D. To solve the problem of missing and sparse observations, a useful solution is to fuse multisourced remote-sensing imagery. For example, Y. Zhang et al. (2021) combined Landsat and Sentinel-2 imagery to track tropical forest disturbances with an overall accuracy of more than 87 %. Therefore, further work will investigate the feasibility of integrating Sentinel-1/2, SPOT, MODIS, and AVHRR imagery as auxiliary datasets to achieve annual land-cover





**Figure 13.** Comparisons of GLC\_FCS30D with two land-cover-dynamics time-series datasets for Paraguay, South America, during 2001–2020. The natural-color imagery were composited from the time-series Landsat imagery.

monitoring before 2000 and further ensure land-cover monitoring quality.

To ensure the stability of GLC\_FCS30D, a spatiotemporal consistency optimization algorithm that has been widely used in impervious surface change optimizations (Li et al., 2015; Zhang et al., 2022) was applied. This makes full use of spatiotemporally neighboring pixels to calculate the land-cover homogeneity and then remove the “salt and pepper” noise caused by the pixel-based classifications. Qualitative comparisons for deforested areas of the Amazon and areas of China that have undergone urban expansion (Fig. S2

in the Supplement) also showed that the spatiotemporal consistency optimization can improve the data quality of GLC\_FCS30D by suppressing salt and pepper noise and optimizing the temporal consistency. Similarly, Yang and Huang (2021) used this algorithm to optimize China’s annual land-cover products during 1990–2019, and found that it improved the mapping accuracy of the land-cover time-series dataset.

GLC\_FCS30D reveals a large number of land-cover changes in the semi-arid regions illustrated in Fig. 5a; these land-cover changes are more influenced by climate factors.



For example, the central region of Australia is a typical semi-arid region, and the dominant land-cover types are grassland, sparse vegetation, shrubland, and bare land. In general, if there is sufficient annual precipitation, the distributions of shrubland and grassland in the area will be more extensive; otherwise, the area will be dominated by bare land and sparse vegetation (Dong et al., 2020; Ge et al., 2022). Recently, some studies suggested suppressing these changes; for example, Bastos et al. (2022) chose to suppress these land-cover changes by fusing these four land-cover types into a single grassland land-cover type for Australia, and Xian et al. (2022) combined grassland and shrubland in the CONUS. Whether these frequent and climate-sensitive land-cover changes should be suppressed will be considered in our further work.

Although we used a global validation dataset to assess the capability of GLC\_FCS30D in the baseline year of 2020 and two third-party regional datasets to assess its variability of the accuracy across the time series in the European Union and the CONUS, the accuracy assessment work should be strengthened. In particular, the classification system differences between GLC\_FCS30D, LUCAS, and LCMAP\_Val cannot be ignored. For example, the impervious surface land-cover type in LUCAS and LCMAP\_Val contains artificial surfaces and their surroundings (such as city greenery) (Stehman et al., 2021; Xian et al., 2022), while GLC\_FCS30D only includes artificial structures (Zhang et al., 2022), so the impervious surface in GLC\_FCS30D had low P.A. when validated with the LUCAS and LCMAP\_Val datasets in Sect. 4.3. The time-series accuracy variability was only analyzed in two regions, so its performance in more complex areas (such as Africa and Asia) needs to be further investigated. Thus, our future work will focus on creating long-term time-series datasets used for validation for more regions and on building a long-time-series global validation dataset based on the existing works in Sect. 2.5.1, after which the accuracy metrics of the pixels with changed land cover and their intra-annual variability will be analyzed for all land-cover types.

## 5 Data availability

The developed GLC\_FCS30D dataset can be freely accessed via <https://doi.org/10.5281/zenodo.8239305> (Liu et al., 2023). To help users to navigate this dataset, it is saved as  $961\ 5^\circ \times 5^\circ$  independent tiles. Each tile is named “GLC\_FCS30D\_yyyyYYYY\_E/W\*\*N/S\*\*\*.tif”, where “E/W\*\*N/S\*\*” represents the longitude and latitude coordinates of the top-left corner and yyyy and YYYY are the start and end years of the land-cover change monitoring. GLC\_FCS30D contains 26 maps for time steps from 1985 to 2022, updated every 5 years before 2000 and annually from 2000 to 2022. It should be noted that GLC\_FCS30D adopted a 5-year cycle before 2000 because of the sparse

availability of Landsat 5 imagery at this early stage; thus, we increased the temporal cycle length to guarantee land-cover mapping accuracy. The first three time steps are saved together and the following 23 time steps are saved separately. For example, GLC\_FCS30D\_19851995\_E115N15.tif and GLC\_FCS30D\_20002022\_E115N15.tif are, respectively, the data for the first three time steps and the data for the following 23 annual time steps from 1985 to 2022 for the region corresponding to  $115\text{--}120^\circ\text{E}$ ,  $10\text{--}15^\circ\text{N}$ .

## 6 Conclusion

Land-cover change is the main cause or driving force of global climate change and has attracted increasing attention in recent decades. Long-time-series global land-cover dynamics monitoring is still a challenging task. In this study, the first global 30 m land-cover dynamics dataset that has a fine classification system (GLC\_FCS30D) containing 35 fine land-cover subcategories and which covers the period from 1985 to 2022 in 26 time steps was generated on the GEE platform. Specifically, we took advantage of the full time series of Landsat observations and the CCD algorithm to capture the time points of changed areas, and then we updated and optimized the changed-land-cover areas based on the local adaptive modeling strategy and a temporal-consistency algorithm. The accuracy assessments indicate that the proposed method can achieve accurate and spatiotemporally consistent land-cover change monitoring and that GLC\_FCS30D achieved an overall accuracy for 2020 of 80.88 % ( $\pm 0.27\%$ ) for the basic classification system's 10 major land-cover types and 73.04 % ( $\pm 0.30\%$ ) for the LCCS level-1 validation system's 17 land-cover types. Therefore, GLC\_FCS30D is the first global land-cover dynamics monitoring product with a 37-year time span, and it has the most diverse classification system. It will be essential for sustainable development, environmental protection, and informed decision-making to address the challenges of a rapidly changing world.

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## References

- Azzari, G. and Lobell, D. B.: Landsat-based classification in the cloud: An opportunity for a paradigm shift in land cover monitoring, *Remote Sens. Environ.*, 202, 64–74, <https://doi.org/10.1016/j.rse.2017.05.025>, 2017.
- Ballin, M., Barcaroli, G., Masselli, M., and Scarnò, M.: Redesign sample for land use/cover area frame survey (LUCAS) 2018, *Eurostat: statistical working papers*, 10, 132365, 2018.
- Ban, Y., Gong, P., and Giri, C.: Global land cover mapping using Earth observation satellite data: Recent progresses and challenges, *ISPRS J. Photogramm. Remote*, 103, 1–6, <https://doi.org/10.1016/j.isprsjprs.2015.01.001>, 2015.
- Bastos, A., Ciais, P., Sitch, S., Aragao, L., Chevallier, F., Fawcett, D., Rosan, T. M., Saunois, M., Gunther, D., Perugini, L., Robert, C., Deng, Z., Pongratz, J., Ganzenmuller, R., Fuchs, R., Winkler, K., Zaehle, S., and Albergel, C.: On the use of Earth Observation to support estimates of national greenhouse gas emissions and sinks for the Global stocktake process: lessons learned from ESA-CCI RECCAP2, *Carbon Balance Manag.*, 17, 15, <https://doi.org/10.1186/s13021-022-00214-w>, 2022.
- Belgiu, M. and Drăguș, L.: Random forest in remote sensing: A review of applications and future directions, *ISPRS J. Photogramm. Remote*, 114, 24–31, <https://doi.org/10.1016/j.isprsjprs.2016.01.011>, 2016.
- Buchhorn, M., Lesiv, M., Tsendbazar, N.-E., Herold, M., Bertels, L., and Smets, B.: Copernicus Global Land Cover Layers – Collection 2, *Remote Sens.*, 12, 1044, <https://doi.org/10.3390/rs12061044>, 2020.
- Bullock, E. L., Woodcock, C. E., and Holden, C. E.: Improved change monitoring using an ensemble of time series algorithms, *Remote Sens. Environ.*, 111165, <https://doi.org/10.1016/j.rse.2019.04.018>, 2019.
- Bullock, E. L., Healey, S. P., Yang, Z., Houborg, R., Gorelick, N., Tang, X., and Andrianirina, C.: Timeliness in forest change monitoring: A new assessment framework demonstrated using Sentinel-1 and a continuous change detection algorithm, *Remote Sens. Environ.*, 276, 113043, <https://doi.org/10.1016/j.rse.2022.113043>, 2022.
- Chen, J., Chen, J., Liao, A., Cao, X., Chen, L., Chen, X., He, C., Han, G., Peng, S., Lu, M., Zhang, W., Tong, X., and Mills, J.: Global land cover mapping at 30m resolution: A POK-based operational approach, *ISPRS J. Photogramm. Remote*, 103, 7–27, <https://doi.org/10.1016/j.isprsjprs.2014.09.002>, 2015.
- d’Andrimont, R., Yordanov, M., Martinez-Sanchez, L., Eiselt, B., Palmieri, A., Dominici, P., Gallego, J., Reuter, H. I., Joebges, C., and Lemoine, G.: Harmonised LUCAS in-situ land cover and use database for field surveys from 2006 to 2018 in the European Union, *Sci. Data*, 7, 352, <https://doi.org/10.1038/s41597-020-00675-z>, 2020.
- Defourny, P., Kirches, G., Brockmann, C., Boettcher, M., Peters, M., Bontemps, S., Lamarche, C., Schlerf, M., and M., S.: Land Cover CCI: Product User Guide Version 2, 2018, 2018.
- DeVries, B., Verbesselt, J., Kooistra, L., and Herold, M.: Robust monitoring of small-scale forest disturbances in a tropical montane forest using Landsat time series, *Remote Sens. Environ.*, 161, 107–121, <https://doi.org/10.1016/j.rse.2015.02.012>, 2015.
- Dong, J., Xiao, X., Kou, W., Qin, Y., Zhang, G., Li, L., Jin, C., Zhou, Y., Wang, J., Biradar, C., Liu, J., and Moore, B.: Tracking the dynamics of paddy rice planting area in 1986–2010 through time series Landsat images and phenology-based algorithms, *Remote Sens. Environ.*, 160, 99–113, <https://doi.org/10.1016/j.rse.2015.01.004>, 2015.
- Dong, S., Shang, Z., Gao, J., and Boone, R. B.: Enhancing sustainability of grassland ecosystems through ecological restoration and grazing management in an era of climate change on Qinghai-Tibetan Plateau, *Agric. Ecosyst. Environ.*, 287, 106684, <https://doi.org/10.1016/j.agee.2019.106684>, 2020.
- Du, P., Samat, A., Waske, B., Liu, S., and Li, Z.: Random Forest and Rotation Forest for fully polarized SAR image classification using polarimetric and spatial features, *ISPRS J. Photogramm. Remote*, 105, 38–53, <https://doi.org/10.1016/j.isprsjprs.2015.03.002>, 2015.
- Foley, J. A., Defries, R., Asner, G. P., Barford, C., Bonan, G., Carpenter, S. R., Chapin, F. S., Coe, M. T., Daily, G. C., Gibbs, H. K., Helkowski, J. H., Holloway, T., Howard, E. A., Kucharik, C. J., Monfreda, C., Patz, J. A., Prentice, I. C., Ramankutty, N., and Snyder, P. K.: Global consequences of land use, *Science*, 309, 570–574, <https://doi.org/10.1126/science.1111772>, 2005.
- Foody, G. M.: Sample size determination for image classification accuracy assessment and comparison, *Int. J. Remote*, 30, 5273–5291, <https://doi.org/10.1080/01431160903130937>, 2009.
- Foody, G. M. and Arora, M. K.: An evaluation of some factors affecting the accuracy of classification by an artificial neural network, *Int. J. Remote*, 18, 799–810, <https://doi.org/10.1080/014311697218764>, 2010.
- Friedl, M. A., Sulla-Menashe, D., Tan, B., Schneider, A., Ramankutty, N., Sibley, A., and Huang, X.: MODIS Collection 5 global land cover: Algorithm refinements and characterization of new datasets, *Remote Sens. Environ.*, 114, 168–182, <https://doi.org/10.1016/j.rse.2009.08.016>, 2010.
- Friedl, M. A., Woodcock, C. E., Olofsson, P., Zhu, Z., Loveland, T., Stanimirova, R., Arevalo, P., Bullock, E., Hu, K.-T., Zhang, Y., Turlej, K., Tarrio, K., McAvoy, K., Gorelick, N., Wang, J. A., Barber, C. P., and Souza, C.: Medium Spatial Resolution Map-

- ping of Global Land Cover and Land Cover Change Across Multiple Decades From Landsat, *Front. Remote Sens.*, 3, 894571, <https://doi.org/10.3389/frsen.2022.894571>, 2022.
- Friedlingstein, P., O'Sullivan, M., Jones, M. W., Andrew, R. M., Hauck, J., Olsen, A., Peters, G. P., Peters, W., Pongratz, J., Sitch, S., Le Quééré, C., Canadell, J. G., Ciais, P., Jackson, R. B., Alin, S., Aragão, L. E. O. C., Arneeth, A., Arora, V., Bates, N. R., Becker, M., Benoit-Cattin, A., Bittig, H. C., Bopp, L., Bultan, S., Chandra, N., Chevallier, F., Chini, L. P., Evans, W., Florentie, L., Forster, P. M., Gasser, T., Gehlen, M., Gilfillan, D., Gkritzalis, T., Gregor, L., Gruber, N., Harris, I., Hartung, K., Haverd, V., Houghton, R. A., Ilyina, T., Jain, A. K., Joetzjer, E., Kadono, K., Kato, E., Kitidis, V., Korsbakken, J. I., Landschützer, P., Lefèvre, N., Lenton, A., Lienert, S., Liu, Z., Lombardozi, D., Marland, G., Metzl, N., Munro, D. R., Nabel, J. E. M. S., Nakaoka, S.-I., Niwa, Y., O'Brien, K., Ono, T., Palmer, P. I., Pierrot, D., Poulter, B., Resplandy, L., Robertson, E., Rödenbeck, C., Schwinger, J., Séférian, R., Skjelvan, I., Smith, A. J. P., Sutton, A. J., Tans, P. P., Tian, H., Tilbrook, B., van der Werf, G., Vuichard, N., Walker, A. P., Wanninkhof, R., Watson, A. J., Willis, D., Wiltshire, A. J., Yuan, W., Yue, X., and Zaehle, S.: Global Carbon Budget 2020, *Earth Syst. Sci. Data*, 12, 3269–3340, <https://doi.org/10.5194/essd-12-3269-2020>, 2020.
- Gallant, A.: The Challenges of Remote Monitoring of Wetlands, *Remote Sens.*, 7, 10938–10950, <https://doi.org/10.3390/rs70810938>, 2015.
- Gao, Y., Liu, L., Zhang, X., Chen, X., Mi, J., and Xie, S.: Consistency Analysis and Accuracy Assessment of Three Global 30-m Land-Cover Products over the European Union using the LUCAS Dataset, *Remote Sens.*, 12, 3479, <https://doi.org/10.3390/rs12213479>, 2020.
- Ge, F., Xu, M., Gong, C., Zhang, Z., Tan, Q., and Pan, X.: Land cover changes the soil moisture response to rainfall on the Loess Plateau, *Hydrol. Process.*, 36, e14714, <https://doi.org/10.1002/hyp.14714>, 2022.
- Giri, C., Pengra, B., Long, J., and Loveland, T. R.: Next generation of global land cover characterization, mapping, and monitoring, *Int. J. Appl. Earth Obs.*, 25, 30–37, <https://doi.org/10.1016/j.jag.2013.03.005>, 2013.
- Gislason, P. O., Benediktsson, J. A., and Sveinsson, J. R.: Random Forests for land cover classification, *Pattern Recogn. Lett.*, 27, 294–300, <https://doi.org/10.1016/j.patrec.2005.08.011>, 2006.
- Gong, P., Li, X., and Zhang, W.: 40-Year (1978–2017) human settlement changes in China reflected by impervious surfaces from satellite remote sensing, *Sci. Bull.*, 64, 756–763, <https://doi.org/10.1016/j.scib.2019.04.024>, 2019a.
- Gong, P., Liu, H., Zhang, M., Li, C., Wang, J., Huang, H., Clinton, N., Ji, L., Li, W., Bai, Y., Chen, B., Xu, B., Zhu, Z., Yuan, C., Ping Suen, H., Guo, J., Xu, N., Li, W., Zhao, Y., Yang, J., Yu, C., Wang, X., Fu, H., Yu, L., Dronova, I., Hui, F., Cheng, X., Shi, X., Xiao, F., Liu, Q., and Song, L.: Stable classification with limited sample: transferring a 30-m resolution sample set collected in 2015 to mapping 10-m resolution global land cover in 2017, *Sci. Bull.*, 64, 370–373, <https://doi.org/10.1016/j.scib.2019.03.002>, 2019b.
- Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., and Moore, R.: Google Earth Engine: Planetary-scale geospatial analysis for everyone, *Remote Sens. Environ.*, 202, 18–27, <https://doi.org/10.1016/j.rse.2017.06.031>, 2017.
- Hansen, M. C. and Loveland, T. R.: A review of large area monitoring of land cover change using Landsat data, *Remote Sens. Environ.*, 122, 66–74, <https://doi.org/10.1016/j.rse.2011.08.024>, 2012.
- Hansen, M. C., Potapov, P. V., Moore, R., Hancher, M., Turubanova, S. A., Tyukavina, A., Thau, D., Stehman, S. V., Goetz, S. J., Loveland, T. R., Kommareddy, A., Egorov, A., Chini, L., Justice, C. O., and Townshend, J. R.: High-resolution global maps of 21st-century forest cover change, *Science*, 342, 850–853, <https://doi.org/10.1126/science.1244693>, 2013.
- Harris, N. L., Gibbs, D. A., Baccini, A., Birdsey, R. A., de Bruin, S., Farina, M., Fatoyinbo, L., Hansen, M. C., Herold, M., Houghton, R. A., Potapov, P. V., Suarez, D. R., Roman-Cuesta, R. M., Saatchi, S. S., Slay, C. M., Turubanova, S. A., and Tyukavina, A.: Global maps of twenty-first century forest carbon fluxes, *Nat. Clim. Change*, 11, 234–240, <https://doi.org/10.1038/s41558-020-00976-6>, 2021.
- Healey, S. P., Cohen, W. B., Yang, Z., Kenneth Brewer, C., Brooks, E. B., Gorelick, N., Hernandez, A. J., Huang, C., Joseph Hughes, M., Kennedy, R. E., Loveland, T. R., Moisen, G. G., Schroeder, T. A., Stehman, S. V., Vogelmann, J. E., Woodcock, C. E., Yang, L., and Zhu, Z.: Mapping forest change using stacked generalization: An ensemble approach, *Remote Sens. Environ.*, 204, 717–728, <https://doi.org/10.1016/j.rse.2017.09.029>, 2018.
- Herold, M., Mayaux, P., Woodcock, C. E., Baccini, A., and Schmullius, C.: Some challenges in global land cover mapping: An assessment of agreement and accuracy in existing 1 km datasets, *Remote Sens. Environ.*, 112, 2538–2556, <https://doi.org/10.1016/j.rse.2007.11.013>, 2008.
- Homer, C., Dewitz, J., Jin, S., Xian, G., Costello, C., Danielson, P., Gass, L., Funk, M., Wickham, J., Stehman, S., Auch, R., and Riitters, K.: Conterminous United States land cover change patterns 2001–2016 from the 2016 National Land Cover Database, *ISPRS J. Photogramm. Remote*, 162, 184–199, <https://doi.org/10.1016/j.isprsjprs.2020.02.019>, 2020.
- Hong, C., Burney, J. A., Pongratz, J., Nabel, J., Mueller, N. D., Jackson, R. B., and Davis, S. J.: Global and regional drivers of land-use emissions in 1961–2017, *Nature*, 589, 554–561, <https://doi.org/10.1038/s41586-020-03138-y>, 2021.
- Huang, C., Goward, S. N., Schleeweis, K., Thomas, N., Masek, J. G., and Zhu, Z.: Dynamics of national forests assessed using the Landsat record: Case studies in eastern United States, *Remote Sens. Environ.*, 113, 1430–1442, <https://doi.org/10.1016/j.rse.2008.06.016>, 2009.
- Jin, H., Stehman, S. V., and Mountrakis, G.: Assessing the impact of training sample selection on accuracy of an urban classification: a case study in Denver, Colorado, *Int. J. Remote Sens.*, 35, 2067–2081, <https://doi.org/10.1080/01431161.2014.885152>, 2014.
- Jin, S., Yang, L., Zhu, Z., and Homer, C.: A land cover change detection and classification protocol for updating Alaska NLCD 2001 to 2011, *Remote Sens. Environ.*, 195, 44–55, <https://doi.org/10.1016/j.rse.2017.04.021>, 2017.
- Jin, S., Dewitz, J., Li, C., Sorenson, D., Zhu, Z., Shogib, M. R. I., Danielson, P., Granneman, B., Costello, C., Case, A., and Gass, L.: National Land Cover Database 2019: A Comprehensive Strategy for Creating the 1986–2019 Forest Disturbance Product, *J. Remote Sens.*, 3, 0021, <https://doi.org/10.34133/remotesensing.0021>, 2023.

- Kennedy, R. E., Cohen, W. B., and Schroeder, T. A.: Trajectory-based change detection for automated characterization of forest disturbance dynamics, *Remote Sens. Environ.*, 110, 370–386, <https://doi.org/10.1016/j.rse.2007.03.010>, 2007.
- Kennedy, R. E., Yang, Z., and Cohen, W. B.: Detecting trends in forest disturbance and recovery using yearly Landsat time series: 1. LandTrendr – Temporal segmentation algorithms, *Remote Sens. Environ.*, 114, 2897–2910, <https://doi.org/10.1016/j.rse.2010.07.008>, 2010.
- Kenny, Q. Y.: Indicator function and its application in two-level factorial designs, *Ann. Stat.*, 31, 984–994, <https://doi.org/10.1214/aos/1056562470>, 2003.
- Li, X., Gong, P., and Liang, L.: A 30-year (1984–2013) record of annual urban dynamics of Beijing City derived from Landsat data, *Remote Sens. Environ.*, 166, 78–90, <https://doi.org/10.1016/j.rse.2015.06.007>, 2015.
- Liu, C., Zhang, Q., Luo, H., Qi, S., Tao, S., Xu, H., and Yao, Y.: An efficient approach to capture continuous impervious surface dynamics using spatial-temporal rules and dense Landsat time series stacks, *Remote Sens. Environ.*, 229, 114–132, <https://doi.org/10.1016/j.rse.2019.04.025>, 2019.
- Liu, H., Gong, P., Wang, J., Wang, X., Ning, G., and Xu, B.: Production of global daily seamless data cubes and quantification of global land cover change from 1985 to 2020 – iMap World 1.0, *Remote Sens. Environ.*, 258, 112364, <https://doi.org/10.1016/j.rse.2021.112364>, 2021.
- Liu, L., Zhang, X., Chen, X., Gao, Y., and Mi, J.: GLC\_FCS30-2020: Global Land Cover with Fine Classification System at 30 m in 2020 (v1.2), Zenodo [data set], <https://doi.org/10.5281/zenodo.4280923>, 2020.
- Liu, L., Zhang, X., Gao, Y., Chen, X., Shuai, X., and Mi, J.: Finer-Resolution Mapping of Global Land Cover: Recent Developments, Consistency Analysis, and Prospects, *J. Remote Sens.*, 2021, 5289697, <https://doi.org/10.34133/2021/5289697>, 2021.
- Liu, L., Zhang, X., and Zhao, T.: GLC\_FCS30D: the first global 30-m land-cover dynamic monitoring product with fine classification system from 1985 to 2022, Zenodo [data set], <https://doi.org/10.5281/zenodo.8239305>, 2023.
- Mellor, A., Boukir, S., Haywood, A., and Jones, S.: Exploring issues of training data imbalance and mislabelling on random forest performance for large area land cover classification using the ensemble margin, *ISPRS J. Photogramm. Remote.*, 105, 155–168, <https://doi.org/10.1016/j.isprsjprs.2015.03.014>, 2015.
- Millard, K. and Richardson, M.: On the Importance of Training Data Sample Selection in Random Forest Image Classification: A Case Study in Peatland Ecosystem Mapping, *Remote Sens.*, 7, 8489–8515, <https://doi.org/10.3390/rs70708489>, 2015.
- Pekel, J. F., Cottam, A., Gorelick, N., and Belward, A. S.: High-resolution mapping of global surface water and its long-term changes, *Nature*, 540, 418–422, <https://doi.org/10.1038/nature20584>, 2016.
- Pengra, B., Gallant, A., Zhu, Z., and Dahal, D.: Evaluation of the Initial Thematic Output from a Continuous Change-Detection Algorithm for Use in Automated Operational Land-Change Mapping by the U.S. Geological Survey, *Remote Sens.*, 8, 811, <https://doi.org/10.3390/rs8100811>, 2016.
- Pontus Olofsson, G. M. F.: Good practices for estimating area and assessing accuracy of land change, *Remote Sens. Environ.*, 148, 42–57, <https://doi.org/10.1016/j.rse.2014.02.015>, 2014.
- Potapov, P., Turubanova, S., Hansen, M. C., Tyukavina, A., Zalles, V., Khan, A., Song, X.-P., Pickens, A., Shen, Q., and Cortez, J.: Global maps of cropland extent and change show accelerated cropland expansion in the twenty-first century, *Nature Food*, 3, 19–28, <https://doi.org/10.1038/s43016-021-00429-z>, 2021.
- Potapov, P., Hansen, M. C., Pickens, A., Hernandez-Serna, A., Tyukavina, A., Turubanova, S., Zalles, V., Li, X., Khan, A., Stolle, F., Harris, N., Song, X.-P., Baggett, A., Kommareddy, I., and Kommareddy, A.: The Global 2000–2020 Land Cover and Land Use Change Dataset Derived From the Landsat Archive: First Results, *Front. Remote Sens.*, 3, 856903, <https://doi.org/10.3389/frsen.2022.856903>, 2022.
- Qin, Y., Xiao, X., Wigneron, J.-P., Ciaï, P., Canadell, J. G., Brandt, M., Li, X., Fan, L., Wu, X., Tang, H., Dubayah, R., Doughty, R., Chang, Q., Crowell, S., Zheng, B., Neal, K., Celis, J. A., and Moore, B.: Annual Maps of Forests in Australia from Analyses of Microwave and Optical Images with FAO Forest Definition, *J. Remote Sens.*, 2021, 9784657, <https://doi.org/10.34133/2021/9784657>, 2021.
- Radoux, J., Lamarche, C., Van Bogaert, E., Bontemps, S., Brockmann, C., and Defourny, P.: Automated Training Sample Extraction for Global Land Cover Mapping, *Remote Sens.*, 6, 3965–3987, <https://doi.org/10.3390/rs6053965>, 2014.
- Roy, D. P., Qin, Y., Kovalsky, V., Vermote, E. F., Ju, J., Egorov, A., Hansen, M. C., Kommareddy, I., and Yan, L.: Conterminous United States demonstration and characterization of MODIS-based Landsat ETM+ atmospheric correction, *Remote Sens. Environ.*, 140, 433–449, <https://doi.org/10.1016/j.rse.2013.09.012>, 2014a.
- Roy, D. P., Wulder, M. A., Loveland, T. R., C.E. W., Allen, R. G., Anderson, M. C., Helder, D., Irons, J. R., Johnson, D. M., Kennedy, R., Scambos, T. A., Schaaf, C. B., Schott, J. R., Sheng, Y., Vermote, E. F., Belward, A. S., Bindaschadler, R., Cohen, W. B., Gao, F., Hipple, J. D., Hostert, P., Huntington, J., Justice, C. O., Kilic, A., Kovalsky, V., Lee, Z. P., Lymburner, L., Masek, J. G., McCorkel, J., Shuai, Y., Trezza, R., Vogelmann, J., Wynne, R. H., and Zhu, Z.: Landsat-8: Science and product vision for terrestrial global change research, *Remote Sens. Environ.*, 145, 154–172, <https://doi.org/10.1016/j.rse.2014.02.001>, 2014b.
- Roy, D. P., Kovalsky, V., Zhang, H. K., Vermote, E. F., Yan, L., Kumar, S. S., and Egorov, A.: Characterization of Landsat-7 to Landsat-8 reflective wavelength and normalized difference vegetation index continuity, *Remote Sens. Environ.*, 185, 57–70, <https://doi.org/10.1016/j.rse.2015.12.024>, 2016.
- Song, X. P., Hansen, M. C., Stehman, S. V., Potapov, P. V., Tyukavina, A., Vermote, E. F., and Townshend, J. R.: Global land change from 1982 to 2016, *Nature*, 560, 639–643, <https://doi.org/10.1038/s41586-018-0411-9>, 2018.
- Stehman, S. V., Pengra, B. W., Horton, J. A., and Wellington, D. F.: Validation of the U.S. Geological Survey’s Land Change Monitoring, Assessment and Projection (LCMAP) Collection 1.0 annual land cover products 1985–2017, *Remote Sens. Environ.*, 265, 112646, <https://doi.org/10.1016/j.rse.2021.112646>, 2021.
- Sulla-Menashe, D., Gray, J. M., Abercrombie, S. P., and Friedl, M. A.: Hierarchical mapping of annual global land cover 2001 to present: The MODIS Collection 6 Land Cover product, *Remote Sens. Environ.*, 222, 183–194, <https://doi.org/10.1016/j.rse.2018.12.013>, 2019.

- Tachikawa, T., Hato, M., Kaku, M., and Iwasaki, A.: Characteristics of ASTER GDEM Version 2, *Geoscience and Remote Sensing Symposium (IGARSS)*, 3657–3660, <https://doi.org/10.1109/IGARSS.2011.6050017>, 2011.
- Venter, Z. S., Barton, D. N., Chakraborty, T., Simensen, T., and Singh, G.: Global 10 m Land Use Land Cover Datasets: A Comparison of Dynamic World, World Cover and Esri Land Cover, *Remote Sens.*, 14, 4101, <https://doi.org/10.3390/rs14164101>, 2022.
- Verbesselt, J., Hyndman, R., Newnham, G., and Culvenor, D.: Detecting trend and seasonal changes in satellite image time series, *Remote Sens. Environ.*, 114, 106–115, 2010.
- Vermote, E.: LEDAPS surface reflectance product description, <https://www.usgs.gov/media/files/landsat-4-7-collection-1-surface-reflectance-code-ledaps-product-guide> (last access: 12 March 2024), 2007.
- Vermote, E. F. and Kotchenova, S.: Atmospheric correction for the monitoring of land surfaces, *J. Geophys. Res.*, 113, D23S90, <https://doi.org/10.1029/2007jd009662>, 2008.
- Wang, N., Zhang, X., Yao, S., Wu, J., and Xia, H.: How Good Are Global Layers for Mapping Rural Settlements? Evidence from China, *Land*, 11, 1308, <https://doi.org/10.3390/land11081308>, 2022.
- Winkler, K., Fuchs, R., Rounsevell, M., and Herold, M.: Global land use changes are four times greater than previously estimated, *Nat. Commun.*, 12, 2501, <https://doi.org/10.1038/s41467-021-22702-2>, 2021.
- Xian, G. Z., Smith, K., Wellington, D., Horton, J., Zhou, Q., Li, C., Auch, R., Brown, J. F., Zhu, Z., and Reker, R. R.: Implementation of the CCDC algorithm to produce the LCMAP Collection 1.0 annual land surface change product, *Earth Syst. Sci. Data*, 14, 143–162, <https://doi.org/10.5194/essd-14-143-2022>, 2022.
- Xiao, Y., Wang, Q., Tong, X., and Atkinson, P. M.: Thirty-meter map of young forest age in China, *Earth Syst. Sci. Data*, 15, 3365–3386, <https://doi.org/10.5194/essd-15-3365-2023>, 2023.
- Xie, S., Liu, L., Zhang, X., and Yang, J.: Mapping the annual dynamics of land cover in Beijing from 2001 to 2020 using Landsat dense time series stack, *ISPRS J. Photogramm. Remote*, 185, 201–218, <https://doi.org/10.1016/j.isprsjprs.2022.01.014>, 2022.
- Yang, J. and Huang, X.: The 30 m annual land cover dataset and its dynamics in China from 1990 to 2019, *Earth Syst. Sci. Data*, 13, 3907–3925, <https://doi.org/10.5194/essd-13-3907-2021>, 2021.
- Ye, S., Rogan, J., Zhu, Z., and Eastman, J. R.: A near-real-time approach for monitoring forest disturbance using Landsat time series: stochastic continuous change detection, *Remote Sens. Environ.*, 252, 112167, <https://doi.org/10.1016/j.rse.2020.112167>, 2021.
- Zhang, H. K. and Roy, D. P.: Using the 500 m MODIS land cover product to derive a consistent continental scale 30 m Landsat land cover classification, *Remote Sens. Environ.*, 197, 15–34, <https://doi.org/10.1016/j.rse.2017.05.024>, 2017.
- Zhang, X., Liu, L., Chen, X., Xie, S., and Gao, Y.: Fine Land-Cover Mapping in China Using Landsat Databcube and an Operational SPECLib-Based Approach, *Remote Sens.*, 11, 1056, <https://doi.org/10.3390/rs11091056>, 2019.
- Zhang, X., Liu, L., Wu, C., Chen, X., Gao, Y., Xie, S., and Zhang, B.: Development of a global 30 m impervious surface map using multisource and multitemporal remote sensing datasets with the Google Earth Engine platform, *Earth Syst. Sci. Data*, 12, 1625–1648, <https://doi.org/10.5194/essd-12-1625-2020>, 2020.
- Zhang, X., Liu, L., Chen, X., Gao, Y., and Jiang, M.: Automatically Monitoring Impervious Surfaces Using Spectral Generalization and Time Series Landsat Imagery from 1985 to 2020 in the Yangtze River Delta, *J. Remote Sens.*, 2021, 1–16, <https://doi.org/10.34133/2021/9873816>, 2021a.
- Zhang, X., Liu, L., Chen, X., Gao, Y., Xie, S., and Mi, J.: GLC\_FCS30: global land-cover product with fine classification system at 30 m using time-series Landsat imagery, *Earth Syst. Sci. Data*, 13, 2753–2776, <https://doi.org/10.5194/essd-13-2753-2021>, 2021b.
- Zhang, X., Liu, L., Zhao, T., Gao, Y., Chen, X., and Mi, J.: GISD30: global 30 m impervious-surface dynamic dataset from 1985 to 2020 using time-series Landsat imagery on the Google Earth Engine platform, *Earth Syst. Sci. Data*, 14, 1831–1856, <https://doi.org/10.5194/essd-14-1831-2022>, 2022.
- Zhang, X., Liu, L., Zhao, T., Chen, X., Lin, S., Wang, J., Mi, J., and Liu, W.: GWL\_FCS30: a global 30 m wetland map with a fine classification system using multi-sourced and time-series remote sensing imagery in 2020, *Earth Syst. Sci. Data*, 15, 265–293, <https://doi.org/10.5194/essd-15-265-2023>, 2023.
- Zhang, Y., Ling, F., Wang, X., Foody, G. M., Boyd, D. S., Li, X., Du, Y., and Atkinson, P. M.: Tracking small-scale tropical forest disturbances: Fusing the Landsat and Sentinel-2 data record, *Remote Sens. Environ.*, 261, 112470, <https://doi.org/10.1016/j.rse.2021.112470>, 2021c.
- Zhao, T., Zhang, X., Gao, Y., Mi, J., Liu, W., Wang, J., Jiang, M., and Liu, L.: Assessing the Accuracy and Consistency of Six Fine-Resolution Global Land Cover Products Using a Novel Stratified Random Sampling Validation Dataset, *Remote Sens.*, 15, 2285, <https://doi.org/10.3390/rs15092285>, 2023.
- Zhu, Z., Wang, S. X., and Woodcock, C. E.: Improvement and expansion of the Fmask algorithm: cloud, cloud shadow, and snow detection for Landsats 4–7, 8, and Sentinel 2 images, *Remote Sens. Environ.*, 159, 269–277, <https://doi.org/10.1016/j.rse.2014.12.014>, 2015.
- Zhu, Z., Gallant, A. L., Woodcock, C. E., Pengra, B., Olofsson, P., Loveland, T. R., Jin, S., Dahal, D., Yang, L., and Auch, R. F.: Optimizing selection of training and auxiliary data for operational land cover classification for the LCMAP initiative, *ISPRS J. Photogramm. Remote*, 122, 206–221, <https://doi.org/10.1016/j.isprsjprs.2016.11.004>, 2016.
- Zhu, Z.: Change detection using landsat time series: A review of frequencies, preprocessing, algorithms, and applications, *ISPRS J. Photogramm. Remote*, 130, 370–384, <https://doi.org/10.1016/j.isprsjprs.2017.06.013>, 2017.
- Zhu, Z. and Woodcock, C. E.: Object-based cloud and cloud shadow detection in Landsat imagery, *Remote Sens. Environ.*, 118, 83–94, <https://doi.org/10.1016/j.rse.2011.10.028>, 2012.
- Zhu, Z. and Woodcock, C. E.: Automated cloud, cloud shadow, and snow detection in multitemporal Landsat data: An algorithm designed specifically for monitoring land cover change, *Remote Sens. Environ.*, 152, 217–234, <https://doi.org/10.1016/j.rse.2014.06.012>, 2014a.
- Zhu, Z. and Woodcock, C. E.: Continuous change detection and classification of land cover using all available Landsat data, *Remote Sens. Environ.*, 144, 152–171, <https://doi.org/10.1016/j.rse.2014.01.011>, 2014b.



Zhu, Z., Woodcock, C. E., and Olofsson, P.: Continuous monitoring of forest disturbance using all available Landsat imagery, *Remote Sens. Environ.*, 122, 75–91, <https://doi.org/10.1016/j.rse.2011.10.030>, 2012.

Zhu, Z., Zhang, J., Yang, Z., Aljaddani, A. H., Cohen, W. B., Qiu, S., and Zhou, C.: Continuous monitoring of land disturbance based on Landsat time series, *Remote Sens. Environ.*, 238, 111116, <https://doi.org/10.1016/j.rse.2019.03.009>, 2019.